ANNEX 3 Methodological Descriptions for Additional Source or Sink Categories

3.1. Methodology for Estimating Emissions of CH_4, N_2O, and Indirect Greenhouse Gases from Stationary Combustion

Estimates of CH_4 and N_2O Emissions

Methane (CH_4) and nitrous oxide (N_2O) emissions from stationary combustion were estimated using IPCC emission factors and methods. Estimates were obtained by multiplying emission factors—by sector and fuel type—by fossil fuel and wood consumption data. This "top-down" methodology is characterized by two basic steps, described below. Data are presented in Table A-68 through Table A-72.

Step 1: Determine Energy Consumption by Sector and Fuel Type

Energy consumption from stationary combustion activities was grouped by sector: industrial, commercial, residential, electric power, and U.S. territories. For CH_4 and N_2O, estimates were based upon consumption of coal, gas, oil, and wood. Energy consumption data for the United States were obtained from EIA's *Monthly Energy Review, September 2006* and Unpublished Supplemental Tables on Petroleum Product detail (EIA 2006a). Wood consumption data for the United States was obtained from EIA's *Annual Energy Review* (EIA 2006b). Because the United States does not include territories in its national energy statistics, fuel consumption data for territories were collected separately from the EIA from Grillot (2006).[21] Fuel consumption for the industrial sector was adjusted to subtract out construction and agricultural use, which is reported under mobile sources.[22] Construction and agricultural fuel use was obtained from EPA (2004). The energy consumption data by sector were then adjusted from higher to lower heating values by multiplying by 0.9 for natural gas and wood and by 0.95 for coal and petroleum fuel. This is a simplified convention used by the International Energy Agency. Table A-68 provides annual energy consumption data for the years 1990 through 2005.

Step 2: Determine the Amount of CH_4 and N_2O Emitted

Activity data for each sector and fuel type were then multiplied by emission factors to obtain emission estimates. Emission factors for the residential, commercial, industrial, and electric power sectors were taken from the *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997). These N_2O emission factors by fuel type (consistent across sectors) were also assumed for U.S. territories. The CH_4 emission factors by fuel type for U.S. territories were estimated based on the emission factor for the primary sector in which each fuel was combusted. Table A-69 provides emission factors used for each sector and fuel type.

Estimates of NO_x, CO, and NMVOC Emissions

Emissions estimates for NO_x, CO, and NMVOCs were obtained from preliminary data (EPA 2005) and disaggregated based on EPA (2003), which, in its final iteration, will be published on the National Emission Inventory (NEI) Air Pollutant Emission Trends web site. Due to lack of data availability, emissions were held constant from 2004 to 2005.

---

[21] U.S. territories data also include combustion from mobile activities because data to allocate territories' energy use were unavailable. For this reason, CH_4 and N_2O emissions from combustion by U.S. territories are only included in the stationary combustion totals.

[22] Though emissions from construction and farm use occur due to both stationary and mobile sources, detailed data was not available to determine the magnitude from each. Currently, these emissions are assumed to be predominantly from mobile sources.

For indirect greenhouse gases, the major source categories included coal, fuel oil, natural gas, wood, other fuels (i.e., bagasse, liquefied petroleum gases, coke, coke oven gas, and others), and stationary internal combustion, which includes emissions from internal combustion engines not used in transportation. EPA periodically estimates emissions of NO_x, CO, and NMVOCs by sector and fuel type using a "bottom-up" estimating procedure. In other words, the emissions were calculated either for individual sources (e.g., industrial boilers) or for many sources combined, using basic activity data (e.g., fuel consumption or deliveries, etc.) as indicators of emissions. The national activity data used to calculate the individual categories were obtained from various sources. Depending upon the category, these activity data may include fuel consumption or deliveries of fuel, tons of refuse burned, raw material processed, etc. Activity data were used in conjunction with emission factors that relate the quantity of emissions to the activity. Table A-70 through Table A-72 present indirect greenhouse gas emission estimates for 1990 through 2005.

The basic calculation procedure for most source categories presented in EPA (2003) and EPA (2005) is represented by the following equation:

$$E_{p,s} = A_s \times EF_{p,s} \times (1 - C_{p,s}/100)$$

Where,

E = Emissions
p = Pollutant
s = Source category
A = Activity level
EF = Emission factor
C = Percent control efficiency

The EPA currently derives the overall emission control efficiency of a category from a variety of sources, including published reports, the 1985 National Acid Precipitation and Assessment Program (NAPAP) emissions inventory, and other EPA databases. The U.S. approach for estimating emissions of NO_x, CO, and NMVOCs from stationary combustion as described above is similar to the methodology recommended by the IPCC (IPCC/UNEP/OECD/IEA 1997).

Table A-68: Fuel Consumption by Stationary Combustion for Calculating CH_4 and N_2O Emissions (TBtu)

Fuel/End-Use Sector	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Coal	**18,056**	**17,984**	**18,162**	**18,933**	**18,996**	**19,144**	**20,088**	**20,534**	**20,751**	**20,779**	**21,767**	**21,165**	**21,231**	**21,654**	**21,817**	**22,187**
Residential	31	25	26	26	21	17	17	16	12	14	11	12	12	12	14	10
Commercial	124	116	117	117	118	117	122	129	93	103	92	97	90	82	101	84
Industrial	1,632	1,586	1,546	1,585	1,586	1,533	1,510	1,474	1,420	1,373	1,433	1,432	1,325	1,331	1,355	1,312
Electric Power	16,261	16,250	16,466	17,196	17,261	17,466	18,429	18,905	19,216	19,279	20,220	19,614	19,783	20,185	20,305	20,737
U.S. Territories	7	8	9	10	10	10	10	10	11	10	10	11	21	44	43	43
Petroleum	**6,968**	**6,553**	**6,660**	**6,312**	**6,354**	**5,674**	**6,258**	**6,142**	**6,014**	**6,095**	**6,285**	**6,946**	**6,186**	**6,907**	**7,033**	**6,919**
Residential	1,382	1,387	1,382	1,358	1,325	1,293	1,436	1,343	1,243	1,377	1,453	1,472	1,366	1,503	1,475	1,369
Commercial	940	892	831	735	738	685	724	661	621	611	696	705	631	752	723	699
Industrial	2,982	2,651	3,012	2,635	2,725	2,480	2,846	2,765	2,398	2,434	2,520	2,860	2,672	2,825	2,961	2,975
Electric Power	1,289	1,198	991	1,124	1,059	755	817	927	1,306	1,211	1,144	1,277	961	1,205	1,212	1,235
U.S. Territories	375	425	445	460	506	462	435	445	445	461	472	632	557	622	662	641
Natural Gas	**18,384**	**18,852**	**19,579**	**20,042**	**20,391**	**21,320**	**21,708**	**21,770**	**21,447**	**21,511**	**22,466**	**21,558**	**22,285**	**21,760**	**21,829**	**21,452**
Residential	4,523	4,697	4,835	5,095	4,988	4,981	5,383	5,118	4,669	4,858	5,126	4,910	5,031	5,247	5,016	4,953
Commercial	2,701	2,813	2,890	2,942	2,979	3,113	3,244	3,302	3,098	3,130	3,265	3,110	3,235	3,284	3,226	3,147
Industrial	7,827	7,943	8,320	8,446	8,424	8,901	9,198	9,204	8,981	8,598	8,746	8,033	8,211	7,938	7,950	7,295
Electric Power	3,332	3,399	3,534	3,560	4,000	4,325	3,883	4,146	4,698	4,926	5,316	5,481	5,785	5,264	5,611	6,033
U.S. Territories	0	0	0	0	0	0	0	0	0	0	13	23	23	27	25	25
Wood	**2,216**	**2,214**	**2,313**	**2,260**	**2,324**	**2,370**	**2,437**	**2,381**	**2,184**	**2,224**	**2,272**	**2,006**	**1,995**	**2,002**	**2,121**	**1,896**
Residential	580	610	640	550	520	520	540	440	380	400	430	370	380	400	410	420
Commercial	66	68	72	76	72	72	76	73	64	67	71	67	69	71	70	70
Industrial	1,442	1,410	1,461	1,484	1,580	1,652	1,683	1,731	1,603	1,620	1,636	1,443	1,396	1,363	1,476	1,238
Electric Power	129	126	140	150	152	125	138	137	137	138	134	126	150	167	165	168
U.S. Territories	NE	NE	NE	NE	NE	NE	NE	NE	NE	NE	NE	NE	NE	NE	NE	NE

NE (Not Estimated)
Note: Totals may not sum due to independent rounding.

Table A-69: CH_4 and N_2O Emission Factors by Fuel Type and Sector (g/GJ)[23]

Fuel/End-Use Sector	CH_4	N_2O
Coal		
Residential	300	1.4
Commercial	10	1.4
Industrial	10	1.4
Electric Power	1	1.4
U.S. Territories	1	1.4
Petroleum		
Residential	10	0.6
Commercial	10	0.6
Industrial	2	0.6
Electric Power	3	0.6
U.S. Territories	5	0.6
Natural Gas		
Residential	5	0.1
Commercial	5	0.1
Industrial	5	0.1
Electric Power	1	0.1
U.S. Territories	1	0.1
Wood		
Residential	300	4.0
Commercial	300	4.0
Industrial	30	4.0
Electric Power	30	4.0
U.S. Territories	NA	NA

NA (Not Applicable)

[23] GJ (Gigajoule) = 10^9 joules. One joule = 9.486×10^{-4} Btu

Table A-70: NO_x Emissions from Stationary Combustion (Gg)

Sector/Fuel Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Electric Power	**6,045**	**5,914**	**5,900**	**6,034**	**5,956**	**5,792**	**5,592**	**5,695**	**5,654**	**5,190**	**4,836**	**4,461**	**4,286**	**3,873**	**3,393**	**3,261**
Coal	5,119	5,043	5,061	5,211	5,113	5,061	5,079	5,118	4,932	4,437	4,130	3,802	3,646	3,295	2,886	2,774
Fuel Oil	200	192	154	163	148	87	107	131	202	179	147	149	143	129	113	108
Natural gas	513	526	526	500	536	510	258	289	346	400	383	332	318	288	252	242
Wood	NA	NA	NA	NA	NA	NA	5	6	24	33	36	37	36	32	28	27
Other Fuels[a]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Internal Combustion	213	152	159	160	159	134	142	150	149	141	140	140	143	130	114	109
Industrial	**2,753**	**2,703**	**2,786**	**2,859**	**2,855**	**2,852**	**2,859**	**2,812**	**2,767**	**2,458**	**2,470**	**2,499**	**1,881**	**1,885**	**1,888**	**1,891**
Coal	530	517	521	534	546	541	490	487	475	475	484	518	390	391	391	392
Fuel Oil	240	215	222	222	219	224	203	196	190	190	166	153	116	116	116	116
Natural gas	1,072	1,134	1,180	1,207	1,210	1,201	1,092	1,078	1,066	880	902	914	688	689	690	692
Wood	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Other Fuels[a]	119	117	115	113	113	111	109	103	104	100	109	116	87	87	87	87
Internal Combustion	792	720	749	783	767	774	965	948	932	813	809	798	601	602	602	603
Commercial	**336**	**333**	**348**	**360**	**365**	**365**	**360**	**369**	**347**	**255**	**256**	**261**	**247**	**247**	**248**	**248**
Coal	36	33	35	37	36	35	30	32	34	23	21	21	19	19	19	20
Fuel Oil	88	80	84	84	86	94	86	88	73	54	52	52	50	50	50	50
Natural gas	181	191	204	211	215	210	224	229	220	156	161	165	157	157	157	157
Wood	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Other Fuels[a]	31	29	25	28	28	27	20	21	21	22	22	23	21	21	21	21
Residential	**749**	**829**	**879**	**827**	**817**	**813**	**726**	**699**	**651**	**441**	**439**	**446**	**422**	**423**	**424**	**424**
Coal[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Fuel Oil[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Natural Gas[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Wood	42	45	48	40	40	44	27	27	27	25	21	22	21	21	21	21
Other Fuels[a]	707	784	831	786	778	769	699	671	624	416	417	424	402	402	403	404
Total	**9,883**	**9,779**	**9,914**	**10,079**	**9,993**	**9,821**	**9,537**	**9,575**	**9,419**	**8,344**	**8,002**	**7,667**	**6,837**	**6,428**	**5,952**	**5,824**

NA (Not Applicable)

[a] "Other Fuels" include LPG, waste oil, coke oven gas, coke, and non-residential wood (EPA 2003, 2005).

[b] Residential coal, fuel oil, and natural gas emissions are included in the "Other Fuels" category (EPA 2003, 2005).

Note: Totals may not sum due to independent rounding.

Table A-71: CO Emissions from Stationary Combustion (Gg)

Sector/Fuel Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Electric Power	**329**	**317**	**318**	**329**	**336**	**337**	**369**	**384**	**410**	**450**	**439**	**439**	**598**	**610**	**622**	**634**
Coal	213	212	214	224	224	227	228	233	220	187	221	220	300	306	312	318
Fuel Oil	18	17	14	15	13	9	11	13	17	36	27	28	38	39	39	40
Natural gas	46	46	47	45	48	49	72	76	88	151	96	92	125	128	130	133
Wood	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Other Fuels[a]	NA	NA	NA	NA	NA	NA	7	8	30	24	31	32	44	45	46	47
Internal Combustion	52	41	43	46	50	52	52	54	55	52	63	67	91	93	95	97
Industrial	**797**	**835**	**866**	**946**	**944**	**958**	**1,078**	**1,055**	**1,044**	**1,100**	**1,106**	**1,137**	**1,163**	**1,177**	**1,192**	**1,206**
Coal	95	92	92	92	91	88	100	99	96	114	118	125	128	130	131	133
Fuel Oil	67	54	58	60	60	64	49	47	46	54	48	45	46	47	47	48
Natural gas	205	257	272	292	306	313	307	307	305	350	355	366	374	379	383	388
Wood	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Other Fuels[a]	253	242	239	259	260	270	316	302	303	286	300	321	329	333	337	341
Internal Combustion	177	189	205	244	228	222	305	299	294	296	285	279	286	289	293	296
Commercial	**205**	**196**	**204**	**207**	**212**	**211**	**122**	**126**	**122**	**151**	**151**	**154**	**190**	**192**	**194**	**197**
Coal	13	13	13	14	13	14	13	13	14	16	14	13	17	17	17	17
Fuel Oil	16	16	16	16	16	17	17	18	15	17	17	17	21	22	22	22
Natural gas	40	40	46	48	49	49	58	59	57	81	83	84	104	106	107	108
Wood	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Other Fuels[a]	136	128	128	129	134	132	34	36	36	36	36	38	47	48	49	49
Residential	**3,668**	**3,965**	**4,195**	**3,586**	**3,515**	**3,877**	**2,364**	**2,362**	**2,353**	**3,323**	**2,644**	**2,648**	**3,274**	**3,313**	**3,353**	**3,394**
Coal[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Fuel Oil[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Natural Gas[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Wood	3,430	3,710	3,930	3,337	3,272	3,629	2,132	2,133	2,133	3,094	2,416	2,424	2,997	3,033	3,070	3,107
Other Fuels[a]	238	255	265	249	243	248	231	229	220	229	228	224	276	280	283	287
Total	**5,000**	**5,313**	**5,583**	**5,069**	**5,007**	**5,383**	**3,933**	**3,927**	**3,928**	**5,024**	**4,340**	**4,377**	**5,224**	**5,292**	**5,361**	**5,431**

NA (Not Applicable)

[a] "Other Fuels" include LPG, waste oil, coke oven gas, coke, and non-residential wood (EPA 2003, 2005).

[b] Residential coal, fuel oil, and natural gas emissions are included in the "Other Fuels" category (EPA 2003, 2005).

Note: Totals may not sum due to independent rounding.

Table A-72: NMVOC Emissions from Stationary Combustion (Gg)

Sector/Fuel Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Electric Power	**43**	**40**	**40**	**41**	**41**	**40**	**44**	**47**	**51**	**49**	**56**	**55**	**45**	**44**	**43**	**42**
Coal	24	24	25	26	26	26	25	26	26	25	27	26	21	21	20	20
Fuel Oil	5	5	4	4	4	2	3	4	5	4	4	4	4	3	3	3
Natural gas	2	2	2	2	2	2	7	7	9	9	12	12	10	10	10	9
Wood	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Other Fuels[a]	NA	NA	NA	NA	NA	NA	+	+	1	2	2	2	1	1	1	1
Internal Combustion	11	9	9	9	9	9	9	10	10	10	11	10	9	8	8	8
Industrial	**165**	**178**	**170**	**169**	**178**	**187**	**163**	**160**	**159**	**156**	**157**	**159**	**138**	**138**	**139**	**139**
Coal	7	5	7	5	7	5	6	6	6	9	9	10	9	9	9	9
Fuel Oil	11	10	11	11	11	11	8	7	7	10	9	9	7	7	7	7
Natural gas	52	54	48	46	57	66	54	54	53	52	53	54	47	47	47	47
Wood	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Other Fuels[a]	46	47	45	46	45	45	33	31	31	26	27	29	26	26	26	26
Internal Combustion	49	61	60	60	58	60	63	62	61	60	58	57	49	49	50	50
Commercial	**18**	**18**	**20**	**22**	**21**	**21**	**22**	**22**	**21**	**25**	**28**	**29**	**46**	**46**	**46**	**46**
Coal	1	1	1	1	1	1	1	1	1	1	1	1	2	2	2	2
Fuel Oil	3	2	3	3	3	3	3	3	3	3	4	4	7	7	7	7
Natural gas	7	8	9	10	10	10	13	13	12	11	14	14	26	26	26	26
Wood	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Other Fuels[a]	8	7	7	8	8	8	5	5	5	10	9	10	11	11	11	11
Residential	**686**	**739**	**782**	**670**	**657**	**725**	**789**	**788**	**786**	**815**	**837**	**836**	**1,504**	**1,506**	**1,508**	**1,510**
Coal[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Fuel Oil[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Natural Gas[b]	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA	NA
Wood	651	704	746	633	621	688	756	756	756	794	809	809	1,455	1,457	1,459	1,461
Other Fuels[a]	35	35	36	36	36	37	33	32	30	21	27	27	49	49	49	49
Total	**912**	**975**	**1,012**	**901**	**897**	**973**	**1,018**	**1,017**	**1,017**	**1,045**	**1,077**	**1,080**	**1,733**	**1,734**	**1,735**	**1,736**

NA (Not Applicable)

[a] "Other Fuels" include LPG, waste oil, coke oven gas, coke, and non-residential wood (EPA 2003, 2005).

[b] Residential coal, fuel oil, and natural gas emissions are included in the "Other Fuels" category (EPA 2003, 2005).

Note: Totals may not sum due to independent rounding.

3.2. Methodology for Estimating Emissions of CH_4, N_2O, and Indirect Greenhouse Gases from Mobile Combustion and Methodology for and Supplemental Information on Transportation-Related GHG Emissions

Estimates of CH_4 and N_2O Emissions

Mobile source emissions of greenhouse gases other than CO_2 are reported by transport mode (e.g., road, rail, aviation, and waterborne), vehicle type, and fuel type. Emissions estimates of CH_4 and N_2O were derived using a methodology similar to that outlined in the *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997).

Activity data were obtained from a number of U.S. government agencies and other publications. Depending on the category, these basic activity data included such information as fuel consumption and vehicle miles traveled (VMT). These estimates were then multiplied by emission factors, expressed as grams per unit of fuel consumed or per vehicle mile.

Methodology for Highway Gasoline and Diesel Vehicles

Step 1: Determine Vehicle Miles Traveled by Vehicle Type, Fuel Type, and Model Year

VMT by vehicle type (e.g., passenger cars, light-duty trucks, heavy-duty trucks,[24] buses, and motorcycles) were obtained from the Federal Highway Administration's (FHWA) *Highway Statistics* (FHWA 1996 through 2006). As these vehicle categories are not fuel-specific, VMT for each vehicle type was disaggregated by fuel type (gasoline, diesel) so that the appropriate emission factors could be applied. VMT from *Highway Statistics* Table VM-1 (FHWA 1996 through 2006) was allocated to fuel types (gasoline, diesel, other) using historical estimates of fuel shares reported in the Appendix to the *Transportation Energy Data Book* (DOE 2006). These fuel shares are drawn from various sources, including the Vehicle Inventory and Use Survey, the National Vehicle Population Profile, and the American Public Transportation Association. The fuel shares were first adjusted proportionately so that the gasoline and diesel shares for each vehicle type summed to 100 percent in order to develop an interim estimate of VMT for each vehicle/fuel type category that summed to the total national VMT estimate. VMT for alternative fuel vehicles (AFVs) was calculated separately, and the methodology is explained in the following section on AFVs. Estimates of VMT from AFVs were then subtracted from the appropriate interim VMT estimates to develop the final VMT estimates by vehicle/fuel type category.[25] The resulting national VMT estimates for gasoline and diesel highway vehicles are presented in Table A- 73 and Table A- 74, respectively.

Total VMT for each highway category (i.e., gasoline passenger cars, light-duty gasoline trucks, heavy-duty gasoline vehicles, diesel passenger cars, light-duty diesel trucks, heavy-duty diesel vehicles, and motorcycles) were distributed across 31 model years shown in Table A- 86 through Table A- 92. This distribution was derived by weighting the appropriate age distribution of the U.S. vehicle fleet according to vehicle registrations (Table A- 77 through Table A- 83) by the average annual age-specific vehicle mileage accumulation of U.S. vehicles). Age distribution values were obtained from EPA's MOBILE6 model for all years before 1999 (EPA 2000) and EPA's MOVES model for years 1999 forward (EPA 2006e).[26] Age-specific vehicle mileage accumulation was obtained from EPA's MOBILE6 model (EPA 2000).

---

[24] The category "heavy-duty trucks" includes vehicles that are sometimes classified as medium-duty trucks (those with a GVWR between 8,500 and 14,000 lbs.). The only exception is **Error! Reference source not found.**, which provides VMT data for medium-duty alternative fuel vehicles.

[25] In Inventories through 2002, gasoline-electric hybrid vehicles were considered part of an "alternative fuel and advanced technology" category. However, vehicles are now only separated into gasoline, diesel, or alternative fuel categories, and gas-electric hybrids are now considered within the gasoline vehicle category.

[26] Age distributions were held constant for the period 1990-1998, and reflect a 25-year vehicle age span. EPA (2006e) provides a variable age distribution and 31-year vehicle age span beginning in year 1999.

Step 2: Allocate VMT Data to Control Technology Type

VMT by vehicle type for each model year was distributed across various control technologies as shown in Table A- 95 through Table A- 98. The categories "EPA Tier 0" and "EPA Tier 1" were used instead of the early three-way catalyst and advanced three-way catalyst categories, respectively, as defined in the *Revised 1996 IPCC Guidelines*. EPA Tier 0, EPA Tier 1, Tier 2, and LEV refer to U.S. emission regulations, rather than control technologies; however, each does correspond to particular combinations of control technologies and engine design. EPA Tier 2 and its predecessors EPA Tier 1 and Tier 0 apply to vehicles equipped with three-way catalysts. The introduction of "early three-way catalysts," and "advanced three-way catalysts," as described in the *Revised 1996 IPCC Guidelines*, roughly correspond to the introduction of EPA Tier 0 and EPA Tier 1 regulations (EPA 1998).[27] EPA Tier 2 regulations affect vehicles produced starting in 2004 and represent a significant decrease in emissions over EPA Tier 1 emissions (EPA 1999b).

Control technology assignments for light and heavy-duty conventional fuel vehicles for model years 1972 (when regulations began to take effect) through 1995 were estimated in EPA (1998). Assignments for 1998 through 2004 were determined using confidential engine family sales data submitted to EPA (EPA 2006b). Vehicle classes and emission standard tiers to which each engine family was certified were taken from annual certification test results and data (EPA 2006a). This information was used to determine the fraction of sales of each class of vehicle that met EPA Tier 0, EPA Tier 1, Tier 2, and LEV standards. Assignments for 1996 and 1997 were estimated based on the fact that EPA Tier 1 standards for light-duty vehicles were fully phased in by 1996. Tier 2 began initial phase-in by 2004.

Step 3: Determine CH_4 and N_2O Emission Factors by Vehicle, Fuel, and Control Technology Type

Emission factors for gasoline and diesel highway vehicles were developed by ICF (2004). These factors were based on EPA and CARB laboratory test results of different vehicle and control technology types. The EPA and CARB tests were designed following the Federal Test Procedure (FTP), which covers three separate driving segments, since vehicles emit varying amounts of GHGs depending on the driving segment. These driving segments are: (1) a transient driving cycle that includes cold start and running emissions, (2) a cycle that represents running emissions only, and (3) a transient driving cycle that includes hot start and running emissions. For each test run, a bag was affixed to the tailpipe of the vehicle and the exhaust was collected; the content of this bag was later analyzed to determine quantities of gases present. The emission characteristics of segment 2 was used to define running emissions, and subtracted from the total FTP emissions to determine start emissions. These were then recombined based upon MOBILE6.2's ratio of start to running emissions for each vehicle class to approximate average driving characteristics.

Step 4: Determine the Amount of CH_4 and N_2O Emitted by Vehicle, Fuel, and Control Technology Type

Emissions of CH_4 and N_2O were then calculated by multiplying total VMT by vehicle, fuel, and control technology type by the emission factors developed in Step 3.

Methodology for Alternative Fuel Vehicles (AFVs)

Step 1: Determine Vehicle Miles Traveled by Vehicle and Fuel Type

VMT for alternative fuel and advanced technology vehicles were calculated from "VMT Projections for Alternative Fueled and Advanced Technology Vehicles through 2025" (Browning 2003). Alternative Fuels include Compressed Natural Gas (CNG), Liquid Natural Gas (LNG), Liquefied Petroleum Gas (LPG), Ethanol, Methanol, and Electric Vehicles (battery powered). Most of the vehicles that use these fuels run on an Internal Combustion Engine (ICE) powered by the alternative fuel, although many of the vehicles can run on either the alternative fuel or gasoline (or diesel), or some combination.[28] The data obtained include vehicle fuel use and total number of vehicles

[27] For further description, see "Definitions of Emission Control Technologies and Standards" section of this annex.

[28] Fuel types used in combination depend on the vehicle class. For light-duty vehicles, gasoline is generally blended with ethanol or methanol; some vehicles are also designed to run on gasoline or an alternative fuel – either natural gas or LPG –

in use from 1992 through 2005. Because AFVs run on different fuel types, their fuel use characteristics are not directly comparable. Accordingly, fuel economy for each vehicle type is expressed in gasoline equivalent terms, i.e., how much gasoline contains the equivalent amount of energy as the alternative fuel. Energy economy ratios (the ratio of the gasoline equivalent fuel economy of a given technology to that of conventional gasoline or diesel vehicles) were taken from full fuel cycle studies done for the California Air Resources Board (Unnasch and Browning, 2000). These ratios were used to estimate fuel economy in miles per gasoline gallon equivalent for each alternative fuel and vehicle type. Energy use per fuel type was then divided among the various weight categories and vehicle technologies that use that fuel. Total VMT per vehicle type for each calendar year was then determined by dividing the energy usage by the fuel economy. Note that for AFVs capable of running on both/either traditional and alternative fuels, the VMT given reflects only those miles driven that were powered by the alternative fuel, as explained in Browning (2003). VMT estimates for AFVs by vehicle category (passenger car, light-duty truck, heavy-duty vehicles) are shown in Table A- 75, while more detailed estimates of VMT by control technology are shown in Table A- 76.

Step 2: Determine CH_4 and N_2O Emission Factors by Vehicle and Alternative Fuel Type

CH_4 and N_2O emission factors for alternative fuel vehicles (AFVs) are calculated according to studies by Argonne National Laboratory (2006) and Lipman & Delucchi (2002), and are reported in ICF (2006a). In these studies, N_2O and CH_4 emissions for AFVs were expressed as a multiplier corresponding to conventional vehicle counterpart emissions. Emission estimates in these studies represent the current AFV fleet and were compared against Tier 1 emissions from light-duty gasoline vehicles to develop new multipliers. Alternative fuel heavy-duty vehicles were compared against gasoline heavy-duty vehicles as most alternative fuel heavy-duty vehicles use catalytic after treatment and perform more like gasoline vehicles than diesel vehicles. These emission factors are shown in Table A- 100.

Step 3: Determine the Amount of CH_4 and N_2O Emitted by Vehicle and Fuel Type

Emissions of CH_4 and N_2O were calculated by multiplying total VMT for each vehicle and fuel type (Step 1) by the appropriate emission factors (Step 2).

Methodology for Non-Highway Mobile Sources

CH_4 and N_2O emissions from non-highway mobile sources were estimated by applying emission factors to the amount of fuel consumed by mode and vehicle type.

Activity data for non-highway vehicles include annual fuel consumption statistics by transportation mode and fuel type, as shown in Table A- 94. Consumption data for ships and boats (i.e., vessel bunkering) were obtained from EIA (1991 through 2006) for distillate fuel, and EIA (2006a) for residual fuel; marine transport fuel consumption data for U.S. territories (EIA 2002b, EIA 2003 through 2004) were added to domestic consumption, and this total was reduced by the amount of fuel used for international bunkers.[29] Gasoline consumption by recreational boats was obtained from EPA's NONROAD model (EPA 2006d). Annual diesel consumption for Class I railroad locomotives was obtained from AAR (2006), diesel consumption from commuter rail totals was obtained from APTA (2006), and consumption by Class II and III railroad locomotives was provided by Benson (2004) and Whorton (2006). Diesel consumption by commuter and intercity rail was obtained from DOE (1993 through 2005). Data on the consumption of jet fuel and aviation gasoline in aircraft were obtained from EIA (2006a), as described in Annex 2.1: Methodology for Estimating Emissions of CO_2 from Fossil Fuel Combustion, and were reduced by the amount allocated to international bunker fuels. Pipeline fuel consumption was obtained from EIA (2004) (note: pipelines are a transportation source but are stationary, not mobile, sources). Data on fuel consumption by all non-

but not at the same time, while other vehicles are designed to run on E85 (85% ethanol) or gasoline, or any mixture of the two. Heavy-duty vehicles are more likely to run on a combination of diesel fuel and either natural gas, LPG, ethanol, or methanol.

[29] See International Bunker Fuels section of the Energy Chapter.

transportation mobile sources[30] were obtained from EPA's NONROAD model (EPA 2006d). Finally, gasoline consumption for trucks used off-road was taken from FHWA (1996 through 2006).

Emissions of CH_4 and N_2O from non-highway mobile sources were calculated by multiplying U.S. default emission factors in the *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997) by activity data for each source type (see Table A- 101). Table A- 102 and Table A- 103 provide complete emissions of CH_4 and N_2O emissions, respectively, for 1990 through 2005.

Estimates of NO_x, CO, and NMVOC Emissions

The emission estimates of NO_x, CO, and NMVOCs for mobile combustion were obtained from preliminary data (EPA (2006c), EPA (2005) and EPA (2003)), which, in final iteration, will be published on the EPA's National Emission Inventory (NEI) Air Pollutant Emission Trends web site. This EPA report provides emission estimates for these gases by fuel type using a procedure whereby emissions were calculated using basic activity data, such as amount of fuel delivered or miles traveled, as indicators of emissions.

Table A- 104 through Table A- 106 provide complete emissions estimates for 1990 through 2005.

Table A- 73: Vehicle Miles Traveled for Gasoline Highway Vehicles (10^9 Miles)

Year	Passenger Cars	Light-Duty Trucks	Heavy-Duty Vehicles	Motorcycles
1990	1,391.3	554.1	25.7	9.6
1991	1,341.8	627.1	25.2	9.2
1992	1,355.0	682.8	25.0	9.6
1993	1,356.7	720.2	24.7	9.9
1994	1,387.6	738.5	25.1	10.2
1995	1,420.8	762.2	24.9	9.8
1996	1,454.9	787.8	24.3	9.9
1997	1,488.8	820.8	23.9	10.1
1998	1,536.9	838.5	23.9	10.3
1999	1,559.3	870.1	24.1	10.6
2000	1,591.9	891.3	23.9	10.5
2001	1,619.7	911.6	23.7	9.6
2002	1,649.6	933.6	23.6	9.6
2003	1,663.1	951.0	24.0	9.6
2004	1,690.8	992.7	24.3	10.1
2005	1,680.9	1,024.0	24.5	10.8

Source: Derived from FHWA (1996 through 2006).

Table A- 74: Vehicle Miles Traveled for Diesel Highway Vehicles (10^9 Miles)

Year	Passenger Cars	Light-Duty Trucks	Heavy-Duty Vehicles
1990	16.9	19.7	125.4
1991	16.3	21.6	129.2
1992	16.5	23.4	133.3
1993	17.9	24.7	140.2
1994	18.3	25.3	150.4
1995	17.3	26.9	158.7
1996	14.7	27.8	164.2
1997	13.5	29.0	173.2
1998	12.4	28.7	178.3
1999	9.4	29.8	184.9
2000	8.0	30.5	187.7
2001	8.1	30.3	190.7
2002	8.3	31.0	196.0
2003	8.4	31.5	198.9

[30] "Non-transportation mobile sources" are defined as any vehicle or equipment not used on the traditional road system, but excluding aircraft, rail and watercraft. This category includes snowmobiles, golf carts, riding lawn mowers, agricultural equipment, and trucks used for off-road purposes, among others.

2004	8.5	32.9	201.4
2005	8.4	34.0	203.0

Source: Derived from FHWA (1996 through 2006).

Table A-75: Vehicle Miles Traveled for Alternative Fuel Highway Vehicles (10^9 Miles)

Year	Passenger Cars	Light-Duty Trucks	Heavy-Duty Vehicles
1990	0.1	0.8	0.9
1991	0.1	0.8	0.9
1992	0.1	0.7	0.8
1993	0.1	0.9	1.1
1994	0.2	0.9	1.0
1995	0.2	0.9	1.0
1996	0.2	0.9	1.1
1997	0.3	1.0	1.2
1998	0.3	1.1	1.3
1999	0.4	1.2	1.3
2000	0.4	1.3	1.5
2001	0.5	1.4	1.8
2002	0.6	1.5	1.8
2003	0.6	1.5	1.8
2004	0.6	1.6	1.9
2005	0.6	1.6	1.9

Source: Derived from Browning (2003).

Table A-76: Detailed Vehicle Miles Traveled for Alternative Fuel Highway Vehicles (10^6 Miles)

Vehicle Type	1990	1995	2000	2001	2002	2003	2004	2005
Light-Duty Cars	**67.4**	**189.8**	**408.7**	**525.7**	**562.1**	**583.0**	**608.5**	**635.6**
Methanol-Flex Fuel ICE	0.0	44.2	14.2	10.8	8.2	0.0	0.0	0.0
Ethanol-Flex Fuel ICE	0.0	0.8	32.4	40.5	47.2	58.6	68.6	78.8
CNG ICE	7.5	25.9	76.4	100.5	106.5	112.1	115.6	118.2
CNG Bi-fuel	15.9	61.4	175.9	232.9	244.9	249.4	254.3	260.4
LPG ICE	5.0	5.2	6.4	6.7	7.0	7.3	8.0	8.8
LPG Bi-fuel	38.9	39.7	47.0	48.0	50.4	52.9	55.8	58.8
NEVs	0.0	11.4	50.9	77.9	88.1	89.6	86.8	82.5
Electric Vehicle	0.0	1.2	5.5	8.4	9.8	13.2	19.4	28.2
Light-Duty Trucks	**845.9**	**851.8**	**1,271.3**	**1,384.8**	**1,471.1**	**1,521.9**	**1,571.6**	**1,592.4**
Ethanol-Flex Fuel ICE	0.0	2.5	104.2	130.3	152.0	188.6	222.2	255.3
CNG ICE	7.0	30.5	100.4	124.1	136.9	141.0	143.2	145.8
CNG Bi-fuel	15.8	45.1	174.6	215.3	237.8	216.8	223.2	220.6
LPG ICE	18.8	18.1	20.7	21.1	22.0	22.9	24.2	25.3
LPG Bi-fuel	804.3	753.5	861.6	879.6	905.4	931.5	934.8	918.9
Electric Vehicle	0.0	2.1	9.7	14.4	17.0	21.0	24.0	26.4
Medium Duty Trucks	**192.9**	**179.2**	**221.3**	**251.9**	**259.6**	**266.9**	**275.4**	**284.1**
CNG Bi-fuel	1.5	4.3	10.5	11.9	12.7	13.0	13.3	13.6
LPG ICE	16.4	15.6	19.6	22.4	23.0	23.5	24.3	25.1
LPG Bi-fuel	174.9	159.3	191.2	217.6	223.9	230.4	237.8	245.4
Heavy-Duty Trucks	**632.7**	**726.3**	**997.2**	**1,175.8**	**1,206.9**	**1,233.7**	**1,265.4**	**1,298.1**
Neat Methanol ICE	0.0	7.5	0.0	0.0	0.0	0.0	0.0	0.0
Neat Ethanol ICE	0.0	2.9	0.1	0.0	0.0	0.0	0.0	0.0
CNG ICE	14.2	51.2	139.2	176.4	174.8	180.0	186.3	192.7
LPG ICE	522.0	575.2	726.1	838.5	860.8	876.9	895.3	913.7
LPG Bi-fuel	96.5	83.9	114.1	136.2	142.7	144.3	147.4	150.5
LNG	0.0	5.6	17.7	24.8	28.5	32.5	36.5	41.3
Buses	**90.5**	**122.7**	**269.9**	**327.7**	**332.4**	**338.0**	**344.6**	**350.9**
Neat Methanol ICE	3.7	3.8	1.8	1.6	0.0	0.0	0.0	0.0
Neat Ethanol ICE	0.1	1.7	0.0	0.0	0.0	0.0	0.0	0.0
CNG ICE	17.1	44.3	160.3	195.2	195.1	197.0	198.9	200.9
LPG ICE	69.7	67.9	92.2	108.9	110.0	111.1	112.2	113.3
LNG	0.0	5.0	15.4	21.7	27.0	29.4	33.0	36.2

Electric	0.0	0.1	0.2	0.3	0.4	0.4	0.5	0.6

Source: Derived from Browning (2003).
[a] Throughout the rest of this Inventory, medium-duty trucks are grouped with heavy-duty trucks; they are reported separately here because these two categories may run on a slightly different range of fuel types.

Table A- 77: Age Distribution by Vehicle/Fuel Type for Highway Vehicles,[a] 1990 to 1998

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC[c]
0	5.3%	5.8%	4.9%	5.3%	5.9%	4.2%	14.4%
1	7.1%	7.6%	8.9%	7.1%	7.4%	7.8%	16.8%
2	7.1%	7.5%	8.1%	7.1%	6.9%	7.2%	13.5%
3	7.1%	7.3%	7.4%	7.1%	6.4%	6.7%	10.9%
4	7.0%	7.1%	6.8%	7.0%	6.0%	6.2%	8.8%
5	7.0%	6.8%	6.2%	7.0%	5.6%	5.8%	7.0%
6	6.9%	6.5%	5.6%	6.9%	5.2%	5.3%	5.6%
7	6.8%	6.1%	5.1%	6.8%	4.8%	5.0%	4.5%
8	6.6%	5.7%	4.7%	6.6%	4.5%	4.6%	3.6%
9	6.3%	5.2%	4.3%	6.3%	4.2%	4.3%	2.9%
10	5.9%	4.7%	3.9%	5.9%	3.9%	4.0%	2.3%
11	5.4%	4.2%	3.6%	5.4%	3.6%	3.7%	9.7%
12	4.6%	3.6%	3.3%	4.6%	3.4%	3.4%	0.0%
13	3.6%	3.1%	3.0%	3.6%	3.2%	3.2%	0.0%
14	2.9%	2.6%	2.7%	2.9%	2.9%	2.9%	0.0%
15	2.3%	2.2%	2.5%	2.3%	2.7%	2.7%	0.0%
16	1.8%	1.8%	2.3%	1.8%	2.5%	2.5%	0.0%
17	1.4%	1.4%	2.1%	1.4%	2.4%	2.4%	0.0%
18	1.1%	1.2%	1.9%	1.1%	2.2%	2.2%	0.0%
19	0.9%	1.1%	1.7%	0.9%	2.1%	2.0%	0.0%
20	0.7%	1.1%	1.6%	0.7%	1.9%	1.9%	0.0%
21	0.6%	1.0%	1.5%	0.6%	1.8%	1.8%	0.0%
22	0.4%	1.0%	1.3%	0.4%	1.7%	1.6%	0.0%
23	0.4%	0.9%	1.2%	0.4%	1.6%	1.5%	0.0%
24	1.0%	4.6%	5.4%	1.0%	7.3%	7.2%	0.0%
Total	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Source: EPA (2000).
[a] The following abbreviations correspond to vehicle types: LDGV (light-duty gasoline vehicles), LDGT (light-duty gasoline trucks), HDGV (heavy-duty gasoline vehicles),LDDV (light-duty diesel vehicles), LDDT (light-duty diesel trucks), HDDV (heavy-duty diesel vehicles), and MC (motorcycles).
[b] Because of a lack of data, all motorcycles 11 of age or older are considered to have the same emissions and travel characteristics, and therefore are presented in aggregate.

Table A- 78: Age Distribution by Vehicle/Fuel Type for Highway Vehicles,[a] 1999

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	8.1%	7.7%	8.1%	8.1%	7.7%	8.1%	9.5%
1	6.1%	7.0%	6.5%	6.1%	7.0%	6.5%	9.3%
2	6.2%	6.5%	5.8%	6.2%	6.5%	5.8%	7.5%
3	5.9%	7.6%	5.6%	5.9%	7.6%	5.6%	6.8%
4	6.9%	8.0%	6.7%	6.9%	8.0%	6.7%	6.1%
5	6.1%	7.6%	6.0%	6.1%	7.6%	6.0%	5.7%
6	6.1%	6.6%	4.8%	6.1%	6.6%	4.8%	5.2%
7	5.6%	5.1%	3.6%	5.6%	5.1%	3.6%	4.3%
8	5.8%	5.0%	3.7%	5.8%	5.0%	3.7%	3.7%
9	5.8%	4.7%	4.7%	5.8%	4.7%	4.7%	3.5%
10	6.1%	5.2%	5.5%	6.1%	5.2%	5.5%	3.4%
11	5.9%	5.1%	5.2%	5.9%	5.1%	5.2%	3.9%
12	5.3%	5.2%	4.5%	5.3%	5.2%	4.5%	4.6%
13	4.8%	5.0%	4.9%	4.8%	5.0%	4.9%	4.2%
14	3.9%	3.2%	4.1%	3.9%	3.2%	4.1%	3.8%
15	3.1%	2.5%	2.9%	3.1%	2.5%	2.9%	3.4%
16	1.9%	1.6%	3.0%	1.9%	1.6%	3.0%	3.1%
17	1.2%	0.9%	1.8%	1.2%	0.9%	1.8%	2.7%
18	1.0%	0.8%	1.7%	1.0%	0.8%	1.7%	2.3%
19	0.9%	0.6%	1.7%	0.9%	0.6%	1.7%	2.0%

20	1.0%	0.7%	1.4%	1.0%	0.7%	1.4%	1.6%
21	0.8%	1.0%	1.7%	0.8%	1.0%	1.7%	1.3%
22	0.6%	0.5%	1.0%	0.6%	0.5%	1.0%	0.9%
23	0.4%	0.5%	0.9%	0.4%	0.5%	0.9%	0.6%
24	0.2%	0.1%	0.8%	0.2%	0.1%	0.8%	0.3%
25	0.3%	0.1%	0.7%	0.3%	0.1%	0.7%	0.0%
26	0.0%	0.3%	0.8%	0.0%	0.3%	0.8%	0.0%
27	0.0%	0.3%	1.5%	0.0%	0.3%	1.5%	0.0%
28	0.0%	0.1%	0.4%	0.0%	0.1%	0.4%	0.0%
29	0.0%	0.2%	0.2%	0.0%	0.2%	0.2%	0.0%
30	0.0%	0.1%	0.1%	0.0%	0.1%	0.1%	0.0%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Source: EPA (2006e).

[a] The following abbreviations correspond to vehicle types: LDGV (light-duty gasoline vehicles), LDGT (light-duty gasoline trucks), HDGV (heavy-duty gasoline vehicles), LDDV (light-duty diesel vehicles), LDDT (light-duty diesel trucks), HDDV (heavy-duty diesel vehicles), and MC (motorcycles).

Table A- 79: Age Distribution by Vehicle/Fuel Type for Highway Vehicles,[a] 2000

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	7.9%	7.7%	7.1%	7.9%	7.7%	7.1%	9.4%
1	7.8%	7.4%	7.8%	7.8%	7.4%	7.8%	9.1%
2	5.8%	6.7%	6.3%	5.8%	6.7%	6.3%	9.0%
3	6.0%	6.3%	5.6%	6.0%	6.3%	5.6%	7.2%
4	5.6%	7.3%	5.4%	5.6%	7.3%	5.4%	6.4%
5	6.6%	7.6%	6.3%	6.6%	7.6%	6.3%	5.7%
6	5.8%	7.1%	5.7%	5.8%	7.1%	5.7%	5.3%
7	5.8%	6.1%	4.5%	5.8%	6.1%	4.5%	4.8%
8	5.2%	4.7%	3.4%	5.2%	4.7%	3.4%	4.0%
9	5.3%	4.6%	3.5%	5.3%	4.6%	3.5%	3.3%
10	5.3%	4.3%	4.4%	5.3%	4.3%	4.4%	3.2%
11	5.5%	4.7%	5.0%	5.5%	4.7%	5.0%	3.0%
12	5.3%	4.6%	4.8%	5.3%	4.6%	4.8%	3.4%
13	4.8%	4.6%	4.1%	4.8%	4.6%	4.1%	4.0%
14	4.2%	4.5%	4.5%	4.2%	4.5%	4.5%	3.6%
15	3.5%	2.8%	3.7%	3.5%	2.8%	3.7%	3.3%
16	2.7%	2.2%	2.6%	2.7%	2.2%	2.6%	2.9%
17	1.6%	1.4%	2.7%	1.6%	1.4%	2.7%	2.6%
18	1.1%	0.7%	1.6%	1.1%	0.7%	1.6%	2.3%
19	0.9%	0.7%	1.5%	0.9%	0.7%	1.5%	1.9%
20	0.7%	0.5%	1.5%	0.7%	0.5%	1.5%	1.6%
21	0.9%	0.6%	1.3%	0.9%	0.6%	1.3%	1.3%
22	0.7%	0.9%	1.5%	0.7%	0.9%	1.5%	1.0%
23	0.5%	0.4%	0.9%	0.5%	0.4%	0.9%	0.8%
24	0.3%	0.4%	0.8%	0.3%	0.4%	0.8%	0.5%
25	0.2%	0.1%	0.7%	0.2%	0.1%	0.7%	0.2%
26	0.2%	0.1%	0.6%	0.2%	0.1%	0.6%	0.0%
27	0.0%	0.2%	0.7%	0.0%	0.2%	0.7%	0.0%
28	0.0%	0.2%	1.3%	0.0%	0.2%	1.3%	0.0%
29	0.0%	0.1%	0.3%	0.0%	0.1%	0.3%	0.0%
30	0.0%	0.2%	0.2%	0.0%	0.2%	0.2%	0.0%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Source: EPA (2006e).

[a] The following abbreviations correspond to vehicle types: LDGV (light-duty gasoline vehicles), LDGT (light-duty gasoline trucks), HDGV (heavy-duty gasoline vehicles), LDDV (light-duty diesel vehicles), LDDT (light-duty diesel trucks), HDDV (heavy-duty diesel vehicles), and MC (motorcycles).

Table A- 80: Age Distribution by Vehicle/Fuel Type for Highway Vehicles,[a] 2001

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	7.2%	7.6%	5.5%	7.2%	7.6%	5.5%	8.7%
1	7.6%	7.4%	7.0%	7.6%	7.4%	7.0%	9.1%
2	7.5%	7.2%	7.6%	7.5%	7.2%	7.6%	8.8%
3	5.6%	6.5%	6.1%	5.6%	6.5%	6.1%	8.6%

4	5.7%	6.0%	5.5%	5.7%	6.0%	5.5%	6.8%
5	5.4%	7.0%	5.2%	5.4%	7.0%	5.2%	6.0%
6	6.3%	7.2%	6.1%	6.3%	7.2%	6.1%	5.4%
7	5.5%	6.7%	5.4%	5.5%	6.7%	5.4%	4.9%
8	5.4%	5.7%	4.3%	5.4%	5.7%	4.3%	4.4%
9	4.8%	4.4%	3.2%	4.8%	4.4%	3.2%	3.6%
10	4.9%	4.3%	3.3%	4.9%	4.3%	3.3%	3.0%
11	4.8%	3.9%	4.1%	4.8%	3.9%	4.1%	2.9%
12	5.0%	4.3%	4.7%	5.0%	4.3%	4.7%	2.7%
13	4.8%	4.1%	4.4%	4.8%	4.1%	4.4%	3.0%
14	4.3%	4.2%	3.8%	4.3%	4.2%	3.8%	3.5%
15	3.8%	4.0%	4.2%	3.8%	4.0%	4.2%	3.2%
16	3.1%	2.5%	3.4%	3.1%	2.5%	3.4%	2.8%
17	2.3%	1.9%	2.4%	2.3%	1.9%	2.4%	2.5%
18	1.4%	1.2%	2.4%	1.4%	1.2%	2.4%	2.2%
19	0.9%	0.6%	1.4%	0.9%	0.6%	1.4%	1.9%
20	0.8%	0.6%	1.3%	0.8%	0.6%	1.3%	1.6%
21	0.6%	0.5%	1.3%	0.6%	0.5%	1.3%	1.3%
22	0.7%	0.5%	1.1%	0.7%	0.5%	1.1%	1.1%
23	0.6%	0.7%	1.3%	0.6%	0.7%	1.3%	0.8%
24	0.4%	0.3%	0.8%	0.4%	0.3%	0.8%	0.6%
25	0.3%	0.4%	0.7%	0.3%	0.4%	0.7%	0.4%
26	0.2%	0.1%	0.6%	0.2%	0.1%	0.6%	0.2%
27	0.2%	0.1%	0.6%	0.2%	0.1%	0.6%	0.0%
28	0.0%	0.2%	0.6%	0.0%	0.2%	0.6%	0.0%
29	0.0%	0.2%	1.1%	0.0%	0.2%	1.1%	0.0%
30	0.0%	0.1%	0.3%	0.0%	0.1%	0.3%	0.0%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Source: EPA (2006e).

[a] The following abbreviations correspond to vehicle types: LDGV (light-duty gasoline vehicles), LDGT (light-duty gasoline trucks), HDGV (heavy-duty gasoline vehicles), LDDV (light-duty diesel vehicles), LDDT (light-duty diesel trucks), HDDV (heavy-duty diesel vehicles), and MC (motorcycles).

Table A- 81: Age Distribution by Vehicle/Fuel Type for Highway Vehicles,[a] 2002

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	6.8%	7.2%	4.9%	6.8%	7.2%	4.9%	8.3%
1	7.0%	7.4%	5.5%	7.0%	7.4%	5.5%	8.5%
2	7.3%	7.2%	6.9%	7.3%	7.2%	6.9%	8.8%
3	7.2%	6.9%	7.5%	7.2%	6.9%	7.5%	8.5%
4	5.4%	6.3%	6.1%	5.4%	6.3%	6.1%	8.2%
5	5.5%	5.8%	5.3%	5.5%	5.8%	5.3%	6.4%
6	5.2%	6.6%	5.0%	5.2%	6.6%	5.0%	5.7%
7	6.1%	6.8%	5.9%	6.1%	6.8%	5.9%	5.0%
8	5.2%	6.2%	5.2%	5.2%	6.2%	5.2%	4.6%
9	5.0%	5.3%	4.1%	5.0%	5.3%	4.1%	4.0%
10	4.5%	4.0%	3.1%	4.5%	4.0%	3.1%	3.2%
11	4.5%	3.9%	3.1%	4.5%	3.9%	3.1%	2.7%
12	4.4%	3.6%	3.9%	4.4%	3.6%	3.9%	2.5%
13	4.6%	3.9%	4.4%	4.6%	3.9%	4.4%	2.4%
14	4.3%	3.7%	4.1%	4.3%	3.7%	4.1%	2.6%
15	3.8%	3.7%	3.6%	3.8%	3.7%	3.6%	3.1%
16	3.4%	3.5%	3.9%	3.4%	3.5%	3.9%	2.7%
17	2.7%	2.2%	3.2%	2.7%	2.2%	3.2%	2.4%
18	2.1%	1.7%	2.2%	2.1%	1.7%	2.2%	2.1%
19	1.2%	1.0%	2.2%	1.2%	1.0%	2.2%	1.8%
20	0.8%	0.5%	1.3%	0.8%	0.5%	1.3%	1.6%
21	0.7%	0.5%	1.2%	0.7%	0.5%	1.2%	1.3%
22	0.5%	0.4%	1.2%	0.5%	0.4%	1.2%	1.1%
23	0.6%	0.4%	1.0%	0.6%	0.4%	1.0%	0.9%
24	0.5%	0.6%	1.2%	0.5%	0.6%	1.2%	0.7%

25	0.3%	0.3%	0.7%	0.3%	0.3%	0.7%	0.5%
26	0.2%	0.3%	0.6%	0.2%	0.3%	0.6%	0.3%
27	0.1%	0.1%	0.6%	0.1%	0.1%	0.6%	0.1%
28	0.1%	0.0%	0.5%	0.1%	0.0%	0.5%	0.0%
29	0.0%	0.2%	0.5%	0.0%	0.2%	0.5%	0.0%
30	0.0%	0.2%	1.1%	0.0%	0.2%	1.1%	0.0%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Source: EPA (2006e).

[a] The following abbreviations correspond to vehicle types: LDGV (light-duty gasoline vehicles), LDGT (light-duty gasoline trucks), HDGV (heavy-duty gasoline vehicles), LDDV (light-duty diesel vehicles), LDDT (light-duty diesel trucks), HDDV (heavy-duty diesel vehicles), and MC (motorcycles).

Table A- 82: Age Distribution by Vehicle/Fuel Type for Highway Vehicles,[a] 2003

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	6.7%	7.2%	5.1%	6.7%	7.2%	5.1%	8.3%
1	6.6%	7.0%	4.8%	6.6%	7.0%	4.8%	8.1%
2	6.8%	7.1%	5.4%	6.8%	7.1%	5.4%	8.2%
3	7.1%	7.0%	6.8%	7.1%	7.0%	6.8%	8.5%
4	7.0%	6.7%	7.5%	7.0%	6.7%	7.5%	8.1%
5	5.2%	6.0%	5.9%	5.2%	6.0%	5.9%	7.7%
6	5.4%	5.5%	5.2%	5.4%	5.5%	5.2%	6.0%
7	5.0%	6.2%	4.9%	5.0%	6.2%	4.9%	5.3%
8	5.7%	6.3%	5.7%	5.7%	6.3%	5.7%	4.6%
9	4.8%	5.7%	5.0%	4.8%	5.7%	5.0%	4.2%
10	4.7%	4.8%	4.0%	4.7%	4.8%	4.0%	3.6%
11	4.1%	3.7%	2.9%	4.1%	3.7%	2.9%	2.9%
12	4.1%	3.6%	3.0%	4.1%	3.6%	3.0%	2.4%
13	4.0%	3.2%	3.7%	4.0%	3.2%	3.7%	2.3%
14	4.1%	3.5%	4.2%	4.1%	3.5%	4.2%	2.1%
15	3.9%	3.3%	3.9%	3.9%	3.3%	3.9%	2.3%
16	3.4%	3.3%	3.4%	3.4%	3.3%	3.4%	2.6%
17	3.0%	3.1%	3.6%	3.0%	3.1%	3.6%	2.3%
18	2.4%	1.9%	2.9%	2.4%	1.9%	2.9%	2.1%
19	1.8%	1.4%	2.0%	1.8%	1.4%	2.0%	1.8%
20	1.1%	0.9%	2.0%	1.1%	0.9%	2.0%	1.5%
21	0.7%	0.5%	1.2%	0.7%	0.5%	1.2%	1.3%
22	0.6%	0.4%	1.1%	0.6%	0.4%	1.1%	1.1%
23	0.4%	0.3%	1.1%	0.4%	0.3%	1.1%	0.9%
24	0.5%	0.3%	0.9%	0.5%	0.3%	0.9%	0.7%
25	0.4%	0.5%	1.1%	0.4%	0.5%	1.1%	0.5%
26	0.3%	0.2%	0.6%	0.3%	0.2%	0.6%	0.4%
27	0.2%	0.2%	0.6%	0.2%	0.2%	0.6%	0.2%
28	0.1%	0.1%	0.5%	0.1%	0.1%	0.5%	0.1%
29	0.1%	0.0%	0.4%	0.1%	0.0%	0.4%	0.0%
30	0.0%	0.2%	0.8%	0.0%	0.2%	0.8%	0.0%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Source: EPA (2006e).

[a] The following abbreviations correspond to vehicle types: LDGV (light-duty gasoline vehicles), LDGT (light-duty gasoline trucks), HDGV (heavy-duty gasoline vehicles), LDDV (light-duty diesel vehicles), LDDT (light-duty diesel trucks), HDDV (heavy-duty diesel vehicles), and MC (motorcycles).

Table A- 83: Age Distribution by Vehicle/Fuel Type for Highway Vehicles,[a] 2004

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	6.6%	7.2%	6.0%	6.6%	7.2%	6.0%	8.2%
1	6.5%	7.0%	5.0%	6.5%	7.0%	5.0%	8.0%
2	6.4%	6.7%	4.7%	6.4%	6.7%	4.7%	7.8%
3	6.6%	6.9%	5.3%	6.6%	6.9%	5.3%	7.9%
4	6.9%	6.8%	6.7%	6.9%	6.8%	6.7%	8.1%
5	6.8%	6.4%	7.2%	6.8%	6.4%	7.2%	7.6%
6	5.1%	5.7%	5.7%	5.1%	5.7%	5.7%	7.3%
7	5.2%	5.2%	5.0%	5.2%	5.2%	5.0%	5.6%
8	4.7%	5.8%	4.6%	4.7%	5.8%	4.6%	4.9%

9	5.3%	5.9%	5.4%	5.3%	5.9%	5.4%	4.2%
10	4.5%	5.3%	4.7%	4.5%	5.3%	4.7%	3.8%
11	4.3%	4.5%	3.7%	4.3%	4.5%	3.7%	3.3%
12	3.8%	3.4%	2.7%	3.8%	3.4%	2.7%	2.6%
13	3.8%	3.2%	2.8%	3.8%	3.2%	2.8%	2.1%
14	3.7%	2.9%	3.4%	3.7%	2.9%	3.4%	2.0%
15	3.7%	3.1%	3.9%	3.7%	3.1%	3.9%	1.8%
16	3.5%	2.9%	3.6%	3.5%	2.9%	3.6%	2.0%
17	3.0%	2.8%	3.1%	3.0%	2.8%	3.1%	2.3%
18	2.6%	2.7%	3.3%	2.6%	2.7%	3.3%	2.0%
19	2.1%	1.6%	2.7%	2.1%	1.6%	2.7%	1.7%
20	1.6%	1.2%	1.8%	1.6%	1.2%	1.8%	1.5%
21	0.9%	0.8%	1.8%	0.9%	0.8%	1.8%	1.3%
22	0.6%	0.4%	1.1%	0.6%	0.4%	1.1%	1.1%
23	0.5%	0.3%	1.0%	0.5%	0.3%	1.0%	0.9%
24	0.4%	0.3%	1.0%	0.4%	0.3%	1.0%	0.7%
25	0.4%	0.3%	0.8%	0.4%	0.3%	0.8%	0.6%
26	0.3%	0.4%	0.9%	0.3%	0.4%	0.9%	0.4%
27	0.2%	0.2%	0.5%	0.2%	0.2%	0.5%	0.3%
28	0.1%	0.2%	0.5%	0.1%	0.2%	0.5%	0.2%
29	0.1%	0.0%	0.4%	0.1%	0.0%	0.4%	0.1%
30	0.1%	0.1%	0.6%	0.1%	0.1%	0.6%	0.0%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Source: EPA (2006e).

[a] The following abbreviations correspond to vehicle types: LDGV (light-duty gasoline vehicles), LDGT (light-duty gasoline trucks), HDGV (heavy-duty gasoline vehicles), LDDV (light-duty diesel vehicles), LDDT (light-duty diesel trucks), HDDV (heavy-duty diesel vehicles), and MC (motorcycles).

Table A- 84: Age Distribution by Vehicle/Fuel Type for Highway Vehicles,[a] 2005

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	6.6%	6.9%	6.4%	6.6%	6.9%	6.4%	8.4%
1	6.4%	7.0%	5.9%	6.4%	7.0%	5.9%	8.0%
2	6.3%	6.8%	4.9%	6.3%	6.8%	4.9%	7.8%
3	6.2%	6.6%	4.6%	6.2%	6.6%	4.6%	7.5%
4	6.4%	6.7%	5.2%	6.4%	6.7%	5.2%	7.5%
5	6.7%	6.5%	6.4%	6.7%	6.5%	6.4%	7.6%
6	6.6%	6.1%	6.9%	6.6%	6.1%	6.9%	7.2%
7	4.9%	5.4%	5.4%	4.9%	5.4%	5.4%	6.8%
8	4.9%	4.8%	4.7%	4.9%	4.8%	4.7%	5.2%
9	4.4%	5.4%	4.4%	4.4%	5.4%	4.4%	4.4%
10	4.9%	5.4%	5.1%	4.9%	5.4%	5.1%	3.8%
11	4.1%	4.9%	4.4%	4.1%	4.9%	4.4%	3.4%
12	4.0%	4.1%	3.5%	4.0%	4.1%	3.5%	2.9%
13	3.5%	3.0%	2.5%	3.5%	3.0%	2.5%	2.3%
14	3.4%	2.9%	2.6%	3.4%	2.9%	2.6%	1.9%
15	3.3%	2.6%	3.2%	3.3%	2.6%	3.2%	1.7%
16	3.3%	2.7%	3.6%	3.3%	2.7%	3.6%	1.6%
17	3.1%	2.5%	3.3%	3.1%	2.5%	3.3%	1.7%
18	2.7%	2.5%	2.8%	2.7%	2.5%	2.8%	1.9%
19	2.3%	2.3%	3.0%	2.3%	2.3%	3.0%	1.6%
20	1.8%	1.4%	2.4%	1.8%	1.4%	2.4%	1.4%
21	1.3%	1.0%	1.6%	1.3%	1.0%	1.6%	1.2%
22	0.8%	0.6%	1.7%	0.8%	0.6%	1.7%	1.0%
23	0.5%	0.3%	0.9%	0.5%	0.3%	0.9%	0.8%
24	0.4%	0.3%	0.9%	0.4%	0.3%	0.9%	0.7%
25	0.3%	0.2%	0.9%	0.3%	0.2%	0.9%	0.6%
26	0.4%	0.2%	0.7%	0.4%	0.2%	0.7%	0.4%
27	0.3%	0.3%	0.8%	0.3%	0.3%	0.8%	0.3%
28	0.2%	0.1%	0.5%	0.2%	0.1%	0.5%	0.2%
29	0.1%	0.1%	0.4%	0.1%	0.1%	0.4%	0.1%

30	0.1%	0.1%	0.6%	0.1%	0.1%	0.6%	0.1%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Source: EPA (2006e).

[a] The following abbreviations correspond to vehicle types: LDGV (light-duty gasoline vehicles), LDGT (light-duty gasoline trucks), HDGV (heavy-duty gasoline vehicles), LDDV (light-duty diesel vehicles), LDDT (light-duty diesel trucks), HDDV (heavy-duty diesel vehicles), and MC (motorcycles).

Table A- 85: Annual Average Vehicle Mileage Accumulation per Vehicle (miles)

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC[a]
0	14,910	19,906	20,218	14,910	26,371	28,787	4,786
1	14,174	18,707	18,935	14,174	24,137	26,304	4,475
2	13,475	17,559	17,100	13,475	22,095	24,038	4,164
3	12,810	16,462	16,611	12,810	20,228	21,968	3,853
4	12,178	15,413	15,560	12,178	18,521	20,078	3,543
5	11,577	14,411	14,576	11,577	16,960	18,351	3,232
6	11,006	13,454	13,655	11,006	15,533	16,775	2,921
7	10,463	12,541	12,793	10,463	14,227	15,334	2,611
8	9,947	11,671	11,987	9,947	13,032	14,019	2,300
9	9,456	10,843	11,231	9,456	11,939	12,817	1,989
10	8,989	10,055	10,524	8,989	10,939	11,719	1,678
11	8,546	9,306	9,863	8,546	10,024	10,716	1,368
12	8,124	8,597	9,243	8,124	9,186	9,799	1,368
13	7,723	7,925	8,662	7,723	8,420	8,962	1,368
14	7,342	7,290	8,028	7,342	7,718	8,196	1,368
15	6,980	6,690	7,610	6,980	7,075	7,497	1,368
16	6,636	6,127	7,133	6,636	6,487	6,857	1,368
17	6,308	5,598	6,687	6,308	5,948	6,273	1,368
18	5,997	5,103	6,269	5,997	5,454	5,739	1,368
19	5,701	4,642	5,877	5,701	5,002	5,250	1,368
20	5,420	4,214	5,510	5,420	4,588	4,804	1,368
21	5,152	3,818	5,166	5,152	4,209	4,396	1,368
22	4,898	3,455	4,844	4,898	3,861	4,023	1,368
23	4,656	3,123	4,542	4,656	3,542	3,681	1,368
24	4,427	2,822	4,259	4,427	3,250	3,369	1,368
25	4,427	2,822	4,259	4,427	3,250	3,369	1,368
26	4,427	2,822	4,259	4,427	3,250	3,369	1,368
27	4,427	2,822	4,259	4,427	3,250	3,369	1,368
28	4,427	2,822	4,259	4,427	3,250	3,369	1,368
29	4,427	2,822	4,259	4,427	3,250	3,369	1,368
30	4,427	2,822	4,259	4,427	3,250	3,369	1,368

Source: EPA (2000).

[a] Because of a lack of data, all motorcycles over 12 years old are considered to have the same emissions and travel characteristics, and therefore are presented in aggregate.

Table A- 86: VMT Distribution by Vehicle Age and Vehicle/Fuel Type, 1990-1998

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	7.51%	9.41%	7.89%	7.51%	11.50%	8.27%	19.39%
1	9.52%	11.56%	13.48%	9.52%	13.07%	14.00%	21.15%
2	9.05%	10.62%	11.11%	9.05%	11.15%	11.86%	15.82%
3	8.59%	9.70%	9.85%	8.59%	9.51%	10.05%	11.82%
4	8.14%	8.80%	8.43%	8.14%	8.11%	8.52%	8.77%
5	7.68%	7.92%	7.21%	7.68%	6.92%	7.22%	6.37%
6	7.22%	7.04%	6.16%	7.22%	5.90%	6.13%	4.60%
7	6.72%	6.19%	5.27%	6.72%	5.04%	5.20%	3.31%
8	6.20%	5.36%	4.51%	6.20%	4.30%	4.41%	2.33%
9	5.64%	4.57%	3.86%	5.64%	3.67%	3.74%	1.62%
10	5.03%	3.82%	3.31%	5.03%	3.13%	3.18%	1.09%
11	4.38%	3.14%	2.83%	4.38%	2.67%	2.70%	3.73%
12	3.54%	2.52%	2.42%	3.54%	2.28%	2.29%	0.00%
13	2.67%	1.99%	2.07%	2.67%	1.95%	1.94%	0.00%
14	2.01%	1.54%	1.76%	2.01%	1.66%	1.65%	0.00%
15	1.52%	1.16%	1.52%	1.52%	1.42%	1.40%	0.00%
16	1.14%	0.87%	1.30%	1.14%	1.21%	1.19%	0.00%

17	0.86%	0.64%	1.12%	0.86%	1.04%	1.01%	0.00%
18	0.65%	0.50%	0.96%	0.65%	0.89%	0.86%	0.00%
19	0.49%	0.43%	0.82%	0.49%	0.76%	0.73%	0.00%
20	0.37%	0.37%	0.70%	0.37%	0.65%	0.62%	0.00%
21	0.28%	0.32%	0.60%	0.28%	0.55%	0.53%	0.00%
22	0.21%	0.27%	0.52%	0.21%	0.47%	0.45%	0.00%
23	0.16%	0.23%	0.44%	0.16%	0.40%	0.38%	0.00%
24	0.43%	1.04%	1.85%	0.43%	1.75%	1.65%	0.00%
25	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%
26	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%
27	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%
28	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%
29	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%
30	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%	0.00%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Note: Estimated by weighting data in Table A- 77 by data in Table A- 85.

Table A- 87: VMT Distribution by Vehicle Age and Vehicle/Fuel Type, 1999

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	11.64%	12.03%	13.39%	11.64%	13.39%	15.49%	16.17%
1	8.27%	10.23%	10.08%	8.27%	11.08%	11.37%	14.92%
2	8.07%	8.94%	8.11%	8.07%	9.45%	9.26%	11.21%
3	7.21%	9.79%	7.55%	7.21%	10.10%	8.11%	9.36%
4	8.06%	9.67%	8.47%	8.06%	9.76%	8.87%	7.74%
5	6.76%	8.52%	7.14%	6.76%	8.43%	7.30%	6.57%
6	6.47%	6.92%	5.36%	6.47%	6.71%	5.34%	5.41%
7	5.66%	5.01%	3.79%	5.66%	4.77%	3.69%	4.03%
8	5.49%	4.60%	3.64%	5.49%	4.31%	3.46%	3.04%
9	5.25%	3.98%	4.30%	5.25%	3.68%	3.99%	2.52%
10	5.26%	4.07%	4.70%	5.26%	3.72%	4.25%	2.01%
11	4.80%	3.69%	4.15%	4.80%	3.34%	3.66%	1.89%
12	4.14%	3.49%	3.42%	4.14%	3.13%	2.95%	2.25%
13	3.53%	3.12%	3.47%	3.53%	2.79%	2.91%	2.06%
14	2.75%	1.82%	2.66%	2.75%	1.62%	2.21%	1.87%
15	2.05%	1.33%	1.80%	2.05%	1.18%	1.44%	1.68%
16	1.19%	0.78%	1.73%	1.19%	0.69%	1.35%	1.50%
17	0.75%	0.37%	0.96%	0.75%	0.33%	0.73%	1.32%
18	0.60%	0.31%	0.85%	0.60%	0.28%	0.63%	1.14%
19	0.47%	0.23%	0.81%	0.47%	0.21%	0.59%	0.97%
20	0.53%	0.22%	0.64%	0.53%	0.20%	0.46%	0.80%
21	0.39%	0.31%	0.70%	0.39%	0.29%	0.48%	0.63%
22	0.28%	0.13%	0.39%	0.28%	0.13%	0.26%	0.46%
23	0.17%	0.13%	0.34%	0.17%	0.13%	0.22%	0.30%
24	0.10%	0.03%	0.28%	0.10%	0.03%	0.18%	0.14%
25	0.12%	0.02%	0.26%	0.12%	0.02%	0.17%	0.00%
26	0.00%	0.07%	0.26%	0.00%	0.07%	0.17%	0.00%
27	0.00%	0.06%	0.51%	0.00%	0.06%	0.33%	0.00%
28	0.00%	0.03%	0.12%	0.00%	0.03%	0.08%	0.00%
29	0.00%	0.05%	0.06%	0.00%	0.04%	0.04%	0.00%
30	0.00%	0.01%	0.05%	0.00%	0.01%	0.03%	0.00%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Note: Estimated by weighting data in Table A- 78 by data in Table A- 85.

Table A- 88: VMT Distribution by Vehicle Age and Vehicle/Fuel Type, 2000

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	11.33%	12.10%	11.77%	11.33%	13.46%	13.60%	15.68%
1	10.59%	10.94%	12.09%	10.59%	11.86%	13.63%	14.27%
2	7.52%	9.28%	8.77%	7.52%	9.82%	10.00%	13.11%
3	7.34%	8.11%	7.59%	7.34%	8.37%	8.15%	9.69%
4	6.55%	8.87%	6.81%	6.55%	8.95%	7.14%	7.96%
5	7.33%	8.66%	7.56%	7.33%	8.57%	7.73%	6.46%

6	6.15%	7.54%	6.32%	6.15%	7.32%	6.30%	5.43%
7	5.82%	6.05%	4.74%	5.82%	5.77%	4.61%	4.38%
8	4.99%	4.33%	3.32%	4.99%	4.06%	3.15%	3.18%
9	4.80%	3.97%	3.19%	4.80%	3.67%	2.95%	2.33%
10	4.58%	3.40%	3.77%	4.58%	3.10%	3.40%	1.86%
11	4.55%	3.46%	4.07%	4.55%	3.14%	3.59%	1.44%
12	4.15%	3.10%	3.60%	4.15%	2.79%	3.10%	1.64%
13	3.54%	2.90%	2.93%	3.54%	2.59%	2.46%	1.93%
14	2.99%	2.58%	2.94%	2.99%	2.30%	2.44%	1.75%
15	2.33%	1.49%	2.31%	2.33%	1.32%	1.85%	1.59%
16	1.71%	1.07%	1.53%	1.71%	0.96%	1.19%	1.41%
17	0.98%	0.62%	1.47%	0.98%	0.55%	1.12%	1.24%
18	0.62%	0.30%	0.80%	0.62%	0.27%	0.60%	1.08%
19	0.49%	0.24%	0.71%	0.49%	0.22%	0.52%	0.93%
20	0.38%	0.18%	0.67%	0.38%	0.17%	0.47%	0.78%
21	0.43%	0.17%	0.53%	0.43%	0.15%	0.37%	0.64%
22	0.31%	0.23%	0.58%	0.31%	0.22%	0.39%	0.50%
23	0.22%	0.10%	0.32%	0.22%	0.10%	0.21%	0.36%
24	0.13%	0.10%	0.28%	0.13%	0.09%	0.18%	0.24%
25	0.08%	0.02%	0.25%	0.08%	0.02%	0.16%	0.11%
26	0.10%	0.02%	0.23%	0.10%	0.02%	0.14%	0.00%
27	0.00%	0.05%	0.23%	0.00%	0.05%	0.15%	0.00%
28	0.00%	0.05%	0.44%	0.00%	0.05%	0.29%	0.00%
29	0.00%	0.02%	0.11%	0.00%	0.02%	0.07%	0.00%
30	0.00%	0.04%	0.07%	0.00%	0.04%	0.04%	0.00%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Note: Estimated by weighting data in Table A- 79 by data in Table A- 85.

Table A- 89: VMT Distribution by Vehicle Age and Vehicle/Fuel Type, 2001

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	10.43%	12.01%	9.26%	10.43%	13.38%	10.74%	14.44%
1	10.41%	11.02%	10.92%	10.41%	11.94%	12.36%	14.05%
2	9.73%	9.95%	10.82%	9.73%	10.52%	12.39%	12.72%
3	6.91%	8.44%	8.44%	6.91%	8.71%	9.10%	11.50%
4	6.74%	7.35%	7.05%	6.74%	7.43%	7.41%	8.36%
5	6.02%	7.95%	6.27%	6.02%	7.87%	6.43%	6.75%
6	6.74%	7.68%	6.89%	6.74%	7.45%	6.89%	5.43%
7	5.59%	6.61%	5.75%	5.59%	6.30%	5.62%	4.47%
8	5.19%	5.24%	4.27%	5.19%	4.92%	4.07%	3.51%
9	4.41%	3.74%	2.99%	4.41%	3.46%	2.78%	2.48%
10	4.23%	3.39%	2.87%	4.23%	3.10%	2.61%	1.75%
11	4.00%	2.89%	3.36%	4.00%	2.62%	2.97%	1.35%
12	3.97%	2.92%	3.63%	3.97%	2.62%	3.14%	1.27%
13	3.58%	2.58%	3.18%	3.58%	2.30%	2.68%	1.43%
14	3.02%	2.40%	2.56%	3.02%	2.14%	2.13%	1.66%
15	2.55%	2.11%	2.63%	2.55%	1.88%	2.11%	1.50%
16	1.97%	1.20%	2.02%	1.97%	1.07%	1.58%	1.35%
17	1.43%	0.86%	1.34%	1.43%	0.76%	1.02%	1.19%
18	0.82%	0.49%	1.27%	0.82%	0.44%	0.95%	1.04%
19	0.52%	0.23%	0.70%	0.52%	0.21%	0.51%	0.89%
20	0.40%	0.19%	0.61%	0.40%	0.17%	0.43%	0.76%
21	0.31%	0.14%	0.57%	0.31%	0.13%	0.40%	0.64%
22	0.34%	0.12%	0.46%	0.34%	0.12%	0.31%	0.51%
23	0.25%	0.18%	0.49%	0.25%	0.17%	0.32%	0.39%
24	0.18%	0.08%	0.27%	0.18%	0.07%	0.17%	0.29%
25	0.11%	0.08%	0.25%	0.11%	0.08%	0.16%	0.18%
26	0.07%	0.02%	0.22%	0.07%	0.02%	0.14%	0.08%
27	0.08%	0.01%	0.20%	0.08%	0.01%	0.13%	0.00%
28	0.00%	0.04%	0.21%	0.00%	0.04%	0.13%	0.00%
29	0.00%	0.04%	0.39%	0.00%	0.04%	0.25%	0.00%
30	0.00%	0.03%	0.11%	0.00%	0.03%	0.07%	0.00%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Note: Estimated by weighting data in Table A- 80 by data in Table A- 85.

Table A- 90: VMT Distribution by Vehicle Age and Vehicle/Fuel Type, 2002

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	9.84%	11.42%	8.28%	9.84%	12.73%	9.66%	13.70%
1	9.65%	11.02%	8.69%	9.65%	11.96%	9.89%	13.07%
2	9.63%	10.09%	9.88%	9.63%	10.68%	11.38%	12.65%
3	9.01%	9.11%	10.53%	9.01%	9.41%	11.41%	11.28%
4	6.39%	7.71%	7.92%	6.39%	7.79%	8.38%	10.03%
5	6.24%	6.64%	6.55%	6.24%	6.58%	6.76%	7.16%
6	5.57%	7.10%	5.76%	5.57%	6.90%	5.80%	5.72%
7	6.17%	6.78%	6.33%	6.17%	6.47%	6.22%	4.51%
8	5.02%	5.76%	5.24%	5.02%	5.41%	5.02%	3.61%
9	4.61%	4.56%	3.89%	4.61%	4.23%	3.64%	2.76%
10	3.91%	3.22%	2.72%	3.91%	2.94%	2.48%	1.88%
11	3.72%	2.91%	2.59%	3.72%	2.63%	2.31%	1.28%
12	3.52%	2.45%	3.03%	3.52%	2.20%	2.63%	1.20%
13	3.45%	2.44%	3.24%	3.45%	2.18%	2.75%	1.11%
14	3.08%	2.15%	2.80%	3.08%	1.92%	2.34%	1.24%
15	2.60%	1.98%	2.31%	2.60%	1.76%	1.87%	1.45%
16	2.17%	1.72%	2.32%	2.17%	1.53%	1.83%	1.29%
17	1.66%	0.97%	1.78%	1.66%	0.86%	1.37%	1.15%
18	1.21%	0.69%	1.17%	1.21%	0.62%	0.88%	1.00%
19	0.68%	0.39%	1.11%	0.68%	0.35%	0.81%	0.86%
20	0.42%	0.18%	0.60%	0.42%	0.17%	0.43%	0.74%
21	0.33%	0.15%	0.53%	0.33%	0.14%	0.37%	0.63%
22	0.25%	0.11%	0.50%	0.25%	0.10%	0.34%	0.52%
23	0.28%	0.09%	0.39%	0.28%	0.09%	0.26%	0.41%
24	0.20%	0.13%	0.42%	0.20%	0.13%	0.27%	0.32%
25	0.15%	0.06%	0.25%	0.15%	0.06%	0.16%	0.23%
26	0.09%	0.06%	0.22%	0.09%	0.06%	0.15%	0.14%
27	0.06%	0.01%	0.20%	0.06%	0.01%	0.13%	0.07%
28	0.06%	0.01%	0.18%	0.06%	0.01%	0.12%	0.00%
29	0.00%	0.03%	0.18%	0.00%	0.03%	0.12%	0.00%
30	0.00%	0.04%	0.38%	0.00%	0.04%	0.25%	0.00%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Note: Estimated by weighting data in Table A- 81 by data in Table A- 85.

Table A- 91: VMT Distribution by Vehicle Age and Vehicle/Fuel Type, 2003

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	9.74%	11.53%	8.76%	9.74%	12.86%	10.27%	13.63%
1	9.13%	10.47%	7.74%	9.13%	11.37%	8.85%	12.43%
2	8.94%	10.09%	7.83%	8.94%	10.69%	9.06%	11.80%
3	8.93%	9.23%	9.59%	8.93%	9.54%	10.43%	11.24%
4	8.35%	8.32%	9.85%	8.35%	8.41%	10.46%	9.86%
5	5.93%	6.96%	7.34%	5.93%	6.89%	7.60%	8.61%
6	5.79%	5.93%	6.01%	5.79%	5.76%	6.07%	6.09%
7	5.12%	6.27%	5.28%	5.12%	5.98%	5.21%	4.76%
8	5.55%	5.91%	5.75%	5.55%	5.55%	5.53%	3.66%
9	4.47%	5.02%	4.75%	4.47%	4.65%	4.46%	2.85%
10	4.11%	3.92%	3.53%	4.11%	3.59%	3.23%	2.10%
11	3.45%	2.76%	2.45%	3.45%	2.50%	2.19%	1.38%
12	3.28%	2.46%	2.33%	3.28%	2.21%	2.03%	1.14%
13	3.06%	2.05%	2.69%	3.06%	1.83%	2.29%	1.06%
14	2.98%	2.04%	2.85%	2.98%	1.81%	2.39%	0.97%
15	2.66%	1.77%	2.52%	2.66%	1.58%	2.04%	1.08%
16	2.22%	1.61%	2.03%	2.22%	1.43%	1.61%	1.25%
17	1.83%	1.38%	2.04%	1.83%	1.23%	1.58%	1.10%
18	1.40%	0.77%	1.55%	1.40%	0.70%	1.17%	0.97%
19	1.01%	0.54%	1.02%	1.01%	0.49%	0.75%	0.83%
20	0.56%	0.30%	0.96%	0.56%	0.28%	0.69%	0.72%

21	0.35%	0.14%	0.52%	0.35%	0.13%	0.36%	0.61%
22	0.27%	0.11%	0.45%	0.27%	0.10%	0.31%	0.51%
23	0.20%	0.08%	0.42%	0.20%	0.08%	0.28%	0.42%
24	0.22%	0.07%	0.33%	0.22%	0.07%	0.22%	0.33%
25	0.17%	0.11%	0.38%	0.17%	0.11%	0.25%	0.25%
26	0.12%	0.05%	0.22%	0.12%	0.05%	0.14%	0.18%
27	0.08%	0.05%	0.20%	0.08%	0.05%	0.13%	0.11%
28	0.04%	0.01%	0.18%	0.04%	0.01%	0.12%	0.05%
29	0.05%	0.01%	0.16%	0.05%	0.01%	0.10%	0.00%
30	0.00%	0.04%	0.28%	0.00%	0.04%	0.18%	0.00%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Note: Estimated by weighting data in Table A- 82 by data in Table A- 85.

Table A- 92: VMT Distribution by Vehicle Age and Vehicle/Fuel Type, 2004

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	9.64%	11.64%	10.36%	9.64%	12.98%	12.15%	13.57%
1	9.05%	10.56%	8.05%	9.05%	11.47%	9.21%	12.40%
2	8.47%	9.58%	6.86%	8.47%	10.14%	7.94%	11.25%
3	8.30%	9.22%	7.47%	8.30%	9.54%	8.13%	10.52%
4	8.29%	8.42%	8.81%	8.29%	8.52%	9.36%	9.85%
5	7.75%	7.50%	8.96%	7.75%	7.43%	9.29%	8.49%
6	5.50%	6.21%	6.61%	5.50%	6.03%	6.69%	7.34%
7	5.32%	5.23%	5.41%	5.32%	4.99%	5.34%	5.08%
8	4.61%	5.46%	4.71%	4.61%	5.13%	4.54%	3.88%
9	4.95%	5.13%	5.13%	4.95%	4.76%	4.82%	2.89%
10	3.98%	4.31%	4.24%	3.98%	3.94%	3.89%	2.17%
11	3.62%	3.36%	3.12%	3.62%	3.05%	2.79%	1.55%
12	3.04%	2.33%	2.16%	3.04%	2.10%	1.89%	1.23%
13	2.86%	2.06%	2.03%	2.86%	1.84%	1.73%	1.01%
14	2.65%	1.71%	2.33%	2.65%	1.52%	1.96%	0.93%
15	2.57%	1.68%	2.52%	2.57%	1.49%	2.04%	0.85%
16	2.27%	1.44%	2.18%	2.27%	1.28%	1.72%	0.94%
17	1.87%	1.29%	1.76%	1.87%	1.15%	1.36%	1.07%
18	1.55%	1.10%	1.75%	1.55%	0.99%	1.32%	0.93%
19	1.17%	0.61%	1.33%	1.17%	0.55%	0.98%	0.81%
20	0.83%	0.42%	0.86%	0.83%	0.39%	0.62%	0.70%
21	0.47%	0.23%	0.81%	0.47%	0.22%	0.57%	0.59%
22	0.29%	0.11%	0.44%	0.29%	0.10%	0.30%	0.50%
23	0.22%	0.08%	0.38%	0.22%	0.08%	0.25%	0.41%
24	0.16%	0.06%	0.35%	0.16%	0.06%	0.23%	0.34%
25	0.19%	0.06%	0.30%	0.19%	0.06%	0.19%	0.26%
26	0.14%	0.09%	0.34%	0.14%	0.09%	0.22%	0.20%
27	0.10%	0.04%	0.20%	0.10%	0.04%	0.13%	0.14%
28	0.06%	0.04%	0.18%	0.06%	0.04%	0.12%	0.09%
29	0.04%	0.01%	0.16%	0.04%	0.01%	0.10%	0.04%
30	0.04%	0.02%	0.22%	0.04%	0.02%	0.14%	0.00%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Note: Estimated by weighting data in Table A- 83 by data in Table A- 85.

Table A- 93: VMT Distribution by Vehicle Age and Vehicle/Fuel Type, 2005

Vehicle Age	LDGV	LDGT	HDGV	LDDV	LDDT	HDDV	MC
0	9.77%	11.22%	11.06%	9.77%	12.52%	12.95%	13.83%
1	8.94%	10.72%	9.46%	8.94%	11.65%	10.80%	12.32%
2	8.39%	9.71%	7.09%	8.39%	10.29%	8.19%	11.20%
3	7.86%	8.80%	6.49%	7.86%	9.10%	7.06%	10.01%
4	7.70%	8.46%	6.81%	7.70%	8.56%	7.23%	9.20%
5	7.69%	7.64%	7.96%	7.69%	7.57%	8.24%	8.46%
6	7.19%	6.72%	8.02%	7.19%	6.54%	8.10%	7.22%
7	5.05%	5.50%	5.91%	5.05%	5.25%	5.83%	6.11%
8	4.78%	4.57%	4.79%	4.78%	4.30%	4.61%	4.13%
9	4.10%	4.77%	4.17%	4.10%	4.42%	3.92%	3.06%
10	4.41%	4.43%	4.54%	4.41%	4.06%	4.16%	2.20%

11	3.51%	3.71%	3.71%	3.51%	3.36%	3.32%	1.60%
12	3.19%	2.86%	2.73%	3.19%	2.57%	2.38%	1.38%
13	2.65%	1.96%	1.87%	2.65%	1.75%	1.59%	1.09%
14	2.47%	1.72%	1.74%	2.47%	1.54%	1.46%	0.88%
15	2.28%	1.41%	2.04%	2.28%	1.26%	1.65%	0.81%
16	2.19%	1.37%	2.16%	2.19%	1.22%	1.71%	0.73%
17	1.92%	1.16%	1.87%	1.92%	1.04%	1.44%	0.80%
18	1.58%	1.04%	1.49%	1.58%	0.93%	1.12%	0.90%
19	1.29%	0.87%	1.48%	1.29%	0.79%	1.09%	0.78%
20	0.96%	0.48%	1.11%	0.96%	0.44%	0.80%	0.67%
21	0.69%	0.33%	0.72%	0.69%	0.30%	0.51%	0.57%
22	0.38%	0.18%	0.68%	0.38%	0.17%	0.46%	0.48%
23	0.23%	0.08%	0.36%	0.23%	0.08%	0.24%	0.40%
24	0.17%	0.06%	0.32%	0.17%	0.06%	0.21%	0.33%
25	0.14%	0.05%	0.31%	0.14%	0.05%	0.20%	0.27%
26	0.15%	0.05%	0.26%	0.15%	0.05%	0.17%	0.21%
27	0.12%	0.07%	0.30%	0.12%	0.07%	0.19%	0.15%
28	0.08%	0.03%	0.17%	0.08%	0.03%	0.11%	0.11%
29	0.05%	0.03%	0.15%	0.05%	0.03%	0.10%	0.07%
30	0.04%	0.01%	0.20%	0.04%	0.01%	0.13%	0.03%
	100%	**100%**	**100%**	**100%**	**100%**	**100%**	**100%**

Note: Estimated by weighting data in Table A- 84 by data in Table A- 85.

Table A- 94: Fuel Consumption for Non-Highway Vehicles by Fuel Type (million gallons)

Vehicle Type/Year	1990	1995	2000	2001	2002	2003	2004	2005
Aircraft								
Gasoline[a]	374.2	329.3	301.9	290.7	280.6	251.4	259.9	294.3
Jet Fuel	18,419.0	17,956.2	20,265.7	19,243.1	18,355.4	18,031.5	18,550.9	19,316.4
Ships and Boats								
Diesel	1,175.2	1,184.8	1,525.5	1,680.4	1,653.8	1,371.6	1,727.4	1,441.7
Gasoline	1,429.0	1,612.7	1,678.0	1,680.2	1,678.4	1,673.0	1,664.8	1,655.0
Residual	2,060.7	2,646.1	3,027.6	1,130.7	2,649.9	2,285.4	2,645.6	3,146.1
Construction/Mining Equipment								
Diesel	3,674.5	4,387.1	5,095.1	5,282.7	5,429.5	5,576.4	5,723.2	5,870.1
Gasoline	167.7	167.7	153.9	151.1	148.7	146.9	143.4	139.7
Agricultural Equipment								
Diesel	2,320.9	2,771.6	3,222.3	3,307.4	3,390.4	3,473.5	3,556.5	3,639.6
Gasoline	62.0	68.4	72.7	73.5	74.3	75.0	75.1	75.2
Locomotives								
Diesel	3,460.6	3,863.5	4,105.9	4,119.0	4,088.7	4,176.1	4,412.9	4,450.5
Other[b]								
Diesel	1,423.3	1,720.0	2,016.0	2,095.1	2,161.0	2,226.9	2,292.8	2,358.7
Gasoline	4,630.2	4,959.5	5,004.1	5,775.2	5,892.8	6,033.7	6,175.1	6,120.5

Sources: AAR (2006), APTA (2006), BEA (1991 through 2005), Benson (2002 through 2004), DOE (1993 through 2006), DESC (2006), DOC (1991 through 2006), DOT (1991 through 2006), EIA (2002a), EIA (2002b), EIA (2006a), EIA (2006b), EIA (2004), EIA (2003 through 2004), EIA (1991 through 2006), EPA (2006e), FAA (2006a and 2006b), and Whorton (2006).

[a] For aircraft, this is aviation gasoline. For all other categories, this is motor gasoline.

[b] "Other" includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment.

Table A- 95: Control Technology Assignments for Gasoline Passenger Cars (Percent of VMT)

Model Years	Non-catalyst	Oxidation	EPA Tier 0	EPA Tier 1	LEV	EPA Tier 2
1973-1974	100%	-	-	-	-	-
1975	20%	80%	-	-	-	-
1976-1977	15%	85%	-	-	-	-
1978-1979	10%	90%	-	-	-	-
1980	5%	88%	7%	-	-	-
1981	-	15%	85%	-	-	-
1982	-	14%	86%	-	-	-
1983	-	12%	88%	-	-	-
1984-1993	-	-	100%	-	-	-

1994	-	-	60%	40%	-	-
1995	-	-	20%	80%	-	-
1996	-	-	1%	97%	2%	-
1997	-	-	0.5%	96.5%	3%	-
1998	-	-	<1%	87%	13%	-
1999	-	-	<1%	67%	33%	-
2000	-	-	-	44%	56%	-
2001	-	-	-	3%	97%	-
2002	-	-	-	1%	99%	-
2003	-	-	-	<1%	87%	13%
2004	-	-	-	<1%	41%	59%
2005	-	-	-	-	38%	62%

Sources: EPA (1998), EPA (2006a), and EPA (2006b).
Note: Detailed descriptions of emissions control technologies are provided in the following section of this annex.
- Not applicable.

Table A- 96: Control Technology Assignments for Gasoline Light-Duty Trucks (Percent of VMT)[a]

Model Years	Non-catalyst	Oxidation	EPA Tier 0	EPA Tier 1	LEV [b]	EPA Tier 2
1973-1974	100%	-	-	-	-	-
1975	30%	70%	-	-	-	-
1976	20%	80%	-	-	-	-
1977-1978	25%	75%	-	-	-	-
1979-1980	20%	80%	-	-	-	-
1981	-	95%	5%	-	-	-
1982	-	90%	10%	-	-	-
1983	-	80%	20%	-	-	-
1984	-	70%	30%	-	-	-
1985	-	60%	40%	-	-	-
1986	-	50%	50%	-	-	-
1987-1993	-	5%	95%	-	-	-
1994	-	-	60%	40%	-	-
1995	-	-	20%	80%	-	-
1996	-	-	-	100%	-	-
1997	-	-	-	100%	-	-
1998	-	-	-	80%	20%	-
1999	-	-	-	57%	43%	-
2000	-	-	-	65%	35%	-
2001	-	-	-	1%	99%	-
2002	-	-	-	10%	90%	-
2003	-	-	-	<1%	53%	47%
2004	-	-	-	-	72%	28%
2005	-	-	-	-	38%	62%

Sources: EPA (1998), EPA (2006a), and EPA (2006b).
[a] Detailed descriptions of emissions control technologies are provided in the following section of this annex.
[b] The proportion of LEVs as a whole has decreased since 2001, as carmakers have been able to achieve greater emission reductions with certain types of LEVs, such as ULEVs. Because ULEVs emit about half the emissions of LEVs, a carmaker can reduce the total number of LEVs they need to build to meet a specified emission average for all of their vehicles in a given model year.
- Not applicable.

Table A- 97: Control Technology Assignments for Gasoline Heavy-Duty Vehicles (Percent of VMT)[a]

Model Years	Uncontrolled	Non-catalyst	Oxidation	EPA Tier 0	EPA Tier 1	LEV [b]	EPA Tier 2
≤1981	100%	-	-	-	-	-	-
1982-1984	95%	-	5%	-	-	-	-
1985-1986	-	95%	5%	-	-	-	-
1987	-	70%	15%	15%	-	-	-
1988-1989	-	60%	25%	15%	-	-	-
1990-1995	-	45%	30%	25%	-	-	-
1996	-	-	25%	10%	65%	-	-
1997	-	-	10%	5%	85%	-	-
1998	-	-	-	-	96%	4%	-
1999	-	-	-	-	78%	22%	-
2000	-	-	-	-	54%	46%	-
2001	-	-	-	-	64%	36%	-

2002	-	-	-	-	69%	31%	-
2003	-	-	-	-	65%	30%	5%
2004	-	-	-	-	5%	37%	59%
2005	-	-	-	-	-	23%	77%

Sources: EPA (1998), EPA (2006a), and EPA (2006b).
[a] Detailed descriptions of emissions control technologies are provided in the following section of this annex.
[b] The proportion of LEVs as a whole has decreased since 2000, as carmakers have been able to achieve greater emission reductions with certain types of LEVs, such as ULEVs. Because ULEVs emit about half the emissions of LEVs, a carmaker can reduce the total number of LEVs they need to build to meet a specified emission average for all of their vehicles in a given model year.
- Not applicable.

Table A- 98: Control Technology Assignments for Diesel Highway Vehicles and Motorcycles

Vehicle Type/Control Technology	Model Years
Diesel Passenger Cars and Light-Duty Trucks	
Uncontrolled	1960-1982
Moderate control	1983-1995
Advanced control	1996-2004
Heavy-Duty Diesel Vehicles	
Uncontrolled	1960-1982
Moderate control	1983-1995
Advanced control	1996-2004
Motorcycles	
Uncontrolled	1960-1995
Non-catalyst controls	1996-2004

Source: EPA (1998) and Browning (2005)
Note: Detailed descriptions of emissions control technologies are provided in the following section of this annex.

Table A- 99: Emission Factors for CH_4 and N_2O for Highway Vehicles

Vehicle Type/Control Technology	N_2O (g/mi)	CH_4 (g/mi)
Gasoline Passenger Cars		
EPA Tier 2	0.0036	0.0173
Low Emission Vehicles	0.0150	0.0105
EPA Tier 1[a]	0.0429	0.0271
EPA Tier 0[a]	0.0647	0.0704
Oxidation Catalyst	0.0504	0.1355
Non-Catalyst Control	0.0197	0.1696
Uncontrolled	0.0197	0.1780
Gasoline Light-Duty Trucks		
EPA Tier 2	0.0066	0.0163
Low Emission Vehicles	0.0157	0.0148
EPA Tier 1[a]	0.0871	0.0452
EPA Tier 0[a]	0.1056	0.0776
Oxidation Catalyst	0.0639	0.1516
Non-Catalyst Control	0.0218	0.1908
Uncontrolled	0.0220	0.2024
Gasoline Heavy-Duty Vehicles		
EPA Tier 2	0.0134	0.0333
Low Emission Vehicles	0.0320	0.0303
EPA Tier 1[a]	0.1750	0.0655
EPA Tier 0[a]	0.2135	0.2630
Oxidation Catalyst[b]	0.1317	0.2356
Non-Catalyst Control	0.0473	0.4181
Uncontrolled	0.0497	0.4604
Diesel Passenger Cars		
Advanced	0.0010	0.0005
Moderate	0.0010	0.0005
Uncontrolled	0.0012	0.0006
Diesel Light-Duty Trucks		
Advanced	0.0015	0.0010
Moderate	0.0014	0.0009
Uncontrolled	0.0017	0.0011

Diesel Heavy-Duty Vehicles		
Advanced	0.0048	0.0051
Moderate	0.0048	0.0051
Uncontrolled	0.0048	0.0051
Motorcycles		
Non-Catalyst Control	0.0069	0.0672
Uncontrolled	0.0087	0.0899

Source: ICF (2006b) and (2004).
[a] The categories "EPA Tier 0" and "EPA Tier 1" were substituted for the early three-way catalyst and advanced three-way catalyst categories, respectively, as defined in the *Revised 1996 IPCC Guidelines*. Detailed descriptions of emissions control technologies are provided at the end of this annex.

Table A-100: Emission Factors for CH_4 and N_2O for Alternative Fuel Vehicles

	N2O (g/mi)	CH4 (g/mi)
Light Duty Vehicles		
Methanol	0.067	0.018
CNG	0.050	0.737
LPG	0.067	0.037
Ethanol	0.067	0.055
Heavy Duty Vehicles		
Methanol	0.175	0.066
CNG	0.175	1.966
LNG	0.175	1.966
LPG	0.175	0.066
Ethanol	0.175	0.197
Buses		
Methanol	0.175	0.066
CNG	0.175	1.966
Ethanol	0.175	0.197

Source: Developed by ICF (2006a) using ANL (2006) and Lipman and Delucchi (2002).

Table A-101: Emission Factors for CH_4 and N_2O Emissions from Non-Highway Mobile Combustion (g gas/kg fuel)

Vehicle Type/Fuel Type	N_2O	CH_4
Ships and Boats		
Residual	0.08	0.230
Distillate	0.08	0.230
Gasoline	0.08	0.230
Locomotives		
Diesel	0.08	0.250
Agricultural Equipment		
Gas	0.08	0.450
Diesel	0.08	0.450
Construction		
Gas	0.08	0.180
Diesel	0.08	0.180
Other Non-Highway		
All "Other" Categories*	0.08	0.180
Aircraft		
Jet Fuel	0.10	0.087
Aviation Gasoline	0.04	2.640

Source: IPCC/UNEP/OECD/IEA (1997).
* "Other" includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment.

Table A-102: CH_4 Emissions from Mobile Combustion (Tg CO_2 Eq.)

Fuel Type/Vehicle Type	1990	1995	2000	2001	2002	2003	2004	2005
Gasoline Highway	**4.2**	**3.8**	**2.8**	**2.6**	**2.4**	**2.2**	**2.1**	**1.9**
Passenger Cars	2.6	2.1	1.6	1.5	1.4	1.2	1.2	1.1
Light-Duty Trucks	1.4	1.4	1.1	1.0	1.0	0.9	0.8	0.8
Heavy-Duty Vehicles	0.2	0.2	0.1	0.1	0.1	0.1	0.1	0.1
Motorcycles	+	+	+	+	+	+	+	+
Diesel Highway	+	+	+	+	+	+	+	+

Passenger Cars	+	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	+	+	+	+	+	+
Heavy-Duty Vehicles	+	+	+	+	+	+	+	+
Alternative Fuel Highway	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**
Non-Highway	**0.5**	**0.5**	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**
Boats	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Locomotives	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Aircraft	0.2	0.1	0.2	0.1	0.1	0.1	0.1	0.1
Farm Equipment	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Construction Equipment	+	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Other*	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Total	**4.7**	**4.3**	**3.5**	**3.2**	**3.1**	**2.9**	**2.8**	**2.6**

+ Less than 0.05 Tg CO_2 Eq.
Note: Totals may not sum due to independent rounding.
[a] "Other" includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment.

Table A-103: N_2O Emissions from Mobile Combustion (Tg CO_2 Eq.)

Fuel Type/Vehicle Type	1990	1995	2000	2001	2002	2003	2004	2005
Gasoline Highway	**40.1**	**49.8**	**48.8**	**45.5**	**42.8**	**39.5**	**36.7**	**33.4**
Passenger Cars	25.4	26.9	24.7	23.2	21.9	20.3	18.8	17.0
Light-Duty Trucks	14.1	22.1	23.3	21.4	20.0	18.2	17.0	15.6
Heavy-Duty Vehicles	0.6	0.7	0.9	0.9	0.9	0.9	0.9	0.8
Motorcycles	+	+	+	+	+	+	+	+
Diesel Highway	**0.2**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**
Passenger Cars	+	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	+	+	+	+	+	+
Heavy-Duty Vehicles	0.2	0.2	0.3	0.3	0.3	0.3	0.3	0.3
Alternative Fuel Highway	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**
Non-Highway	**3.4**	**3.6**	**4.0**	**3.8**	**3.9**	**3.9**	**4.0**	**4.1**
Boats	0.4	0.4	0.5	0.3	0.5	0.4	0.5	0.5
Locomotives	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.4
Aircraft	1.7	1.7	1.9	1.8	1.7	1.7	1.7	1.8
Farm Equipment	0.2	0.3	0.3	0.3	0.3	0.3	0.4	0.4
Construction Equipment	0.3	0.4	0.4	0.5	0.5	0.5	0.5	0.5
Other*	0.4	0.5	0.5	0.6	0.6	0.6	0.6	0.6
Total	**43.7**	**53.7**	**53.2**	**49.7**	**47.1**	**43.8**	**41.2**	**38.0**

+ Less than 0.05 Tg CO_2 Eq.
Note: Totals may not sum due to independent rounding.
[a] "Other" includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment.

Table A-104: NO_x Emissions from Mobile Combustion (Gg)

Fuel Type/Vehicle Type	1990	1995	2000	2001	2002	2003	2004	2005
Gasoline Highway	**5,746**	**4,560**	**3,812**	**3,715**	**3,887**	**3,710**	**3,541**	**3,380**
Passenger Cars	3,847	2,752	2,084	2,027	2,120	2,024	1,932	1,844
Light-Duty Trucks	1,364	1,325	1,303	1,285	1,344	1,283	1,225	1,169
Heavy-Duty Vehicles	515	469	411	390	408	389	372	355
Motorcycles	20	14	13	14	14	14	13	12
Diesel Highway	**2,956**	**3,493**	**3,803**	**3,338**	**3,492**	**3,333**	**3,181**	**3,037**
Passenger Cars	39	19	7	6	6	6	6	5
Light-Duty Trucks	20	12	6	5	6	5	5	5
Heavy-Duty Vehicles	2,897	3,462	3,791	3,326	3,480	3,322	3,171	3,026
Alternative Fuel Highway	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**
Non-Highway	**2,218**	**2,569**	**2,695**	**2,766**	**2,940**	**2,868**	**2,797**	**2,728**
Boats	402	488	506	544	581	567	553	539
Locomotives	338	433	451	485	518	506	493	481
Aircraft	25	31	40	39	42	41	40	39
Farm Equipment	437	478	484	480	508	495	483	471
Construction Equipment	641	697	697	690	730	712	694	677
Other*	376	442	518	529	561	548	534	521

Total	**10,920**	**10,622**	**10,310**	**9,819**	**10,319**	**9,911**	**9,520**	**9,145**

IE = Included Elsewhere
[a] NO_x emissions from alternative fuel highway vehicles are included under gasoline and diesel highway.
[b] Aircraft estimates include only emissions related to LTO cycles, and therefore do not include cruise altitude emissions.
[c] "Other" includes gasoline powered recreational, industrial, lawn and garden, light commercial, logging, airport service, other equipment; and diesel powered recreational, industrial, lawn and garden, light construction, airport service.
Note: Totals may not sum due to independent rounding.

Table A-105: CO Emissions from Mobile Combustion (Gg)

Fuel Type/Vehicle Type	1990	1995	2000	2001	2002	2003	2004	2005
Gasoline Highway	**98,328**	**74,673**	**60,657**	**56,716**	**56,253**	**53,488**	**50,859**	**48,360**
Passenger Cars	60,757	42,065	32,867	31,600	31,342	29,801	28,336	26,944
Light-Duty Trucks	29,237	27,048	24,532	22,574	22,390	21,290	20,243	19,248
Heavy-Duty Vehicles	8,093	5,404	3,104	2,411	2,391	2,274	2,162	2,056
Motorcycles	240	155	154	131	130	124	118	112
Diesel Highway	**1,696**	**1,424**	**1,088**	**869**	**862**	**820**	**779**	**741**
Passenger Cars	35	18	7	6	6	6	5	5
Light-Duty Trucks	22	16	6	5	5	5	5	4
Heavy-Duty Vehicles	1,639	1,391	1,075	858	851	809	769	732
Alternative Fuel Highway	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**
Non-Highway	**19,457**	**21,658**	**21,935**	**22,387**	**20,267**	**20,448**	**20,630**	**20,814**
Boats	1,678	1,906	1,946	1,952	1,767	1,783	1,799	1,815
Locomotives	85	93	90	90	81	82	83	84
Aircraft	217	224	245	233	211	213	215	217
Farm Equipment	581	628	626	621	562	567	572	578
Construction Equipment	1,090	1,132	1,047	1,041	943	951	959	968
Other*	15,805	17,676	17,981	18,449	16,702	16,851	17,001	17,153
Total	**119,480**	**97,755**	**83,680**	**79,972**	**77,382**	**74,756**	**72,269**	**69,915**

IE = Included Elsewhere
[a] CO emissions from alternative fuel highway vehicles are included under gasoline and diesel highway.
[b] Aircraft estimates include only emissions related to LTO cycles, and therefore do not include cruise altitude emissions.
[c] "Other" includes gasoline powered recreational, industrial, lawn and garden, light commercial, logging, airport service, other equipment; and diesel powered recreational, industrial, lawn and garden, light construction, airport service.
Note: Totals may not sum due to independent rounding.

Table A-106: NMVOCs Emissions from Mobile Combustion (Gg)

Fuel Type/Vehicle Type	1990	1995	2000	2001	2002	2003	2004	2005
Gasoline Highway	**8,110**	**5,819**	**4,615**	**4,285**	**4,033**	**3,819**	**3,616**	**3,424**
Passenger Cars	5,120	3,394	2,610	2,393	2,252	2,133	2,019	1,912
Light-Duty Trucks	2,374	2,019	1,750	1,664	1,566	1,483	1,404	1,330
Heavy-Duty Vehicles	575	382	232	206	194	184	174	165
Motorcycles	42	24	23	22	21	20	19	18
Diesel Highway	**406**	**304**	**216**	**207**	**195**	**185**	**175**	**166**
Passenger Cars	16	8	3	3	3	3	3	2
Light-Duty Trucks	14	9	4	4	3	3	3	3
Heavy-Duty Vehicles	377	286	209	201	189	179	169	160
Alternative Fuel Highway	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**	**IE**
Non-Highway	**2,415**	**2,622**	**2,398**	**2,379**	**2,380**	**2,299**	**2,220**	**2,144**
Boats	608	739	744	730	730	705	681	658
Locomotives	33	36	35	35	35	34	33	32
Aircraft	28	28	24	19	19	19	18	17
Farm Equipment	85	86	76	72	72	70	68	65
Construction Equipment	149	152	130	125	125	121	117	113
Other*	1,512	1,580	1,390	1,397	1,398	1,350	1,304	1,259
Total	**10,932**	**8,745**	**7,230**	**6,872**	**6,608**	**6,302**	**6,011**	**5,734**

IE = Included Elsewhere
[a] NMVOC emissions from alternative fuel highway vehicles are included under gasoline and diesel highway.
[b] Aircraft estimates include only emissions related to LTO cycles, and therefore do not include cruise altitude emissions.
[c] "Other" includes gasoline powered recreational, industrial, lawn and garden, light commercial, logging, airport service, other equipment; and diesel powered recreational, industrial, lawn and garden, light construction, airport service.
Note: Totals may not sum due to independent rounding.

Definitions of Emission Control Technologies and Standards

The N_2O and CH_4 emission factors used depend on the emission standards in place and the corresponding level of control technology for each vehicle type. Table A- 95 through Table A- 98 show the years in which these technologies or standards were in place and the penetration level for each vehicle type. These categories are defined below.

Uncontrolled

Vehicles manufactured prior to the implementation of pollution control technologies are designated as uncontrolled. Gasoline light-duty cars and trucks (pre-1973), gasoline heavy-duty vehicles (pre-1984), diesel vehicles (pre-1983), and motorcycles (pre-1996) are assumed to have no control technologies in place.

Gasoline Emission Controls

Below are the control technologies and emissions standards applicable to gasoline vehicles.

Non-catalyst

These emission controls were common in gasoline passenger cars and light-duty gasoline trucks during model years (1973-1974) but phased out thereafter, in heavy-duty gasoline vehicles beginning in the mid-1980s, and in motorcycles beginning in 1996. This technology reduces hydrocarbon (HC) and carbon monoxide (CO) emissions through adjustments to ignition timing and air-fuel ratio, air injection into the exhaust manifold, and exhaust gas recirculation (EGR) valves, which also helps meet vehicle NO_x standards.

Oxidation Catalyst

This control technology designation represents the introduction of the catalytic converter, and was the most common technology in gasoline passenger cars and light-duty gasoline trucks made from 1975 to 1980 (cars) and 1975 to 1985 (trucks). This technology was also used in some heavy-duty gasoline vehicles between 1982 and 1997. The two-way catalytic converter oxidizes HC and CO, significantly reducing emissions over 80 percent beyond non-catalyst-system capacity. One reason unleaded gasoline was introduced in 1975 was due to the fact that oxidation catalysts cannot function properly with leaded gasoline.

EPA Tier 0

This emission standard from the Clean Air Act was met through the implementation of early "three-way" catalysts, therefore this technology was used in gasoline passenger cars and light-duty gasoline trucks sold beginning in the early 1980s, and remained common until 1994. This more sophisticated emission control system improves the efficiency of the catalyst by converting CO and HC to CO_2 and H_2O, reducing NO_x to nitrogen and oxygen, and using an on-board diagnostic computer and oxygen sensor. In addition, this type of catalyst includes a fuel metering system (carburetor or fuel injection) with electronic "trim" (also known as a "closed-loop system"). New cars with three-way catalysts met the Clean Air Act's amended standards (enacted in 1977) of reducing HC to 0.41 g/mile by 1980, CO to 3.4 g/mile by 1981 and NO_x to 1.0 g/mile by 1981.

EPA Tier 1

This emission standard created through the 1990 amendments to the Clean Air Act limited passenger car NO_x emissions to 0.4 g/mi, and HC emissions to 0.25 g/mi. These bounds respectively amounted to a 60 and 40 percent reduction from the EPA Tier 0 standard set in 1981. For light-duty trucks, this standard set emissions at 0.4 to 1.1 g/mi for NO_x, and 0.25 to 0.39 g/mi for HCs, depending on the weight of the truck. Emission reductions were met through the use of more advanced emission control systems, and applied to light-duty gasoline vehicles beginning in 1994. These advanced emission control systems included advanced three-way catalysts, electronically controlled fuel injection and ignition timing, EGR, and air injection.

EPA Tier 2

This emission standard was specified in the 1990 amendments to the Clean Air Act, limiting passenger car NO_x emissions to 0.07 g/mi on average and aligning emissions standards for light-duty cars and trucks. Manufacturers can meet this average emission level by producing vehicles in 11 emission "Bins", the three highest of which expire in 2006. These new emission levels represent a 77 to 95% reduction in emissions from the EPA Tier 1 standard set in 1994. Emission reductions were met through the use of more advanced emission control systems and lower sulfur fuels and are applied to vehicles beginning in 2004. These advanced emission control systems include improved combustion, advanced three-way catalysts, electronically controlled fuel injection and ignition timing, EGR, and air injection.

Low Emission Vehicles (LEV)

This emission standard requires a much higher emission control level than the Tier 1 standard. Applied to light-duty gasoline passenger cars and trucks beginning in small numbers in the mid-1990s, LEV includes multi-port fuel injection with adaptive learning, an advanced computer diagnostics systems and advanced and close coupled catalysts with secondary air injection. LEVs as defined here include transitional low-emission vehicles (TLEVs), low emission vehicles, ultra-low emission vehicles (ULEVs) and super ultra-low emission vehicles (SULEVs). In this analysis, all categories of LEVs are treated the same due to the fact that there are very limited CH_4 or N_2O emission factor data for LEVs to distinguish among the different types of vehicles. Zero emission vehicles (ZEVs) are incorporated into the alternative fuel and advanced technology vehicle assessments.

Diesel Emission Controls

Below are the two levels of emissions control for diesel vehicles.

Moderate control

Improved injection timing technology and combustion system design for light- and heavy-duty diesel vehicles (generally in place in model years 1983 to 1995) are considered moderate control technologies. These controls were implemented to meet emission standards for diesel trucks and buses adopted by the EPA in 1985 to be met in 1991 and 1994.

Advanced control

EGR and modern electronic control of the fuel injection system are designated as advanced control technologies. These technologies provide diesel vehicles with the level of emission control necessary to comply with standards in place from 1996 through 2005.

Supplemental Information on GHG Emissions from Transportation and Other Mobile Sources

This section of this Annex includes supplemental information on the contribution of transportation and other mobile sources to U.S. greenhouse gas emissions. In the main body of the Inventory report, emission estimates are generally presented by greenhouse gas, with separate discussions of the methodologies used to estimate CO_2, N_2O, CH_4, and HFC emissions. Although the inventory is not required to provide detail beyond what is contained in the body of this report, the IPCC allows presentation of additional data and detail on emission sources. The purpose of this subannex, within the annex that details the calculation methods and data used for non-CO_2 calculations, is to provide all transportation estimates presented throughout the repot in one place.

This section of this Annex reports total greenhouse gas emissions from transportation and other (non-transportation) mobile sources in CO_2 equivalents, with information on the contribution by greenhouse gas and by mode, vehicle type, and fuel type. In order to calculate these figures, additional analyses were conducted to develop estimates of CO_2 from non-transportation mobile sources (e.g., agricultural equipment, construction equipment, recreational vehicles), and to provide more detailed breakdowns of emissions by source. This section also summarizes the methodology used to apportion CO_2 emissions to transportation modes.

Methodology for Apportioning CO_2 Emissions to Transportation Modes

Transportation-related CO_2 emissions, as presented in Table 3-7 of Chapter 3, were calculated using the methodology described in Annex 2.1. This section provides information on the methodology for apportioning CO_2 emissions to individual transportation modes and vehicle types. As noted in Annex 2.1, CO_2 emissions estimates for the transportation sector as a whole were developed for all fuel types except diesel (i.e., motor gasoline, jet fuel, aviation gasoline, residual fuel oil, natural gas, LPG, and electricity) based on transportation fuel consumption estimates from the Energy Information Administration (EIA 2004 and EIA 2006). An interagency group including EPA, EIA and FHWA that met in 2005 compared EIA estimates of transportation diesel fuel consumption with calculations using "bottom-up" (VMT) data, and determined that the "bottom-up" suggested additional diesel fuel consumption by transportation sources. As a result, since the previous inventory, the "bottom-up" data has been used directly to compute transportation diesel CO_2 emissions. Since the total diesel consumption estimate from EIA is considered to be accurate at the national level, the diesel consumption totals for the residential, commercial, and industrial sectors were adjusted downward proportionately. For all non-diesel fuels, CO_2 emissions by fuel type were apportioned to individual transportation modes (e.g., automobiles, light-duty trucks) on the basis of "bottom up" fuel consumption estimates from various data sources.

The EIA transportation fuel consumption estimates are generally consistent with the "bottom up" fuel consumption estimates, with the exceptions of diesel and jet fuel. As noted above, CO_2 from transportation diesel fuel consumption was calculated directly from the "bottom up" figures. For jet fuel, the EIA fuel consumption estimates exceed the fuel consumption estimates drawn from the following "bottom up" sources: DOT (1991 through 2006) and FAA (2006b) for commercial aircraft, FAA (2006a) for general aviation aircraft, and DESC (2006) for military aircraft. Data from these sources were used directly to calculate CO_2 from these sources.

The methodology for developing "bottom up" fuel consumption estimates for other sources is as follows. For highway vehicles, annual estimates of fuel consumption by vehicle category were taken from FHWA's *Highway Statistics'* annual editions, Table VM-1 (FHWA 1996 through 2006). For each vehicle category, the percent gasoline, diesel, and other (e.g., CNG, LPG) fuel consumption was estimated using data from the Appendix to DOE's *Transportation Energy Data Book* (DOE 2006). The highway gas and diesel fuel consumption estimates by vehicle type were then adjusted for each year so that the sum of gasoline and diesel fuel consumption across all vehicle categories matched with the fuel consumption estimates in *Highway Statistics'* Table MF-21 (FHWA 1996 through 2006). Estimates of gasoline fuel consumption from recreational boats were taken from EPA's NONROAD Model (EPA 2006d). Estimates of natural gas and LPG fuel consumption by vehicle type and mode were taken from DOE (1993 through 2006).

Estimates of diesel fuel consumption from locomotives were taken from the Association of American Railroads (AAR 2006) for Class I railroads, the Upper Great Plains Transportation Institute (Benson 2002 through 2004) and Whorton (2006) for Class II and III railroads, and DOE's *Transportation Energy Data Book* (DOE 1993 through 2006) for passenger rail. Diesel and residual fuel consumption from ships and boats were taken from EIA's *Fuel Oil and Kerosene Sales* (1991 through 2006).

Apportionment of CO_2 to Other Mobile Sources

The estimates of N_2O and CH_4 from fuel combustion presented in the Energy chapter of the inventory include both transportation sources and other mobile sources. Other mobile sources include construction equipment, agricultural equipment, vehicles used off-road, and other sources that have utility associated with their movement but do not have a primary purpose of transporting people or goods (e.g., snowmobiles, riding lawnmowers, etc.). Estimates of CO_2 from non-transportation mobile sources, based on EIA fuel consumption data, are included in the agricultural, industrial, and commercial sectors. In order to provide comparable information on transportation and mobile sources, Table A- 107 provides estimates of CO_2 from these other mobile sources. These estimates were developed using the same data sources utilized in developing the N_2O and CH_4 estimates and using the methodology for estimating CO_2 described in Annex 2.1.

Table A- 107: CO_2 Emissions from Non-Transportation Mobile Sources (Tg CO_2 Eq.)

Fuel Type/Vehicle Type	1990	1995	2000	2001	2002	2003	2004	2005
Farm Equipment	30.8	36.3	38.4	40.6	41.7	42.8	45.7	46.5
Construction Equipment	41.6	48.5	54.7	59.4	61.1	62.9	64.7	65.7
Total Other	55.5	61.1	64.4	72.0	73.7	75.6	77.6	77.8
Total	**127.9**	**145.9**	**157.6**	**172.1**	**176.6**	**181.3**	**188.1**	**190.0**

Contribution of Transportation and Mobile Sources to Greenhouse Gas Emissions, by Mode/Vehicle Type/Fuel Type

Table A- 108 presents estimates of greenhouse gas emissions from all transportation and other mobile sources in CO_2 equivalents. In total, transportation and mobile sources emitted 2,203.5 Tg CO_2 Eq. in 2005, an increase of 33 percent from 1990. These estimates were generated using the estimates of CO_2 emissions from transportation sources reported in Table 3-6, CH_4 emissions reported in Table 3-21, and N_2O emissions reported in Table 3-22 of Chapter 3; information on HFCs from mobile air conditioners, comfort cooling for trains and buses, and refrigerated transportation from Chapter 4; and estimates of CO_2 emitted from non-transportation mobile sources reported in Table A- 107 above.

Although all emissions reported here are based on estimates reported throughout this Inventory, some additional calculations were performed in order to provide a detailed breakdown of emissions by mode and vehicle category. In the case of N_2O and CH_4, additional calculations were performed to develop emissions estimates by type of aircraft and type of heavy-duty vehicle (i.e., heavy-duty trucks or buses) to match the level of detail for CO_2 emissions. Nitrous oxide and CH_4 estimates were developed for individual aircraft types by multiplying the emissions estimates for aircraft for each fuel type (jet fuel and aviation gasoline) by the portion of fuel used by each aircraft type (from FAA 2006a and 2006b). Similarly, N_2O and CH_4 estimates were developed for heavy-duty trucks and buses by multiplying the emission estimates for heavy-duty vehicles for each fuel type (gasoline, diesel) from Table 3-21 and Table 3-22 of Chapter 3 by the portion of fuel used by each vehicle type (from DOE 1993-2006). Otherwise, the table and figure are drawn directly from emission estimates presented elsewhere in the inventory, and are dependent on the methodologies presented in Annex 2.1 (for CO_2), Chapter 4, and Annex 3.8 (for HFCs), and earlier in this Annex (for CH_4 and N_2O).

Transportation sources include highway (on-road) vehicles, aircraft, boats and ships, rail, and pipelines (note: pipelines are a transportation source but are stationary, not mobile sources). In addition, transportation-related greenhouse gas emissions also include HFC released from mobile air conditioners and refrigerated transportation, and the release of CO_2 from lubricants (such as motor oil) used in transportation. Together, transportation sources were responsible for 2,015.3 Tg CO_2 Eq. in 2005.

On-road vehicles were responsible for about 73 percent of transportation GHG emissions in 2005. Although passenger cars make up the largest component of on-road vehicle greenhouse gas emissions, light-duty and heavy-duty trucks have been the primary sources of growth in on-road vehicle emissions. Between 1990 and 2005, greenhouse gas emissions from passenger cars decreased 2 percent, while emissions from light-duty trucks increased 72 percent, largely due to the increased use of sport-utility vehicles and other light-duty trucks. Meanwhile, greenhouse gas emissions from heavy-duty trucks increased 69 percent, reflecting the increased volume of total freight movement and an increasing share transported by trucks.

In contrast to other transportation sources, aircraft saw only a modest (four percent) increase in GHG emissions between 1990 and 2005, despite a substantial rise in passenger miles traveled. The small increase reflected a large decline in emissions from military aircraft (50 percent) and a 16 percent increase in emissions from commercial aircraft. Greenhouse gas emissions from commercial aircraft rose 21 percent between 1990 and 2000, but then declined in 2001, 2002 and 2003, due largely to a decrease in air travel following the September 11, 2001 terrorist attacks.

Non-transportation mobile sources, such as construction equipment, agricultural equipment, and industrial/commercial equipment, emitted approximately 191.8 Tg CO_2 Eq. in 2005. Together, these sources emitted more greenhouse gases than boats and ships (domestic travel in the United States), rail, and pipelines combined. Emissions from non-transportation mobile sources increased rapidly, growing approximately 49 percent between 1990 and 2005.

Contribution of Transportation and Mobile Sources to Greenhouse Gas Emissions, by Gas

Table A- 109 presents estimates of greenhouse gas emissions from transportation and other mobile sources broken down by greenhouse gas. As this table shows, CO_2 accounts for the vast majority of transportation greenhouse gas emissions (approximately 95 percent in 2005). Emissions of CO_2 from transportation and mobile sources increased by 493.0 Tg CO_2 Eq. between 1990 and 2005. In contrast, the combined emissions of CH_4 and N_2O decreased by 7.8 Tg CO_2 Eq. over the same period, due largely to the introduction of control technologies

designed to reduce criteria pollutant emissions.[31] Meanwhile, HFC emissions from mobile air conditioners and refrigerated transport increased from virtually no emissions in 1990 to 66.7 Tg CO_2 Eq. in 2005 as these chemicals were phased in as substitutes for ozone depleting substances. It should be noted, however, that the ozone depleting substances that HFCs replaced are also powerful greenhouse gases, but are not included in national greenhouse gas inventories due to their mandated phase out.

Greenhouse Gas Emissions from Freight and Passenger Transportation

Table A- 110 and Table A- 111 present greenhouse gas estimates from transportation, broken down into the passenger and freight categories. Passenger modes include light-duty vehicles, buses, passenger rail, aircraft (general and commercial aviation), recreational boats, and mobile air conditioners, and are illustrated in Table A- 110. Freight modes include heavy-duty trucks, freight rail, refrigerated transport, waterborne freight vessels, and pipelines, and are illustrated in Table A- 111. Note that although aircraft do carry some freight, separating out the emissions associated with freight versus passenger aircraft travel is difficult, and so general and commercial aviation are counted as passenger transportation for purposes of this analysis. The remaining transportation and mobile emissions were from sources not considered to be either freight or passenger modes (e.g., construction and agricultural equipment, lubricants).

The estimates in these tables are drawn from the estimates in Table A- 108. In addition, estimates of fuel consumption from DOE (1993 through 2004) were used to allocate rail emissions between passenger and freight categories.

In 2005, passenger transportation modes emitted 1,462.9 Tg CO_2 Eq., while freight transportation modes emitted 525.0 Tg CO_2 Eq. The rate of growth of greenhouse gas emissions from freight sources, however, was more than twice as fast, due largely to the rapid increase in emissions associated with heavy-duty trucks.

Table A- 108: Total U.S. Greenhouse Gas Emissions from Transportation and Mobile Sources (Tg CO_2 Eq.)

Mode / Vehicle Type / Fuel Type	1990	1995	2000	2001	2002	2003	2004	2005	Percent of total	Percent Change 1990-2005
Transportation Total[a]	**1,526.1**	**1,680.4**	**1,906.7**	**1,879.8**	**1,934.6**	**1,932.4**	**1,987.0**	**2,015.6**	**91%**	**32%**
On-Road Vehicles	**1,210.6**	**1,336.7**	**1,499.4**	**1,501.0**	**1,541.3**	**1,549.6**	**1,586.7**	**1,602.5**	**73%**	**32%**
Passenger Cars	643.2	628.6	658.3	659.4	672.8	650.7	648.7	632.9	29%	-2%
Gasoline	635.4	620.8	654.7	655.7	669.0	646.5	644.3	628.5	28%	-1%
Diesel	7.8	7.7	3.6	3.7	3.7	4.2	4.3	4.4	<0.5%	-44%
AFVs	+	0.1	+	+	+	+	+	+	<0.5%	793%
Light-Duty Trucks	329.4	425.2	483.6	485.1	497.6	529.8	551.4	566.7	26%	72%
Gasoline	317.5	410.0	466.0	467.7	479.7	507.6	527.6	541.2	25%	70%
Diesel	11.3	14.7	17.4	17.1	17.5	21.9	23.4	25.0	1%	121%
AFVs	0.6	0.5	0.3	0.3	0.4	0.4	0.5	0.4	<0.5%	-23%
Medium/Heavy-Duty Trucks	227.7	271.9	344.5	344.5	359.3	356.7	369.8	385.8	17%	69%
Gasoline	38.4	36.2	36.2	35.1	35.7	30.8	31.2	31.3	1%	-18%
Diesel	188.5	235.1	307.8	308.8	323.0	325.1	337.8	353.7	16%	88%
AFVs	0.9	0.6	0.5	0.6	0.6	0.7	0.8	0.8	<0.5%	-11%
Buses	8.5	9.3	11.2	10.3	9.9	10.8	15.1	15.3	1%	79%
Gasoline	0.6	0.6	0.5	0.5	0.4	0.4	0.5	0.5	<0.5%	-10%
Diesel	7.9	8.6	10.2	9.3	8.8	9.6	13.8	14.0	1%	76%
AFVs	+	0.1	0.5	0.5	0.7	0.8	0.8	0.8	<0.5%	13115%
Motorcycles	1.7	1.8	1.9	1.7	1.7	1.6	1.7	1.8	<0.5%	6%
Gasoline	1.7	1.8	1.9	1.7	1.7	1.6	1.7	1.8	<0.5%	6%
Aircraft	**181.9**	**176.4**	**198.5**	**188.5**	**179.8**	**176.6**	**181.6**	**189.2**	**9%**	**4%**
General Aviation Aircraft	9.6	8.2	11.9	11.6	12.0	11.1	11.6	12.2	1%	28%
Jet Fuel	6.4	5.4	9.3	9.2	9.6	8.9	9.4	9.7	<0.5%	52%
Aviation Gasoline	3.2	2.8	2.6	2.5	2.4	2.1	2.2	2.5	<0.5%	-21%

[31] The decline in CFC emissions is not captured in the official transportation estimates.

Commercial Aircraft	137.6	144.2	165.9	154.2	147.3	145.4	148.7	159.7	7%	16%
Jet Fuel	137.6	144.2	165.9	154.2	147.3	145.4	148.7	159.7	7%	16%
Military Aircraft (Jet Fuel)	34.6	24.0	20.7	22.7	20.6	20.1	21.3	17.3	1%	-50%
Jet Fuel	34.6	24.0	20.7	22.7	20.6	20.1	21.3	17.3	1%	-50%
Boats and Ships	**47.2**	**55.9**	**64.4**	**43.4**	**61.2**	**53.8**	**61.7**	**64.8**	**3%**	**37%**
Gasoline	12.5	14.1	14.6	14.6	14.5	14.4	14.4	14.4	1%	15%
Distillate Fuel	10.8	11.0	14.6	16.2	15.9	13.0	16.6	13.7	1%	26%
Residual Fuel	23.9	30.8	35.3	12.7	30.8	26.4	30.7	36.6	2%	53%
Rail	**38.5**	**42.6**	**45.5**	**45.5**	**45.3**	**47.0**	**49.6**	**50.8**	**2%**	**32%**
Distillate Fuel	35.4	39.6	42.0	42.2	41.9	42.8	45.2	45.6	2%	29%
Electricity	3.0	3.1	3.4	3.3	3.4	4.3	4.4	5.2	<0.5%	70%
Pipelines[b]	**36.1**	**38.3**	**35.2**	**34.4**	**36.6**	**32.7**	**31.5**	**31.1**	**1%**	**-14%**
Natural Gas	36.1	38.3	35.2	34.4	36.6	32.7	31.5	31.1	1%	-14%
Other Transportation	**11.9**	**30.5**	**63.7**	**66.9**	**70.4**	**72.7**	**75.9**	**77.2**	**3%**	**552%**
Mobile AC	+	16.8	41.6	44.9	47.7	50.0	52.2	53.1	2%	NA
Comfort Cooling for Trains and Buses	+	+	0.2	0.2	0.2	0.2	0.3	0.3	<0.5%	NA
Refrigerated Transport	+	2.3	9.8	10.8	11.5	12.3	13.1	13.6	1%	NA
Lubricants	11.9	11.3	12.1	11.1	10.9	10.1	10.2	10.2	<0.5%	-14%
Non-Trans Mobile Total	**129.1**	**147.3**	**159.1**	**173.7**	**178.2**	**183.0**	**189.8**	**191.8**	**9%**	**49%**
Agricultural Equipment	**31.1**	**36.7**	**38.9**	**41.1**	**42.2**	**43.2**	**46.2**	**47.0**	**2%**	**51%**
Gasoline	7.3	8.3	5.8	7.2	7.4	7.6	9.8	9.6	<0.5%	32%
Diesel	23.8	28.4	33.0	33.9	34.8	35.6	36.5	37.3	2%	57%
Construction Equipment	**42.0**	**48.9**	**55.2**	**59.9**	**61.6**	**63.5**	**65.3**	**66.3**	**3%**	**58%**
Gasoline	4.4	4.0	3.1	5.9	6.1	6.4	6.7	6.2	<0.5%	43%
Diesel	37.6	44.9	52.2	54.1	55.6	57.1	58.6	60.1	3%	60%
Other Equipment[c]	**56.0**	**61.7**	**65.0**	**72.7**	**74.4**	**76.3**	**78.3**	**78.5**	**4%**	**40%**
Gasoline	41.5	44.1	44.4	51.3	52.3	53.5	54.9	54.4	2%	31%
Diesel	14.6	17.6	20.6	21.4	22.1	22.8	23.4	24.1	1%	65%
Total Transportation & Mobile	**1,655.2**	**1,827.6**	**2,065.8**	**2,053.4**	**2,112.8**	**2,115.4**	**2,176.8**	**2,207.4**	**100%**	**33%**

[a] Not including emissions from international bunker fuels.
[b] Includes only CO_2 from natural gas used to power pipelines; does not include emissions from electricity use or non-CO_2 gases.
[c] "Other" includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment.
+ Less than 0.05 Tg CO_2 Eq.
NA = Not Applicable, as there were no HFC emissions allocated to the transport sector in 1990, and thus a growth rate cannot be calculated.

Table A-109: Transportation and Mobile Source Emissions by Gas (Tg CO_2 Eq.)

Gas	1990	1995	2000	2001	2002	2003	2004	2005
CO_2	1,606.8	1,750.4	1,957.5	1,944.7	2,003.2	2,006.1	2,067.2	2,099.7
N_2O	43.7	53.7	53.2	49.7	47.1	43.8	41.2	38.0
CH_4	4.7	4.3	3.5	3.2	3.1	2.9	2.8	2.6
HFC	+	19.2	51.6	55.8	59.4	62.5	65.6	67.1
Total	**1,655.2**	**1,827.6**	**2,065.6**	**2,053.3**	**2,112.6**	**2,115.1**	**2,176.5**	**2,207.1**

NA = Not Applicable, as there were no HFC emissions allocated to the transport sector in 1990, and thus a growth rate cannot be calculated.

Figure A-4: Domestic Greenhouse Gas Emissions by Mode and Vehicle Type, 1990 to 2005 (Tg CO_2 Eq.)

Table A-110: Greenhouse Gas Emissions from Passenger Transportation (Tg CO_2 Eq.)

Vehicle Type	1990	1995	2000	2001	2002	2003	2004	2005	% Change 1990-2005

Highway Vehicles	**982.8**	**1,064.8**	**1,155.0**	**1,156.5**	**1,182.0**	**1,193.0**	**1,216.9**	**1,216.7**	**24%**
Passenger Cars	643.2	628.6	658.3	659.4	672.8	650.7	648.7	632.9	-2%
Light-duty Trucks	329.4	425.2	483.6	485.1	497.6	529.8	551.4	566.7	72%
Buses	8.5	9.3	11.2	10.3	9.9	10.8	15.1	15.3	79%
Motorcycles	1.7	1.8	1.9	1.7	1.7	1.6	1.7	1.8	6%
Aircraft	**147.2**	**152.4**	**177.8**	**165.8**	**159.2**	**156.4**	**160.3**	**172.0**	**17%**
General Aviation	9.6	8.2	11.9	11.6	12.0	11.1	11.6	12.2	28%
Commercial Aviation	137.7	144.2	165.9	154.2	147.3	145.4	148.7	159.7	16%
Recreational Boats	**12.5**	**14.1**	**14.6**	**14.6**	**14.5**	**14.4**	**14.4**	**14.4**	**15%**
Passenger Rail	**4.3**	**4.4**	**5.1**	**5.1**	**5.0**	**5.8**	**5.9**	**6.7**	**54%**
Mobile Air Conditioners	**-**	**16.8**	**41.6**	**44.9**	**47.7**	**50.0**	**52.2**	**53.1**	**NA**
Comfort Cooling for Trains and Buses	**+**	**+**	**0.2**	**0.2**	**0.2**	**0.2**	**0.3**	**0.3**	**NA**
Total	**1,146.9**	**1,252.6**	**1,394.2**	**1,387.0**	**1,408.7**	**1,419.8**	**1,450.0**	**1,462.2**	**28%**

Note: Data from DOE (1993 through 2006) were used to disaggregate emissions from rail and buses.

Table A- 111: Greenhouse Gas Emissions from Domestic Freight Transportation (Tg CO_2 Eq.)

By Mode	1990	1995	2000	2001	2002	2003	2004	2005	% Change 1990-2005
Trucking	227.7	271.9	344.5	344.5	359.3	356.7	369.8	385.8	69%
Freight Rail	34.1	38.2	40.4	40.5	40.3	41.3	43.7	44.1	29%
Waterborne	34.7	41.8	49.8	28.8	46.6	39.4	47.3	50.3	45%
Refrigerated Transport	+	2.3	9.8	10.8	11.5	12.3	13.1	13.6	NA
Pipelines	36.1	38.2	35.2	34.4	36.6	32.7	31.5	31.1	-14%
Total	**332.7**	**392.3**	**479.7**	**459.0**	**494.4**	**482.3**	**505.4**	**525.0**	**52%**

+ Less than 0.05 Tg CO_2 Eq.
NA = Not Applicable, as there were no HFC emissions allocated to the transport sector in 1990, and thus a growth rate cannot be calculated.
Note: Data from DOE (1993 through 2006) were used to allocate the passenger/freight split of rail emissions.

3.3. Methodology for Estimating CH_4 Emissions from Coal Mining

The methodology for estimating CH_4 emissions from coal mining consists of two distinct steps. The first step addresses emissions from underground mines. For these mines, emissions are estimated on a mine-by-mine basis and then are summed to determine total emissions. The second step of the analysis involves estimating CH_4 emissions for surface mines and post-mining activities. In contrast to the methodology for underground mines, which uses mine-specific data, the surface mine and post-mining activities analysis consists of multiplying basin-specific coal production by basin-specific emission factors.

Step 1: Estimate CH_4 Liberated and CH_4 Emitted from Underground Mines

Underground mines generate CH_4 from ventilation systems and from degasification systems. Some mines recover and use CH_4 generated from degasification systems, thereby reducing emissions to the atmosphere. Total CH_4 emitted from underground mines equals the CH_4 liberated from ventilation systems, plus the CH_4 liberated from degasification systems, minus CH_4 recovered and used.

Step 1.1: Estimate CH_4 Liberated from Ventilation Systems

All coal mines with detectable CH_4 emissions[32] use ventilation systems to ensure that CH_4 levels remain within safe concentrations. Many coal mines do not have detectable levels of CH_4, while others emit several million cubic feet per day (MMCFD) from their ventilation systems. On a quarterly basis, the U.S. Mine Safety and Health Administration (MSHA) measures CH_4 emissions levels at underground mines. MSHA maintains a database of measurement data from all underground mines with detectable levels of CH_4 in their ventilation air. Based on the four quarterly measurements, MSHA estimates average daily CH_4 liberated at each of the underground mines with detectable emissions.

For the years 1990 through 1996 and 1998 through 2005, MSHA emissions data were obtained for a large but incomplete subset of all mines with detectable emissions. This subset includes mines emitting at least 0.1 MMCFD for some years and at least 0.5 MMCFD for other years, as shown in Table A- 112. Well over 90 percent of all ventilation emissions were concentrated in these subsets. For 1997, the complete MSHA database for all 586 mines with detectable CH_4 emissions was obtained. These mines were assumed to account for 100 percent of CH_4 liberated from underground mines. Using the complete database from 1997, the proportion of total emissions accounted for by mines emitting less than 0.1 MMCFD or 0.5 MMCFD was estimated (see Table A- 112). The proportion was then applied to the years 1990 through 2005 to account for the less than 10 percent of ventilation emissions coming from mines without MSHA data.

For 1990 through 1999, average daily CH_4 emissions were multiplied by the number of days in the year (i.e., coal mine assumed in operation for all four quarters) to determine the annual emissions for each mine. For 2000 through 2005, MSHA provided quarterly emissions. The average daily CH_4 emissions were multiplied by the number of days corresponding to the number of quarters the mine vent was operating. For example, if the mine vent was operational in one out of the four quarters, the average daily CH_4 emissions were multiplied by 92 days. Total ventilation emissions for a particular year were estimated by summing emissions from individual mines.

[32] MSHA records coal mine methane readings with concentrations of greater than 50 ppm (parts per million) methane. Readings below this threshold are considered non-detectable.

Table A- 112: Mine-Specific Data Used to Estimate Ventilation Emissions

Year	Individual Mine Data Used
1990	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
1991	1990 Emissions Factors Used Instead of Mine-Specific Data
1992	1990 Emissions Factors Used Instead of Mine-Specific Data
1993	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
1994	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
1995	All Mines Emitting at Least 0.5 MMCFD (Assumed to Account for 94.1% of Total)*
1996	All Mines Emitting at Least 0.5 MMCFD (Assumed to Account for 94.1% of Total)*
1997	All Mines with Detectable Emissions (Assumed to Account for 100% of Total)
1998	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
1999	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
2000	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
2001	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
2002	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
2003	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
2004	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*
2005	All Mines Emitting at Least 0.1 MMCFD (Assumed to Account for 97.8% of Total)*

* Factor derived from a complete set of individual mine data collected for 1997.

Step 1.2: Estimate CH_4 Liberated from Degasification Systems

Coal mines use several different types of degasification systems to remove CH_4, including vertical wells and horizontal boreholes to recover CH_4 prior to mining of the coal seam. Gob wells and cross-measure boreholes recover CH_4 from the overburden (i.e., GOB area) after mining of the seam (primarily in longwall mines).

MSHA collects information about the presence and type of degasification systems in some mines, but does not collect quantitative data on the amount of CH_4 liberated. Thus, the methodology estimated degasification emissions on a mine-by-mine basis based on other sources of available data. Many of the coal mines employing degasification systems have provided EPA with information regarding CH_4 liberated from their degasification systems. For these mines, this reported information was used as the estimate. In other cases in which mines sell CH_4 recovered from degasification systems to a pipeline, gas sales were used to estimate CH_4 liberated from degasification systems (see Step 1.3). Finally, for those mines that do not sell CH_4 to a pipeline and have not provided information to EPA, CH_4 liberated from degasification systems was estimated based on the type of system employed. For example, for coal mines employing gob wells and horizontal boreholes, the methodology assumes that degasification emissions account for 40 percent of total CH_4 liberated from the mine.

Step 1.3: Estimate CH_4 Recovered from Degasification Systems and Used (Emissions Avoided)

In 2005, fifteen active coal mines had CH_4 recovery and use projects, thirteen sold the recovered CH_4 to a pipeline and one used the CH_4 on site to heat mine ventilation air. One coal mine also used some recovered CH_4 in a thermal dryer in addition to selling gas to a pipeline. In order to calculate emissions avoided from pipeline sales, information was needed regarding the amount of gas recovered and the number of years in advance of mining that wells were drilled. Several state agencies provided gas sales data, which were used to estimate emissions avoided for these projects. Additionally, coal mine operators provided information on gas sales and/or the number of years in advance of mining. Emissions avoided were attributed to the year in which the coal seam was mined. For example, if a coal mine recovered and sold CH_4 using a vertical well drilled five years in advance of mining, the emissions avoided associated with those gas sales (cumulative production) were attributed to the well up to the time it was mined through (e.g., five years of gas production). Where individual well data is not available, estimated percentages of the operator's annual gas sales within the field around the coal mine are attributed to emissions avoidance. For some mines, individual well data were used to assign gas sales to the appropriate emissions avoided year. In most cases, coal mine operators provided this information, which was then used to estimate emissions avoided for a particular year. Additionally, several state agencies provided production data for individual wells.

Step 2: Estimate CH_4 Emitted from Surface Mines and Post-Mining Activities

Mine-specific data were not available for estimating CH_4 emissions from surface coal mines or for post-mining activities. For surface mines and post-mining activities, basin-specific coal production was multiplied by a basin-specific emission factor to determine CH_4 emissions.

Step 2.1: Define the Geographic Resolution of the Analysis and Collect Coal Production Data

The first step in estimating CH_4 emissions from surface mining and post-mining activities was to define the geographic resolution of the analysis and to collect coal production data at that level of resolution. The analysis was conducted by coal basin as defined in Table A- 113, which presents coal basin definitions by basin and by state.

The Energy Information Administration's (EIA) Annual Coal Report includes state- and county-specific underground and surface coal production by year. To calculate production by basin, the state level data were grouped into coal basins using the basin definitions listed in Table A- 113. For two states—West Virginia and Kentucky—county-level production data was used for the basin assignments because coal production occurred from geologically distinct coal basins within these states. Table A- 114 presents the coal production data aggregated by basin.

Step 2.2: Estimate Emissions Factors for Each Emissions Type

Emission factors for surface mined coal were developed from the *in situ* CH_4 content of the surface coal in each basin. Based on an analysis presented in EPA (1993), surface mining emission factors were estimated to be from 1 to 3 times the average *in situ* CH_4 content in the basin. For this analysis, the surface mining emission factor was determined to be twice the *in situ* CH_4 content in the basin. Furthermore, the post-mining emission factors used were estimated to be 25 to 40 percent of the average *in situ* CH_4 content in the basin. For this analysis, the post-mining emission factor was determined to be 32.5 percent of the *in situ* CH_4 content in the basin. Table A- 115 presents the average *in situ* content for each basin, along with the resulting emission factor estimates.

Step 2.3: Estimate CH_4 Emitted

The total amount of CH_4 emitted was calculated by multiplying the coal production in each basin by the appropriate emission factors.

Total annual CH_4 emissions are equal to the sum of underground mine emissions plus surface mine emissions plus post-mining emissions. Table A- 116 and Table A- 117 present estimates of CH_4 liberated, used, and emitted for 1990 through 2005. Table A- 118 provides emissions by state.

Table A- 113: Coal Basin Definitions by Basin and by State

Basin	States
Northern Appalachian Basin	Maryland, Ohio, Pennsylvania, West Virginia North
Central Appalachian Basin	Kentucky East, Tennessee, Virginia, West Virginia South
Warrior Basin	Alabama, Mississippi
Illinois Basin	Illinois, Indiana, Kentucky West
South West and Rockies Basin	Arizona, California, Colorado, New Mexico, Utah
North Great Plains Basin	Montana, North Dakota, Wyoming
West Interior Basin	Arkansas, Iowa, Kansas, Louisiana, Missouri, Oklahoma, Texas
Northwest Basin	Alaska, Washington
State	**Basin**
Alabama	Warrior Basin
Alaska	Northwest Basin
Arizona	South West and Rockies Basin
Arkansas	West Interior Basin
California	South West and Rockies Basin
Colorado	South West and Rockies Basin
Illinois	Illinois Basin
Indiana	Illinois Basin
Iowa	West Interior Basin
Kansas	West Interior Basin

Kentucky East	Central Appalachian Basin
Kentucky West	Illinois Basin
Louisiana	West Interior Basin
Maryland	Northern Appalachian Basin
Mississippi	Warrior Basin
Missouri	West Interior Basin
Montana	North Great Plains Basin
New Mexico	South West and Rockies Basin
North Dakota	North Great Plains Basin
Ohio	Northern Appalachian Basin
Oklahoma	West Interior Basin
Pennsylvania.	Northern Appalachian Basin
Tennessee	Central Appalachian Basin
Texas	West Interior Basin
Utah	South West and Rockies Basin
Virginia	Central Appalachian Basin
Washington	Northwest Basin
West Virginia South	Central Appalachian Basin
West Virginia North	Northern Appalachian Basin
Wyoming	North Great Plains Basin

Table A-114: Annual Coal Production (Thousand Short Tons)

Basin	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Underground Coal Production	**423,556**	**406,344**	**406,335**	**351,056**	**399,102**	**396,249**	**409,850**	**420,657**	**417,729**	**391,791**	**372,766**	**380,627**	**357,384**	**352,785**	**367,531**	**368,611**
N. Appalachia	103,865	103,450	105,220	77,032	100,122	98,103	106,729	112,135	116,718	107,575	105,374	107,025	98,643	98,369	106,915	111,151
Cent. Appalachia	198,412	181,873	177,777	164,845	170,893	166,495	171,845	177,720	171,279	157,058	150,584	152,457	137,224	130,724	128,560	123,083
Warrior	17,531	17,062	15,944	15,557	14,471	17,605	18,217	18,505	17,316	14,799	15,895	15,172	14,916	15,375	16,114	13,295
Illinois	69,167	69,947	73,154	55,967	69,050	69,009	67,046	64,728	64,463	63,529	53,720	54,364	54,016	51,780	56,319	59,180
S. West/Rockies	32,754	31,568	31,670	35,409	41,681	42,994	43,088	44,503	45,983	46,957	45,742	51,193	52,121	56,111	59,012	60,865
N. Great Plains	1,722	2,418	2,511	2,146	2,738	2,018	2,788	2,854	1,723	1,673	1,210	0	0	32	201	572
West Interior	105	26	59	100	147	25	137	212	247	200	241	416	464	394	410	465
Northwest	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0	0
Surface Coal Production	**602,753**	**587,143**	**588,944**	**594,372**	**634,401**	**636,726**	**654,007**	**669,271**	**699,608**	**708,639**	**700,608**	**745,306**	**735,912**	**717,689**	**743,553**	**762,191**
N. Appalachia	60,761	51,124	50,512	48,641	44,960	39,372	39,788	40,179	41,043	33,928	34,908	35,334	30,008	27,370	28,174	28,873
Cent. Appalachia	94,343	91,785	95,163	94,433	106,129	106,250	108,869	113,275	108,345	107,507	110,479	116,983	111,340	99,419	103,968	112,222
Warrior	11,413	10,104	9,775	9,211	8,795	7,036	6,420	5,963	5,697	4,723	4,252	4,796	6,320	8,437	9,742	11,599
Illinois	72,000	63,483	58,814	50,535	51,868	40,376	44,754	46,862	47,715	40,474	33,631	40,894	39,380	36,675	34,016	33,702
S. West/Rockies	43,863	42,985	46,052	48,765	49,119	46,643	43,814	48,374	49,635	50,349	49,587	52,180	50,006	41,237	42,558	42,756
N. Great Plains	249,356	259,194	258,281	275,873	308,279	331,367	343,404	349,612	385,438	407,683	407,670	438,367	441,346	444,007	466,224	474,056
West Interior	64,310	61,889	63,562	60,574	58,791	59,116	60,912	59,061	57,951	58,309	54,170	50,613	50,459	53,411	51,706	52,263
Northwest	6,707	6,579	6,785	6,340	6,460	6,566	6,046	5,945	5,982	5,666	5,911	6,138	6,973	7,313	7,165	6,720
Total Coal Production	**1,026,309**	**993,487**	**995,279**	**945,428**	**1,033,503**	**1,032,975**	**1,063,857**	**1,089,928**	**1,118,132**	**1,093,975**	**1,073,374**	**1,127,689**	**1,093,296**	**1,070,654**	**1,111,111**	**1,130,802**
N. Appalachia	164,626	154,574	155,732	125,673	145,082	137,475	146,517	152,314	157,761	141,145	140,282	142,360	128,731	125,739	135,089	140,024
Cent. Appalachia	292,755	273,658	272,940	259,278	277,022	272,745	280,714	290,995	279,624	262,660	261,063	269,440	248,564	230,143	232,528	235,305
Warrior	28,944	27,166	25,719	24,768	23,266	24,641	24,637	24,468	23,013	19,499	20,147	19,967	21,236	23,812	25,856	24,894
Illinois	141,167	133,430	131,968	106,502	120,918	109,385	111,800	111,590	110,176	103,966	87,351	95,258	93,396	88,455	90,335	92,882
S. West/Rockies	76,617	74,553	77,722	84,174	90,800	89,637	86,902	92,877	95,618	96,207	95,239	103,373	102,127	97,348	101,570	103,621
N. Great Plains	251,078	261,612	260,792	278,019	311,017	333,385	346,192	352,466	387,161	406,324	408,880	438,367	441,346	444,039	466,425	474,628
West Interior	64,415	61,915	63,621	60,674	58,938	59,141	61,049	59,273	58,198	58,509	54,411	51,028	50,923	53,805	52,116	52,728
Northwest	6,707	6,579	6,785	6,340	6,460	6,566	6,046	5,945	5,982	5,665	5,911	6,138	6,973	7,313	7,165	6,720

Source for 1990-2005 data: EIA (1990 through 06), Annual Coal Report. U.S. Department of Energy, Washington, DC, Table 3.
Note: Totals may not sum due to independent rounding.

Table A-115: Coal Surface and Post-Mining CH_4 Emission Factors (ft^3 per Short Ton)

Basin	Surface Average *in situ* Content	Underground Average *In situ* Content	Surface Mine Factors	Post-Mining Surface Factors	Post Mining Underground
Northern Appalachia	59.5	138.4	119.0	19.3	45.0
Central Appalachia (WV)	24.9	136.8	49.8	8.1	44.5
Central Appalachia (VA)	24.9	399.1	49.8	8.1	129.7
Central Appalachia (E KY)	24.9	61.4	49.8	8.1	20.0

Warrior	30.7	266.7	61.4	10.0	86.7
Illinois	34.3	64.3	68.6	11.1	20.9
Rockies (Piceance Basin)	33.1	196.4	66.2	10.8	63.8
Rockies (Uinta Basin)	16.0	99.4	32.0	5.2	32.3
Rockies (San Juan Basin)	7.3	104.8	14.6	2.4	34.1
Rockies (Green River Basin)	33.1	247.2	66.2	10.8	80.3
Rockies (Raton Basin)	33.1	127.9	66.2	10.8	41.6
N. Great Plains	5.6	15.8	11.2	1.8	5.1
West Interior (Forest City, Cherokee Basins)	34.3	64.3	68.6	11.1	20.9
West Interior (Arkoma Basin)	74.5	331.2	149.0	24.2	107.6
West Interior (Gulf Coast Basin)	33.1	127.9	66.2	10.8	41.6
Northwest (AK)	5.6	160.0	11.2	1.8	52.0
Northwest (WA)	5.6	47.3	11.2	1.8	18.9

Source: 1986 USBM Circular 9067, *Results of the Direct Method Determination of the Gas Contents of U.S. Coal Basins*, 1983 U.S. DOE Report (DOE/METC/83-76), *Methane Recovery from Coalbeds: A Potential Energy Source*, 1986-88 Gas Research Institute Topical Reports, A Geologic Assessment of Natural Gas from Coal Seams.

Table A- 116: Underground Coal Mining CH_4 Emissions (Billion Cubic Feet)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Ventilation Output	112	NA	NA	95	96	97	90	96	94	92	87	84	79	76	82	76
Adjustment Factor for Mine Data*	97.8%	NA	NA	97.8%	97.8%	91.4%	91.4%	100%	97.8%	97.8%	97.8%	97.8%	97.8%	97.8%	97.8%	97.8%
Adjusted Ventilation Output	114	NA	NA	97	98	106	99	96	96	94	89	86	80	77	84	78
Degasification System Liberated	54	NA	NA	45	46	47	49	42	49	40	45	49	50	50	50	48
Total Underground Liberated	168	164	162	142	144	153	148	139	146	134	134	135	131	127	134	126
Recovered & Used	(14)	(14)	(16)	(23)	(27)	(31)	(35)	(28)	(35)	(31)	(37)	(41)	(43)	(38)	(40)	(37)
Total	**154**	**150**	**146**	**119**	**117**	**122**	**113**	**110**	**110**	**103**	**97**	**94**	**88**	**89**	**94**	**88**

* Refer to Table A- 112.

Note: Totals may not sum due to independent rounding.

Table A- 117: Total Coal Mining CH_4 Emissions (Billion Cubic Feet)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Underground Mining	154	149	144	119	117	122	113	110	110	103	97	94	88	89	94	88
Surface Mining	26	24	24	23	23	22	23	23	23	22	22	23	22	21	21	22
Post-Mining (Underground)	19	18	18	16	17	17	18	18	18	17	17	17	16	16	16	16
Post-Mining (Surface)	4	4	4	4	4	4	4	4	4	4	4	4	4	3	4	4
Total	**203**	**196**	**191**	**162**	**162**	**165**	**157**	**156**	**156**	**146**	**139**	**138**	**129**	**129**	**135**	**130**

Note: Totals may not sum due to independent rounding.

Table A- 118: Total Coal Mining CH_4 Emissions by State (Million Cubic Feet)

State	1990		1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Alabama	32,272		29,618	33,615	29,556	26,002	26,429	25,677	23,282	21,896	18,686	19,288	22,499	16,372
Alaska	22		20	22	19	19	17	20	21	20	15	14	20	19
Arizona	192		222	203	177	199	192	200	223	228	217	205	216	205
Arkansas	7		8	5	4	3	4	4	2	2	2	1	1	1
California	1		0	0	0	0	0	0	0	0	0	0	0	0
Colorado	10,325		9,192	8,663	5,972	9,189	9,181	9,390	10,784	11,117	12,082	13,216	12,554	13,608
Illinois	10,502		10,585	11,084	10,876	8,534	7,847	7,810	8,521	7,270	5,972	4,744	5,784	6,586
Indiana	2,795		2,495	1,866	2,192	2,742	2,878	2,650	2,231	3,373	3,496	3,821	3,527	3,702
Iowa	30		4	0	0	0	0	0	0	0	0	0	0	0
Kansas	57		23	23	19	29	27	33	16	14	16	12	6	14
Kentucky	10,956		11,259	9,748	8,987	10,451	10,005	9,561	9,056	9,363	8,464	8,028	7,916	8,498
Louisiana	245		267	286	248	273	247	227	284	286	293	310	293	320
Maryland	519		237	237	259	267	251	225	331	340	401	391	411	421
Mississippi	-		0	0	0	0	0	1	57	43	165	264	256	254
Missouri	211		67	44	57	32	30	31	35	29	20	43	46	48
Montana	490		542	514	492	534	558	535	449	510	487	481	519	524
New Mexico	451		679	466	408	459	489	497	464	630	1,280	1,864	2,047	3,001
North Dakota	380		420	392	389	385	389	405	407	397	401	401	390	390
Ohio	5,065		4,583	4,029	4,068	4,349	4,350	3,914	3,515	3,619	2,831	2,649	3,180	3,385
Oklahoma	285		359	323	286	385	395	469	453	620	660	620	847	877
Pennsylvania	22,735		24,024	26,995	26,440	30,026	29,491	23,626	22,253	22,253	19,667	24,649	19,980	18,297
Tennessee	296		101	112	143	148	116	119	99	142	142	124	136	140
Texas	4,291		4,028	4,054	4,245	4,104	4,047	4,084	3,732	3,466	3,482	3,657	3,530	3,535
Utah	3,587		2,616	2,410	2,810	3,566	3,859	3,633	2,811	2,081	2,709	3,408	5,240	4,787
Virginia	46,137		26,742	19,820	19,771	16,851	13,978	13,321	11,981	11,506	11,227	11,906	11,299	8,981
Washington	65		64	63	59	59	60	53	56	60	76	81	74	69
West Virginia	49,039		30,588	36,657	36,384	33,572	36,962	35,416	31,311	33,745	31,981	30,070	31,166	30,790
Wyoming	2,385		3,065	3,419	3,604	3,652	4,080	4,376	4,408	4,801	4,859	4,899	5,162	5,261
Total	**203,340**		**161,807**	**165,061**	**157,466**	**155,830**	**155,884**	**145,764**	**138,807**	**137,900**	**129,132**	**129,306**	**135,424**	**130,083**

+ Does not exceed 0.5 Million Cubic Feet

Note: The emission estimates provided above are inclusive of emissions from underground mines, surface mines and post-mining activities. The following states have neither underground nor surface mining and thus report no emissions as a result of coal mining: Connecticut, Delaware, Florida, Georgia, Hawaii, Idaho, Maine, Massachusetts, Michigan, Minnesota, Nebraska, Nevada, New Hampshire, New Jersey, New York, North Carolina, Oregon, Rhode Island, South Carolina, South Dakota, Vermont, and Wisconsin.

3.4. Methodology for Estimating CH_4 Emissions from Natural Gas Systems

The following steps were used to estimate CH_4 and non-energy CO_2 emissions from natural gas systems.

Step 1: Calculate Emission Estimates for Base Year 1992 Using GRI/EPA Study

The first step in estimating CH_4 and non-energy related (i.e., fugitive, vented and flared) CO_2 emissions from natural gas systems was to develop a detailed base year estimate of emissions. The study by EPA/GRI (1996 a-d) divides the industry into four stages to construct a detailed emission inventory for the year 1992. These stages include: field production, processing, transmission and storage (i.e., both underground and liquefied gas storage), and distribution. This study produced emission factors and activity data for over 80 different emission sources within the natural gas system. Emissions for 1992 were estimated by multiplying activity levels by emission factors for each system component and then summing by stage. Since publication, the EPA has updated activity data for some of the components in the system. Table A- 119 displays the 1992 GRI/EPA activity levels and CH_4 emission factors for the natural gas distribution stage, and the current EPA activity levels and emission factors. These data are shown to illustrate the kind of data used to calculate CH_4 and non-energy CO_2 emissions from all stages. For most sources, the CH_4 emission factors were adjusted for CO_2 content when estimating fugitive and vented non-energy CO_2 emissions. In the case of non-energy CO_2 emissions from flared sources, acid gas removal units and condensate tanks, specific industry data related to those sources was used to derive their respective emission factors.

Step 2: Collect Aggregate Statistics on Main Driver Variables

As detailed data on each of the over 80 sources were not available for the period 1990 through 2005, activity levels were estimated using aggregate statistics on key drivers, including: number of producing wells (EIA 2006a-b, New Mexico 2006a-b, Texas 2005a-b), number of gas plants (AGA 1991-1998; OGJ 1997-2006), number of shallow and deep offshore platforms (MMS 2006a-e), miles of transmission pipeline (OPS 2006a), miles of distribution pipeline (OPS 2006b), miles of distribution services (OPS 2006b), energy consumption (EIA 2006d). Data on the distribution of gas mains and services by material type was not available for 1990 through 1992 from OPS. For those years, the distribution by type was back calculated from 1993 using compound growth rates determined for the years 1993 through 2000. Table A- 120 provides the activity levels of some of the key drivers in the natural gas analysis.

Step 3: Estimate Emissions for Each Year and Stage

Emissions from each stage of the natural gas industry were estimated by multiplying the activity factors by the appropriate emission factors, summing all sources for each stage and then accounting for CH_4 reductions reported to the Natural Gas STAR Program.

Industry partners report CH_4 emission reductions by project to the Natural Gas STAR Program. The reductions are estimates using actual measurement data or equipment-specific emission factors. Before incorporating the reductions into the Inventory, quality assurance and quality control checks are undertaken to identify errors, inconsistencies, or irregular data. The checks include matching Natural Gas STAR reported reductions to specific inventory sources to make sure that a reported reduction for one source is not greater than the emission estimate for that source. Total emissions were estimated by adding the emission estimates from each stage. The base year of the inventory is 1992; therefore any reductions reported for 1992 or earlier are considered to be already included in the base-year emission factors and are not subtracted from the inventory estimate. If the reported reduction occurred between 1990 and 1992, then the reduction is added back into the estimate for the appropriate year(s). The reductions are also adjusted to remove the sunsetting time period, which removes reductions from the accounting system after a set time period. In future inventories, the sunsetting may be replaced with a technological lifetime. Methane emissions reductions from the Natural Gas STAR Program do not apply to the non-energy CO_2 emission estimates.

Methane emission reductions from the Natural Gas STAR Program beyond the efforts reflected in the 1992 base year are summarized in Table A- 121. Table A- 122 illustrates emission estimates from the natural gas distribution stage. Table A- 123 presents total natural gas production and associated CH_4 emissions.

The same procedure for estimating CH_4 emissions holds true for estimating non-energy related CO_2 emissions. The only difference is that GRI/EPA emission factors are adjusted for CO_2 content in each sector. Table A- 124 shows the CO_2 content for the different well types in the production sector of the natural gas system.

Table A- 119: 1992 Data and CH_4 Emissions (Gg) for the Natural Gas Distribution Stage

	GRI/EPA Values			*EPA Adjusted Values*		
Activity	**Activity Data**	**Emission Factor**	**Emissions**	**Activity Data**	**Emission Factor**	**Emissions**
Pipeline Leaks						
Mains—Cast Iron	55,288 miles	238.70 Mscf/mile-yr	254,178.95	55,288 miles	238.70 Mscf/mile-yr	254,178.95
Mains—Unprotected steel	174,657 equiv leaks	51.80 Mscf/leak-yr	174,249.70	82,109 miles	110.53 Mscf/mile-yr	174,794.23
Mains—Protected steel	68,308 equiv leaks	20.30 Mscf/leak-yr	26,706.93	444,768 miles	3.13 Mscf/mile-yr	26,790.38
Mains—Plastic	49,226 equiv leaks	99.80 Mscf/leak-yr	94,619.66	254,595 miles	9.91 Mscf/mile-yr	48,593.70
Services—Unprotected steel	458,476 equiv leaks	20.20 Mscf/leak-yr	178,371.00	5,446,393 services	1.71 Mscf/service	178,928.41
Services Protected steel	390,628 equiv leaks	9.20 Mscf/leak-yr	69,216.16	20,352,983 services	0.18 Mscf/service	69,432.46
Services—Plastic	68,903 equiv leaks	2.39 Mscf/leak-yr	3,171.70	17,681,238 services	0.01 Mscf/service	3,181.61
Services—Copper	7,720 equiv leaks	7.68 Mscf/leak-yr	1,141.92	233,246 services	0.25 Mscf/service	1,145.49
Meter/Regulator (City Gates)						
M&R >300	3,460 stations	179.80 scfh/station	104,960.57	3,580 stations	179.80 scfh/station	108,615.98
M&R 100-300	13,335 stations	95.60 scfh/station	215,085.58	13,799 stations	95.60 scfh/station	222,576.26
M&R <100	7,127 stations	4.31 scfh/station	5,182.56	7,375 stations	4.31 scfh/station	5,363.05
Reg >300	3,995 stations	161.90 scfh/station	109,124.94	4,134 stations	161.90 scfh/station	112,925.38
R-Vault >300	2,346 stations	1.30 scfh/station	514.55	2,428 stations	1.30 scfh/station	532.48
Reg 100-300	12,273 stations	40.50 scfh/station	83,862.18	12,700 stations	40.50 scfh/station	86,782.81
R-Vault 100-300	5,514 stations	0.18 scfh/station	167.46	5,706 stations	0.18 scfh/station	173.29
Reg 40-100	36,328 stations	1.04 scfh/station	6,374.34	37,593 stations	1.04 scfh/station	6,596.34
R-Vault 40-100	32,215 stations	0.09 scfh/station	470.15	33,337 stations	0.09 scfh/station	486.52
Reg <40	15,377 stations	0.13 scfh/station	345.05	15,913 stations	0.13 scfh/station	357.07
Customer Meters						
Residential	40,049,306 outdoor meters	138.50 scfy/meter	106,831.92	40,049,306 outdoor meters	143.70 scfy/meter	110,842.94
Commercial/Industry	4,608,000 meters	47.90 scfy/meter	4,251.13	4,607,983 meters	47.90 scfy/meter	4,251.11
Routine Maintenance						
Pressure Relief Valve Releases	836,760 mile main	0.05 Mscf/mile	805.80	836,760 mile main	0.05 Mscf/mile	805.80
Pipeline Blowdown	1,297,569 miles	0.10 Mscfy/mile	2,549.10	1,297,569 miles	0.10 Mscfy/mile	2,549.10
Upsets						
Mishaps (Dig-ins)	1,297,569 miles	1.59 mscfy/mile	39,735.97	1,297,569 miles	1.59 mscfy/mile	39,735.97

Table A- 120: Key Activity Data Drivers

Variable	Units	1990	1999	2000	2001	2002	2003	2004	2005
Transmission Pipelines Length	**miles**	**291,990**	**296,114**	**298,957**	**290,460**	**303,528**	**297,928**	**302,917**	**290,680**
Wells									
NE—Associated Gas Wells*	# wells	68,261	59,330	58,671	54,727	52,928	47,803	47,412	46,471
NE—Non-associated Gas Wells*	# wells	124,241	127,279	143,922	149,436	154,590	156,320	155,257	153,103
MC—Associated Gas Wells*	# wells	64,379	68,476	67,880	67,278	65,786	65,864	65,902	64,861
MC—Non-associated Gas Wells*	# wells	53,940	53,243	51,217	63,595	67,861	70,377	72,809	76,411
RM—Associated Gas Wells*	# wells	13,749	12,941	12,328	12,148	12,446	12,495	12,723	13,495
RM—Non-associated Gas Wells*	# wells	24,339	43,375	64,539	70,450	72,438	71,239	70,770	72,025
SW—Associated Gas Wells*	# wells	69,339	58,704	54,830	57,188	60,623	60,315	59,381	59,409
SW—Non-associated Gas Wells*	# wells	24,217	29,407	32,346	33,936	35,025	36,648	37,219	34,606
WC—Associated Gas Wells*	# wells	20,672	19,332	20,494	20,808	22,503	22,263	22,584	21,562
WC—Non-associated Gas Wells*	# wells	1,292	1,270	1,338	1,434	1,415	1,459	1,388	1,410
GC—Associated Gas Wells*	# wells	36,279	34,109	32,497	32,549	29,880	28,078	27,675	28,104
GC—Non-associated Gas Wells*	# wells	41,753	47,846	48,316	51,182	53,198	54,245	54,544	57,600
Platforms									
Gulf of Mexico and Pacific OCS Off-shore Platforms	# platforms	3,939	4,022	4,027	4,075	4,057	4,013	3,940	3,913
GoM and Pacific OCS Deep Water Platforms	# platforms	17	36	38	40	44	46	50	59
Gas Plants	**# gas plants**	**761**	**581**	**585**	**570**	**590**	**574**	**572**	**566**
Distribution Services	**# of services**	**47,883,083**	**54,317,439**	**56,761,042**	**57,461,795**	**58,876,416**	**58,537,395**	**61,089,889**	**58,556,335**
Steel—Unprotected	# of services	7,633,526	5,751,250	5,675,520	5,449,653	5,186,134	4,840,347	4,791,652	5,308,375
Steel—Protected	# of services	19,781,581	18,310,719	17,855,560	17,911,402	17,778,463	17,258,710	18,147,587	15,883,423
Plastic	# of services	18,879,865	28,796,952	31,795,871	32,706,753	34,547,274	35,071,961	36,811,107	36,152,277
Copper	# of services	1,588,111	1,458,518	1,434,091	1,393,987	1,364,545	1,366,377	1,339,543	1,212,260
Distribution Mains	**miles**	**944,157**	**1,004,907**	**1,048,485**	**1,099,137**	**1,133,625**	**1,104,683**	**1,151,995**	**1,093,909**
Cast Iron	miles	58,292	45,865	44,750	44,283	42,025	41,091	40,600	37,371
Steel—Unprotected	miles	108,941	84,534	82,800	81,291	78,119	74,042	75,859	69,291
Steel—Protected	miles	465,538	459,298	471,510	475,329	480,982	483,782	495,861	461,459
Plastic	miles	311,386	415,210	449,425	498,234	532,499	505,768	539,675	525,788

* NEMS (National Energy Modeling System) projects the production, imports, conversion, consumption, and prices of energy, subject to assumptions on macroeconomic and financial factors, world energy markets, resource availability and costs, behavioral and technological choice criteria, cost and performance characteristics of energy technologies, and demographics.

Table A- 121: CH_4 reductions derived from the Natural Gas STAR program (Gg)

Process	1992		1999	2000	2001	2002	2003	2004	2005
Production	0		255	324	386	416	523	670	799
Processing	0		11	17	22	28	58	49	109
Transmission and Storage	0		233	271	349	365	360	456	561
Distribution	0		33	28	33	163	112	96	40

Note: These reductions will not match the Natural Gas STAR program reductions. These numbers are adjusted for reductions prior to the 1992 base year, and do not include a sunsetting period.

Table A- 122: CH_4 Emission Estimates from the Natural Gas Distribution Stage (Gg)

Activity	1990		1999	2000	2001	2002	2003	2004	2005
Pipeline Leaks									
Mains—Cast Iron	280.00		220.31	214.95	212.71	201.86	197.37	195.02	179.50
Mains—Unprotected steel	191.15		148.33	145.28	142.64	137.07	129.92	133.10	121.58
Mains—Protected steel	26.59		26.23	26.93	27.15	27.47	27.63	28.32	26.35
Mains—Plastic	56.61		75.49	81.71	90.58	96.81	91.95	98.12	95.59
Services—Unprotected steel	191.33		144.16	142.26	136.60	129.99	121.32	120.10	133.05
Services Protected steel	69.57		64.40	62.80	62.99	62.53	60.70	63.82	55.86
Services—Plastic	2.90		4.43	4.89	5.03	5.31	5.39	5.66	5.56
Services—Copper	1.16		1.07	1.05	1.02	1.00	1.00	0.98	0.89
Meter/Regulator (City Gates)									
M&R >300	101.70		109.45	115.70	110.49	113.22	117.62	113.13	112.04
M&R 100-300	208.40		224.28	237.10	226.42	232.02	241.03	231.83	229.60
M&R <100	5.02		5.40	5.71	5.46	5.59	5.81	5.59	5.53
Reg >300	105.73		113.79	120.29	114.87	117.72	122.29	117.62	116.49
R-Vault >300	0.50		0.54	0.57	0.54	0.56	0.58	0.55	0.55
Reg 100-300	81.26		87.45	92.44	88.28	90.46	93.98	90.39	89.52
R-Vault 100-300	0.16		0.17	0.18	0.18	0.18	0.19	0.18	0.18
Reg 40-100	6.18		6.65	7.03	6.71	6.88	7.14	6.87	6.80
R-Vault 40-100	0.46		0.49	0.52	0.49	0.51	0.53	0.51	0.50
Reg <40	0.33		0.36	0.38	0.36	0.37	0.39	0.37	0.37
Customer Meters									
Residential	103.78		111.69	118.07	112.76	115.54	120.03	115.45	114.34
Commercial/Industry	3.97		4.58	4.66	4.27	4.38	4.25	4.28	3.95
Routine Maintenance									
Pressure Relief Valve Releases	0.86		0.91	0.95	1.00	1.03	1.01	1.05	1.00
Pipeline Blowdown	2.39		2.57	2.72	2.59	2.66	2.76	2.66	2.63
Upsets									
Mishaps (Dig-ins)	37.20		40.04	42.33	40.42	41.42	43.03	41.39	40.99

Table A- 123: U.S. Total Natural Gas Production (Trillion Ft^3/yr) and Associated CH_4 Emissions (Gg)

Activity	1990		1999	2000	2001	2002	2003	2004	2005
Production	17.8		18.8	19.2	19.6	18.9	19.1	18.8	18.2
CH_4 Emissions from Production	1,497		1,936	2,182	2,377	2,467	2,501	2,510	2,502

Table A- 124: U.S. Production Sector CO_2 Content in Natural Gas by NEMS Region and Natural Gas Well type

	U.S. Region						
Well Types	North East	Mid-Central	Gulf Coast	South West	Rocky Mountain	West Coast	Lower-48 States
Conventional	0.92%	0.79%	2.17%	3.81%	7.95%	0.16%	3.41%
Un-conventional	7.42%	0.31%	0.23%	NA	0.64%	NA	4.83%
All types	3.04%	0.79%	2.17%	3.81%	7.58%	0.16%	3.45%

3.5. Methodology for Estimating CH_4 Emissions from Petroleum Systems

The methodology for estimating CH_4 emissions from petroleum systems is based on the 1999 EPA draft report, *Estimates of Methane Emissions from the U.S. Oil Industry* (EPA 1999) and the study, *Methane Emissions from the U.S. Petroleum Industry* (EPA/GRI 1996). Sixty-four activities that emit CH_4 from petroleum systems were examined from these reports. Most of the activities analyzed involve crude oil production field operations, which accounted for about 97 percent of total oil industry emissions. Crude transportation and refining accounted for the remaining emissions of less than one and just over two percent, respectively. The following steps were taken to estimate CH_4 emissions from petroleum systems.

Step 1: Determine Emission Factors for all Activities

The emission factors for the majority of the activities for 1995 are taken from the 1999 EPA draft report, which contained the most recent and comprehensive determination of CH_4 emission factors for the 64 CH_4-emitting activities in the oil industry at that time. Emission factors for pneumatic devices in the production sector were recalculated in 2002 using emissions data in the EPA/GRI 1996 study. The gas engine emission factor is taken from (EPA/GRI 1996b). The oil tank venting emission factor is taken from the API E&P Tank Calc average for API gravity less than 44 deg. Offshore emissions from shallow water and deep water oil platforms are taken from analysis of the Gulf-wide Offshore Activity Data System (GOADS) report (EPA 2006, MMS 2005c). The emission factors determined for 1995 were assumed to be representative of emissions from each source type over the period 1990 through 2005. Therefore, the same emission factors are used for each year throughout this period.

Step 2: Determine Activity Levels for Each Year

Activity levels change from year to year. Some factors change in proportion to crude oil rates: production, transportation, refinery runs. Some change in proportion to the number of facilities: oil wells, petroleum refineries. Some factors change proportional to both rate and number of facilities.

For most sources, activity levels found in the EPA/GRI 1996 for the 1995 base year are extrapolated to other years using publicly-available data sources. For the remaining sources, the activity levels are obtained directly from publicly available data and are not extrapolated from the 1995 base year.

For both sets of data, a determination was made on a case-by-case basis as to which measure of petroleum industry activity best reflects the change in annual activity. Publicly-reported data from the Minerals Management Service (MMS), Energy Information Administration (EIA), American Petroleum Institute (API), and the Oil & Gas Journal (O&GJ) were used to extrapolate the activity levels from the base year to each year between 1990 and 2005. Data used include total domestic crude oil production, number of domestic crude oil wells, total imports and exports of crude oil, and total petroleum refinery crude runs. The activity data for the transportation sector are not available. In this case, all the crude oil that was transported was assumed to go to refineries. Therefore, the activity data for the refining sector was used also for the transportation sector. For a small number of sources, 2005 data were not yet available; in these cases, the 2004 activity factors were used. In the few cases where no data was located, such as average stripper well production, oil industry data based on expert judgment was used.

Step 3: Estimate Methane Emissions for Each Activity for Each Year

Annual emissions from each of the 64 petroleum system activities were estimated by multiplying the activity data for each year by the corresponding emission factor. These annual emissions for each activity were then summed to estimate the total annual CH_4 emissions. Table A- 125, Table A- 126, and Table A- 127 provide 2005 activity factors, emission factors, and emission estimates. Table A- 128 provides a summary of emission estimates for the years 1990 through 2005.

Table A- 125: 2005 CH_4 Emissions from Petroleum Production Field Operations

Activity/Equipment	Emission Factor	Units	Activity Factor	Units	Emissions (Bcf/yr)
Vented Emissions					**62.702**
Oil Tanks	5.28	scf of CH_4/bbl crude	1,310	MMbbl/yr (non stripper wells)	6.918

Activity/Equipment	Emission Factor	Units	Activity Factor	Units	Emissions (Bcf/yr)
Pneumatic Devices, High Bleed	330	scfd CH_4/device	135,341	No. of high-bleed devices	16.322
Pneumatic Devices, Low Bleed	52	scfd CH_4/device	251,348	No. of low-bleed devices	4.771
Chemical Injection Pumps	248	scfd CH_4/pump	27,309	No. of pumps	2.473
Vessel Blowdowns	78	scfy CH_4/vessel	176,618	No. of vessels	0.014
Compressor Blowdowns	3,775	scf/yr of CH_4/compressor	2,393	No. of compressors	0.009
Compressor Starts	8,443	scf/yr. of CH_4/compressor	2,393	No. of compressors	0.020
Stripper wells	2,345	scf/yr of CH_4/stripper well	317,890	No. of stripper wells vented	0.745
Well Completion Venting	733	scf/completion	9409	Oil well completions	0.007
Well Workovers	96	scf CH_4/workover	38,250	Oil well workovers	0.004
Pipeline Pigging	2.40	scfd of CH_4/pig station	0	No. of crude pig stations	0.000
Offshore Platforms, Shallow water Oil, fugitive, vented and combusted	54,795	scfd CH_4/platform	1,465	No. of oil platforms	29.290
Offshore Platforms, Deepwater oil, fugitive, vented and combusted	260,274	scfd CH_4/platform	22	No. of oil platforms	2.130
Fugitive Emissions					**2.401**
Oil Wellheads (heavy crude)	0.13	scfd/well	13,544	No. of hvy. crude wells	0.001
Oil Wellheads (light crude)	16.6	scfd/well	178,567	No. of lt. crude wells	1.084
Separators (heavy crude)	0.15	scfd CH_4/separator	10,399	No. of hvy. crude seps.	0.001
Separators (light crude)	14	scfd CH_4/separator	94,637	No. of lt. crude seps.	0.479
Heater/Treaters (light crude)	19	scfd CH_4/heater	71,582	No. of heater treaters	0.501
Headers (heavy crude)	0.08	scfd CH_4/header	13,303	No. of hvy. crude hdrs.	0.000
Headers (light crude)	11	scfd CH_4/header	41,242	No. of lt. crude hdrs.	0.163
Floating Roof Tanks	338,306	scf CH_4/floating roof tank/yr.	24	No. of floating roof tanks	0.008
Compressors	100	scfd CH_4/compressor	2,393	No. of compressors	0.087
Large Compressors	16,360	scfd CH_4/compressor	0	No. of large comprs.	0.000
Sales Areas	41	scf CH_4/loading	1,577,388	Loadings/year	0.063
Pipelines	0	scfd of CH_4/mile of pipeline	15,193	Miles of gathering line	0.000
Well Drilling	0	scfd of CH_4/oil well drilled	10,596	No. of oil wells drilled	0.000
Battery Pumps	0.24	scfd of CH_4/pump	153,000	No. of battery pumps	0.013
Combustion Emissions					**3.648**
Gas Engines	0.24	scf CH_4/HP-hr	15,074	MMHP-hr	3.618
Heaters	0.52	scf CH_4/bbl	1,869.2	MBbl/yr	0.001
Well Drilling	2,453	scf CH_4/well drilled	10,596	Oil wells drilled	0.026
Flares	20	scf CH_4/Mcf flared	152,186	Mcf flared/yr	0.003
Process Upset Emissions					**0.094**
Pressure Relief Valves	35	scf/yr/PR valve	156,131	No. of PR valves	0.005
Well Blowouts Onshore	2.5	MMscf/blowout	35.3	No. of blowouts/yr	0.088
Total					**68.84**

Table A-126: 2005 CH_4 Emissions from Petroleum Transportation

Activity/Equipment	Emission Factor	Units	Activity Factor	Units	Emissions (Bcf/yr)
Vented Emissions					**0.220**
Tanks	0.021	scf CH_4/yr/bbl of crude delivered to refineries	5,550	MMbbl crude feed/yr	0.114
Truck Loading	0.520	scf CH_4/yr/bbl of crude transported by truck	54.6	MMbbl trans. by truck	0.028
Marine Loading	2.544	scf CH_4/1000 gal. crude marine loadings	24,210,946	1,000 gal./yr loaded	0.062
Rail Loading	0.520	scf CH_4/yr/bbl of crude transported by rail	4.9	MMbbl. Crude by rail/yr	0.003
Pump Station Maintenance	36.80	scf CH_4/station/yr	507	No. of pump stations	0.000
Pipeline Pigging	39	scfd of CH_4/pig station	1,015	No. of pig stations	0.013
Fugitive Emissions					**0.050**
Pump Stations	25	scfCH_4/mile/yr.	50,749	No. of miles of crude p/l	0.001
Pipelines	0	scf CH_4/bbl crude transported by pipeline	6,612	MM bbl crude piped	0.000
Floating Roof Tanks	58,965	scf CH_4/floating roof tank/yr.	824	No. of floating roof tanks	0.049
Combustion Emissions					**0.000**

Activity/Equipment	Emission Factor	Units	Activity Factor	Units	Emissions (Bcf/yr)
Pump Engine Drivers	0.24	scf CH_4/hp-hr	NE	No. of hp-hrs	NE
Heaters	0.521	scf CH_4/bbl.burned	NE	No. of bbl. Burned	NE
Total					**0.270**

NE: Not estimated for lack of activity factor data

Table A-127: 2005 CH_4 Emissions from Petroleum Refining

Activity/Equipment	Emission Factor	Units	Activity Factor	Units	Emissions (Bcf/yr)
Vented Emissions					**1.252**
Tanks	20.6	scfCH_4/Mbbl	1,951	Mbbl/cd heavy crude feed	0.015
System Blowdowns	137	scfCH_4/Mbbl	15,204	Mbbl/cd refinery feed	0.760
Asphalt Blowing	2,555	scfCH_4/Mbbl	510	Mbbl/cd production	0.477
Fugitive Emissions					**0.086**
Fuel Gas System	439	McfCH_4/refinery/yr	142	Refineries	0.062
Floating Roof Tanks	587	scf CH_4/floating roof tank/yr.	767	No. of floating roof tanks	0.000
Wastewater Treating	1.88	scfCH_4/Mbbl	15,204	Mbbl/cd refinery feed	0.010
Cooling Towers	2.36	scfCH_4/Mbbl	15,204	Mbbl/cd refinery feed	0.013
Combustion Emissions					**0.094**
Atmospheric Distillation	3.61	scfCH_4/Mbbl	15,479	Mbbl/cd refinery feed	0.021
Vacuum Distillation	3.61	scfCH_4/Mbbl	6,966	Mbbl/cd feed	0.009
Thermal Operations	6.02	scfCH_4/Mbbl	2,181	Mbbl/cd feed	0.005
Catalytic Cracking	5.17	scfCH_4/Mbbl	5,239	Mbbl/cd feed	0.010
Catalytic Reforming	7.22	scfCH_4/Mbbl	3,222	Mbbl/cd feed	0.008
Catalytic Hydrocracking	7.22	scfCH_4/Mbbl	1,340	Mbbl/cd feed	0.003
Hydrorefining	2.17	scfCH_4/Mbbl	2,167	Mbbl/cd feed	0.002
Hydrotreating	6.50	scfCH_4/Mbbl	9,941	Mbbl/cd feed	0.023
Alkylation/Polymerization	12.6	scfCH_4/Mbbl	1,111	Mbbl/cd feed	0.005
Aromatics/Isomeration	1.80	scfCH_4/Mbbl	992	Mbbl/cd feed	0.001
Lube Oil Processing	0.00	scfCH_4/Mbbl	185	Mbbl/cd feed	0.000
Engines	0.006	scfCH_4/hp-hr	1,187	MMhp-hr/yr	0.007
Flares	0.189	scfCH_4/Mbbl	15,204	Mbbl/cd refinery feed	0.001
Total					**1.432**

Table A-128: Summary of CH_4 Emissions from Petroleum Systems (Gg)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Production Field Operations	**1,609**	**1,607**	**1,552**	**1,501**	**1,477**	**1,450**	**1,433**	**1,410**	**1,382**	**1,326**	**1,292**	**1,271**	**1,242**	**1,196**	**1,176**	**1,324**
Pneumatic device venting	489	498	482	472	469	463	462	461	452	437	428	425	424	412	408	406
Tank venting	179	179	173	167	161	161	161	165	163	153	154	154	151	150	142	133
Combustion & process upsets	88	90	86	84	83	82	82	82	80	76	76	75	75	73	72	72
Misc. venting & fugitives	827	813	786	754	740	719	703	676	662	635	612	594	570	540	533	692
Wellhead fugitives	26	26	24	24	24	25	25	26	26	24	22	22	23	22	21	21
Crude Oil Transportation	**7**	**6**	**6**	**6**	**6**	**6**	**6**	**6**	**6**	**6**	**5**	**5**	**5**	**5**	**5**	**5**
Refining	**25**	**24**	**24**	**25**	**25**	**25**	**26**	**27**	**27**	**27**	**28**	**27**	**27**	**27**	**28**	**28**
Total	**1,640**	**1,638**	**1,582**	**1,532**	**1,508**	**1,482**	**1,465**	**1,442**	**1,415**	**1,358**	**1,325**	**1,303**	**1,275**	**1,229**	**1,209**	**1,357**

3.6. Methodology for Estimating CO_2 and N_2O Emissions from Municipal Solid Waste Combustion

Emissions of CO_2 from municipal solid waste (MSW) combustion include CO_2 generated by the combustion of plastics, synthetic rubber and synthetic fibers in MSW, and combustion of synthetic rubber and C black in tires. Combustion of MSW also results in emissions of N_2O. The methodology for calculating emissions from each of these waste combustion sources is described in this Annex.

CO_2 from Plastics Combustion

In the *Characterization of Municipal Solid Waste in the United States* reports (EPA 1997, 1998, 1999, 2000b, 2002, 2003, 2005a, 2006b), the flows of plastics in the U.S. waste stream are reported for seven resin categories. For 2005, the most recent year for which these data are reported, the quantity generated, recovered, and discarded for each resin is shown in Table A- 129. The data set for 1990 through 2005 is incomplete, and several assumptions were employed to bridge the data gaps. The EPA reports do not provide estimates for individual materials landfilled and combusted, although they do provide such an estimate for the waste stream as a whole. To estimate the quantity of plastics landfilled and combusted, total discards were apportioned based on the proportions of landfilling and combustion for the entire U.S. waste stream for each year in the time series. For those years when distribution by resin category was not reported (1990-1994), total values were apportioned according to 1995 (the closest year) distribution ratios. Generation and recovery figures for 2002 and 2004 were linearly interpolated between surrounding years' data.

Table A- 129: 2005 Plastics in the Municipal Solid Waste Stream by Resin (Gg)

Waste Pathway	PET	HDPE	PVC	LDPE/ LLDPE	PP	PS	Other	Total
Generation	2,860	5,890	1,640	6,450	4,000	2,590	5,480	**28,910**
Recovery	540	520	0	190	10	0	390	**1,650**
Discard	2,320	5,370	1,640	6,260	3,990	2,590	5,090	**27,260**
Landfill	1,855	4,294	1,311	5,006	3,191	2,071	4,070	**21,799**
Combustion	465	1,076	329	1,254	799	519	1,020	**5,461**
Recovery*	19%	9%	0%	3%	0%	0%	7%	**6%**
Discard*	81%	91%	100%	97%	100%	100%	93%	**94%**
Landfill*	65%	73%	80%	78%	80%	80%	74%	**75%**
Combustion*	16%	18%	20%	19%	20%	20%	19%	**19%**

*As a percent of waste generation.
Note: Totals may not sum due to independent rounding. Abbreviations: PET (polyethylene terephthalate), HDPE (high density polyethylene), PVC (polyvinyl chloride), LDPE/LLDPE (linear low density polyethylene), PP (polypropylene), PS (polystyrene).

Fossil fuel-based CO_2 emissions were calculated as the product of plastic combusted, C content, and fraction oxidized (see Table A- 130). The C content of each of the six types of plastics is listed, with the value for "other plastics" assumed equal to the weighted average of the six categories. The fraction oxidized was assumed to be 98 percent.

Table A- 130: 2005 Plastics Combusted (Gg), Carbon Content (%), Fraction Oxidized (%) and Carbon Combusted (Gg)

Factor	PET	HDPE	PVC	LDPE/ LLDPE	PP	PS	Other	Total
Quantity Combusted	465	1,076	328	1,254	799	519	1,019	**5,460**
Carbon Content of Resin	63%	86%	38%	86%	86%	92%	66% [a]	**-**
Fraction Oxidized	98%	98%	98%	98%	98%	98%	98%	**-**
Carbon in Resin Combusted	285	904	124	1054	671	469	658	**4,164**
Emissions (Tg CO_2 Eq.)	**0.9**	**3.0**	**0.4**	**3.5**	**2.2**	**1.6**	**2.2**	**13.9**

[a] Weighted average of other plastics produced.
Note: Totals may not sum due to independent rounding.

CO_2 from Combustion of Synthetic Rubber and Carbon Black in Tires

Emissions from tire combustion require two pieces of information: the amount of tires combusted and the C content of the tires. *U.S. Scrap Tire Markets in the United States 2005* (RMA 2006) reports that 155.1 million of the 299.2 million scrap tires generated in 2005 (approximately 52 percent of generation) were used for fuel purposes.

Using RMA's Scrap Tire Management Council (STMC) estimates of average tire composition and weight, the mass of synthetic rubber and C black in scrap tires was determined:

- Synthetic rubber in tires was estimated to be 90 percent C by weight, based on the weighted average C contents of the major elastomers used in new tire consumption.[33] Table A- 131 shows consumption and C content of elastomers used for tires and other products in 2002, the most recent year for which data are available.

- C black is 100 percent C (Miller 1999).

Multiplying the mass of scrap tires combusted by the total C content of the synthetic rubber and C black portions of scrap tires and by a 98 percent oxidation factor yielded CO_2 emissions, as shown in Table A- 132. The disposal rate of rubber in tires (0.3 Tg C/yr) is smaller than the consumption rate for tires based on summing the elastomers listed in Table A- 131 (1.3 Tg/yr); this is due to the fact that much of the rubber is lost through tire wear during the product's lifetime and may also reflect the lag time between consumption and disposal of tires. Tire production and fuel use for 1990 through 2001 were taken from RMA 2004; when data were not reported, they were linearly interpolated between bracketing years' data or, for the ends of time series, set equal to the closest year with reported data.

Table A- 131: Elastomers Consumed in 2002 (Gg)

Elastomer	Consumed	Carbon Content	Carbon Equivalent
Styrene butadiene rubber solid	**768**	**91%**	**700**
For Tires	660	91%	602
For Other Products*	108	91%	98
Polybutadiene	**583**	**89%**	**518**
For Tires	408	89%	363
For Other Products	175	89%	155
Ethylene Propylene	**301**	**86%**	**258**
For Tires	6	86%	5
For Other Products	295	86%	253
Polychloroprene	**54**	**59%**	**32**
For Tires	0	59%	0
For Other Products	54	59%	32
Nitrile butadiene rubber solid	**84**	**77%**	**65**
For Tires	1	77%	1
For Other Products	83	77%	64
Polyisoprene	**58**	**88%**	**51**
For Tires	48	88%	42
For Other Products	10	88%	9
Others	**367**	**88%**	**323**
For Tires	184	88%	161
For Other Products	184	88%	161
Total	**2,215**	**-**	**1,950**
For Tires	**1,307**	**-**	**1,174**

*Used to calculate C content of non-tire rubber products in municipal solid waste.
- Not applicable
Note: Totals may not sum due to independent rounding.

Table A- 132: Scrap Tire Constituents and CO_2 Emissions from Scrap Tire Combustion in 2005

Material	Weight of Material (Tg)	Fraction Oxidized	Carbon Content	Emissions (Tg CO_2 Eq.)
Synthetic Rubber	0.4	98%	90%	1.2
Carbon Black	0.4	98%	100%	1.6
Total	**0.8**	**-**	**-**	**2.8**

- Not applicable

[33] The carbon content of tires (1,158 Gg C) divided by the mass of rubber in tires (1,285 Gg) equals 90 percent.

CO_2 from Combustion of Synthetic Rubber in Municipal Solid Waste

Similar to the methodology for scrap tires, CO_2 emissions from synthetic rubber in MSW were estimated by multiplying the amount of rubber combusted by an average rubber C content. The amount of rubber in the MSW stream was estimated from data provided in the *Characterization of Municipal Solid Waste in the United States* reports (EPA 1996, 1997, 1998, 1999, 2000b, 2002, 2003, 2005a, 2006b; Schneider 2007). The reports divide rubber found in MSW into three product categories: other durables (not including tires), non-durables (which includes clothing and footwear and other non-durables), and containers and packaging. Since there was negligible recovery for these product types, all the waste generated can be considered discarded. Similar to the plastics method, discards were apportioned into landfilling and combustion based on their relative proportions, for each year, for the entire U.S. waste stream. The report aggregates rubber and leather in the MSW stream; an assumed synthetic rubber content was assigned to each product type, as shown in Table A-133.[34] A C content of 85 percent was assigned to synthetic rubber for all product types (based on the weighted average C content of rubber consumed for non-tire uses), and a 98 percent fraction oxidized was assumed.

Table A-133: Rubber and Leather in Municipal Solid Waste in 2005

Product Type	Combustion (Gg)	Synthetic Rubber (%)	Carbon Content (%)	Fraction Oxidized (%)	Emissions (Tg CO_2 Eq.)
Durables (not Tires)	530.7	100%	85%	98%	1.7
Non-Durables	71.34	-			0.2
Clothing and Footwear	31.8	25%	85%	98%	0.1
Other Non-Durables	39.5	75%	85%	98%	0.1
Containers and Packaging	5.5	100%	85%	98%	+
Total	**607.5**	**-**	**-**	**-**	**1.9**

+ Less than 0.05 Tg CO_2 Eq.
- Not applicable

CO_2 from Combustion of Synthetic Fibers

CO_2 emissions from synthetic fibers were estimated as the product of the amount of synthetic fiber discarded annually and the average C content of synthetic fiber. Fiber in the MSW stream was estimated from data provided in the *Characterization of Municipal Solid Waste in the United States* (EPA 2000b, 2002, 2003, 2005a, 2006b) reports for textiles. Production data for the synthetic fibers was based on data from the American Chemical Society (FEB 2006). The amount of synthetic fiber in MSW was estimated by subtracting (a) the amount recovered from (b) the waste generated (see Table A-134). As with the other materials in the MSW stream, discards were apportioned based on the annually variable proportions of landfilling and combustion for the entire U.S. waste stream. It was assumed that approximately 55 percent of the fiber was synthetic in origin, based on information received from the Fiber Economics Bureau (DeZan 2000). An average C content of 70 percent was assigned to synthetic fiber using the production-weighted average of the C contents of the four major fiber types (polyester, nylon, olefin, and acrylic) produced in 2005 (see Table A-135). The equation relating CO_2 emissions to the amount of textiles combusted is shown below.

$$CO_2 \text{ Emissions from the Combustion of Synthetic Fibers} = \text{Annual Textile Combustion (Gg)} \times (\text{Percent of Total Fiber that is Synthetic}) \times (\text{Average C Content of Synthetic Fiber}) \times (44\text{g } CO_2/12 \text{ g C})$$

[34] As a sustainably harvested biogenic material, the combustion of leather is assumed to have no net CO_2 emissions.

Table A-134: Textiles in MSW (Gg)

Year	Generation	Recovery	Discards	Combustion
1990	2,884	328	2,557	474
1991	3,008	347	2,661	530
1992	3,286	387	2,899	556
1993	3,386	397	2,988	578
1994	3,604	432	3,172	619
1995	3,674	447	3,227	723
1996	3,832	472	3,361	789
1997	4,090	526	3,564	810
1998	4,269	556	3,713	781
1999	4,498	611	3,887	789
2000	4,686	640	4,046	823
2001	4,870	715	4,155	846
2002	5,093	740	4,354	863
2003	5,257	755	4,503	915
2004	5,371	849	4,522	910
2005	5,530	844	4,686	939

Table A-135: Synthetic Fiber Production in 2005

Fiber	Production (Tg)	Carbon Content
Polyester	1.4	63%
Nylon	1.1	64%
Olefin	1.4	86%
Acrylic	0.1	68%
Total	**3.9**	**70%**

N_2O from Municipal Solid Waste Combustion

Estimates of N_2O emissions from MSW combustion in the United States are based on the methodology outlined in the EPA's Compilation of Air Pollutant Emission Factors (EPA 1995) and presented in the *Characterization* reports (2000a, 2002, 2003, 2005a, 2006b; Schneider 2007). According to this methodology, emissions of N_2O from MSW combustion are the product of the mass of MSW combusted, an emission factor of N_2O emitted per unit mass of waste combusted, and an N_2O emissions control removal efficiency. The mass of waste combusted was derived from the information published in *BioCycle* (Simmons et al 2006). For MSW combustion in the United States, an emission factor of 44 g N_2O/metric ton MSW (the average of the values provided for hearth/grate combustors as listed in the IPCC *Good Practice Guidance*, 2000) and an estimated emissions control removal efficiency of zero percent were used. No information was available on the mass of waste combusted in 2005, so the value was assumed to remain constant at the 2004 level.

Despite the differences in methodology and data sources, the two series of references (EPA's and BioCycle's) provide estimates of total solid waste combusted that are relatively consistent (see Table A-136).

Table A-136: U.S. Municipal Solid Waste Combusted, as Reported by EPA and BioCycle (Metric Tons)

Year	EPA	BioCycle
1990	28,939,680	30,632,057
1991	30,209,760	25,462,836
1992	29,656,368	29,086,574
1993	29,865,024	27,838,884
1994	29,474,928	29,291,583
1995	32,241,888	29,639,040
1996	32,740,848	29,707,171
1997	33,294,240	27,798,368
1998	31,216,752	25,489,893
1999	30,881,088	24,296,249
2000	30,599,856	25,974,978
2001	30,481,920	29,732,255[a]
2002	30,255,120	33,723,677
2003	30,527,280	33,970,308
2004	30,962,736	34,181,035[b]
2005	30,300,480	NA

NA (Not Available)

[a] Interpolated between 2000 and 2002 values.

3.7. Methodology for Estimating Emissions from International Bunker Fuels used by the U.S. Military

Bunker fuel emission estimates for the Department of Defense (DoD) are developed using data generated by the Defense Energy Support Center for aviation and naval fuels. The Defense Energy Support Center (DESC) of the Defense Logistics Agency (DLA) prepared a special report based on data in the Fuels Automated System (FAS), a database that recently replaced the Defense Fuels Automated Management System (DFAMS). Data for intermediate fuel oil, however, currently remains in the original DFAMS database. DFAMS/FAS contains data for 1995 through 2005, but the data set was not complete for years prior to 1995. Fuel quantities for 1990 to 1994 were estimated based on a back-calculation of the 1995 DFAMS values using DLA aviation and marine fuel procurement data. The back-calculation was refined in 1999 to better account for the jet fuel conversion from JP4 to JP8 that occurred within the DoD between 1992 and 1995.

Step 1: Omit Extra-Territorial Fuel Deliveries

Beginning with the complete DFAMS data set for each year, the first step in the development of DoD-related emissions from international bunker fuels was to identify data that would be representative of international bunker fuel consumption as that term is defined by decisions of the UNFCCC (i.e., fuel sold to a vessel, aircraft, or installation within the United States or its territories and used in international maritime or aviation transport). Therefore, fuel data were categorized by the location of fuel delivery in order to identify and omit all extra-territorial fuel transactions/deliveries (i.e., sales abroad).

Step 2: Allocate JP-8 between Aviation and Land-based Vehicles

As a result of DoD[35] and NATO[36] policies on implementing the Single Fuel For the Battlefield concept, DoD activities have been increasingly replacing diesel fuel with JP8 (a type of jet fuel) in compression ignition and turbine engines in land-based equipment. Based on this concept and examination of all data describing jet fuel used in land-based vehicles, it was determined that a portion of JP8 consumption should be attributed to ground vehicle use. Based on available Service data and expert judgment, it was determined that a small fraction of the total JP8 use should be reallocated from the aviation subtotal to a new land-based jet fuel category for 1997 and subsequent years. As a result of this reallocation, the JP8 use reported for aviation will be reduced and the total fuel use for land-based equipment will increase. DoD's total fuel use will not change.

Table A-137 displays DoD's consumption of fuels that remain at the completion of Step 1, summarized by fuel type. Table A-137 reflects the adjustments for jet fuel used in land-based equipment, as described above.

Step 3: Omit Land-Based Fuels

Navy and Air Force land-based fuels (i.e., fuel not used by ships or aircraft) were also omitted for the purpose of calculating international bunker fuels. The remaining fuels, listed below, were considered potential DoD international bunker fuels.

- Marine: naval distillate fuel (F76), marine gas oil (MGO), and intermediate fuel oil (IFO).
- Aviation: jet fuels (JP8, JP5, JP4, JAA, JA1, and JAB).

---

[35] DoD Directive 4140.43, Fuel Standardization, 1998; DoD Directive 4140.25, DoD Management Policy for Energy Commodities and Related Services, 1999.

[36] NATO Standard Agreement NATO STANAG 4362, Fuels for Future Ground Equipments Using Compression Ignition or Turbine Engines, 1987.

Step 4: Omit Fuel Transactions Received by Military Services that are not Considered to be International Bunker Fuels

Next, the records were sorted by Military Service. The following assumptions were used regarding bunker fuel use by Service, leaving only the Navy and Air Force as users of military international bunker fuels.

- Only fuel delivered to a ship, aircraft, or installation in the United States was considered a potential international bunker fuel. Fuel consumed in international aviation or marine transport was included in the bunker fuel estimate of the country where the ship or aircraft was fueled. Fuel consumed entirely within a country's borders was not considered a bunker fuel.
- Based on discussions with the Army staff, only an extremely small percentage of Army aviation emissions, and none of its watercraft emissions, qualified as bunker fuel emissions. The magnitude of these emissions was judged to be insignificant when compared to Air Force and Navy emissions. Based on this, Army bunker fuel emissions were assumed to be zero.
- Marine Corps aircraft operating while embarked consumed fuel reported as delivered to the Navy. Bunker fuel emissions from embarked Marine Corps aircraft were reported in the Navy bunker fuel estimates. Bunker fuel emissions from other Marine Corps operations and training were assumed to be zero.
- Bunker fuel emissions from other DoD and non-DoD activities (i.e., other federal agencies) that purchased fuel from DESC were assumed to be zero.

Step 5: Determine Bunker Fuel Percentages

Next it was necessary to determine what percent of the marine and aviation fuels were used as international bunker fuels. Military aviation bunkers include international operations (i.e., sorties that originate in the United States and end in a foreign country), operations conducted from naval vessels at sea, and operations conducted from U.S. installations principally over international water in direct support of military operations at sea (e.g., anti-submarine warfare flights). For the Air Force, a bunker fuel weighted average was calculated based on flying hours by major command. International flights were weighted by an adjustment factor to reflect the fact that they typically last longer than domestic flights. In addition, a fuel use correction factor was used to account for the fact that transport aircraft burn more fuel per hour of flight than most tactical aircraft. The Air Force bunker fuel percentage was determined to be 13.2 percent. This percentage was multiplied by total annual Air Force aviation fuel delivered for U.S. activities, producing an estimate for international bunker fuel consumed by the Air Force. The Naval Aviation bunker fuel percentage of total fuel was calculated using flying hour data from Chief of Naval Operations Flying Hour Projection System Budget for fiscal year 1998, and estimates of bunker fuel percent of flights provided by the fleet. The Naval Aviation bunker fuel percentage, determined to be 40.4 percent, was multiplied by total annual Navy aviation fuel delivered for U.S. activities, yielding total Navy aviation bunker fuel consumed.

For marine bunkers, fuels consumed while ships were underway were assumed to be bunker fuels. In 2000, the Navy reported that 79 percent of vessel operations were underway, while the remaining 21 percent of operations occurred in port (i.e., pierside). Therefore, the Navy maritime bunker fuel percentage was determined to be 79 percent. The percentage of time underway may vary from year-to-year. For example, for years prior to 2000, the bunker fuel percentage was 87 percent. Table A-138 and Table A-139 display DoD bunker fuel use totals for the Navy and Air Force.

Step 6: Calculate Emissions from International Bunker Fuels

Bunker fuel totals were multiplied by appropriate emission factors to determine GHG emissions. CO_2 emissions from Aviation Bunkers and distillate Marine Bunkers are the total of military aviation and marine bunker fuels, respectively.

The rows labeled "U.S. Military" and "U.S. Military Naval Fuels" within Table 3-50 and Table 3-51 in the Energy Chapter were based on the international bunker fuel totals provided in Table A-138 and Table A-139, below. CO_2 emissions from aviation bunkers and distillate marine bunkers presented in Table A-142, and are based on emissions from fuels tallied in Table A-138 and Table A-139.

Table A-137: Transportation Fuels from Domestic Fuel Deliveries[a] (Million Gallons)

Vehicle Type/Fuel	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Aviation	**4,598.4**	**4,562.8**	**3,734.5**	**3,610.8**	**3,246.2**	**3,099.9**	**2,941.9**	**2,685.6**	**2,741.4**	**2,635.2**	**2,664.4**	**2,900.6**	**2,609.8**	**2,615.0**	**2,703.1**	**2,147.4**
Total Jet Fuels	4,598.4	4,562.8	3,734.5	3,610.8	3,246.2	3,099.9	2,941.9	2,685.6	2,741.4	2,635.2	2,664.4	2,900.6	2,609.6	2,614.9	2,703.1	2,147.3
JP8	285.7	283.5	234.5	989.4	1,598.1	2,182.8	2,253.1	2,072.0	2,122.5	2,066.5	2,122.7	2,326.2	2,091.4	2,094.3	2,126.2	1,786.4
JP5	1,025.4	1,017.4	832.7	805.1	723.8	691.2	615.8	552.8	515.6	505.5	472.1	503.2	442.2	409.1	433.7	308.4
Other Jet Fuels	3,287.3	3,261.9	2,667.3	1,816.3	924.3	225.9	72.9	60.9	103.3	63.3	69.6	71.2	76.1	111.4	143.2	52.5
Aviation Gasoline	+	+	+	+	+	+	+	+	+	+	+	+	0.1	0.1	+	0.1
Marine	**686.8**	**632.6**	**646.2**	**589.4**	**478.6**	**438.9**	**493.3**	**639.8**	**674.2**	**598.9**	**454.4**	**418.4**	**455.8**	**609.1**	**704.5**	**593.9**
Middle Distillate (MGO)	0.0	0.0	0.0	0.0	0.0	0.0	38.5	47.5	51.1	49.2	48.3	33.0	41.2	88.1	71.2	108.5
Naval Distillate (F76)	686.8	632.6	646.2	589.4	478.6	438.9	449.0	583.4	608.4	542.9	398.0	369.1	395.1	460.9	583.5	446.2
Intermediate Fuel Oil (IFO)[b]	0.0	0.0	0.0	0.0	0.0	0.0	5.9	9.0	14.7	6.7	8.1	16.3	19.5	60.2	49.9	39.1
Other[c]	**717.1**	**590.4**	**491.7**	**415.1**	**356.1**	**310.9**	**276.9**	**263.3**	**256.8**	**256.0**	**248.2**	**109.8**	**211.1**	**221.2**	**170.9**	**128.4**
Diesel	93.0	97.9	103.0	108.3	113.9	119.9	126.1	132.6	139.5	146.8	126.6	26.6	57.7	60.8	46.4	36.2
Gasoline	624.1	492.5	388.7	306.8	242.1	191.1	150.8	119.0	93.9	74.1	74.8	24.7	27.5	26.5	19.4	16.5
Jet Fuel[d]	0.0	0.0	0.0	0.0	0.0	0.0	0.0	11.7	23.4	35.0	46.7	58.4	125.9	133.9	105.1	75.7
Total (Including Bunkers)	**6,002.4**	**5,785.9**	**4,872.3**	**4,615.3**	**4,080.9**	**3,849.8**	**3,712.1**	**3,588.8**	**3,672.4**	**3,490.1**	**3,367.0**	**3,428.8**	**3,276.7**	**3,445.3**	**3,578.5**	**2,869.6**

Note: Totals may not sum due to independent rounding.

[a] Includes fuel consumption in the United States and U.S. Territories.

[b] Intermediate fuel oil (IFO 180 and IFO 380) is a blend of distillate and residual fuels. IFO is used by the Military Sealift Command.

[c] Prior to 2001, gasoline and diesel fuel totals were estimated using data provided by the military Services for 1990 and 1996. The 1991 through 1995 data points were interpolated from the Service inventory data. The 1997 through 1999 gasoline and diesel fuel data were initially extrapolated from the 1996 inventory data. Growth factors used for other diesel and gasoline were 5.2 and -21.1 percent, respectively. However, prior diesel fuel estimates from 1997 through 2000 were reduced according to the estimated consumption of jet fuel that is assumed to have replaced the diesel fuel consumption in land-based vehicles. Data sets for other diesel and gasoline consumed by the military in 2000 were estimated based on ground fuels consumption trends. This method produced a result that was more consistent with expected consumption for 2000. In 2001, other gasoline and diesel fuel totals were generated by DESC.

[d] The fraction of jet fuel consumed in land-based vehicles was estimated using Service data, DESC data, and expert judgment.

+ Does not exceed 0.05 million gallons.

Table A-138: Total U.S. Military Aviation Bunker Fuel (Million Gallons)

Fuel Type/Service	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
JP8	**56.7**	**56.3**	**46.4**	**145.3**	**224.0**	**300.4**	**308.8**	**292.0**	**306.4**	**301.4**	**307.6**	**341.2**	**309.5**	**305.1**	**309.8**	**271.6**
Navy	56.7	56.3	46.1	44.6	40.1	38.3	39.8	46.9	53.8	55.5	53.4	73.8	86.6	76.3	79.2	63.1
Air Force	+	+	0.3	100.8	183.9	262.2	269.0	245.1	252.6	245.9	254.2	267.4	222.9	228.7	230.6	208.5
JP5	**370.5**	**367.7**	**300.9**	**291.0**	**261.6**	**249.8**	**219.4**	**194.2**	**184.4**	**175.4**	**160.3**	**169.7**	**158.3**	**146.1**	**157.9**	**61.5**
Navy	365.3	362.5	296.7	286.8	257.9	246.3	216.1	191.2	181.4	170.6	155.6	163.7	153.0	141.3	153.8	53.5
Air Force	5.3	5.2	4.3	4.1	3.7	3.5	3.3	3.0	3.0	4.8	4.7	6.1	5.3	4.9	4.1	8.0
JP4	**420.8**	**417.5**	**341.4**	**229.6**	**113.1**	**21.5**	**1.1**	**0.1**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**
Navy	+	+	+	+	+	+	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	+
Air Force	420.8	417.5	341.4	229.6	113.1	21.5	1.1	0.1	+	+	+	+	+	+	+	+
JAA	**13.7**	**13.6**	**11.1**	**10.8**	**9.7**	**9.2**	**10.3**	**9.4**	**10.8**	**10.8**	**12.5**	**12.6**	**13.7**	**21.7**	**30.0**	**9.7**
Navy	8.5	8.4	6.9	6.6	6.0	5.7	6.6	5.9	6.6	6.3	7.9	8.0	9.8	15.5	21.5	7.5
Air Force	5.3	5.2	4.3	4.1	3.7	3.5	3.7	3.5	4.2	4.5	4.5	4.6	3.8	6.2	8.6	2.3
JA1	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**0.1**	**0.6**	**0.2**	**0.5**	**0.1**
Navy	+	+	+	+	+	+	+	+	+	0.0	+	+	+	+	+	0.1
Air Force	+	+	+	+	+	+	+	+	+	0.0	+	0.1	0.6	0.2	0.5	+
JAB	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**
Navy	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Air Force	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Navy Subtotal	**430.5**	**427.2**	**349.6**	**338.1**	**303.9**	**290.2**	**262.5**	**244.0**	**241.8**	**232.4**	**216.9**	**245.5**	**249.4**	**233.1**	**254.4**	**124.2**
Air Force Subtotal	**431.3**	**427.9**	**350.2**	**338.6**	**304.4**	**290.7**	**277.0**	**251.7**	**259.9**	**255.2**	**263.5**	**278.1**	**232.7**	**239.9**	**243.7**	**218.8**
Total	**861.8**	**855.1**	**699.9**	**676.7**	**608.4**	**580.9**	**539.5**	**495.6**	**501.7**	**487.5**	**480.4**	**523.6**	**482.1**	**473.0**	**498.1**	**342.9**

+ Does not exceed 0.05 million gallons.
Note: Totals may not sum due to independent rounding.

Table A-139: Total U.S. DoD Maritime Bunker Fuel (Million Gallons)

Marine Distillates	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Navy—MGO	0.0	0.0	0.0	0.0	0.0	0.0	30.3	35.6	31.9	39.7	23.8	22.5	27.1	63.7	56.2	77.1
Navy—F76	522.4	481.2	491.5	448.3	364.0	333.8	331.9	441.7	474.2	466.0	298.6	282.6	305.6	347.8	434.4	319.2
Navy—IFO	0.0	0.0	0.0	0.0	0.0	0.0	4.6	7.1	11.6	5.3	6.4	12.9	15.4	47.5	39.4	30.9
Total	**522.4**	**481.2**	**491.5**	**448.3**	**364.0**	**333.8**	**366.8**	**484.3**	**517.7**	**511.0**	**328.8**	**318.0**	**348.2**	**459.0**	**530.0**	**427.3**

+ Does not exceed 0.005 million gallons.
Note: Totals may not sum due to independent rounding.

Table A-140: Aviation and Marine Carbon Contents (Tg Carbon/QBtu) and Fraction Oxidized

Mode (Fuel)	Carbon Content Coefficient	Fraction Oxidized
Aviation (Jet Fuel)	variable	1.00
Marine (Distillate)	19.93	1.00
Marine (Residual)	21.49	1.00

Source: EIA (2006) and IPCC (2006)

Table A-141: Annual Variable Carbon Content Coefficient for Jet Fuel (Tg Carbon/QBtu)

Fuel	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Jet Fuel	19.40	19.40	19.39	19.37	19.35	19.34	19.33	19.33	19.33	19.33	19.33	19.33	19.33	19.33	19.33	19.33

Source: EIA (2006)

Table A-142: Total U.S. DoD CO_2 Emissions from Bunker Fuels (Tg CO_2 Eq.)

Mode	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Aviation	8.1	8.0	6.5	6.4	5.8	5.5	5.1	4.7	4.8	4.6	4.6	5.0	4.6	4.5	4.7	3.3
Marine	5.3	4.9	5.0	4.5	3.7	3.4	3.7	4.9	5.3	5.2	3.3	3.2	3.5	4.7	5.4	4.3
Total	**13.4**	**12.9**	**11.5**	**10.9**	**9.5**	**8.9**	**8.9**	**9.6**	**10.0**	**9.8**	**7.9**	**8.2**	**8.1**	**9.2**	**10.1**	**7.6**

Note: Totals may not sum due to independent rounding.

3.8. Methodology for Estimating HFC and PFC Emissions from Substitution of Ozone Depleting Substances

The Vintaging Model was developed as a tool for estimating the annual chemical emissions from industrial sectors that have historically used ODS in their products. Under the terms of the Montreal Protocol and the United States' Clean Air Act Amendments of 1990, the domestic U.S. production of ODS—chlorofluorocarbons (CFCs), halons, carbon tetrachloride, methyl chloroform, and hydrochlorofluorocarbons (HCFCs)—has been drastically reduced, forcing these industrial sectors to transition to more ozone friendly chemicals. As these industries have moved toward ODS alternatives such as hydrofluorocarbons (HFCs) and perfluorocarbons (PFCs), the Vintaging Model has evolved into a tool for estimating the rise in consumption and emissions of these alternatives, and the decline of ODS consumption and emissions.

The Vintaging Model estimates emissions from five ODS substitute end-use sectors; air-conditioning and refrigeration, foams, aerosols, solvents, and fire-extinguishing. Within these sectors, there are over 40 independently modeled end-uses. The model requires information on the market growth for each of the end-uses, as well as a history of the market transition from ODS to alternatives. As ODS are phased out, a percentage of the market share originally filled by the ODS is allocated to each of its substitutes.

The model, named for its method of tracking the emissions of annual "vintages" of new equipment that enter into service, is a "bottom-up" model. It models the consumption of chemicals based on estimates of the quantity of equipment or products sold, serviced, and retired each year, and the amount of the chemical required to manufacture and/or maintain the equipment. The Vintaging Model makes use of this market information to build an inventory of the in-use stocks of the equipment and ODS and ODS substitute in each of the end-uses. The simulation is considered to be a "business-as-usual" baseline case, and does not incorporate measures to reduce or eliminate the emissions of these gases other than those regulated by U.S. law or otherwise common in the industry. Emissions are estimated by applying annual leak rates, service emission rates, and disposal emission rates to each population of equipment. By aggregating the emission and consumption output from the different end-uses, the model produces estimates of total annual use and emissions of each chemical.

The Vintaging Model synthesizes data from a variety of sources, including data from the ODS Tracking System maintained by the Stratospheric Protection Division and information from submissions to EPA under the Significant New Alternatives Policy (SNAP) program. Published sources include documents prepared by the United Nations Environment Programme (UNEP) Technical Options Committees, reports from the Alternative Fluorocarbons Environmental Acceptability Study (AFEAS), and conference proceedings from the International Conferences on Ozone Protection Technologies and Earth Technologies Forums. EPA also coordinates extensively with numerous trade associations and individual companies. For example, the Alliance for Responsible Atmospheric Policy, the Air-Conditioning and Refrigeration Institute, the Association of Home Appliance Manufacturers, the American Automobile Manufacturers Association, and many of their member companies, have provided valuable information over the years. In some instances the unpublished information that the EPA uses in the model is classified as Confidential Business Information (CBI). The annual emissions inventories of chemicals are aggregated in such a way that CBI cannot be inferred. Full public disclosure of the inputs to the Vintaging Model would jeopardize the security of the CBI that has been entrusted to the EPA.

The following sections discuss the forms of the emission estimating equations used in the Vintaging Model for each broad end-use category. These equations are applied separately for each chemical used within each of the different end-uses. In the majority of these end-uses, more than one ODS substitute chemical is used.

In general, the modeled emissions are a function of the amount of chemical consumed in each end-use market. Estimates of the consumption of ODS alternatives can be inferred by extrapolating forward in time from the amount of regulated ODS used in the early 1990s. Using data gleaned from a variety of sources, assessments are made regarding which alternatives will likely be used, and what fraction of the ODS market in each end-use will be captured by a given alternative. By combining this with estimates of the total end-use market growth, a consumption value can be estimated for each chemical used within each end-use.

Methodology

The Vintaging Model estimates the use and emissions of ODS alternatives by taking the following steps:

1. *Gather historical emissions data*. The Vintaging Model is populated with information on each end-use, taken from published sources and industry experts.

2. *Simulate the implementation of new, non-ODS technologies*. The Vintaging model uses detailed characterizations of the existing uses of the ODSs, as well as data on how the substitutes are replacing the ODSs, to simulate the implementation of new technologies that ensure compliance with ODS phase-out policies. As part of this simulation, the ODS substitutes are introduced in each of the end-uses over time as needed to comply with the ODS phase-out.

3. *Estimate emissions of the ODS substitutes*. The chemical use is estimated from the amount of substitutes that are required each year for the manufacture, installation, use, or servicing of products. The emissions are estimated from the emission profile for each vintage of equipment or product in each end-use. By aggregating the emissions from each vintage, a time profile of emissions from each end-use is developed.

Each set of end uses is discussed in more detail in the following sections.

Refrigeration and Air-Conditioning

For refrigeration and air conditioning products, emission calculations are split into two categories: emissions during equipment lifetime, which arise from annual leakage and service losses, and disposal emissions, which occur at the time of discard. Two separate steps are required to calculate the lifetime emissions from leakage and service, and the emissions resulting from disposal of the equipment. These lifetime emissions and disposal emissions are summed to calculate the total emissions from refrigeration and air-conditioning. As new technologies replace older ones, it is generally assumed that there are improvements in their leak, service, and disposal emission rates.

Step 1: Calculate lifetime emissions

Emissions from any piece of equipment include both the amount of chemical leaked during equipment operation and the amount emitted during service. Emissions from leakage and servicing can be expressed as follows:

$$Es_j = (l_a + l_s) \times \sum Qc_{j-i+1} \quad for\ i = 1 \rightarrow k$$

Where:

Es = Emissions from Equipment Serviced. Emissions in year j from normal leakage and servicing (including recharging) of equipment.

l_a = Annual Leak Rate. Average annual leak rate during normal equipment operation (expressed as a percentage of total chemical charge).

l_s = Service Leak Rate. Average leakage during equipment servicing (expressed as a percentage of total chemical charge).

Qc = Quantity of Chemical in New Equipment. Total amount of a specific chemical used to charge new equipment in a given year by weight.

i = Counter, runs from 1 to lifetime (k).

j = Year of emission.

k = Lifetime. The average lifetime of the equipment.

Step 2: Calculate disposal emissions

The disposal emission equations assume that a certain percentage of the chemical charge will be emitted to the atmosphere when that vintage is discarded. Disposal emissions are thus a function of the quantity of chemical contained in the retiring equipment fleet and the proportion of chemical released at disposal:

$$Ed_j = Qc_{j-k+1} \times [1 - (rm \times rc)]$$

Where:

Ed = Emissions from Equipment Disposed. Emissions in year *j* from the disposal of equipment.

Qc = Quantity of Chemical in New Equipment. Total amount of a specific chemical used to charge new equipment in year *j-k+1*, by weight.

rm = Chemical Remaining. Amount of chemical remaining in equipment at the time of disposal (expressed as a percentage of total chemical charge).

rc = Chemical Recovery Rate. Amount of chemical that is recovered just prior to disposal (expressed as a percentage of chemical remaining at disposal *(rm)*).

j = Year of emission.

k = Lifetime. The average lifetime of the equipment.

Step 3: Calculate total emissions

Finally, lifetime and disposal emissions are summed to provide an estimate of total emissions.

$$E_j = Es_j + Ed_j$$

Where:

E = Total Emissions. Emissions from refrigeration and air conditioning equipment in year *j*.

Es = Emissions from Equipment Serviced. Emissions in year *j* from normal leakage and servicing (recharging) of equipment.

Ed = Emissions from Equipment Disposed. Emissions in year *j* from the disposal of equipment.

j = Year of emission.

Assumptions

The assumptions used by the Vintaging Model to trace the transition of each type of equipment away from ODS are presented in Table A- 143, below. As new technologies replace older ones, it is generally assumed that there are improvements in their leak, service, and disposal emission rates. Additionally, the market for each equipment type is assumed to grow independently, according to annual growth rates, presented in Table A- 143.

Table A- 143. Refrigeration and Air-Conditioning Market Transition Assumptions

Initial Market Segment	Primary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Secondary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Growth Rate
Mobile Air Conditioners									
CFC-12	HFC-134a	1992	1994	100%	None				2.6%
CFC-12	HFC-134a	1994	1997	99.5%	None				2.6%
	HCFC-22	1994	1995	0.5%	HFC-134a	2006	2007	100%	
HCFC-22	HFC-134a	2002	2009	50%	None				2.6%
	R-407C	2002	2009	50%	None				
HCFC-22	HFC-134a	1995	2009	100%	None				2.6%
Chillers									
CFC-11	HCFC-123	1993	1993	45%	HFC-134a	2015	2019	75%	0.5%
					HFC-245fa	2015	2019	25%	
	HCFC-22	1991	1993	16%	HFC-134a	2000	2010	70%	
					R-407C	2000	2010	30%	
	HFC-134a	1992	1993	39%	None				
CFC-12	HFC-134a	1992	1994	53%	None				0.5%
	HCFC-22	1991	1994	16%	HFC-134a	2000	2009	70%	
					R-407C	2000	2009	30%	
	HCFC-123	1993	1994	31%	HFC-134a	2015	2019	75%	
					HFC-245fa	2015	2019	25%	
R-500	HFC-134a	1992	1994	53%	None				0.5%

Initial Market Segment	Primary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Secondary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Growth Rate
	HCFC-22	1991	1994	16%	HFC-134a	2000	2009	70%	
					R-407C	2000	2009	30%	
	HCFC-123	1993	1994	31%	HFC-134a	2015	2019	75%	
					HFC-245fa	2015	2019	25%	
HCFC-22*	HCFC-22	1992	1993	100%	HFC-134a	2000	2009	7%	0.5%
					R-407C	2000	2009	3%	
					HFC-134a	2009	2010	70%	
					R-407C	2009	2010	30%	
CFC-114	HFC-236fa	1993	1994	100%	HFC-134a	1998	2009	100%	0.2%
Cold Storage									
CFC-12	HCFC-22	1990	1993	65%	R-404A	1996	2010	75%	2.5%
					R-507	1996	2010	25%	
	HFC-134a	1994	1996	35%	HFC-134a	2005	2005	100%	
HCFC-22	HCFC-22	1992	1993	100%	R-404A	1996	2009	8%	2.5%
					R-507	1996	2009	3%	
					R-404A	2009	2010	68%	
					R-507	2009	2010	23%	
R-502	HCFC-22	1990	1993	40%	R-404A	1996	2010	38%	2.5%
					R-507	1996	2010	12%	
					Non-ODP/GWP	1996	2010	50%	
	R-404A	1993	1996	45%	R-404A	2010	2010	100%	
	R-507	1994	1996	15%	R-507	2010	2010	100%	
Commercial Unitary Air Conditioners									
HCFC-22	R-410A	2000	2006	3%	None				2.5%
	R-410A	2006	2009	18%	None				
	R-410A	2000	2006	8%	None				
	R-410A	2006	2009	71%	None				
HCFC-22	R-410A	2001	2005	1%	None				2.5%
	R-407C	2006	2009	9%	None				
	R-410A	2006	2009	5%					
	R-407C	2009	2010	81%	None				
	R-410A	2009	2010	3%	None				
Dehumidifiers									
HCFC-22	HFC-134a	1997	1997	89%	None				0.5%
	R-410A	2007	2010	11%	None				
Ice Makers									
CFC-12	HFC-134a	1993	1995	100%	None				
Industrial Process Refrigeration									
CFC-11	HCFC-123	1992	1994	70%	HFC-134a	2015	2019	100%	
	HFC-134a	1992	1994	15%	None				
	HCFC-22	1991	1994	15%	HFC-134a	2010	2010	100%	
CFC-12	HCFC-22	1991	1994	10%	HFC-134a	1995	2010	15%	2.5%
					R-404A	1995	2010	50%	
					R-410A	1999	2010	20%	
					R-507	1995	2010	15%	
	HCFC-123	1992	1994	35%	HFC-134a	2015	2019	100%	
	HFC-134a	1992	1994	50%	None				
	R-401A	1995	1996	5%	HFC-134a	1997	2000	100%	
HCFC-22	HCFC-22	1992	1993	100%	HFC-134a	1995	2009	2%	2.5%
					R-404A	1995	2009	5%	
					R-410A	1999	2009	2%	
					R-507	1995	2009	2%	
					HFC-134a	2009	2010	14%	
					R-404A	2009	2010	45%	
					R-410A	2009	2010	18%	
					R-507	2009	2010	14%	
Refrigerated Appliances									

Initial Market Segment	Primary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Secondary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Growth Rate
CFC-12	HFC-134a	1994	1995	100%	None				0.5%
Residential Unitary Air Conditioners									
HCFC-22	R-410A	2007	2010	29%	None				1.9%
	R-407C	2010	2010	14%	None				
	R-410A	2011	2015	100%	None				
	R-410A	2010	2010	57%	None				
	R-410A	2007	2010	29%	None				
	R-410A	2000	2005	5%	R-410	2006	2006	100%	
	R-410A	2000	2006	5%	None				
	R-410A	2006	2006	20%	None				
Retail Food									
CFC-12	HCFC-22	1990	1993	70%	R-404A	1995	2000	18%	1.7%
					R-507	1995	2000	8%	
					R-404A	2000	2005	32%	
					R-507	2000	2005	14%	
					R-404A	2005	2010	18%	
					R-507	2005	2010	12%	
HCFC-22	HCFC-22	1990	1993	70%	R-404A	1995	2000	18%	1.7%
					R-507	1995	2000	8%	
					R-404A	2000	2005	32%	
					R-507	2000	2005	14%	
					R-404A	2005	2010	18%	
					R-507	2005	2010	12%	
R-502	HCFC-22	1990	1993	40%	R-404	2000	2010	75%	1.7%
					R-507	2000	2010	25%	
	R-404	1993	1996	40%	R-404	2005	2005	100%	
	R-507	1994	1996	10%	R-507	2005	2005	100%	
	HFC-134a	1996	1996	10%	None				
Transport Refrigeration									
CFC-12	HFC-134a	1993	1995	98%	None				2.5%
	HCFC-22	1993	1995	2%	HFC-134a	1995	1999	100%	
R-502	HFC-134a	1993	1995	55%	None				2.5%
	R-404A	1993	1995	45%	R-404A	2005	2005	100%	
Water-Source, Ground-Source and Unitary Heat Pumps; Packaged Terminal Air Conditioners and Heat Pumps									
HCFC-22	R-407C	2000	2006	5%	None				2.5%
	R-410A	2000	2006	5%	None				
	HFC-134a	2000	2009	2%	None				
	R-407C	2006	2009	3%	None				
	R-410A	2006	2009	5%	None				
	HFC-134a	2009	2010	18%	None				
	R-407C	2009	2010	23%	None				
	R-410A	2009	2010	41%	None				
HCFC-22	R-410A	2006	2009	10%	None				2.5%
	R-410A	2006	2010	90%	None				
Window Units									
HCFC-22	R-407C	2003	2009	3%	None				5.0%
	R-407C	2009	2010	35%	None				
	R-410A	2003	2009	7%	None				
	R-410A	2009	2010	55%	None				

* HCFC-22 Chillers has a tertiary substitution; R-407C is substituted with R-407C (60%) and R-410A (40%). Substitution begins in 2010, with 100% penetration in new equipment by 2020.

Table A- 144 presents the average equipment lifetimes for each end use assumed by the Vintaging Model.

Table A- 144. Refrigeration and Air-conditioning Lifetime Assumptions

End Use	Lifetime (Years)
Mobile Air Conditioners	5-12
Chillers	20 - 27

Retail Food	15 - 20
Cold Storage	20 - 25
Industrial Process Refrigeration	25
Transport Refrigeration	12
Dehumidifiers	11
Ice Makers	20
Refrigerated Appliances	20
Residential Unitary A/C	15
Commercial Unitary A/C	15
Water & Ground Source Heat Pumps	20
PTAC/PTHP	12
Window Units	12

Aerosols

ODSs, HFCs and many other chemicals are used as propellant aerosols. Pressurized within a container, a nozzle releases the chemical, which allows the product within the can to also be released. Two types of aerosol products are modeled, including metered dose inhalers and consumer aerosols. In the United States, the use of ODSs in consumer aerosols was banned in 1977, and many products transitioned to "not-in-kind" technologies, such as solid deodorants and finger-pump hair sprays.

All HFCs and PFCs used in aerosols are assumed to be emitted in the year of manufacture. Since there is currently no aerosol recycling, it is assumed that all of the annual production of aerosol propellants is released to the atmosphere. The following equation describes the emissions from the aerosols sector.

$$E_j = Qc_j$$

Where:

E = Emissions. Total emissions of a specific chemical in year j from use in aerosol products, by weight.

Qc = Quantity of Chemical. Total quantity of a specific chemical contained in aerosol products sold in year j, by weight.

j = Year of emission.

Assumptions

Transition assumptions and growth rates for those items that use ODSs or HFCs as propellants, including vital medical devices and specialty consumer products, are presented in Table A- 145.

Table A- 145. Aerosol Product Transition Assumptions

Initial Market Segment	Primary Substitute	Start Date	Date of Full Penetration in New Products	Maximum Market Penetration	Secondary Substitute	Start Date	Date of Full Penetration in New Products	Maximum Market Penetration	Growth Rate
MDI Aerosols									
CFC-11	HFC-134a	1997	2008	9%	None				1.5%
	HFC-227ea	1997	2008	1%	None				
	HFC-134a	2009	2009	27%	None				
	HFC-227ea	2009	2009	3%	None				
	HFC-134a	2009	2015	54%	None				
	HFC-227ea	2009	2015	6%	None				
CFC-12	HFC-134a	1997	2008	9%	None				1.5%
	HFC-227ea	1997	2008	1%	None				
	HFC-134a	2009	2009	27%	None				
	HFC-227ea	2009	2009	3%	None				
	HFC-134a	2009	2015	54%	None				
	HFC-227ea	2009	2015	6%	None				
CFC-114	HFC-134a	1997	2008	9%	None				1.5%
	HFC-227ea	1997	2008	1%	None				

	HFC-134a	2009	2009	27%	None				
	HFC-227ea	2009	2009	3%	None				
	HFC-134a	2009	2015	54%	None				
	HFC-227ea	2009	2015	6%	None				
Consumer Aerosols									
NA*	HFC-152a	1990	1991	50%	None				2.0%
	HFC-134a	1995	1995	50%	HFC-152a	1997	1998	44%	
					HFC-134a	1997	1998	56%	

*Consumer Aerosols transitioned away from ODS prior to the beginning of the Vintaging Model, which begins in 1985. The portion of the market that is now using HFC propellants is modeled.

Solvents

ODSs, HFCs, PFCs and other chemicals are used as solvents to clean items. For example, electronics may need to be cleaned after production to remove any manufacturing process oils or residues left. Solvents are applied by moving the item to be cleaned within a bath or stream of the solvent. Generally, most solvents are assumed to remain in the liquid phase and are not emitted as gas. Thus, emissions are considered "incomplete," and are a fixed percentage of the amount of solvent consumed in a year. The remainder of the consumed solvent is assumed to be reused or disposed without being released to the atmosphere. The following equation calculates emissions from solvent applications.

$$E_j = l \times Qc_j$$

Where:

E = Emissions. Total emissions of a specific chemical in year j from use in solvent applications, by weight.

l = Percent Leakage. The percentage of the total chemical that is leaked to the atmosphere, assumed to be 90 percent.

Qc = Quantity of Chemical. Total quantity of a specific chemical sold for use in solvent applications in the year j, by weight.

j = Year of emission.

Assumptions

The transition assumptions and growth rates used within the Vintaging Model for electronics cleaning, metals cleaning, precision cleaning, and adhesives, coatings and inks, are presented in Table A- 146.

Table A- 146. Solvent Market Transition Assumptions

Initial Market Segment	Primary Substitute	Start Date	Date of Full Penetration in New Uses	Maximum Market Penetration	Secondary Substitute	Start Date	Date of Full Penetration in New Uses	Maximum Market Penetration	Growth Rate
Electronics Cleaning									
CFC-113	Non-ODP/GWP	1992	1996	40.3%	None				
	Non-ODP/GWP	1992	1996	5.7%	None				2.0%
	HCFC-225ca/cb	1994	1995	0.2%	None				
	Non-ODP/GWP	1994	1995	52.5%	None				
	HFE-7100	1994	1995	0.7%	None				
	HFC-4310mee	1995	1996	0.7%	None				
MCF	Non-ODP/GWP	1996	1997	28.5%	None				2.0%
	Non-ODP/GWP	1996	1997	6.5%	None				
	Non-ODP/GWP	1996	1997	8.5%	None				
	PFC/PFPE	1996	1997	28.5%	Non-ODP/GWP	2000	2003	90%	
					Non-ODP/GWP	2005	2009	10%	
	Non-ODP/GWP	1996	1997	56.3%	None				
Metals Cleaning									
MCF	Non-ODP/GWP	1992	1996	100%	None				2.0%
CFC-113	Non-ODP/GWP	1992	1996	100%	None				2.0%
CCl4	Non-ODP/GWP	1992	1996	100%	None				2.0%
Precision Cleaning									

MCF	Non-ODP/GWP	1995	1996	14.5%	None					2.0%
	Non-ODP/GWP	1995	1996	9.6%						
	Non-ODP/GWP	1995	1996	29.4%					e	
	Non-ODP/GWP	1995	1996	11.7%						
	HFC-4310mee	1995	1996	0.6%	None					
	PFC/PFPE	1995	1996	0.1%	Non-ODP/GWP	2000	2003	90%		
					Non-ODP/GWP	2005	2009	10%		
	Non-ODP/GWP	1995	1996	34.1%						
CFC-113	Non-ODP/GWP	1995	1996	90.2%	None					2.0%
	HCFC-225ca/cb	1995	1996	1.0%	None					
	HFE-7100	1995	1996	3.3%	None					
	Non-ODP/GWP	1995	1996	5.5%	None					
Adhesives, Coatings, Inks										
MCF	Non-ODP/GWP	1994	1995	100%	None					2.0%

MCF= Methyl Chloroform, also known as TCA or 1,1,1-Trichloroethane
Non-ODP/GWP includes chemicals with 0 ODP and low GWP, such as hydrocarbons and ammonia, as well as not-in-kind alternatives such as "no clean" technologies.

Fire Extinguishing

ODSs, HFCs, PFCs and other chemicals are used as fire-extinguishing agents, in both hand-held "streaming" applications as well as in built-up "flooding" equipment similar to water sprinkler systems. Although these systems are generally built to be leak-tight, some leaks do occur and of course emissions occur when the agent is released. Total emissions from fire extinguishing are assumed, in aggregate, to equal a percentage of the total quantity of chemical in operation at a given time. For modeling purposes, it is assumed that fire extinguishing equipment leaks at a constant rate for an average equipment lifetime, as shown in the equation below. In streaming systems, emissions are assumed to be 2 percent of all chemical in use in each year, while in flooding systems 1.5 percent of the installed base of chemical is assumed to leak annually. The equation is applied for a single year, accounting for all fire protection equipment in operation in that year. Each fire protection agent is modeled separately. In the Vintaging Model, streaming applications have a 10-year lifetime and flooding applications have a 20-year lifetime.

$$E_j = \mathrm{r} \times \sum Qc_{j-i+1} \quad for\ i=1 \rightarrow k$$

Where:

E = Emissions. Total emissions of a specific chemical in year j for streaming fire extinguishing equipment, by weight.

r = Percent Released. The percentage of the total chemical in operation that is released to the atmosphere.

Qc = Quantity of Chemical. Total amount of a specific chemical used in new fire extinguishing equipment in a given year, j-i+1, by weight.

i = Counter, runs from 1 to lifetime (k).

j = Year of emission.

k = Lifetime. The average lifetime of the equipment.

Assumptions

Transition assumptions and growth rates for these two fire extinguishing types are presented in Table A-147.

Table A-147. Fire Extinguishing Market Transition Assumptions

Initial Market Segment	Primary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Secondary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Growth Rate
Streaming Agents									

Halon 1211	HFC-236fa	1997	1999	4%	None				3.0%
	Blends	1995	1999	4%	Non-ODP/GWP	2015	2015	25%	
					HFC-236fa	2015	2015	75%	
	Non-ODP/GWP	1993	1994	88%	None				
Flooding Agents									
Halon 1301	HFC-23	1994	1999	0.5%	None				2.2%
	HFC-227ea	1994	1999	28.0%	Novec 1230	2003	2010	10%	
					HFC-125	2001	2008	10%	
	Non-ODP/GWP	1994	1999	66.0%	Novec 1230	2003	2010	10%	
	C4F10	1994	1999	0.5%	Novec 1230	2003	2010	10%	
	HFC-125	1997	2006	1.0%	None				

Foam Blowing

ODSs, HFCs, and other chemicals are used to produce foams, including such items as the foam insulation panels around refrigerators, insulation sprayed on buildings, etc. The chemical is used to create pockets of gas within a substrate, increasing the insulating properties of the item. Foams are given emission profiles depending on the foam type (open cell or closed cell). Open cell foams are assumed to be 100 percent emissive in the year of manufacture. Closed cell foams are assumed to emit a portion of their total HFC or PFC content upon manufacture, a portion at a constant rate over the lifetime of the foam, and a portion at disposal.

Step 1: Calculate emissions from open-cell foam

Emissions from open-cell foams are calculated using the following equation.

$$E_j = Qc_j$$

Where:

E = Emissions. Total emissions of a specific chemical in year j used for open-cell foam blowing, by weight.

Qc = Quantity of Chemical. Total amount of a specific chemical used for open-cell foam blowing in year j, by weight.

j = Year of emission.

Step 2: Calculate emissions from closed-cell foam

Emissions from foams occur at many different stages, including manufacturing, lifetime, disposal and post-disposal. Manufacturing emissions occur in the year of foam manufacture, and are calculated as presented in the following equation.

$$Em_j = lm \times Qc_j$$

Where:

Em_j = Emissions from manufacturing. Total emissions of a specific chemical in year j due to manufacturing losses, by weight.

lm = Loss Rate. Percent of original blowing agent emitted during foam manufacture.

Qc = Quantity of Chemical. Total amount of a specific chemical used to manufacture closed-cell foams in a given year.

j = Year of emission.

Lifetime emissions occur annually from closed cell foams throughout the lifetime of the foam, as calculated as presented in the following equation.

$$Eu_j = \text{lu} \times \sum Qc_{j-i+1} \quad for\ i=1 \rightarrow k$$

Where:

Eu_j = Emissions from Lifetime Losses. Total emissions of a specific chemical in year j due to lifetime losses during use, by weight.

lu = Leak Rate. Percent of original blowing agent emitted during lifetime use.

Qc = Quantity of Chemical. Total amount of a specific chemical used to manufacture closed-cell foams in a given year..

i = Counter, runs from 1 to lifetime (k).

j = Year of emission.

k = Lifetime. The average lifetime of foam product.

Disposal emissions occur in the year the foam is disposed, and are calculated as presented in the following equation.

$$Ed_j = ld \times Qc_{j+k}$$

Where:

Ed_j = Emissions from disposal. Total emissions of a specific chemical in year j at disposal, by weight.

ld = Loss Rate. Percent of original blowing agent emitted at disposal.

Qc = Quantity of Chemical. Total amount of a specific chemical used to manufacture closed-cell foams in a given year.

j = Year of emission.

k = Lifetime. The average lifetime of foam product.

Post-Disposal emissions occur in the years after the foam is disposed, and are assumed to occur while the disposed foam is in a landfill. Currently, the only foam type assumed to have post-disposal emissions is polyurethane appliance foam, which is expected to continue to emit for 32 years post-disposal, and are calculated as presented in the following equation.

$$Ep_j = lp \times \sum Qc_{j-m} \quad for\ m=k \rightarrow k+32$$

Where:

Ep_j = Emissions from post disposal. Total emissions of a specific chemical in year j, by weight.

lp = Leak Rate. Percent of original blowing agent emitted post disposal.

Qc = Quantity of Chemical. Total amount of a specific chemical used to manufacture closed-cell foams in a given year.

k = Lifetime. The average lifetime of foam product.

m = Counter. Runs from lifetime (k) to (k+32).

j = Year of emission.

To calculate total emissions from foams in any given year, emissions from all foam stages must be summed, as presented in the following equation.

$$E_j = Em_j + Eu_j + Eu_j + Ed_j + Ep_j$$

Where:

E_j = Total Emissions. Total emissions of a specific chemical in year *j*, by weight.

Em = Emissions from manufacturing. Total emissions of a specific chemical in year *j* due to manufacturing losses, by weight.

Eu_j = Emissions from Lifetime Losses. Total emissions of a specific chemical in year j due to lifetime losses during use, by weight.

Ed_j = Emissions from disposal. Total emissions of a specific chemical in year j at disposal, by weight.

Ep_j = Emissions from post disposal. Total emissions of a specific chemical in year j, by weight.

Assumptions

The Vintaging Model contains 13 foam types, whose transition assumptions away from ODS and growth rates are presented in Table A- 148. The emission profiles of the foam types estimating in the Vintaging Model are shown in Table A- 149.

Table A- 148. Foam Blowing Market Transition Assumptions

Initial Market Segment	Primary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Secondary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Growth Rate
Commercial Refrigeration Foam									
CFC-11	HCFC-141b	1989	1996	40%	HFC-245fa	2002	2003	80%	6.0%
					Non-ODP/GWP	2002	2003	20%	
	HCFC-142b	1989	1996	8%	Non-ODP/GWP	2009	2010	80%	
					HFC-245fa	2009	2010	20%	
	HCFC-22	1989	1996	52%	Non-ODP/GWP	2009	2010	80%	
					HFC-245fa	2009	2010	20%	
Flexible Polyurethane Foam									
CFC-11	Non-ODP/GWP	1992	1992	100%	None				2.0%
One Component Foam									
CFC-12	Blend	1989	1996	70%	Non-ODP/GWP	2009	2010	80%	4.0%
					HFC-134a	2009	2010	10%	
					HFC-152a	2009	2010	10%	
	HCFC-22	1989	1996	30%	Non-ODP/GWP	2009	2010	80%	
					HFC-134a	2009	2010	10%	
					HFC-152a	2009	2010	10%	
Phenolic Foam									
CFC-11	HCFC-141b	1989	1990	100%	Non-ODP/GWP	1992	1992	100%	2.0%
Polyisocyanurate Boardstock Foam									
CFC-11	HCFC-141b	1993	1996	100%	Non-ODP/GWP	2000	2003	95%	6.0%
					Blend	2000	2003	5%	
Polyolefin Foam									
CFC-114	HFC-152a	1989	1993	10%	Non-ODP/GWP	2005	2010	100%	2.0%
	HCFC-142b	1989	1993	90%	Non-ODP/GWP	1994	1996	100%	
Polystyrene Boardstock Foam									
CFC-12	Blend	1989	1994	10%	HFC-134a	2009	2010	70%	2.5%
					HFC-152a	2009	2010	10%	
					CO_2	2009	2010	10%	
					Non-ODP/GWP	2009	2010	10%	
	HCFC-142b	1989	1994	90%	HFC-134a	2009	2010	70%	
					HFC-152a	2009	2010	10%	
					CO_2	2009	2010	10%	
					Non-ODP/GWP	2009	2010	10%	
Polystyrene Sheet/Insulation Board Foam									
CFC-12	CO_2	1989	1994	1%	None				2.0%
	Non-ODP/GWP	1989	1994	99%	CO_2	1995	1999	9%	
					HFC-152a	1995	1999	10%	
Polyurethane Appliance Foam									
CFC-11	HCFC-141b	1993	1996	89%	HFC-134a	1996	2003	10%	3.0%

Initial Market Segment	Primary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Secondary Substitute	Start Date	Date of Full Penetration in New Equipment	Maximum Market Penetration	Growth Rate
					HFC-245fa	2002	2003	85%	
					Non-ODP/GWP	2002	2003	5%	
	Blend	1993	1996	1%	HFC-245fa	2009	2010	50%	
					HFC-134a	2009	2010	50%	
	HCFC-22	1993	1996	10%	HFC-134a	2009	2010	100%	
Polyurethane Integral Skin Foam									
CFC-11	HCFC-141b	1989	1990	100%	HFC-134a	1993	1996	25%	2.0%
					HFC-134a	1994	1996	25%	
					CO_2	1993	1996	25%	
					CO_2	1994	1996	25%	
Polyurethane Panel Foam*									
CFC-11	HCFC-141b	1989	1996	82%	Blend	2001	2003	20%	6.0%
					Blend	2002	2004	20%	
					Non-ODP/GWP	2001	2004	40%	
					HFC-134a	2002	2004	20%	
	HCFC-22	1989	1996	18%	Blend	2009	2010	40%	
					Non-ODP/GWP	2009	2010	20%	
					CO_2	2009	2010	20%	
					HFC-134a	2009	2010	20%	
Polyurethane Slabstock and Other Foam**									
CFC-11	HCFC-141b	1989	1996	100%	CO_2	1999	2003	45%	2.0%
					Non-ODP/GWP	2001	2003	45%	
					HCFC-22	2003	2003	10%	
Polyurethane Spray Foam									
CFC-11	HCFC-141b	1989	1996	100%	HFC-245fa	2002	2003	30%	6.0%
					Blend	2002	2003	60%	
					Non-ODP/GWP	2001	2003	10%	

* Polyurethane Panel Foam has a tertiary substitution; the first blend is assumed to contain HCFCs, and is thus substituted with a 50/50 mixture of another blend and a non-ODP/GWP substitute in 2009, with 100% penetration in new equipment by 2010.
** Polyurethane Slabstock and Other Foam has a tertiary substitution; HCFC-22 is substituted with a non-ODP/GWP substitute in 2009, with 100% penetration in new equipment in 2010.

Table A-149. Emission profile for the foam end-uses

Foam End-Use	Loss at Manufacturing (%)	Annual Leakage Rate (%)	Leakage Lifetime (years)	Loss at Disposal (%)	Total* (%)
Flexible PU	100	0	1	0	100
Polyisocyanurate Boardstock	6	1	50	44	100
Rigid PU Integral Skin	95	2.5	2	0	100
Rigid PU Appliance	4	0.25	20	27.3	36
Rigid PU Commercial Refrigeration	6	0.25	15	90.25	100
Rigid PU Spray	15	1.5	56	1	100
One Component	100	0	1	0	100
Rigid PU Slabstock and Other	37.5	0.75	15	51.25	100
Phenolic	23	0.875	32	49	100
Polyolefin	95	2.5	2	0	100
XPS Sheet/Insulation Board*	40	2	25	0	90
XPS Boardstock	25	0.75	50	37.5	100
PU Sandwich Panels	5.5	0.5	50	69.5	100

PU (Polyurethane)
XPS (Extruded Polystyrene)
*In general, total emissions from foam end-uses are assumed to be 100 percent, although work is underway to investigate that assumption. In the XPS Sheet/Insulation Board end-use, the source of emission rates and lifetimes did not yield 100 percent emission; it is unclear at this time whether that was intentional. In the Rigid PU Appliance Foam end-use, the source of emission rates and lifetimes did not yield 100 percent emission; the remainder is anticipated to be emitted at a rate of 2.0%/year post-disposal for the next 32 years.

Sterilization

Sterilization is used to control microorganisms and pathogens during the growing, collecting, storing and distribution of flowers as well as various foods including grains, vegetables and fruits. Currently, the Vintaging Model assumes that the sterilization sector has not transitioned to any HFC or PFC as an ODS substitute, however, the modeling methodology is provided below for completeness.

The sterilization sector is modeled as a single end-use. For sterilization applications, all chemicals that are used in the equipment in any given year are assumed to be emitted in that year, as shown in the following equation.

$$E_j = Qc_j$$

Where:

E = Emissions. Total emissions of a specific chemical in year j from use in sterilization equipment, by weight.

Qc = Quantity of Chemical. Total quantity of a specific chemical used in sterilization equipment in year j, by weight.

j = Year of emission.

Model Output

By repeating these calculations for each year, the Vintaging Model creates annual profiles of use and emissions for ODS and ODS substitutes. The results can be shown for each year in two ways: 1) on a chemical-by-chemical basis, summed across the end-uses, or 2) on an end-use basis. Values for use and emissions are calculated both in metric tons and in teragrams of CO_2 equivalents (Tg CO_2 Eq.). The conversion of metric tons of chemical to Tg CO_2 Eq. is accomplished through a linear scaling of tonnage by the global warming potential (GWP) of each chemical.

Throughout its development, the Vintaging Model has undergone annual modifications. As new or more accurate information becomes available, the model is adjusted in such a way that both past and future emission estimates are often altered.

3.9. Methodology for Estimating CH_4 Emissions from Enteric Fermentation

Methane emissions from enteric fermentation were estimated for five livestock categories: cattle, horses, sheep, swine, and goats. Emissions from cattle represent the majority of U.S. emissions; consequently, the more detailed IPCC Tier 2 methodology was used to estimate emissions from cattle and the IPCC Tier 1 methodology was used to estimate emissions from the other types of livestock.

Estimate Methane Emissions from Cattle

This section describes the process used to estimate methane emissions from cattle enteric fermentation. A model based on recommendations provided in IPCC/UNEP/OECD/IEA (1997) and IPCC (2000) was developed that uses information on population, energy requirements, digestible energy, and methane conversion rates to estimate methane emissions.[37] The emission methodology consists of the following three steps: (1) characterize the cattle population to account for animal population categories with different emission profiles; (2) characterize cattle diets to generate information needed to estimate emission factors; and (3) estimate emissions using these data and the IPCC Tier 2 equations.

Step 1: Characterize U.S. Cattle Population

Each stage in the cattle lifecycle was modeled to simulate the cattle population from birth to slaughter. This level of detail accounts for the variability in CH_4 emissions associated with each life stage. Given that a stage can last less than one year (e.g., beef calves are weaned at 7 months), each is modeled on a per-month basis. The type of cattle also impacts CH_4 emissions (e.g., beef versus dairy). Consequently, cattle life stages were modeled for several categories of dairy and beef cattle. These categories are listed in Table A- 150.

Table A- 150: Cattle Population Categories Used for Estimating CH_4 Emissions

Dairy Cattle	Beef Cattle
Calves	Calves
Heifer Replacements	Heifer Replacements
Cows	Heifer and Steer Stockers
	Animals in Feedlots (Heifers & Steers)
	Cows
	Bulls

The key variables tracked for each of these cattle population categories (except bulls[38]) are as follows:

- *Calving rates:* The number of animals born on a monthly basis was used to initiate monthly cohorts and to determine population age structure. The number of calves born each month was obtained by multiplying annual births by the percentage of births by month. Annual birth information for each year was taken from USDA (2006a, 2005a, 2004a, 2003a, 2002a, 2001a, 2000a, 1999a, 1995a). Average percentages of births by month for beef from USDA (USDA/APHIS/VS 1998, 1994, 1993) were used for 1990 through 2005. For dairy animals, birth rates were assumed constant throughout the year. To determine whether calves were born to dairy or beef cows, the dairy cow calving rate (USDA/APHIS/VS 2002, USDA/APHIS/VS 1996) was multiplied by the total dairy cow population to determine the number of births attributable to dairy cows, with the remainder assumed to be attributable to beef cows.

- *Average weights and weight gains:* Average weights were tracked for each monthly age group using starting weight and monthly weight gain estimates. Weight gain (i.e., pounds per month) was estimated based on weight gain needed to reach a set target weight, divided by the number of months remaining before target weight was achieved. Birth weight was assumed to be 88 pounds for both beef and dairy animals. Weaning weights were estimated to range from 480 to 575 pounds. Other reported target weights were available for 12,

[37] Additional information on the Cattle Enteric Fermentation Model can be found in ICF (2006)

[38] Only published population statistics and national average emission factors were used to estimate methane emissions from the bull population.

15, 24, and 36 month-old animals. Live slaughter weights were derived from dressed slaughter weight data for each year (USDA 2006c, 2005c, 2004c, 2003c, 2002c, 2001c, 2000c, 1999a, 1995a). Live slaughter weight was estimated as dressed weight divided by 0.63 USDA (1999c). This ratio represents the dressed weight (i.e. weight of the carcass after removal of the internal organs), to the live weight, (i.e. weight taken immediately before slaughter). Table A- 151 provides the target weights used to track average weights of cattle by animal type.

Table A- 151: Target Weights for Use in Estimating Average Weights and Weight Gains (lbs)

Cattle Type	Typical Weights (lbs)
Beef Replacement Heifer Data	
Replacement Weight at 15 months	715
Replacement Weight at 24 months	1,078
Mature Weight at 36 months	1,172
Dairy Replacement Heifer Data	
Replacement Weight at 15 months	800
Replacement Weight at 24 months	1,225
Mature Weight at 36 months	1,350
Stockers Data—Grazing/Forage Based Only	
Steer Weight Gain/Month to 12 months	45
Steer Weight Gain/Month to 24 months	35
Heifer Weight Gain/Month to 12 months	35
Heifer Weight Gain/Month to 24 months	30

Source: Feedstuffs (1998), Western Dairyman (1998), Johnson (1999), NRC (1999).

- *Feedlot placements:* Feedlot placement statistics were available that specify placement of animals from the stocker population into feedlots on a monthly basis by weight class. The model used these data to shift a sufficient number of animals from the stocker cohorts into the feedlot populations to match the reported placement data. After animals are placed in feedlots they progress through two steps. First, animals spend time on a step-up diet to become acclimated to the new feed type. Animals are then switched to a finishing diet for a period of time before they are slaughtered. The length of time an animal spends in a feedlot depends on the start weight (i.e., placement weight), the rate of weight gain during the start-up and finishing phase of diet, and the end weight (as determined by weights at slaughter). Weight gain during start-up diets is estimated to be 2.8 to 3 pounds per day. Weight gain during finishing diets is estimated to be 3 to 3.3 pounds per day (Johnson 1999). All animals are estimated to spend 25 days in the step-up diet phase (Johnson 1999). Length of time finishing was calculated based on start weight, weight gain per day, and target slaughter weight. Table A- 152 provides a summary of the reported feedlot placement statistics for 2005. Since only the total placement data for 1990 to 1995 were available, the placements for each weight category (displayed in Table A- 152) are based on the average of monthly placements from the 1996 to 1998 reported figures.

Table A- 152: Feedlot Placements in the United States for 2005 (Number of animals placed in Thousand Head)

Weight When Placed	Jan	Feb	Mar	Apr	May	Jun	Jul	Aug	Sep	Oct	Nov	Dec	Total
< 600 lbs	380	277	299	340	435	412	400	418	565	880	660	575	**5,641**
600 – 700 lbs	458	308	320	225	390	347	338	385	505	790	650	554	**5,270**
700 – 800 lbs	631	526	609	500	663	480	465	540	550	593	420	430	**6,407**
> 800 lbs	419	412	522	595	735	530	475	650	735	525	315	325	**6,238**
Total	**1,888**	**1,523**	**1,750**	**1,660**	**2,223**	**1,769**	**1,678**	**1,993**	**2,355**	**2,788**	**2,045**	**1,884**	**23,556**

Source: USDA (2006f, 2005f, 2004f, 2003f, 2002f, 2001f, 2000f, 1999a, 1995a).
Note: Totals may not sum due to independent rounding.

- *Pregnancy and lactation:* Energy requirements and hence, composition of diets, level of intake, and emissions for particular animals, are greatly influenced by whether the animal is pregnant or lactating. Information is therefore needed on the percentage of all mature animals that are pregnant each month, as well as milk production, to estimate CH_4 emissions. A weighted average percent of pregnant cows each month was estimated using information on births by month and average pregnancy term. For beef cattle, a weighted average total milk production per animal per month was estimated using information on typical lactation cycles and amounts (NRC 1999), and data on births by month. This process results in a range of weighted monthly lactation estimates expressed as lbs/animal/month. The monthly estimates from January to December are 3.3, 5.1, 8.7, 12.0, 13.6, 13.3, 11.7, 9.3, 6.9, 4.4, 3.0, and 2.8 lbs milk/animal/day. Monthly estimates for dairy cattle

were taken from USDA monthly milk production statistics. Dairy lactation estimates for 1990 through 2005 are shown in Table A- 153.

Table A- 153: Dairy Lactation by Region (lbs· year/cow)*

Year	California	West	Northern Great Plains	Southcentral	Northeast	Midwest	Southeast
1990	18,443	17,293	13,431	13,399	14,557	14,214	12,852
1991	18,522	17,615	13,525	13,216	14,985	14,446	13,053
1992	18,709	18,083	13,998	13,656	15,688	14,999	13,451
1993	18,839	18,253	14,090	14,027	15,602	15,086	13,739
1994	20,190	18,802	14,686	14,395	15,732	15,276	14,111
1995	19,559	18,708	14,807	14,294	16,254	15,680	14,318
1996	19,148	19,076	15,040	14,402	16,271	15,651	14,232
1997	19,815	19,537	15,396	14,330	16,519	16,116	14,517
1998	19,437	19,814	15,919	14,722	16,864	16,676	14,404
1999	20,767	20,477	16,325	14,990	17,246	16,966	14,840
2000	21,116	20,781	17,205	15,363	17,482	17,426	15,176
2001	20,890	20,775	17,242	14,952	17,603	17,217	15,304
2002	21,263	21,073	18,079	15,746	18,001	17,576	15,451
2003	20,979	21,132	18,550	16,507	17,727	18,048	15,113
2004	21,125	21,140	18,746	17,567	17,720	18,176	15,696
2005	21,389	21,724	19,627	18,589	18,446	18,839	16,045

Source: USDA (2006d, 2005d, 2004d, 2003d, 2002d, 2001d, 2000d, 1999a, 1995a).
* Beef lactation data were developed using the methodology described in Step 1.

- *Death rates:* This factor is applied to all heifer and steer cohorts to account for death loss within the model on a monthly basis. The death rates are estimated by determining the death rate that results in model estimates of the end-of-year population for cows that match the published end-of-year population census statistics from the U. S. Department of Agriculture (USDA) National Agricultural Statistics Service (NASS) reports.

- *Number of animals per category each month:* The population of animals per category is calculated based on number of births (or graduates) into the monthly age group minus those animals that die or are slaughtered and those that graduate to the next category (including feedlot placements). These monthly age groups are tracked in the enteric fermentation model to estimate emissions by animal type on a regional basis.

- Cattle population data were taken from U.S. Department of Agriculture (USDA) National Agricultural Statistics Service (NASS) reports. A summary of the annual average populations upon which all livestock-related emissions are based is provided in Table A- 161 of the Manure Management Annex. The USDA publishes monthly, annual, and multi-year livestock population and production estimates. Multi-year reports include revisions to earlier published data. Cattle and calf populations, feedlot placement statistics (e.g., number of animals placed in feedlots by weight class), slaughter numbers, and lactation data were obtained from USDA (2006a, 2005a, 2004a, 2004c, 2003a, 2003c, 2002a, 2002c, 2001a, 2002c, 2000a, 2000c, 1999a, 1995a). Beef calf birth percentages were obtained from the National Animal Health Monitoring System (NAHMS) (USDA/APHIS/VS 1998, 1994, 1993).

Step 2: Characterize U.S. Cattle Population Diets

To support development of digestible energy (DE, the percent of gross energy intake digested by the animal) and CH_4 conversion rate (Y_m, the fraction of gross energy converted to CH_4) values for each of the cattle population categories, data were collected on diets considered representative of different regions. For both grazing animals and animals being fed mixed rations, representative regional diets were estimated using information collected from state livestock specialists and from USDA (1996). The data for each of the diets (e.g., proportions of different feed constituents, such as hay or grains) were used to determine feed chemical composition for use in estimating digestible energy and Y_m for each animal type. Additional detail on the regional diet characterization is provided in EPA (2000).

Digestible energy and Y_m vary by diet and animal type. The IPCC recommends Y_m values of 3.5 to 4.5 percent for feedlot cattle and 5.5 to 6.5 percent for all other cattle. Given the availability of detailed diet information

for different regions and animal types in the United States, digestible energy and Y_m values unique to the United States[39] were developed. Digestible energy and Y_m values were estimated for each cattle population category, for each year in the time series based on physiological modeling, published values, and/or expert opinion.

Digestible energy and Y_m values for dairy cows were estimated using a model (Donovan and Baldwin 1999) that represents physiological processes in the ruminant animals. The three major categories of input required by the model are animal description (e.g., cattle type, mature weight), animal performance (e.g., initial and final weight, age at start of period), and feed characteristics (e.g., chemical composition, habitat, grain or forage USDA (1996)). Data used to simulate ruminant digestion is provided for a particular animal that is then used to represent a group of animals with similar characteristics. The model accounts for differing diets (i.e., grain-based, forage-based, range-based), so that Y_m values for the variable feeding characteristics within the U.S. cattle population can be estimated.

To calculate the digestible energy values for grazing beef cattle, the diet descriptions were used to estimate weighted digestible energy values for a combination of forage only and supplemented diets. Where DE values were not available for specific feed types, total digestible nutrients (TDN) as a percent of dry matter (DM) intake was used as a proxy for DE. For forage diets, two separate regional DE values were used to account for the generally lower forage quality in the western United States. For non-western grazing animals, the forage DE was an average of the seasonal "TDN percent DM" for Grass Pasture diets listed in Appendix Table 1 of the NRC (2000). This average digestible energy for the non-western grazing animals was 64.7 percent. This value was used for all regions except the west. For western grazing animals, the forage digestible energy was calculated as the average "TDN percent DM" for meadow and range diets listed in Appendix Table 1 of the NRC (2000). The calculated DE for western grazing animals was 58.5 percent. The DE values of supplemental diets were estimated for each specific feed component, as shown in Table A- 154, along with the percent of each feed type in each region. Finally, weighted averages were developed for DE values for each region using both the supplemented diet and the forage diet.[40] For beef cows, the DE value was adjusted downward by two percent to reflect the lower digestibility diets of the mature beef cow based on Johnson (2002). The percent of each diet that is assumed to be supplemental and the DE values for each region are shown in Table A- 155 . Y_m values for all grazing beef cattle were set at 6.5 percent based on Johnson (2002).

For feedlot animals, DE and Y_m values for 1996 through 2005 were taken from Johnson (1999). Values for 1990 through 1995 were linearly extrapolated from the 1996 value based on Johnson (1999). In response to peer reviewer comments (Johnson 2000), values for dairy replacement heifers are based on EPA (1993).

Table A- 156 shows the regional DE, the Y_m, and percent of total U.S. cattle population in each region based on 2005 data.

[39] In some cases, the Y_m values used for this analysis extend beyond the range provided by the IPCC. However, EPA believes that these values are representative for the U.S. due to research conducted to characterize the diets of U.S. cattle and assess the Y_m values associated with different animal performance and feed characteristics in the United States.

[40] For example, in California the forage DE of 64.7 was used for 95 percent of the grazing cattle diet and a supplemented diet DE of 65.2 percent was used for five percent of the diet, for a total weighted DE of 64.9 percent.

Table A- 154: DE Values and Representative Regional Diets (Percent of Diet for each Region) for the Supplemental Diet of Grazing Beef Cattle

Feed	Source of TDN (NRC 2000)	Unweighted TDN or DE	California	West	Northern Great Plains	Southcentral	Northeast	Midwest	Southeast
Alfalfa Hay	Table 11-1, feed #4	59.6%	65%	30%	30%	29%	12%	30%	
Barley	Table 11-1, feed #12	86.3%	10%	15%					
Bermuda	Table 11-1, feed #17	48.5%							35%
Bermuda Hay	Table 11-1, feed #17	48.5%				40%			
Corn	Table 11-1, feed #38	88.1%	10%	10%	25%	11%	13%	13%	
Corn Silage	Table 11-1, feed #39	71.2%			25%		20%	20%	
Cotton Seed Meal	Table 11-1, feed #42	74.4%				7%			
Grass Hay	Table 1a, feed #129, 147, 148	53.7%		40%				30%	
Orchard	Table 11-1, feed #61	53.5%							40%
Soybean Meal Supplement	Table 11-1, feed #70	83.1%		5%	5%				5%
Sorghum	Table 11-1, feed #67	81.3%							20%
Soybean Hulls	Table 11-1, feed #69	76.4%						7%	
Timothy Hay	Table 11-1, feed #77	55.5%					50%		
Whole Cotton Seed	Table 11-1, feed #41	89.2%	5%				5%		
Wheat Middlings	Table 1a, feed #433	83.0%			15%	13%			
Wheat	Table 11-1, feed #83	87.2%	10%						
Weighted Total			**69%**	**65%**	**74%**	**62%**	**65%**	**65%**	**59%**

Source of representative regional diets: Donovan (1999).

Table A- 155 : Percent of Each Diet that is Supplemental, and the Resulting DE Values for Each Region

Region[a]	Percent Supplement	Percent Forage	Calculated Weighted Average DE
West	10	90	59
Northeast	15	85	65
Southcentral	10	90	64
Midwest	15	85	65
Northern Great Plains	15	85	66
Southeast	5	95	64
California	5	95	65

Source of percent of total diet that is supplemental diet: Donovan (1999).

[a] The Western region includes AK, WA, OR, ID, NV, UT, AZ, HI, and NM; the Northeastern region includes PA, NY, MD, DE, NJ, CT, RI, MA, VT, NH, ME, and WV; the Southcentral region includes AR, LA, OK, and TX; the Midwestern region includes MO, IL, IN, OH, MN, WI, MI, and IA; the Northern Great Plains include MT, WY, ND, SD, NE, KS, and CO; and the Southeastern region includes VA, NC, KY, TN, MS, AL, GA, SC, and FL.

Table A-156: Regional Digestible Energy (DE), CH_4 Conversion Rates (Y_m), and Population Percents for Cattle in 2005

Animal Type	Data	California	West	Northern Great Plains	Southcentral	Northeast	Midwest	Southeast
Beef Repl. Heif.	DE[a]	65%	59%	66%	64%	65%	65%	64%
	Y_m[b]	6.5%	6.5%	6.5%	6.5%	6.5%	6.5%	6.5%
	Pop.[c]	2%	9%	29%	25%	2%	14%	18%
Dairy Repl. Heif.	DE	66%	66%	66%	64%	68%	66%	66%
	Y_m	5.9%	5.9%	5.6%	6.4%	6.3%	5.6%	6.9%
	Pop.	18%	14%	4%	4%	18%	35%	6%
Steer Stockers	DE	65%	59%	66%	64%	65%	65%	64%
	Y_m	6.5%	6.5%	6.5%	6.5%	6.5%	6.5%	6.5%
	Pop.	4%	8%	39%	24%	2%	18%	4%
Heifer Stockers	DE	65%	59%	66%	64%	65%	65%	64%
	Y_m	6.5%	6.5%	6.5%	6.5%	6.5%	6.5%	6.5%
	Pop.	2%	7%	49%	22%	1%	15%	4%
Steer Feedlot	DE	85%	85%	85%	85%	85%	85%	85%
	Y_m	3.0%	3.0%	3.0%	3.0%	3.0%	3.0%	3.0%
	Pop.	4%	8%	48%	22%	1%	16%	0%
Heifer Feedlot	DE	85%	85%	85%	85%	85%	85%	85%
	Y_m	3.0%	3.0%	3.0%	3.0%	3.0%	3.0%	3.0%
	Pop.	4%	8%	48%	22%	1%	16%	0%
Beef Cows	DE	63%	57%	64%	62%	63%	63%	62%
	Y_m	6.5%	6.5%	6.5%	6.5%	6.5%	6.5%	6.5%
	Pop.	2%	8%	27%	27%	2%	15%	19%
Dairy Cows	DE	69%	66%	69%	68%	69%	69%	68%
	Y_m	4.8%	5.8%	5.8%	5.7%	5.8%	5.8%	5.6%
	Pop.	19%	15%	5%	5%	17%	31%	7%
Steer Step-Up	DE	59%	60%	54%	51%	51%	52%	44%
	Y_m	74%	74%	74%	74%	74%	74%	74%
Heifer Step-Up	DE	74%	74%	74%	74%	74%	74%	74%
	Y_m	4.8%	4.8%	4.8%	4.8%	4.8%	4.8%	4.8%

[a] Digestible Energy in units of percent GE (MJ/Day).
[b] Methane Conversion Rate is the fraction of GE in feed converted to methane.
[c] Percent of each subcategory population present in each region.

Step 3: Estimate CH_4 Emissions from Cattle

Emissions were estimated in three steps: a) determine gross energy (GE) intake using the IPCC (2000) equations, b) determine an emissions factor using the GE values and other factors, and c) sum the daily emissions for each animal type. The necessary data values include:

- Body Weight (kg)
- Weight Gain (kg/day)
- Net Energy for Activity (C_a, MJ/day)[41]
- Standard Reference Weight[42] (Dairy = 1,324 lbs; Beef = 1,195 lbs)
- Milk Production (kg/day)
- Milk Fat (percent of fat in milk = 4)
- Pregnancy (percent of population that is pregnant)
- DE (percent of gross energy intake digestible)
- Y_m (the fraction of gross energy converted to CH_4)

41 Zero for feedlot conditions, 0.17 for high quality confined pasture conditions, 0.36 for extensive open range or hilly terrain grazing conditions. C_a factor for dairy cows is weighted to account for the fraction of the population in the region that grazes during the year.

42 Standard Reference Weight is used in the model to account for breed potential.

Step 3a: Gross Energy, GE

As shown in the following equation, gross energy (GE) is derived based on the net energy estimates and the feed characteristics. Only variables relevant to each animal category are used (e.g., estimates for feedlot animals do not require the NE_l factor). All net energy equations are provided in IPCC (2000).

$$GE = [((NE_m + NE_{mobilized} + NE_a + NE_l + NE_p) / \{NE_{ma}/DE\}) + (NE_g / \{NE_{ga}/DE\})] / (DE / 100)$$

Where,

GE	= Gross energy (MJ/day)
NE_m	= Net energy required by the animal for maintenance (MJ/day)
$NE_{mobilized}$	= Net energy due to weight loss (mobilized) (MJ/day)
NE_a	= Net energy for animal activity (MJ/day)
NE_l	= Net energy for lactation (MJ/day)
NE_p	= Net energy required for pregnancy (MJ/day)
$\{NE_{ma}/DE\}$	= Ratio of net energy available in a diet for maintenance to digestible energy consumed
NE_g	= Net energy needed for growth (MJ/day)
$\{NE_{ga}/DE\}$	= Ratio of net energy available for growth in a diet to digestible energy consumed
DE	= Digestible energy expressed as a percentage of gross energy (percent)

Step 3b: Emission Factor

The emission factor (DayEmit) was determined using the gross energy value and the methane conversion factor (Y_m) for each category. This relationship is shown in the following equation:

$$DayEmit = [GE \times Y_m] / [55.65 \text{ MJ/kg } CH_4]$$

Where,

DayEmit	= Emission factor (kg CH_4/head/day)
GE	= Gross energy intake (MJ/head/day)
Y_m	= CH_4 conversion rate, which is the fraction of gross energy in feed converted to CH_4 (percent)

The daily emission factors were estimated for each animal type, weight, and region.

Step 3c: Estimate Total Emissions

Emissions were summed for each month and for each population category using the daily emission factor for a representative animal and the number of animals in the category. The following equation was used:

$$Emissions = DayEmit \times Days/Month \times SubPop$$

Where,

DayEmit	= The emission factor for the subcategory (kg CH_4/head/day)
Days/Month	= The number of days in the month
SubPop	= The number of animals in the subcategory during the month

This process was repeated for each month, and the totals for each subcategory were summed to achieve an emission estimate for the entire year. The estimates for each of the 10 subcategories of cattle are listed in Table A-157. The emissions for each subcategory were then summed to estimate total emissions from beef cattle and dairy cattle for the entire year.

Table A- 157: CH_4 Emissions from Cattle (Gg)

Cattle Type	1990	1995	2000	2001	2002	2003	2004	2005
Dairy	**1,375**	**1,320**	**1,283**	**1,280**	**1,288**	**1,299**	**1,285**	**1,319**
Cows	1,142	1,088	1,058	1,053	1,060	1,070	1,058	1,086
Replacements 7-11 months	49	49	48	48	49	48	48	50
Replacements 12-23 months	184	183	177	179	179	181	178	182
Beef	**3,859**	**4,160**	**3,869**	**3,825**	**3,821**	**3,832**	**3,730**	**3,772**
Cows	2,428	2,628	2,507	2,492	2,472	2,461	2,453	2,457
Replacements 7-11 months	52	61	53	54	54	53	54	56
Replacements 12-23 months	190	232	198	200	200	201	198	204
Steer Stockers	431	480	362	352	355	361	325	337
Heifer Stockers	232	250	207	203	205	210	188	199
Feedlot Cattle	412	382	426	408	421	429	399	406
Bulls	116	127	116	116	115	115	113	114
Total	**5,234**	**5,480**	**5,153**	**5,105**	**5,110**	**5,131**	**5,014**	**5,091**

Note: Totals may not sum due to independent rounding.

Emission Estimates from Other Livestock

All livestock population data, except for horses, were taken from U.S. Department of Agriculture (USDA) National Agricultural Statistics Service (NASS) reports. Table A- 161 of the Manure Management Annex shows the population data for all livestock species that were used for estimating all livestock-related emissions. For each animal category, the USDA publishes monthly, annual, and multi-year livestock population and production estimates. Multi-year reports include revisions to earlier published data. Recent reports were obtained from the USDA Economics and Statistics System, while historical data were downloaded from the USDA-NASS. The Food and Agriculture Organization (FAO) publishes horse population data. These data were accessed from the FAOSTAT database (FAO 2006). Methane emissions from sheep, goats, swine, and horses were estimated by multiplying published national population estimates by the IPCC emission factor for each year. Table A- 158 shows the emission factors used for these other livestock.

A complete time series of enteric fermentation emissions from all livestock types is shown in Table A- 159 (Tg CO_2 Eq.) and Table A- 160 (Gg).

Table A- 158: Emission Factors for Other Livestock (kg CH_4/head/year)

Livestock Type	Emission Factor
Sheep	8
Goats	5
Horses	18
Swine	1.5

Source: IPCC (2000).

Table A- 159: CH_4 Emissions from Enteric Fermentation (Tg CO_2 Eq.)

Livestock Type	1990	1995	2000	2001	2002	2003	2004	2005
Beef Cattle	81.0	87.4	81.3	80.3	80.2	80.5	78.3	79.2
Dairy Cattle	28.9	27.7	27.0	26.9	27.1	27.3	27.0	27.7
Horses	1.9	1.9	2.0	2.0	2.0	2.0	2.0	2.0
Sheep	1.9	1.5	1.2	1.2	1.1	1.1	1.0	1.0
Swine	1.7	1.9	1.9	1.9	1.9	1.9	1.9	1.9
Goats	0.3	0.2	0.3	0.3	0.3	0.3	0.3	0.3
Total	**115.7**	**120.6**	**113.5**	**112.5**	**112.6**	**113.0**	**110.5**	**112.1**

Table A- 160: CH_4 Emissions from Enteric Fermentation (Gg)

Livestock Type	1990	1995	2000	2001	2002	2003	2004	2005
Beef Cattle	3,859	4,160	3,869	3,825	3,821	3,832	3,730	3,772
Dairy Cattle	1,375	1,320	1,283	1,280	1,288	1,299	1,285	1,319
Horses	91	92	94	95	95	95	95	95
Sheep	91	72	56	55	53	51	49	49
Swine	81	88	88	88	90	90	91	91
Goats	13	12	12	12	13	13	13	13
Total	**5,510**	**5,744**	**5,404**	**5,356**	**5,361**	**5,379**	**5,262**	**5,340**

3.10. Methodology for Estimating CH_4 and N_2O Emissions from Manure Management

The following steps were used to estimate CH_4 and N_2O emissions from the management of livestock manure. Nitrous oxide emissions associated with pasture, range, or paddock systems and daily spread systems, as well as indirect N_2O emissions from manure management, are included in the emission estimates for Agricultural Soil Management (see sub-annex 3.11).

Step 1: Livestock Population Characterization Data

Annual animal population data for 1990 through 2005 for all livestock types, except horses and goats, were obtained from the USDA National Agricultural Statistics Service (USDA 1994a-b, 1995a-b, 1998a-b, 1999a-c, 2000a, 2004a-e, 2005a-d, 2006a-e). The population data used in the emissions calculations for cattle and swine were downloaded from the USDA National Agricultural Statistics Service Population Estimates Data Base <http://www.usda.gov/nass/>. Horse population data were obtained from the FAOSTAT database (FAO 2006). Goat population data for 1992, 1997, and 2002 were obtained from the Census of Agriculture (USDA 2005e). Additional data sources used and adjustments to these data sets are described below.

Beef Cattle: Additional information regarding the percent of beef steers and heifers in feedlots was obtained from contacts with the national USDA office (Milton 2000). Data for beef feedlots were also obtained from EPA's Office of Water (ERG 2000a, EPA 2002a, 2002b).

For all beef cattle groups (cows, heifers, steers, bulls, and calves), the USDA data provide cattle inventories from January (for each state) and July (as a U.S. total only) of each year. Cattle inventories change over the course of the year, sometimes significantly, as new calves are born and as fattened cattle are slaughtered; therefore, to develop the best estimate for the annual animal population, the average inventory of cattle by state was calculated. In order to estimate average annual populations by state, a "scaling factor" was developed that adjusts the January state-level data to reflect July inventory changes. This factor equals the average of the U.S. January and July data divided by the January data. The scaling factor is derived for each cattle group and is then applied to the January state-level data to arrive at the state-level annual population estimates.

Swine: The USDA data provides quarterly data for each swine subcategory: breeding, market under 60 pounds (less than 27 kg), market 60 to 119 pounds (27 to 54 kg), market 120 to 179 pounds (54 to 81 kg), and market 180 pounds and over (greater than 82 kg). The average of the quarterly data was used in the emissions calculations. For states where only December inventory is reported, the December data were used directly.

Sheep: Population data for lamb and sheep on feed are not available after 1993. The number of lamb and sheep on feed for 1994 through 2005 were calculated using the average of the percent of lamb and sheep on feed from 1990 through 1993. In addition, all of the sheep and lamb "on feed" are not necessarily on "feedlots;" they may be on pasture/crop residue supplemented by feed. Data for those animals on feed that are in feedlots versus pasture/crop residue were provided only for lamb in 1993. To calculate the populations of sheep and lamb in feedlots for all years, it was assumed that the percentage of sheep and lamb on feed that are in feedlots versus pasture/crop residue is the same as that for lambs in 1993 (Anderson 2000).

Goats: Annual goat population data by state were available for only 1992, 1997, and 2002 (USDA 2005e). The data for 1992 were used for 1990 through 1992 and the data for 2002 were used for 2002 through 2005. Data for 1993 through 1996 were estimated based on the 1992 and 1997 Census data. Data for 1998 through 2001 were extrapolated using the 1997 and 2002 Census data.

Poultry: Recently, USDA changed the classification and reporting of poultry populations of hens versus pullets, so the ratio of hens to pullets has changed for all years of the inventory. However, the total number of hens and pullets combined is the same. The annual population data for boilers and turkeys were adjusted for turnover (i.e., slaughter) rate (Lange 2000). All poultry population data were adjusted to account for states that report non-disclosed populations to USDA National Agricultural Statistics Service. The combined populations of the states reporting non-disclosed populations are reported as "other" states. State populations were estimated for the non-disclosed states by equally distributing the population attributed to "other" states to the non-disclosed states.

Horses: The Food and Agriculture Organization (FAO) publishes annual total U.S. horse population, which were accessed from the FAOSTAT database at < http://faostat.fao.org> (FAO 2006). State horse population

data were estimated using state population distributions from the 1992, 1997, and 2002 Census of Agriculture and the FAO national population data.

A summary of the livestock population characterization data used to calculate CH_4 and N_2O emissions is presented in Table A- 161.

Step 2: Waste Characteristics Data

Methane and N_2O emissions calculations are based on the following animal characteristics for each relevant livestock population:

- Volatile solids excretion rate (VS)
- Maximum methane producing capacity (B_o) for U.S. animal waste
- Nitrogen excretion rate (Nex)
- Typical animal mass (TAM)

Table A-162 presents a summary of the waste characteristics used in the emissions estimates. Published sources were reviewed for U.S.-specific livestock waste characterization data that would be consistent with the animal population data discussed in Step 1. The USDA's Agricultural Waste Management Field Handbook (USDA 1996a) is one of the primary sources of waste characteristics. In some cases, data from the American Society of Agricultural Engineers, Standard D384.1 (ASAE 1999) were used to supplement the USDA data. The dairy cow population is assumed to be comprised of both lactating and dry cows, 17 percent of a dairy herd is assumed to be dry and 83 percent is assumed to be lactating. Nex rates were collected from the sources indicated in Table A-162 and are based on measurement data from excreted manure. The VS and Nex data for breeding swine are from a combination of the types of animals that make up this animal group, namely gestating and farrowing swine and boars. It is assumed that a group of breeding swine is typically broken out as 80 percent gestating sows, 15 percent farrowing swine, and 5 percent boars (Safley 2000). Due to the change in USDA reporting of hens and pullets, new nitrogen and volatile solids excretion rates were calculated for the combined population of hens and pullets. A weighted average rate was calculated based on hen and pullet population data from 1990 to 2004 and the 2004 inventory VS and Nex values of 10.8 (hens) and 9.7 (pullets) kg VS/day per 1,000 kg mass, and 0.83 (hens) and 0.62 (pullets) kg Total Kjeldahl Nitrogen/day per 1,000 kg mass. The new hen/pullet VS and Nex values were applied to hens and pullets for all years of the inventory.

The method for calculating VS production from beef and dairy cows, heifers, and steers is based on the relationship between animal diet and energy utilization, which is modeled in the enteric fermentation portion of the inventory. VS content of manure equals the fraction of the diet consumed by cattle that is not digested and thus excreted as fecal material which, when combined with urinary excretions, constitutes manure. The enteric fermentation model requires the estimation of gross energy intake and its fractional digestibility in the process of estimating enteric methane emissions (see Annex 3.9 for details on the enteric energy model). These two inputs are used to calculate the indigestible energy per animal unit as gross energy minus digestible energy plus an additional 2 percent of gross energy for urinary energy excretion per animal unit. This value is then converted to VS production per animal unit using the typical conversion of dietary gross energy to dry organic matter of 20.1 MJ/kg (Garrett and Johnson 1983). The equation used for calculating volatile solids is as follows:

$$\text{VS production (kg)} = [GE - DE + (0.02 \times GE)] / 20.1 \text{ (MJ/kg)}$$

Where,

GE = Gross energy intake (MJ)
DE = Digestible energy (MJ)

This equation is used to calculate volatile solids rates for each region, cattle type, and year. State-specific volatile solids excretion rates are based on the calculated regional rates (Pederson and Pape 2006).

Table A- 163 presents the state-specific volatile solids production rates used for 2005.

Step 3: Waste Management System Usage Data

Table A- 164 through Table A- 169 summarize 2005 manure distribution data among waste management systems at beef feedlots, dairies, dairy heifer facilities, and swine, layer, broiler, and turkey operations. Manure from beef cattle not on feed, sheep, horses, and goats is managed on pasture, range, or paddocks, on drylot, or with solids storage systems. Additional information on the development of the manure distribution estimates for each animal type is presented below.

Beef Cattle and Dairy Heifers: The beef feedlot and dairy heifer waste management system data were developed using information from EPA's Office of Water's engineering cost analyses conducted to support the development of effluent limitations guidelines for Concentrated Animal Feeding Operations (EPA 2002b). Based on EPA site visits and state contacts supporting this work and contacts with the national USDA office to estimate the percent of beef steers and heifers in feedlots (Milton 2000), feedlot manure is almost exclusively managed in drylots. Therefore, for these animal groups, the percent of manure deposited in drylots is assumed to be 100 percent. In addition, there is a small amount of manure contained in runoff, which may or may not be collected in runoff ponds. The runoff from feedlots was calculated by region in *Calculations: Percent Distribution of Manure for Waste Management Systems* (ERG 2000b) and was used to estimate the percentage of manure managed in runoff ponds in addition to drylots; this percentage ranges from 0.4 to 1.3 percent. The percentage of manure generating emissions from beef feedlots is therefore greater than 100 percent. The remaining population categories of beef cattle outside of feedlots are managed through pasture/range/paddock systems, which are utilized for the majority of the population of beef cattle in the country.

Dairy Cows: The waste management system data for dairy cows were developed using data from the Census of Agriculture, EPA's Office of Water, USDA, and expert sources. Farm-size distribution data are reported in the 1992, 1997, and 2002 Census of Agriculture (USDA 2005e). It was assumed that the data provided for 1992 were the same as that for 1990 and 1991, and data provided for 2002 were the same as that for 2003 through 2005. Data for 1993 through 1996 and 1998 through 2001 were extrapolated using the 1992, 1997, and 2002 data. The percent of waste by system was estimated using the USDA data broken out by geographic region and farm size.

Based on EPA site visits and state contacts, manure from dairy cows at medium (200 through 700 head) and large (greater than 700 head) operations are managed using either flush systems or scrape/slurry systems. In addition, they may have a solids separator in place prior to their storage component. Estimates of the percent of farms that use each type of system (by geographic region) were developed by EPA's Office of Water, and were used to estimate the percent of waste managed in lagoons (flush systems), liquid/slurry systems (scrape systems), and solid storage (separated solids) (EPA 2002b). Manure management system data for small (fewer than 200 head) dairies were obtained from USDA (2000b). These operations are more likely to use liquid/slurry and solid storage management systems than anaerobic lagoon systems. The reported manure management systems were deep pit, liquid/slurry (also includes slurry tank, slurry earth-basin, and aerated lagoon), anaerobic lagoon, and solid storage (also includes manure pack, outside storage, and inside storage).

Data regarding the use of daily spread and pasture, range, or paddock systems for dairy cattle were obtained from personal communications with personnel from several organizations. These organizations include state NRCS offices, state extension services, state universities, USDA National Agricultural Statistics Service (NASS), and other experts (Deal 2000, Johnson 2000, Miller 2000, Stettler 2000, Sweeten 2000, and Wright 2000). Contacts at Cornell University provided survey data on dairy manure management practices in New York (Poe et al. 1999). Census of Agriculture population data for 1992, 1997, and 2002 (USDA 2005e) were used in conjunction with the state data obtained from personal communications to determine regional percentages of total dairy cattle and dairy waste that are managed using these systems. These percentages were applied to the total annual dairy cow and heifer state population data for 1990 through 2005, which were obtained from the National Agricultural Statistics Service (USDA 1995a, 1999a, 2004a, 2005a, 2006a-b).

Of the dairies using systems other than daily spread and pasture, range, or paddock systems, some dairies reported using more than one type of manure management system. Due to limitations in how USDA reports the manure management data, the total percent of systems for a region and farm size is greater than 100 percent. However, manure is typically partitioned to use only one manure management system, rather than transferred between several different systems. Emissions estimates are only calculated for the final manure management system used for each portion of manure. To avoid double counting emissions, the reported percentages of systems in use were adjusted to equal a total of 100 percent, using the same distribution of systems. For example, if USDA

reported that 65 percent of dairies use deep pits to manage manure and 55 percent of dairies use anaerobic lagoons to manage manure, it was assumed that 54 percent (i.e., 65 percent divided by 120 percent) of the manure is managed with deep pits and 46 percent (i.e., 55 percent divided by 120 percent) of the manure is managed with anaerobic lagoons (ERG 2000a).

Swine: Based on data collected during site visits for EPA's Office of Water (ERG 2000a), manure from swine at large (greater than 2000 head) and medium (200 through 2000 head) operations are primarily managed using deep pit systems, liquid/slurry systems, or anaerobic lagoons. Manure management system data were obtained from USDA (USDA 1998c). It was assumed those operations with less than 200 head use pasture, range, or paddock systems. The percent of waste by system was estimated using the USDA data broken out by geographic region and farm size. Farm-size distribution data reported in the 1992, 1997, and 2002 Census of Agriculture (USDA 2005e) were used to determine the percentage of all swine utilizing the various manure management systems. It was assumed that the swine farm size data provided for 1992 were the same as that for 1990 and 1991, and data provided for 2002 were the same as that for 2003 through 2005. Data for 1993 through 1996 and 1998 through 2001 were extrapolated using the 1992, 1997, and 2002 data. The reported manure management systems were deep pit, liquid/slurry (also includes above- and below-ground slurry), anaerobic lagoon, and solid storage (also includes solids separated from liquids).

Some swine operations reported using more than one management system; therefore, the total percent of systems reported by USDA for a region and farm size is greater than 100 percent. Typically, this means that some of the manure at a swine operation is handled in one system (e.g., liquid system), and some of the manure is handled in another system (e.g., dry system). However, it is unlikely that the same manure is moved from one system to another.

Sheep: It was assumed that all sheep waste not deposited in feedlots was deposited on pasture, range, or paddock lands (Anderson 2000).

Goats/Horses: Waste management system data for 1990 to 2005 were obtained from Appendix H of *Global Methane Emissions from Livestock and Poultry Manure* (EPA 1992). It was assumed that all manure not deposited in pasture, range, or paddock lands were managed in dry systems.

Poultry—Layers: Waste management system data for 1992 were obtained from *Global Methane Emissions from Livestock and Poultry Manure* (EPA 1992). These data were also used to represent 1990 and 1991. The percentage of layer operations using a shallow pit flush house with anaerobic lagoon or high-rise house without bedding was obtained for 1999 from a United Egg Producers voluntary survey (UEP 1999). These data were augmented for key poultry states (AL, AR, CA, FL, GA, IA, IN, MN, MO, NC, NE, OH, PA, TX, and WA) with USDA data (USDA 2000c). It was assumed that the change in system usage between 1990 and 1999 is proportionally distributed among those years of the inventory. It was assumed that system usage in 2000 through 2005 was equal to that estimated for 1999. Data collected for EPA's Office of Water, including site visits (EPA 2002b), were used to estimate the distribution of waste by management system and animal type.

Poultry—Broilers/Turkeys: The percentage of turkeys and broilers on pasture was obtained from *Global Methane Emissions from Livestock and Poultry Manure* (EPA 1992). It was assumed that one percent of poultry waste is deposited in pastures, ranges, and paddocks (EPA 1992). The remainder of waste is assumed to be deposited in operations with bedding management.

Step 4: Emission Factor Calculations

Methane conversion factors (MCFs) and N_2O emission factors (EFs) used in the emission calculations were determined using the methodologies presented below:

Methane Conversion Factors (MCFs)

Climate-based IPCC default MCFs (IPCC 2006) were used for all dry systems, while a country-specific methodology was used to develop MCFs for all lagoon and liquid systems.

Based on annual average temperature data, each state and year in the inventory was assigned a climate classification of cool, temperate or warm, and the appropriate IPCC default MCF was applied for animal waste managed in dry systems.

A climate-based approach based on the van't Hoff-Arrhenius equation was developed to estimate MCFs for anaerobic lagoons and other liquid systems that reflects the seasonal changes in temperatures, and also accounts for long-term retention time. This approach is consistent with the recently revised guidelines from IPCC (IPCC 2006). The van't Hoff-Arrhenius equation, with a base temperature of 30°C, is shown in the following equation (Safley and Westerman 1990):

$$f = \exp\left[\frac{E(T_2 - T_1)}{RT_1T_2}\right]$$

Where,

T_1 = 303.15K
T_2 = Ambient temperature (K) for climate zone (in this case, a weighted value for each state)
E = Activation energy constant (15,175 cal/mol)
R = Ideal gas constant (1.987 cal/K mol)

The factor *f* represents the proportion of volatile solids that are biologically available for conversion to methane based on the temperature of the system. For those animal populations using liquid manure management systems or manure runoff ponds (i.e., dairy cow, dairy heifer, layers, beef in feedlots, and swine) monthly average state temperatures were based on the counties where the specific animal population resides (i.e., the temperatures were weighted based on the percent of animals located in each county). The average county and state temperature data were obtained from the National Climate Data Center (NOAA 2006). County population data were calculated from state-level population data from NASS and county-state distribution data from the 1992, 1997, and 2002 Census data (USDA 2005e). County population distribution data for 1990 and 1991 were assumed to be the same as 1992; county population distribution data for 1993 through 1996 were extrapolated based on 1992 and 1997 data; county population data for 1998 through 2001 were extrapolated based on 1997 and 2002 data; and county population data for 2003 to 2005 were assumed to be the same as 2002.

Annual MCFs for liquid systems are calculated as follows for each animal type, state, and year of the inventory:

- The weighted-average temperature for a state is calculated using the population estimates and average monthly temperature in each county.
- Monthly temperatures are used to calculate a monthly van't Hoff-Arrhenius "f" factor, using the equation presented above. A minimum temperature of 5°C is used for uncovered anaerobic lagoons and 7.5°C is used for liquid/slurry and deep pit systems.
- Monthly production of volatile solids that are added to the system is estimated based on the number of animals present and, for lagoon systems, adjusted for a management and design practices factor. This factor accounts for other mechanisms by which volatile solids are removed from the management system prior to conversion to methane, such as solids being removed from the system for application to cropland. This factor, equal to 0.8, has been estimated using currently available methane measurement data from anaerobic lagoon systems in the United States (ERG 2001).
- The amount of volatile solids available for conversion to methane is assumed to be equal to the amount of volatile solids produced during the month (from Step 3). For anaerobic lagoons, the amount of volatile solids available also includes volatile solids that may remain in the system from previous months.
- The amount of volatile solids consumed during the month is equal to the amount available for conversion multiplied by the "f" factor.
- For anaerobic lagoons, the amount of volatile solids carried over from one month to the next is equal to the amount available for conversion minus the amount consumed. Lagoons are also modeled to have a solids clean-out once per year, occurring after the month of September.
- The estimated amount of CH_4 generated during the month is equal to the monthly volatile solids consumed multiplied by the maximum methane potential of the waste (B_o).

The annual MCF is then calculated as:

$$\text{MCF (annual)} = CH_4 \text{ generated (annual)} / (\text{VS produced (annual)} \times B_o)$$

Where,

$\text{MCF}_{\text{(annual)}}$	= Methane conversion factor
$\text{VS produced}_{\text{(annual)}}$	= Volatile solids excreted annually
B_o	= Maximum CH_4 producing potential of the waste

In order to account for the carry-over of volatile solids from the year prior to the inventory year for which estimates are calculated, it is assumed in the MCF calculation for lagoons that a portion of the volatile solids from October, November, and December of the year prior to the inventory year are available in the lagoon system starting January of the inventory year. It is assumed that liquid-slurry systems have a retention time less than 30 days, so the liquid-slurry MCF calculation doesn't reflect the volatile solids carry-over.

Following this procedure, the resulting MCF for lagoons accounts for temperature variation throughout the year, residual volatile solids in a system (carry-over), and management and design practices that may reduce the volatile solids available for conversion to CH_4. The MCFs presented in Table A- 170 by state and waste management system represent the average MCF for 2005 by state for all animal groups located in that state.

Nitrous Oxide Emission Factors

Nitrous oxide emission factors for manure management systems were set equal to the most recent default IPCC factors (IPCC 2006).

Step 5: Weighted Emission Factors

For beef cattle, dairy cattle, swine, and poultry, the emission factors for both CH_4 and N_2O were weighted to incorporate the distribution of waste by management system for each state. The following equation was used to determine the weighted MCF for a particular animal type in a particular state:

$$MCF_{animal,\,state} = \sum_{system} (MCF_{system,\,state} \times \%Manure_{animal,\,system,\,state})$$

Where,

$\text{MCF}_{\text{animal, state}}$	= Weighted MCF for that animal group and state
$\text{MCF}_{\text{system, state}}$	= MCF for that system and state (see Step 4)
$\%\ \text{Manure}_{\text{animal, system, state}}$	= Percent of manure managed in the system for that animal group in that state (expressed as a decimal)

The weighted nitrous oxide emission factor for a particular animal type in a particular state was determined in the same manner:

$$EF_{animal,\,state} = \sum_{system} (EF_{system} \times \%\ Manure_{animal,\,system,\,state})$$

Where,

$\text{EF}_{\text{animal, state}}$	= Weighted emission factor for that animal group and state
$\text{EF}_{\text{system}}$	= Emission factor for that system (see Step 4)
$\%\ \text{Manure}_{\text{animal, system, state}}$	= Percent of manure managed in the system for that animal group in that state (expressed as a decimal)

A summary of the weighted MCFs used to calculate beef feedlot, dairy cow and heifer, swine, and poultry emissions for 2005 are presented Table A- 171. For certain animal groups (beef cattle not on feed, horses, sheep,

and goats), the emission factors do not vary for the management systems used. In these cases, a weighted emission factor was not necessary. A summary of the weighted N_2O emission factors is presented in Table A- 172.

Step 6: CH_4 and N_2O Emission Calculations

Methane emissions were calculated for each animal group as follows:

$$Methane_{\,animal\,group} = \sum_{state} (Population \times VS \times B_o \times MCF_{animal,\,state} \times 0.662)$$

Where,

$Methane_{animal\ group}$ = CH_4 emissions for that animal group (kg CH_4/yr)
Population = Annual average state animal population for that animal group (head)
VS = Total volatile solids produced annually per animal (kg/yr/head)
B_o = Maximum CH_4 producing capacity per kilogram of VS (m^3 CH_4/kg VS)
$MCF_{animal,\ state}$ = Weighted MCF for the animal group and state (see Step 5)
0.662 = Conversion factor of m^3 CH_4 to kilograms CH_4 (kg CH_4 /m^3 CH_4)

Nitrous oxide emissions were calculated for each animal group as follows:

$$\text{Nitrous Oxide}_{\,\text{animal group}} = \sum_{state} (Population \times N_{ex} \times EF_{animal,\,state} \times 44/28)$$

Where,

Nitrous $Oxide_{animal\ group}$ = N_2O emissions for that animal group (kg/yr)
Population = Annual average state animal population for that animal group (head)
N_{ex} = Total Kjeldahl nitrogen excreted annually per animal (kg/yr/head)
$EF_{animal,\ state}$ = Weighted N_2O emission factor for the animal group and state, kg N_2O-N/kg N excreted (see Step 5)
44/28 = Conversion factor of N_2O-N to N_2O

Emission estimates are summarized in Table A- 173 and Table A- 174.

Table A- 161: Livestock Population (1,000 Head)

Animal Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Dairy Cattle	**14,143**	**13,980**	**13,830**	**13,767**	**13,566**	**13,502**	**13,305**	**13,138**	**12,992**	**13,023**	**13,066**	**12,964**	**13,005**	**12,978**	**12,805**	**12,937**
Dairy Cows	10,007	9,883	9,714	9,679	9,504	9,491	9,410	9,309	9,200	9,139	9,216	9,136	9,128	9,121	8,995	9,027
Dairy Heifer	4,135	4,097	4,116	4,088	4,062	4,011	3,895	3,829	3,793	3,884	3,850	3,828	3,877	3,857	3,810	3,909
Swine	**53,941**	**56,478**	**58,532**	**58,016**	**59,951**	**58,899**	**56,220**	**58,728**	**61,989**	**60,238**	**58,864**	**58,913**	**60,028**	**59,817**	**60,717**	**60,946**
Market <60 lbs.	18,359	19,212	19,851	19,434	20,157	19,656	18,851	19,886	20,692	19,928	19,574	19,659	19,863	19,929	20,216	20,175
Market 60-119 lbs.	11,734	12,374	12,839	12,656	13,017	12,836	12,157	12,754	13,551	13,255	12,926	12,900	13,284	13,138	13,400	13,509
Market 120-179 lbs.	9,440	9,840	10,253	10,334	10,671	10,545	10,110	10,480	11,234	11,041	10,748	10,708	11,013	11,046	11,228	11,333
Market >180 lbs.	7,510	7,822	8,333	8,435	8,824	8,937	8,463	8,768	9,671	9,641	9,385	9,465	9,738	9,701	9,921	9,950
Breeding	6,899	7,231	7,255	7,157	7,282	6,926	6,639	6,840	6,841	6,373	6,231	6,181	6,129	6,004	5,952	5,978
Beef Cattle	**86,087**	**87,267**	**88,548**	**90,321**	**92,571**	**94,391**	**94,269**	**92,290**	**90,730**	**90,034**	**89,220**	**88,621**	**87,927**	**87,040**	**86,349**	**87,032**
Feedlot Steers	7,338	7,920	7,581	7,984	7,797	7,763	7,380	7,644	7,845	7,805	8,338	8,622	8,423	7,944	8,174	8,289
Feedlot Heifers	3,621	4,035	3,626	3,971	3,965	4,047	3,999	4,396	4,459	4,587	4,899	5,066	4,852	4,571	4,633	4,584
NOF Bulls[2]	2,180	2,198	2,220	2,239	2,306	2,392	2,392	2,325	2,235	2,241	2,197	2,187	2,172	2,174	2,128	2,160
NOF Calves[2]	23,909	23,853	24,118	24,209	24,586	25,170	25,042	24,363	24,001	23,895	23,508	22,958	22,577	22,273	22,005	22,125
NOF Heifers[2]	8,872	8,938	9,520	9,850	10,469	10,680	10,869	10,481	9,998	9,716	9,326	9,194	9,212	9,336	9,205	9,393
NOF Steers[2]	7,490	7,364	8,031	7,935	8,346	8,693	9,077	8,452	8,050	7,840	7,190	6,946	7,249	7,451	7,075	7,149
NOF Cows[2]	32,677	32,960	33,453	34,132	35,101	35,645	35,509	34,629	34,143	33,950	33,763	33,649	33,442	33,292	33,131	33,333
Sheep	**11,358**	**11,174**	**10,797**	**10,201**	**9,836**	**8,989**	**8,465**	**8,024**	**7,825**	**7,247**	**7,036**	**6,908**	**6,623**	**6,321**	**6,105**	**6,135**
Sheep NOF	10,178	10,062	9,612	9,008	8,823	8,082	7,594	7,171	6,990	6,444	6,232	6,111	5,871	5,584	5,415	5,436
Sheep on Feed	1,180	1,112	1,185	1,193	1,013	907	871	852	836	803	804	797	752	737	690	699
Goats	**2,516**	**2,516**	**2,516**	**2,463**	**2,410**	**2,357**	**2,304**	**2,252**	**2,307**	**2,363**	**2,419**	**2,475**	**2,530**	**2,530**	**2,530**	**2,530**
Poultry	**1,537,074**	**1,594,944**	**1,649,998**	**1,707,422**	**1,769,135**	**1,826,977**	**1,882,078**	**1,926,790**	**1,965,312**	**2,008,632**	**2,033,123**	**2,060,398**	**2,097,691**	**2,085,268**	**2,130,877**	**2,151,030**
Hens >1 yr.	273,467	280,121	284,500	290,626	298,527	299,071	303,922	312,137	321,828	330,180	333,593	340,317	340,209	340,979	343,922	347,917
Pullets1	73,167	76,616	79,870	81,774	79,853	81,369	81,572	90,344	95,845	97,562	95,159	95,656	95,289	100,346	101,429	96,610
Chickens	6,545	6,857	7,113	7,240	7,369	7,637	7,243	7,549	7,682	9,661	8,088	8,126	8,353	8,439	8,248	8,289
Broilers	1,066,209	1,115,845	1,164,089	1,217,147	1,275,916	1,331,940	1,381,229	1,411,673	1,442,593	1,481,165	1,506,127	1,525,413	1,562,015	1,544,155	1,589,209	1,612,791
Turkeys	117,685	115,504	114,426	110,635	107,469	106,960	108,112	105,088	97,365	90,064	90,155	90,887	91,826	91,349	88,069	85,423
Horses	**5,069**	**5,100**	**5,121**	**5,130**	**5,110**	**5,130**	**5,150**	**5,170**	**5,237**	**5,170**	**5,240**	**5,300**	**5,300**	**5,300**	**5,300**	**5,300**

Note: Totals may not sum due to independent rounding.

[1]Pullets includes laying pullets, pullets younger than 3 months, and pullets older than 3 months.

[2]NOF = Not on Feed

Table A-162: Waste Characteristics Data

Animal Group	Average TAM (kg)	Source	Total Kjeldahl Nitrogen, N_{ex} (kg/day per 1,000 kg mass)	Source	Maximum Methane Generation Potential, B_o (m^3 CH_4/kg VS added)	Source	Volatile Solids, VS (kg/day per 1,000 kg mass)	Source
Dairy Cows	604	Safley 2000	0.44	USDA 1996a	0.24	Morris 1976	Table A- 163	Pederson and Pape, 2006
Dairy Heifers	476	Safley 2000	0.31	USDA 1996a	0.17	Bryant et. al. 1976	Table A- 163	Pederson and Pape, 2006
Feedlot Steers	420	USDA 1996a	0.30	USDA 1996a	0.33	Hashimoto 1981	Table A- 163	Pederson and Pape, 2006
Feedlot Heifers	420	USDA 1996a	0.30	USDA 1996a	0.33	Hashimoto 1981	Table A- 163	Pederson and Pape, 2006
NOF Bulls	750	Safley 2000	0.31	USDA 1996a	0.17	Hashimoto 1981	6.04	USDA 1996a
NOF Calves	118	ERG 2003	0.30	USDA 1996a	0.17	Hashimoto 1981	6.41	USDA 1996a
NOF Heifers	420	USDA 1996a	0.31	USDA 1996a	0.17	Hashimoto 1981	Table A- 163	Pederson and Pape, 2006
NOF Steers	318	Safley 2000	0.31	USDA 1996a	0.17	Hashimoto 1981	Table A- 163	Pederson and Pape, 2006
NOF Cows	533	NRC 2000	0.33	USDA 1996a	0.17	Hashimoto 1981	Table A- 163	Pederson and Pape, 2006
Market Swine <60 lbs.	16	Safley 2000	0.60	USDA 1996a	0.48	Hashimoto 1984	8.80	USDA 1996a
Market Swine 60-119 lbs.	41	Safley 2000	0.42	USDA 1996a	0.48	Hashimoto 1984	5.40	USDA 1996a
Market Swine 120-179 lbs.	68	Safley 2000	0.42	USDA 1996a	0.48	Hashimoto 1984	5.40	USDA 1996a
Market Swine >180 lbs.	91	Safley 2000	0.42	USDA 1996a	0.48	Hashimoto 1984	5.40	USDA 1996a
Breeding Swine	198	Safley 2000	0.24	USDA 1996a	0.48	Hashimoto 1984	2.60	USDA 1996a
Feedlot Sheep	25	EPA 1992	0.42	ASAE 1999	0.36	EPA 1992	9.20	EPA 1992
NOF Sheep	80	EPA 1992	0.42	ASAE 1999	0.19	EPA 1992	9.20	EPA 1992
Goats	64	ASAE 1999	0.45	ASAE 1999	0.17	EPA 1992	9.50	EPA 1992
Horses	450	ASAE 1999	0.30	ASAE 1999	0.33	EPA 1992	10.0	EPA 1992
Hens >/= 1 yr	1.8	ASAE 1999	0.83	USDA 1996a	0.39	Hill 1982	10.8	USDA 1996a
Pullets	1.8	ASAE 1999	0.62	USDA 1996a	0.39	Hill 1982	9.7	USDA 1996a
Other Chickens	1.8	ASAE 1999	0.83	USDA 1996a	0.39	Hill 1982	10.8	USDA 1996a
Broilers	0.9	ASAE 1999	1.10	USDA 1996a	0.36	Hill 1984	15.0	USDA 1996a
Turkeys	6.8	ASAE 1999	0.74	USDA 1996a	0.36	Hill 1984	9.7	USDA 1996a

Table A-163: Estimated Volatile Solids Production Rate by State for 2005

State	Dairy Cow	Dairy Heifers	Beef NOF Cow	Beef NOF Heifers	Beef NOF Steer	Beef OF Heifers	Beef OF Steer
Alabama	8.76	6.81	6.74	7.21	7.76	3.91	3.78
Alaska	11.03	6.81	8.71	9.47	10.27	3.90	3.77
Arizona	11.03	6.81	8.71	9.53	10.27	3.90	3.77
Arkansas	9.19	7.56	6.72	7.19	7.74	3.94	3.81
California	9.47	6.81	6.57	7.06	7.55	3.89	3.76
Colorado	8.97	6.81	6.19	6.66	7.08	3.92	3.78
Connecticut	8.62	6.13	6.62	7.09	7.62	3.89	3.76
Delaware	8.62	6.13	6.62	7.13	7.62	3.89	3.76
Florida	8.76	6.81	6.74	7.19	7.76	3.91	3.78
Georgia	8.76	6.81	6.74	7.22	7.76	3.91	3.78
Hawaii	11.03	6.81	8.71	9.49	10.27	3.90	3.77
Idaho	11.03	6.81	8.71	9.58	10.27	3.90	3.77
Illinois	8.74	6.81	6.63	7.14	7.62	3.92	3.79
Indiana	8.74	6.81	6.63	7.13	7.62	3.92	3.79
Iowa	8.74	6.81	6.63	7.16	7.62	3.92	3.79
Kansas	8.97	6.81	6.19	6.67	7.08	3.92	3.78
Kentucky	8.76	6.81	6.74	7.23	7.76	3.91	3.78
Louisiana	9.19	7.56	6.72	7.18	7.74	3.94	3.81
Maine	8.62	6.13	6.62	7.08	7.62	3.89	3.76
Maryland	8.62	6.13	6.62	7.11	7.62	3.89	3.76
Massachusetts	8.62	6.13	6.62	7.07	7.62	3.89	3.76
Michigan	8.74	6.81	6.63	7.13	7.62	3.92	3.79
Minnesota	8.74	6.81	6.63	7.14	7.62	3.92	3.79
Mississippi	8.76	6.81	6.74	7.21	7.76	3.91	3.78
Missouri	8.74	6.81	6.63	7.11	7.62	3.92	3.79
Montana	8.97	6.81	6.19	6.59	7.08	3.92	3.78
Nebraska	8.97	6.81	6.19	6.66	7.08	3.92	3.78
Nevada	11.03	6.81	8.71	9.54	10.27	3.90	3.77
New Hampshire	8.62	6.13	6.62	7.08	7.62	3.89	3.76
New Jersey	8.62	6.13	6.62	7.10	7.62	3.89	3.76
New Mexico	11.03	6.81	8.71	9.55	10.27	3.90	3.77
New York	8.62	6.13	6.62	7.13	7.62	3.89	3.76
North Carolina	8.76	6.81	6.74	7.20	7.76	3.91	3.78
North Dakota	8.97	6.81	6.19	6.63	7.08	3.92	3.78
Ohio	8.74	6.81	6.63	7.11	7.62	3.92	3.79
Oklahoma	9.19	7.56	6.72	7.23	7.74	3.94	3.81
Oregon	11.03	6.81	8.71	9.54	10.27	3.90	3.77
Pennsylvania	8.62	6.13	6.62	7.12	7.62	3.89	3.76
Rhode Island	8.62	6.13	6.62	7.08	7.62	3.89	3.76
South Carolina	8.76	6.81	6.74	7.21	7.76	3.91	3.78
South Dakota	8.97	6.81	6.19	6.64	7.08	3.92	3.78
Tennessee	8.76	6.81	6.74	7.21	7.76	3.91	3.78
Texas	9.19	7.56	6.72	7.24	7.74	3.94	3.81
Utah	11.03	6.81	8.71	9.55	10.27	3.90	3.77
Vermont	8.62	6.13	6.62	7.10	7.62	3.89	3.76
Virginia	8.76	6.81	6.74	7.23	7.76	3.91	3.78
Washington	11.03	6.81	8.71	9.59	10.27	3.90	3.77
West Virginia	8.62	6.13	6.62	7.09	7.62	3.89	3.76
Wisconsin	8.74	6.81	6.63	7.12	7.62	3.92	3.79
Wyoming	8.97	6.81	6.19	6.62	7.08	3.92	3.78

Table A-164: 2005 Manure Distribution Among Waste Management Systems at Beef Feedlots (Percent)

State	Pasture	Daily Spread	Solid Storage	Dry Lot[a]	Liquid/ Slurry[a]	Anaerobic Lagoon	Deep Pit	Poultry with Litter	Poultry without Litter
Alabama	0	0	0	100	1.3	0	0	0	0
Alaska	0	0	0	100	1.3	0	0	0	0
Arizona	0	0	0	100	0.4	0	0	0	0
Arkansas	0	0	0	100	1.3	0	0	0	0
California	0	0	0	100	1.3	0	0	0	0
Colorado	0	0	0	100	0.4	0	0	0	0
Connecticut	0	0	0	100	1.0	0	0	0	0
Delaware	0	0	0	100	1.0	0	0	0	0
Florida	0	0	0	100	1.3	0	0	0	0
Georgia	0	0	0	100	1.3	0	0	0	0
Hawaii	0	0	0	100	1.3	0	0	0	0
Idaho	0	0	0	100	0.4	0	0	0	0
Illinois	0	0	0	100	0.6	0	0	0	0
Indiana	0	0	0	100	0.6	0	0	0	0
Iowa	0	0	0	100	0.6	0	0	0	0
Kansas	0	0	0	100	0.6	0	0	0	0
Kentucky	0	0	0	100	1.0	0	0	0	0
Louisiana	0	0	0	100	1.3	0	0	0	0
Maine	0	0	0	100	1.0	0	0	0	0
Maryland	0	0	0	100	1.0	0	0	0	0
Massachusetts	0	0	0	100	1.0	0	0	0	0
Michigan	0	0	0	100	0.6	0	0	0	0
Minnesota	0	0	0	100	0.6	0	0	0	0
Mississippi	0	0	0	100	1.3	0	0	0	0
Missouri	0	0	0	100	0.6	0	0	0	0
Montana	0	0	0	100	0.4	0	0	0	0
Nebraska	0	0	0	100	0.6	0	0	0	0
Nevada	0	0	0	100	0.4	0	0	0	0
New Hampshire	0	0	0	100	1.0	0	0	0	0
New Jersey	0	0	0	100	1.0	0	0	0	0
New Mexico	0	0	0	100	0.4	0	0	0	0
New York	0	0	0	100	1.0	0	0	0	0
North Carolina	0	0	0	100	1.0	0	0	0	0
North Dakota	0	0	0	100	0.6	0	0	0	0
Ohio	0	0	0	100	0.6	0	0	0	0
Oklahoma	0	0	0	100	0.4	0	0	0	0
Oregon	0	0	0	100	1.3	0	0	0	0
Pennsylvania	0	0	0	100	1.0	0	0	0	0
Rhode Island	0	0	0	100	1.0	0	0	0	0
South Carolina	0	0	0	100	1.3	0	0	0	0
South Dakota	0	0	0	100	0.6	0	0	0	0
Tennessee	0	0	0	100	1.0	0	0	0	0
Texas	0	0	0	100	0.4	0	0	0	0
Utah	0	0	0	100	0.4	0	0	0	0
Vermont	0	0	0	100	1.0	0	0	0	0
Virginia	0	0	0	100	1.0	0	0	0	0
Washington	0	0	0	100	1.3	0	0	0	0
West Virginia	0	0	0	100	1.0	0	0	0	0
Wisconsin	0	0	0	100	0.6	0	0	0	0
Wyoming	0	0	0	100	0.4	0	0	0	0

[a] Because manure at beef feedlots may be managed for long periods of time in multiple systems (i.e., both drylot and runoff collection pond), the percent of manure that generates emissions is greater than 100 percent.

Table A-165: 2005 Manure Distribution Among Waste Management Systems at Dairies (Percent)

State	Pasture	Daily Spread	Solid Storage	Dry Lot	Liquid/ Slurry	Anaerobic Lagoon	Deep Pit	Poultry with Litter	Poultry without Litter
Alabama	45	17	11	0	10	17	0	0	0
Alaska	4	6	28	0	24	30	8	0	0
Arizona	0	10	9	0	20	61	0	0	0
Arkansas	57	15	11	0	7	10	1	0	0
California	1	11	9	0	21	58	0	0	0
Colorado	1	1	12	0	24	62	1	0	0
Connecticut	6	43	17	0	20	12	2	0	0
Delaware	6	44	19	0	19	10	2	0	0
Florida	17	22	8	0	15	39	0	0	0
Georgia	40	18	10	0	11	21	0	0	0
Hawaii	1	0	11	0	21	67	0	0	0
Idaho	0	1	12	0	23	63	1	0	0
Illinois	5	8	43	0	26	13	5	0	0
Indiana	8	13	35	0	24	16	3	0	0
Iowa	6	10	41	0	25	14	4	0	0
Kansas	3	5	28	0	33	29	3	0	0
Kentucky	61	14	14	0	6	2	2	0	0
Louisiana	60	14	10	0	6	9	1	0	0
Maine	7	45	20	0	17	9	2	0	0
Maryland	7	44	23	0	15	8	3	0	0
Massachusetts	7	45	24	0	15	6	3	0	0
Michigan	3	6	32	0	33	22	4	0	0
Minnesota	6	10	44	0	24	12	5	0	0
Mississippi	57	15	10	0	7	10	0	0	0
Missouri	8	14	48	0	18	6	5	0	0
Montana	3	4	25	0	26	36	6	0	0
Nebraska	4	6	35	0	30	21	4	0	0
Nevada	0	0	11	0	24	64	0	0	0
New Hampshire	7	44	21	0	16	9	3	0	0
New Jersey	8	45	24	0	14	6	3	0	0
New Mexico	0	10	9	0	19	61	0	0	0
New York	7	45	20	0	16	10	2	0	0
North Carolina	54	15	12	0	11	6	1	0	0
North Dakota	6	11	45	0	22	12	4	0	0
Ohio	8	14	41	0	23	11	4	0	0
Oklahoma	0	6	25	0	23	40	6	0	0
Oregon	20	0	13	0	21	44	2	0	0
Pennsylvania	9	47	25	0	12	5	2	0	0
Rhode Island	9	47	26	0	12	4	3	0	0
South Carolina	54	15	8	0	9	13	0	0	0
South Dakota	5	8	38	0	28	18	4	0	0
Tennessee	59	15	12	0	9	4	2	0	0
Texas	0	8	13	0	24	53	2	0	0
Utah	1	1	17	0	26	51	3	0	0
Vermont	7	44	19	0	17	10	2	0	0
Virginia	57	15	12	0	9	4	2	0	0
Washington	17	0	11	0	22	49	1	0	0
West Virginia	7	45	23	0	16	7	3	0	0
Wisconsin	7	12	42	0	24	12	4	0	0
Wyoming	7	12	22	0	23	30	6	0	0

Table A- 166: 2005 Manure Distribution Among Waste Management Systems at Dairy Heifer Facilities (Percent)

State	Pasture	Daily Spread	Solid Storage	Dry Lot[1]	Liquid/ Slurry[1]	Anaerobic Lagoon	Deep Pit	Poultry with Litter	Poultry without Litter
Alabama	0	0	0	100	1.0	0	0	0	0
Alaska	0	0	0	100	1.0	0	0	0	0
Arizona	0	0	0	100	0.3	0	0	0	0
Arkansas	0	0	0	100	1.0	0	0	0	0
California	0	0	0	100	1.0	0	0	0	0
Colorado	0	0	0	100	0.3	0	0	0	0
Connecticut	0	0	0	100	0.8	0	0	0	0
Delaware	0	0	0	100	0.8	0	0	0	0
Florida	0	0	0	100	1.0	0	0	0	0
Georgia	0	0	0	100	1.0	0	0	0	0
Hawaii	0	0	0	100	1.0	0	0	0	0
Idaho	0	0	0	100	0.3	0	0	0	0
Illinois	0	0	0	100	0.5	0	0	0	0
Indiana	0	0	0	100	0.5	0	0	0	0
Iowa	0	0	0	100	0.5	0	0	0	0
Kansas	0	0	0	100	0.5	0	0	0	0
Kentucky	0	0	0	100	0.8	0	0	0	0
Louisiana	0	0	0	100	1.0	0	0	0	0
Maine	0	0	0	100	0.8	0	0	0	0
Maryland	0	0	0	100	0.8	0	0	0	0
Massachusetts	0	0	0	100	0.8	0	0	0	0
Michigan	0	0	0	100	0.5	0	0	0	0
Minnesota	0	0	0	100	0.5	0	0	0	0
Mississippi	0	0	0	100	1.0	0	0	0	0
Missouri	0	0	0	100	0.5	0	0	0	0
Montana	0	0	0	100	0.3	0	0	0	0
Nebraska	0	0	0	100	0.5	0	0	0	0
Nevada	0	0	0	100	0.3	0	0	0	0
New Hampshire	0	0	0	100	0.8	0	0	0	0
New Jersey	0	0	0	100	0.8	0	0	0	0
New Mexico	0	0	0	100	0.3	0	0	0	0
New York	0	0	0	100	0.8	0	0	0	0
North Carolina	0	0	0	100	0.8	0	0	0	0
North Dakota	0	0	0	100	0.5	0	0	0	0
Ohio	0	0	0	100	0.5	0	0	0	0
Oklahoma	0	0	0	100	0.3	0	0	0	0
Oregon	0	0	0	100	1.0	0	0	0	0
Pennsylvania	0	0	0	100	0.8	0	0	0	0
Rhode Island	0	0	0	100	0.8	0	0	0	0
South Carolina	0	0	0	100	1.0	0	0	0	0
South Dakota	0	0	0	100	0.5	0	0	0	0
Tennessee	0	0	0	100	0.8	0	0	0	0
Texas	0	0	0	100	0.3	0	0	0	0
Utah	0	0	0	100	0.3	0	0	0	0
Vermont	0	0	0	100	0.8	0	0	0	0
Virginia	0	0	0	100	0.8	0	0	0	0
Washington	0	0	0	100	1.0	0	0	0	0
West Virginia	0	0	0	100	0.8	0	0	0	0
Wisconsin	0	0	0	100	0.5	0	0	0	0
Wyoming	0	0	0	100	0.3	0	0	0	0

[1] Because manure from dairy heifers may be managed for long periods of time in multiple systems (i.e., both drylot and runoff collection pond), the percent of manure that generates emissions is greater than 100 percent.

Table A-167: 2005 Manure Distribution Among Waste Management Systems at Swine Operations (Percent)

State	Pasture	Daily Spread	Solid Storage	Dry Lot	Liquid/ Slurry	Anaerobic Lagoon	Deep Pit	Poultry with Litter	Poultry without Litter
Alabama	5	0	4	0	8	51	32	0	0
Alaska	53	0	2	0	13	10	22	0	0
Arizona	14	0	3	0	5	50	27	0	0
Arkansas	4	0	4	0	12	46	35	0	0
California	13	0	3	0	8	47	29	0	0
Colorado	2	0	5	0	26	17	50	0	0
Connecticut	57	0	2	0	12	9	20	0	0
Delaware	12	0	4	0	24	18	42	0	0
Florida	71	0	1	0	8	6	14	0	0
Georgia	8	0	4	0	10	47	32	0	0
Hawaii	23	0	3	0	18	21	34	0	0
Idaho	46	0	3	0	15	10	26	0	0
Illinois	2	0	5	0	28	15	50	0	0
Indiana	3	0	5	0	28	15	50	0	0
Iowa	1	0	4	0	12	48	35	0	0
Kansas	2	0	5	0	28	13	51	0	0
Kentucky	5	0	4	0	12	45	34	0	0
Louisiana	54	0	2	0	13	10	22	0	0
Maine	73	0	1	0	7	6	13	0	0
Maryland	21	0	4	0	21	16	38	0	0
Massachusetts	31	0	3	0	19	14	32	0	0
Michigan	5	0	5	0	25	17	48	0	0
Minnesota	2	0	5	0	26	18	49	0	0
Mississippi	2	0	4	0	6	57	31	0	0
Missouri	3	0	5	0	28	14	51	0	0
Montana	6	0	5	0	25	17	47	0	0
Nebraska	3	0	5	0	28	15	50	0	0
Nevada	35	0	2	0	4	39	20	0	0
New Hampshire	57	0	2	0	12	9	20	0	0
New Jersey	31	0	3	0	19	14	33	0	0
New Mexico	92	0	0	0	2	2	4	0	0
New York	20	0	4	0	21	15	40	0	0
North Carolina	0	0	4	0	6	58	31	0	0
North Dakota	9	0	5	0	24	17	45	0	0
Ohio	7	0	4	0	27	15	47	0	0
Oklahoma	1	0	4	0	6	58	31	0	0
Oregon	58	0	2	0	12	9	20	0	0
Pennsylvania	5	0	5	0	25	18	47	0	0
Rhode Island	56	0	2	0	12	9	21	0	0
South Carolina	6	0	4	0	9	50	32	0	0
South Dakota	3	0	5	0	26	18	48	0	0
Tennessee	10	0	4	0	12	41	33	0	0
Texas	7	0	3	0	6	54	29	0	0
Utah	1	0	6	0	26	17	51	0	0
Vermont	86	0	1	0	4	3	7	0	0
Virginia	3	0	4	0	7	55	31	0	0
Washington	37	0	3	0	17	12	31	0	0
West Virginia	58	0	2	0	11	8	21	0	0
Wisconsin	13	0	4	0	24	17	42	0	0
Wyoming	3	0	5	0	26	17	49	0	0

Table A-168: 2005 Manure Distribution Among Waste Management Systems at Layer Operations (Percent)

State	Pasture	Daily Spread	Solid Storage	Dry Lot	Liquid/ Slurry	Anaerobic Lagoon	Deep Pit	Poultry with Litter	Poultry without Litter
Alabama	0	0	0	0	0	42	0	0	58
Alaska	0	0	0	0	0	25	0	0	75
Arizona	0	0	0	0	0	60	0	0	40
Arkansas	0	0	0	0	0	0	0	0	100
California	0	0	0	0	0	12	0	0	88
Colorado	0	0	0	0	0	60	0	0	40
Connecticut	0	0	0	0	0	5	0	0	95
Delaware	0	0	0	0	0	5	0	0	95
Florida	0	0	0	0	0	42	0	0	58
Georgia	0	0	0	0	0	42	0	0	58
Hawaii	0	0	0	0	0	25	0	0	75
Idaho	0	0	0	0	0	60	0	0	40
Illinois	0	0	0	0	0	2	0	0	98
Indiana	0	0	0	0	0	0	0	0	100
Iowa	0	0	0	0	0	0	0	0	100
Kansas	0	0	0	0	0	2	0	0	98
Kentucky	0	0	0	0	0	5	0	0	95
Louisiana	0	0	0	0	0	60	0	0	40
Maine	0	0	0	0	0	5	0	0	95
Maryland	0	0	0	0	0	5	0	0	95
Massachusetts	0	0	0	0	0	5	0	0	95
Michigan	0	0	0	0	0	2	0	0	98
Minnesota	0	0	0	0	0	0	0	0	100
Mississippi	0	0	0	0	0	60	0	0	40
Missouri	0	0	0	0	0	0	0	0	100
Montana	0	0	0	0	0	60	0	0	40
Nebraska	0	0	0	0	0	2	0	0	98
Nevada	0	0	0	0	0	0	0	0	100
New Hampshire	0	0	0	0	0	5	0	0	95
New Jersey	0	0	0	0	0	5	0	0	95
New Mexico	0	0	0	0	0	60	0	0	40
New York	0	0	0	0	0	5	0	0	95
North Carolina	0	0	0	0	0	42	0	0	58
North Dakota	0	0	0	0	0	2	0	0	98
Ohio	0	0	0	0	0	0	0	0	100
Oklahoma	0	0	0	0	0	60	0	0	40
Oregon	0	0	0	0	0	25	0	0	75
Pennsylvania	0	0	0	0	0	0	0	0	100
Rhode Island	0	0	0	0	0	5	0	0	95
South Carolina	0	0	0	0	0	60	0	0	40
South Dakota	0	0	0	0	0	2	0	0	98
Tennessee	0	0	0	0	0	5	0	0	95
Texas	0	0	0	0	0	12	0	0	88
Utah	0	0	0	0	0	60	0	0	40
Vermont	0	0	0	0	0	5	0	0	95
Virginia	0	0	0	0	0	5	0	0	95
Washington	0	0	0	0	0	12	0	0	88
West Virginia	0	0	0	0	0	5	0	0	95
Wisconsin	0	0	0	0	0	2	0	0	98
Wyoming	0	0	0	0	0	60	0	0	40

Table A- 169: 2005 Manure Distribution Among Waste Management Systems at Broiler and Turkey Operations (Percent)

State	Pasture	Daily Spread	Solid Storage	Dry Lot	Liquid/ Slurry	Anaerobic Lagoon	Deep Pit	Poultry with Litter	Poultry without Litter
Alabama	1	0	0	0	0	0	0	99	0
Alaska	1	0	0	0	0	0	0	99	0
Arizona	1	0	0	0	0	0	0	99	0
Arkansas	1	0	0	0	0	0	0	99	0
California	1	0	0	0	0	0	0	99	0
Colorado	1	0	0	0	0	0	0	99	0
Connecticut	1	0	0	0	0	0	0	99	0
Delaware	1	0	0	0	0	0	0	99	0
Florida	1	0	0	0	0	0	0	99	0
Georgia	1	0	0	0	0	0	0	99	0
Hawaii	1	0	0	0	0	0	0	99	0
Idaho	1	0	0	0	0	0	0	99	0
Illinois	1	0	0	0	0	0	0	99	0
Indiana	1	0	0	0	0	0	0	99	0
Iowa	1	0	0	0	0	0	0	99	0
Kansas	1	0	0	0	0	0	0	99	0
Kentucky	1	0	0	0	0	0	0	99	0
Louisiana	1	0	0	0	0	0	0	99	0
Maine	1	0	0	0	0	0	0	99	0
Maryland	1	0	0	0	0	0	0	99	0
Massachusetts	1	0	0	0	0	0	0	99	0
Michigan	1	0	0	0	0	0	0	99	0
Minnesota	1	0	0	0	0	0	0	99	0
Mississippi	1	0	0	0	0	0	0	99	0
Missouri	1	0	0	0	0	0	0	99	0
Montana	1	0	0	0	0	0	0	99	0
Nebraska	1	0	0	0	0	0	0	99	0
Nevada	1	0	0	0	0	0	0	99	0
New Hampshire	1	0	0	0	0	0	0	99	0
New Jersey	1	0	0	0	0	0	0	99	0
New Mexico	1	0	0	0	0	0	0	99	0
New York	1	0	0	0	0	0	0	99	0
North Carolina	1	0	0	0	0	0	0	99	0
North Dakota	1	0	0	0	0	0	0	99	0
Ohio	1	0	0	0	0	0	0	99	0
Oklahoma	1	0	0	0	0	0	0	99	0
Oregon	1	0	0	0	0	0	0	99	0
Pennsylvania	1	0	0	0	0	0	0	99	0
Rhode Island	1	0	0	0	0	0	0	99	0
South Carolina	1	0	0	0	0	0	0	99	0
South Dakota	1	0	0	0	0	0	0	99	0
Tennessee	1	0	0	0	0	0	0	99	0
Texas	1	0	0	0	0	0	0	99	0
Utah	1	0	0	0	0	0	0	99	0
Vermont	1	0	0	0	0	0	0	99	0
Virginia	1	0	0	0	0	0	0	99	0
Washington	1	0	0	0	0	0	0	99	0
West Virginia	1	0	0	0	0	0	0	99	0
Wisconsin	1	0	0	0	0	0	0	99	0
Wyoming	1	0	0	0	0	0	0	99	0

Table A- 170: Methane Conversion Factors By State for Liquid Systems[43] for 2005 (percent)

State	Liquid/Slurry and Deep Pit	Anaerobic Lagoon
Alabama	32	76
Alaska	13	50
Arizona	38	78
Arkansas	31	76
California	28	75
Colorado	19	66
Connecticut	22	71
Delaware	27	75
Florida	43	78
Georgia	31	76
Hawaii	51	77
Idaho	18	65
Illinois	25	73
Indiana	24	73
Iowa	22	71
Kansas	26	74
Kentucky	27	74
Louisiana	39	78
Maine	17	63
Maryland	26	75
Massachusetts	20	69
Michigan	21	70
Minnesota	20	69
Mississippi	34	76
Missouri	27	75
Montana	16	61
Nebraska	23	72
Nevada	21	69
New Hampshire	19	66
New Jersey	25	74
New Mexico	23	72
New York	20	69
North Carolina	28	75
North Dakota	18	65
Ohio	23	72
Oklahoma	30	76
Oregon	17	63
Pennsylvania	23	72
Rhode Island	22	71
South Carolina	32	76
South Dakota	21	70
Tennessee	28	74
Texas	36	77
Utah	18	65
Vermont	18	66
Virginia	25	73
Washington	18	64
West Virginia	23	72
Wisconsin	20	69
Wyoming	17	63

[43] As defined by IPCC (IPCC 2006). MCFs represent weighted average of multiple animal types.

Table A- 171: Weighted Methane Conversion Factors for 2005[a] (Percent)

State	Beef Feedlot-Heifer	Beef Feedlot-Steers	Dairy Cow	Dairy Heifer	Swine—Market	Swine—Breeding	Layer	Broiler	Turkey
Alabama	2.0	1.5	18.1	1.9	53.6	53.4	32.6	1.5	1.5
Alaska	1.2	1.0	20.8	1.2	10.5	10.5	13.7	1.5	1.5
Arizona	1.7	1.5	60.7	1.7	50.7	50.2	46.7	1.5	1.5
Arkansas	2.0	1.5	11.3	1.9	52.7	52.8	1.5	1.5	1.5
California	2.0	1.5	50.8	1.8	45.5	46.2	10.3	1.5	1.5
Colorado	1.1	1.0	45.0	1.1	28.8	28.7	39.0	1.5	1.5
Connecticut	1.3	1.0	14.3	1.2	14.7	14.7	5.0	1.5	1.5
Delaware	1.3	1.0	14.9	1.3	33.5	33.5	5.2	1.5	1.5
Florida	2.2	1.5	38.2	2.0	16.4	16.4	33.5	1.5	1.5
Georgia	2.0	1.5	21.2	1.9	51.2	51.2	32.3	1.5	1.5
Hawaii	2.3	1.5	64.9	2.1	48.1	48.1	20.4	1.5	1.5
Idaho	1.1	1.0	47.0	1.1	15.2	15.2	38.6	1.5	1.5
Illinois	1.2	1.0	19.7	1.1	33.4	33.4	2.9	1.5	1.5
Indiana	1.2	1.0	20.2	1.1	31.9	31.9	1.5	1.5	1.5
Iowa	1.2	1.0	18.1	1.1	45.7	45.7	1.5	1.5	1.5
Kansas	1.2	1.0	32.9	1.2	33.9	33.9	3.0	1.5	1.5
Kentucky	1.3	1.0	5.1	1.3	47.9	47.8	5.2	1.5	1.5
Louisiana	2.1	1.5	11.3	2.0	24.5	24.5	47.2	1.5	1.5
Maine	1.2	1.0	10.5	1.2	8.2	8.2	4.6	1.5	1.5
Maryland	1.3	1.0	11.9	1.2	29.7	29.7	5.2	1.5	1.5
Massachusetts	1.2	1.0	9.0	1.2	22.4	22.4	4.9	1.5	1.5
Michigan	1.1	1.0	24.8	1.1	30.0	30.0	2.9	1.5	1.5
Minnesota	1.1	1.0	15.7	1.1	30.0	30.0	1.5	1.5	1.5
Mississippi	2.0	1.5	12.5	1.9	58.1	58.2	46.5	1.5	1.5
Missouri	1.2	1.0	13.2	1.2	34.2	34.2	1.5	1.5	1.5
Montana	1.1	1.0	27.6	1.1	23.2	23.2	37.3	1.5	1.5
Nebraska	1.2	1.0	25.4	1.1	31.1	31.1	2.9	1.5	1.5
Nevada	1.1	1.0	50.8	1.1	33.4	31.8	1.5	1.5	1.5
New Hampshire	1.2	1.0	10.9	1.2	13.7	13.7	4.7	1.5	1.5
New Jersey	1.3	1.0	9.9	1.2	25.4	25.4	5.1	1.5	1.5
New Mexico	1.1	1.0	50.4	1.1	3.8	3.8	42.4	1.5	1.5
New York	1.2	1.0	11.5	1.2	24.1	24.1	4.9	1.5	1.5
North Carolina	1.3	1.0	9.4	1.2	58.8	58.7	32.2	1.5	1.5
North Dakota	1.1	1.0	14.3	1.1	25.1	25.1	2.8	1.5	1.5
Ohio	1.2	1.0	15.8	1.1	30.4	30.4	1.5	1.5	1.5
Oklahoma	1.1	1.0	42.4	1.6	56.2	56.4	46.7	1.5	1.5
Oregon	1.3	1.0	32.4	1.2	12.5	12.4	17.0	1.5	1.5
Pennsylvania	1.3	1.0	7.9	1.2	32.1	32.1	1.5	1.5	1.5
Rhode Island	1.3	1.0	7.4	1.2	15.3	15.3	5.0	1.5	1.5
South Carolina	2.0	1.5	14.8	1.9	53.7	53.7	46.4	1.5	1.5
South Dakota	1.2	1.0	21.1	1.1	30.6	30.6	2.9	1.5	1.5
Tennessee	1.3	1.0	7.1	1.2	46.5	46.2	5.1	1.5	1.5
Texas	1.6	1.5	53.2	1.6	56.8	56.8	10.7	1.5	1.5
Utah	1.1	1.0	39.9	1.1	25.9	26.0	39.2	1.5	1.5
Vermont	1.2	1.0	11.4	1.2	4.9	4.9	4.7	1.5	1.5
Virginia	1.3	1.0	7.2	1.2	52.5	52.6	5.1	1.5	1.5
Washington	1.3	1.0	35.7	1.2	17.6	17.6	8.9	1.5	1.5
West Virginia	1.3	1.0	10.5	1.2	14.7	14.7	5.0	1.5	1.5
Wisconsin	1.1	1.0	15.7	1.1	27.2	27.2	2.9	1.5	1.5
Wyoming	1.1	1.0	23.1	1.1	25.5	25.4	37.9	1.5	1.5

[a] MCFs are weighted by the distribution of waste management systems for each animal type.

Table A- 172: Weighted N_2O Emission Factors for 2005[a] (kg N_2O-N/kg Kjdl N)

State	Beef Feedlot-Heifer	Beef Feedlot-Steers	Dairy Cow	Dairy Heifer	Swine—Market	Swine—Breeding	Layer	Broiler	Turkey
Alabama	0.0201	0.0201	0.0027	0.0201	0.0013	0.0013	0.0006	0.0010	0.0010
Alaska	0.0201	0.0201	0.0031	0.0201	0.0025	0.0025	0.0008	0.0010	0.0010
Arizona	0.0200	0.0200	0.0016	0.0200	0.0011	0.0011	0.0004	0.0010	0.0010
Arkansas	0.0201	0.0201	0.0032	0.0201	0.0015	0.0015	0.0010	0.0010	0.0010
California	0.0201	0.0201	0.0017	0.0201	0.0013	0.0013	0.0009	0.0010	0.0010
Colorado	0.0200	0.0200	0.0018	0.0200	0.0026	0.0026	0.0004	0.0010	0.0010
Connecticut	0.0201	0.0201	0.0037	0.0200	0.0025	0.0025	0.0010	0.0010	0.0010
Delaware	0.0201	0.0201	0.0038	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Florida	0.0201	0.0201	0.0018	0.0201	0.0025	0.0025	0.0006	0.0010	0.0010
Georgia	0.0201	0.0201	0.0025	0.0201	0.0014	0.0014	0.0006	0.0010	0.0010
Hawaii	0.0201	0.0201	0.0016	0.0201	0.0023	0.0023	0.0008	0.0010	0.0010
Idaho	0.0200	0.0200	0.0018	0.0200	0.0026	0.0026	0.0004	0.0010	0.0010
Illinois	0.0200	0.0200	0.0041	0.0200	0.0027	0.0027	0.0010	0.0010	0.0010
Indiana	0.0200	0.0200	0.0038	0.0200	0.0027	0.0027	0.0010	0.0010	0.0010
Iowa	0.0200	0.0200	0.0040	0.0200	0.0015	0.0015	0.0010	0.0010	0.0010
Kansas	0.0200	0.0200	0.0033	0.0200	0.0027	0.0027	0.0010	0.0010	0.0010
Kentucky	0.0201	0.0201	0.0043	0.0200	0.0015	0.0015	0.0010	0.0010	0.0010
Louisiana	0.0201	0.0201	0.0032	0.0201	0.0025	0.0025	0.0004	0.0010	0.0010
Maine	0.0201	0.0201	0.0039	0.0200	0.0025	0.0025	0.0010	0.0010	0.0010
Maryland	0.0201	0.0201	0.0040	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Massachusetts	0.0201	0.0201	0.0042	0.0200	0.0025	0.0025	0.0010	0.0010	0.0010
Michigan	0.0200	0.0200	0.0037	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Minnesota	0.0200	0.0200	0.0041	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Mississippi	0.0201	0.0201	0.0031	0.0201	0.0011	0.0011	0.0004	0.0010	0.0010
Missouri	0.0200	0.0200	0.0044	0.0200	0.0027	0.0027	0.0010	0.0010	0.0010
Montana	0.0200	0.0200	0.0029	0.0200	0.0026	0.0026	0.0004	0.0010	0.0010
Nebraska	0.0200	0.0200	0.0037	0.0200	0.0027	0.0027	0.0010	0.0010	0.0010
Nevada	0.0200	0.0200	0.0018	0.0200	0.0011	0.0011	0.0010	0.0010	0.0010
New Hampshire	0.0201	0.0201	0.0039	0.0200	0.0025	0.0025	0.0010	0.0010	0.0010
New Jersey	0.0201	0.0201	0.0042	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
New Mexico	0.0200	0.0200	0.0016	0.0200	0.0025	0.0025	0.0004	0.0010	0.0010
New York	0.0201	0.0201	0.0038	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
North Carolina	0.0201	0.0201	0.0038	0.0200	0.0011	0.0011	0.0006	0.0010	0.0010
North Dakota	0.0200	0.0200	0.0041	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Ohio	0.0200	0.0200	0.0042	0.0200	0.0027	0.0027	0.0010	0.0010	0.0010
Oklahoma	0.0200	0.0200	0.0027	0.0200	0.0011	0.0011	0.0004	0.0010	0.0010
Oregon	0.0201	0.0201	0.0022	0.0201	0.0025	0.0025	0.0008	0.0010	0.0010
Pennsylvania	0.0201	0.0201	0.0043	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Rhode Island	0.0201	0.0201	0.0043	0.0200	0.0025	0.0025	0.0010	0.0010	0.0010
South Carolina	0.0201	0.0201	0.0028	0.0201	0.0013	0.0013	0.0004	0.0010	0.0010
South Dakota	0.0200	0.0200	0.0039	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Tennessee	0.0201	0.0201	0.0040	0.0200	0.0016	0.0016	0.0010	0.0010	0.0010
Texas	0.0200	0.0200	0.0020	0.0200	0.0011	0.0011	0.0009	0.0010	0.0010
Utah	0.0200	0.0200	0.0023	0.0200	0.0026	0.0026	0.0004	0.0010	0.0010
Vermont	0.0201	0.0201	0.0038	0.0200	0.0025	0.0025	0.0010	0.0010	0.0010
Virginia	0.0201	0.0201	0.0040	0.0200	0.0012	0.0012	0.0010	0.0010	0.0010
Washington	0.0201	0.0201	0.0020	0.0201	0.0026	0.0026	0.0009	0.0010	0.0010
West Virginia	0.0201	0.0201	0.0041	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Wisconsin	0.0200	0.0200	0.0041	0.0200	0.0026	0.0026	0.0010	0.0010	0.0010
Wyoming	0.0200	0.0200	0.0030	0.0200	0.0026	0.0026	0.0004	0.0010	0.0010

[a] EFs are weighted by the distribution of waste management systems for each animal type.

Table A- 173: CH_4 Emissions from Livestock Manure Management (Gg)

Animal Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Dairy Cattle	**568**	**580**	**569**	**561**	**605**	**634**	**606**	**635**	**677**	**707**	**748**	**789**	**818**	**839**	**814**	**851**
Dairy Cows	561	573	562	554	598	628	599	628	670	700	741	782	811	833	807	844
Dairy Heifer	7	7	7	7	7	7	7	7	7	7	7	7	7	7	7	7
Swine	**623**	**674**	**637**	**678**	**739**	**762**	**728**	**781**	**889**	**845**	**830**	**849**	**873**	**821**	**815**	**852**
Market Swine	484	523	500	534	584	608	581	626	720	692	679	696	720	691	689	727
Market <60 lbs.	102	110	103	108	119	121	116	125	140	132	131	134	137	131	130	135
Market 60-119 lbs.	101	111	104	110	119	123	117	127	144	138	135	138	144	135	136	142
Market 120-179 lbs.	137	147	140	151	165	170	164	176	201	194	189	192	200	196	195	210
Market >180 lbs.	144	156	152	164	182	193	185	198	234	228	225	232	240	228	228	240
Breeding Swine	139	151	138	145	156	154	147	156	169	153	151	153	153	136	132	133
Beef Cattle	**120**	**121**	**118**	**118**	**121**	**122**	**119**	**117**	**120**	**119**	**114**	**117**	**114**	**113**	**110**	**111**
Feedlot Steers	18	19	16	17	15	14	12	13	13	13	14	15	14	13	12	12
Feedlot Heifers	9	10	8	7	8	7	7	7	8	8	8	9	8	8	8	8
NOF Bulls	5	5	5	5	5	5	5	5	5	5	5	5	5	5	5	5
NOF Calves	9	9	9	9	9	9	9	9	9	9	9	9	8	8	8	8
NOF Heifers	13	13	13	14	15	15	15	15	14	14	13	13	13	13	13	13
NOF Steers	8	8	9	8	9	9	10	9	9	9	8	8	8	8	8	8
NOF Cows	58	58	58	59	61	62	61	60	61	61	58	60	58	57	57	57
Sheep	**7**	**7**	**7**	**6**	**6**	**5**	**5**	**5**	**5**	**4**	**4**	**4**	**4**	**4**	**4**	**4**
Goats	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**	**1**
Poultry	**131**	**131**	**127**	**131**	**130**	**128**	**125**	**127**	**130**	**125**	**125**	**129**	**127**	**127**	**126**	**125**
Hens >1 yr.	73	72	70	73	72	69	68	67	70	66	66	70	67	68	65	65
Total Pullets	25	26	23	23	23	22	21	23	23	21	22	22	22	22	23	22
Chickens	4	4	4	4	4	4	3	3	4	3	3	3	4	4	3	3
Broilers	19	20	20	21	22	23	24	25	25	26	27	27	28	27	28	28
Turkeys	10	10	10	10	9	9	9	9	8	8	8	8	8	8	8	7
Horses	**22**	**22**	**21**	**21**	**21**	**21**	**21**	**21**	**22**	**22**	**22**	**22**	**22**	**22**	**22**	**22**

Table A- 174: N_2O Emissions from Livestock Manure Management (Gg)

Animal Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Dairy Cattle	**7.9**	**7.8**	**7.8**	**7.8**	**7.8**	**7.9**	**7.8**	**7.7**	**7.7**	**7.9**	**7.9**	**8.0**	**8.1**	**8.1**	**8.0**	**8.1**
Dairy Cows	3.5	3.4	3.3	3.3	3.3	3.3	3.3	3.2	3.2	3.2	3.2	3.1	3.1	3.1	3.1	3.1
Dairy Heifer	4.4	4.4	4.4	4.5	4.6	4.6	4.5	4.5	4.5	4.7	4.8	4.8	5.0	4.9	4.9	5.1
Swine	**1.5**	**1.6**	**1.6**	**1.6**	**1.6**	**1.6**	**1.5**	**1.5**	**1.6**	**1.6**	**1.5**	**1.5**	**1.5**	**1.5**	**1.5**	**1.5**
Market Swine	1.1	1.2	1.2	1.2	1.2	1.2	1.1	1.2	1.3	1.2	1.2	1.2	1.2	1.2	1.2	1.2
Market <60 lbs.	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2
Market 60-119 lbs.	0.2	0.3	0.3	0.3	0.3	0.3	0.2	0.2	0.3	0.3	0.2	0.2	0.2	0.2	0.2	0.3
Market 120-179 lbs.	0.3	0.3	0.3	0.3	0.4	0.3	0.3	0.3	0.4	0.3	0.3	0.3	0.3	0.3	0.4	0.4
Market >180 lbs.	0.3	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4
Breeding Swine	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.4	0.3	0.3	0.3	0.3	0.3	0.3	0.3
Beef Cattle	**15.9**	**17.3**	**16.2**	**17.3**	**17.0**	**17.1**	**16.5**	**17.4**	**17.8**	**17.9**	**19.2**	**19.8**	**19.2**	**18.1**	**18.5**	**18.6**
Feedlot Steers	10.6	11.5	11.0	11.6	11.3	11.2	10.7	11.1	11.4	11.3	12.1	12.5	12.2	11.5	11.8	12.0
Feedlot Heifers	5.2	5.8	5.2	5.7	5.7	5.9	5.8	6.4	6.5	6.6	7.1	7.3	7.0	6.6	6.7	6.6
Sheep	**0.4**	**0.4**	**0.4**	**0.4**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**	**0.2**	**0.2**	**0.2**	**0.2**
Goats	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**
Poultry	**1.5**	**1.5**	**1.5**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**	**1.4**
Hens >1 yr.	0.4	0.4	0.4	0.3	0.3	0.3	0.3	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2	0.2
Total Pullets	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Chickens	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0	0.0
Broilers	0.6	0.6	0.7	0.7	0.7	0.7	0.8	0.8	0.8	0.8	0.8	0.9	0.9	0.9	0.9	0.9
Turkeys	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.3	0.2
Horses	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**	**0.6**	**0.7**	**0.7**	**0.7**	**0.7**	**0.7**

+ Emission estimate is less than 0.1 Gg.

3.11. Methodology for Estimating N_2O Emissions from Agricultural Soil Management

Nitrous oxide (N_2O) emissions from agricultural soils on managed lands result from the interaction of the natural processes of denitrification and nitrification with the management practices that add or release mineral nitrogen (N) in the soil profile. Emissions can occur directly in the soil where the N is made available or can be transported to another location following volatilization, leaching, or runoff, and then converted into N_2O. The *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997) focused on five specific practices: synthetic mineral N fertilization, organic amendments, crop residue N added to soils, cultivation of N fixing crops, and drainage and cultivation of organic cropland soils. The *IPCC Good Practice Guidance for LULUCF* (IPCC 2003) also recommended accounting for emissions resulting from decomposition of soil organic matter, particularly following land-use change. More recently, the *2006 IPCC Guidelines* (IPCC 2006) have taken the guidance a step further by recommending reporting of all emissions from managed land, because of close linkage between natural and anthropogenic factors and the difficulty of separating out the portion of emissions only influenced by anthropogenic activity (IPCC 2006, see Volume IV, Chapter 1). Given the recent advancements and improvements in methods and emission factors for soil N_2O, this inventory has adopted recommendations from IPCC (2006), which includes a number of changes from IPCC/UNEP/OECD/IEA (1997). For example, the emission factor for direct N_2O emissions from N inputs in IPCC/UNEP/OECD/IEA (1997) was too high according to a recent global analysis of experimental measurements (Bouwmann 2002; Rochette et al. 2004). Thus, the approaches described in this section are based on the original methods from IPCC/UNEP/OECD/IEA (1997), as amended in the IPCC Good Practice Guidance Documents (IPCC 2000, 2003), but also further refined in IPCC (2006).

A combination of Tier 1 and Tier 3 approaches was used to estimate direct and indirect N_2O emissions from agricultural soils. The process-based biogeochemical model DAYCENT (a Tier 3 approach) was applied to estimate N_2O emissions resulting from mineral soil croplands that were used to produce major crops, while the IPCC Tier 1 methodology was applied to estimate N_2O emissions for non-major crop types on mineral soils. The Tier 1 method was also used to estimate direct N_2O emissions due to drainage and cultivation of organic cropland soils. Direct N_2O emissions from grasslands were estimated by using a combination of DAYCENT and IPCC Tier 1 methods. A combination of DAYCENT and Tier 1 methods was also used to estimate indirect emissions from all managed lands. Specifically, the amount of N volatilized from soils and managed manure systems and leached or transported off-site in surface runoff waters was computed by DAYCENT for the direct emission analyses, while IPCC default factors were used to estimate N transport for the analyses using the Tier 1 methodology. The indirect N_2O emissions resulting from off-site transport of N were then computed using the Tier 1 method. Overall, the Tier 3 approach is used to estimate approximately 90% of total soil N_2O emissions associated with agricultural soil management in the United States.

DAYCENT (Del Grosso et al. 2001, Parton et al. 1998) simulates biogeochemical N fluxes between the atmosphere, vegetation, and soil, allowing for a more complete estimation of N_2O emissions than IPCC Tier 1 methods by accounting for the influence of environmental conditions including soil characteristics and weather patterns, specific crop and forage qualities that influence the N cycle, and management practices at a daily time step. For example, plant growth is controlled by nutrient availability, water, and temperature stress; moreover, growth removes mineral N from the soil before it can potentially be converted into N_2O. Nutrient supply is a function of external nutrient additions as well as soil organic matter (SOM) decomposition rates, and increasing decomposition can lead to greater N_2O emissions by enhancing mineral N availability in soils. In this model-based assessment framework, daily maximum/minimum temperature and precipitation, timing and description of management events (e.g., fertilization, tillage, harvest), and soil texture data are model inputs to DAYCENT, which form the basis to simulate key processes and generate a robust estimate of N_2O emissions from soils. Key processes simulated within sub-models of DAYCENT include plant production, organic matter formation and decomposition, soil water and soil temperature regimes by layer, nitrification and denitrification processes, and methane (CH_4) oxidation. Comparison of model results and plot level data show that DAYCENT reliably simulates crop yields, soil organic matter levels, and trace gas fluxes for a number of native and managed systems (Del Grosso et al. 2001, 2005).

There are five steps in estimating direct N_2O emissions from mineral cropland soils, drained and cultivated organic cropland soils, and grassland soils, in addition to indirect N_2O emissions from volatilization, leaching, and runoff from all managed lands. First, the activity data are derived from a combination of land-use, livestock, crop, and grassland management records, as well as expert knowledge. In the second, third, and fourth steps, direct N_2O emissions from croplands and grasslands, in addition to indirect N_2O emissions from all managed lands, are

estimated using DAYCENT and/or the Tier 1 method. In the fifth step, total emissions are computed by summing each component. The remainder of this sub-annex describes the methods underlying each step.

Step 1: Derive Activity Data

The activity data requirements vary for major crops, non-major crops, grasslands, and organic cropland soils. Activity data were derived for direct and indirect N_2O emission calculations as described below.

Step 1a: Activity Data for Direct Emissions from Crop Production on Mineral Soils

Nitrous oxide emissions from mineral cropland soils include emissions from both major and non-major cropping systems and were estimated using a Tier 3 and a Tier 1 approach, respectively.

Major Crop Types: Tier 3 DAYCENT Simulations

The activity data requirements for estimating N_2O emissions from major crop types (corn, soybeans, wheat, alfalfa hay, other hay, sorghum, and cotton) include the following: (1) crop-specific mineral N fertilizer rates and timing, (2) crop-specific manure amendment N rates and timing, (3) other N inputs, (4) crop-specific land management information, (5) native vegetation, (6) daily weather data for every county, (7) sub-county-level soil texture data, and (8) county-level crop areas. The United States was divided into 63 agricultural regions based on common cropping practices as defined by McCarl et al. (1993), and data were assembled and provided as inputs to the DAYCENT biogeochemical ecosystem model.

Unlike in the Tier 1 approach, N inputs from biological fixation and crop residues are not considered activity data in the DAYCENT analysis because N availability from these sources is simulated by the model based on environmental and management conditions, such as temperature, precipitation, and edaphic characteristics. That is, while the model accounts for the contribution of N from fixation and crop residues to the soil profile and subsequent N_2O emissions, these sources of mineral soil N are not activity data in the sense that they are not model inputs. Similarly, N from mineralization of soil organic matter and asymbiotic fixation are also simulated by the model.

Synthetic N Fertilizer Application: Data on N fertilizer rates were obtained primarily from U.S. Department of Agriculture–Economic Research Service 1995 Cropping Practices Survey (ERS 1997). In this survey, data on inorganic N fertilization rates were collected for major crops (corn, cotton, soybeans, potatoes, winter wheat, durum wheat, and other spring wheat) in the high production states during 1995. It is assumed that fertilization practices have not changed much during the inventory reporting period, and therefore the rates and uncertainties reflected in the 1995 survey data are considered representative for 1990 through 2005. Note that all wheat data were combined into one category and assumed to represent small grains in aggregate. Estimates for sorghum fertilizer rates were derived from corn fertilizer rates using a ratio of national average corn fertilizer rates to national average sorghum fertilizer rates derived from additional publications (NASS 2004, 1999, 1992; ERS 1988; Grant and Krenz 1985; USDA 1966, 1957, 1954).

The ERS survey parameter "TOT N" (total amount of N applied per acre), with a small number of records deleted as outliers, was used in determining the fraction of crop acres receiving fertilizer and the average fertilizer rates for each region. Mean fertilizer rates and standard deviations for irrigated and rainfed crops were produced for each state for which a minimum of 15 data points existed for irrigated and rainfed, respectively. If a state was not surveyed for a particular crop or if fewer than 15 data points existed for one of the categories, then data at the USDA Farm Production Region level was substituted (Farm Production Regions are groups of states in the United States with similar agricultural commodities). If Farm Production Region data were not available, then U.S. level estimates (all major states surveyed) were used in the simulation for the particular crop in the state lacking sufficient data. Standard deviations for fertilizer rates were used to construct probability distribution functions (PDFs) with log-normal densities in order to address uncertainties in application rates (see Step 2a for discussion of uncertainty methods). The total fertilizer application is found in Table A- 175.

Table A- 175: Synthetic Fertilizer N Added to Major Crops (Gg N)

Fertilizer Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Fertilizer N	8,323	8,293	8,631	8,190	8,495	8,068	8,694	8,489	8,450	8,199	8,354	8,098	8,269	7,775	7,805	7,719

Prior to 1990, estimates for crop-specific regional fertilizer rates were based largely on extrapolation/interpolation of fertilizer rates from the years with available data. For crops in some agricultural regions, little or no data were available, and, therefore, a geographic regional mean was used to simulate N fertilization rates (e.g., no data were available from Alabama later than 1970 for corn fertilization rates, and, therefore, mean values from the southeastern United States were used to simulate fertilization to corn fields in this state).

Managed Livestock Manure[44] *N Amendment Rates and Timing:* County-level manure addition estimates have been derived from manure N addition rates developed by the Natural Resources Conservation Service (NRCS, Edmonds et al. 2003). Working with the farm-level crop and animal data from the 1997 Census of Agriculture, NRCS has coupled estimates of manure N produced with estimates of manure N recoverability by animal waste management system to produce county-level estimates of manure N applied to cropland and pasture. Edmonds et al. (2003) defined a hierarchy that included 24 crops, cropland used as pasture, and permanent pasture. They estimated the area amended with manure and application rates in 1997 for both manure-producing farms and manure-receiving farms within a county and for two scenarios—before implementation of Comprehensive Nutrient Management Plans (baseline) and after implementation. For DAYCENT simulations, the baseline scenario estimates have been used and the rates for manure-producing farms and manure-receiving farms have been area-weighted and combined to produce a single county-level estimate for the amount of land amended with manure and the manure N application rate for each crop in each county. Several of the crops in Edmonds et al. (2003) have been area-weighted and combined into broader crop categories. For example, all small grain crops have been combined into one category. In order to address uncertainty in these data, uniform probability distributions were constructed based on the proportion of land receiving manure versus the amount not receiving manure for each crop type and pasture. For example, if the 20 percent of land producing corn in a county was amended with manure, randomly drawing a value equal to or greater than 0 and less than 20 would lead to simulation with a manure amendment, while drawing a value greater than or equal to 20 and less than 100 would lead to no amendment in the simulation (see Step 2a for further discussion of uncertainty methods).

Edmonds et al. (2003) only provide manure application rate data for 1997 and the amount of managed manure available for soil application changes annually, so the area amended with manure was adjusted relative to 1997 to account for all the manure produced in any other given year. Specifically, the manure N production in other years was divided by the manure N production in 1997. If the ratio was greater than 1, there was more manure N production in that county relative to the amount in 1997, and so it was assumed a larger area was amended with manure. In contrast, ratios less than one implied less area was amended with manure because there was a lower amount produced in the year compared to 1997. The amendment area in each county for 1997 was then multiplied by the ratio to reflect the impact of more or less manure N production on the area amended.

The amount of managed manure for each livestock type was calculated by determining the population of animals that were on feedlots or otherwise housed in order to collect and manage the manure. Annual animal population data for all livestock types, except horses and goats, were obtained for all years from the U.S. Department of Agriculture-National Agricultural Statistics Service (NASS) (USDA 1994b-c, 1995a-b, 1998a-b, 1999a-c, 2000a, 2004a-e, 2005a-d, 2006a). Horse population data were obtained from the FAOSTAT database (FAO 2006). Goat population data for 1992, 1997, and 2002 were obtained from the *Census of Agriculture* (USDA 2005g); these data were interpolated and extrapolated to derive estimates for the other years. Information regarding poultry turnover (i.e., slaughter) rate was obtained from state Natural Resource Conservation Service personnel (Lange 2000). Additional population data for different farm size categories for dairy and swine were obtained from the 1992 and 1997 *Census of Agriculture* (USDA 2005g). Information regarding the percentage of manure handled using various manure management systems for dairy cattle, beef cattle, and sheep was obtained from

[44] For purposes of the Inventory, total livestock manure is divided into two general categories: (1) managed manure, and (2) unmanaged manure. Managed manure includes manure that is stored in manure management systems such as pits and lagoons, as well as manure applied to soils through daily spread operations. Unmanaged manure encompasses all manure deposited on soils by animals on PRP.

communications with personnel from state Natural Resource Conservation Service offices, state universities, NASS, and other experts (Poe et al. 1999, Anderson 2000, Deal 2000, Johnson 2000, Miller 2000, Milton 2000, Stettler 2000, Sweeten 2000, Wright 2000), as well as the EPA Office of Water's engineering cost analyses conducted to support the development of effluent limitation guidelines for concentrated animal feeding operations (EPA 2002). Information regarding the percentage of manure handled using various manure management systems for swine, poultry, goats, and horses was obtained from site visits for EPA's Office of Water (ERG 2000a), EPA's *Global Methane Emissions from Livestock and Poultry Manure* (EPA 1992), and communications with other experts (Safley et al. 1992, Anderson 2000). A more detailed discussion of manure management system usage is provided in Annex 3.10. Once the animal populations for each livestock type and management system were estimated, these populations were multiplied by an average animal mass constant (USDA 1996, ASAE 1999; NRC 2000, ERG 2003, EPA 1992, Safley 2000) to derive total animal mass for each animal type in each management system. Total Kjeldahl N[45] excreted per year for each livestock type and management system was then calculated using daily rates of N excretion per unit of animal mass (USDA 1996, ASAE 1999). The annual amounts of Kjeldahl N were then summed over all livestock types and management systems to derive estimates of the annual managed manure N produced (Table A- 176). These data were then used to modify the area amended with manure relative to 1997, as discussed previously.

To estimate C inputs associated with manure N application rates derived from Edmonds et al. (2003), carbon-nitrogen (C:N) ratios for livestock-specific manure types were adapted from the Agricultural Waste Management Field Handbook (USDA 1996), On-Farm Composting Handbook (NRAES 1992), and recoverability factors provided by Edmonds et al (2003). The C:N ratios were applied to EPA county-level estimates of manure N excreted by animal type and management system to produce a weighted county average C:N ratio for manure amendments. The average C:N ratio was used to determine the associated C input for crop amendments derived from Edmonds et al. (2003).

To account for the common practice of reducing inorganic N fertilizer inputs when manure is added to a cropland soil, a set of crop-specific reduction factors were derived from mineral fertilization data for land amended with manure versus land not amended with manure in the ERS 1995 Cropping Practices Survey (ERS 1997). In the simulations, mineral N fertilization rates were reduced for crops receiving manure N based on a fraction of the amount of manure N applied, depending on the crop and whether it was irrigated or rainfed. The reduction factors were selected from PDFs with normal densities in order to address uncertainties in the dependence between manure amendments and mineral fertilizer application.

The amount of managed manure N available for application to soils was considerably greater than the amount of managed manure N that was actually applied, according to data provided by Edmonds et al. (2003). The remaining manure N that was not applied to soils was assumed to be volatilized or lost through leaching and runoff of water during storage, treatment, and transport, and thus contributes to indirect emissions. The fate of manure N is summarized in Table A- 176.

Table A- 176: Fate of Livestock Manure Nitrogen (Gg N)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Managed Manure N Applied to Major Crops	1,146	1,172	1,165	1,188	1,192	1,199	1,195	1,206	1,232	1,229	1,260	1,272	1,244	1,230	1,267	1,270
Manure N Lost from Management Systems[a]	1,539	1,584	1,581	1,613	1,628	1,642	1,627	1,677	1,692	1,700	1,718	1,739	1,780	1,754	1,737	1,752
Pasture, Range, & Paddock Manure N	3,879	3,895	3,977	4,017	4,116	4,167	4,159	4,030	3,941	3,887	3,817	3,780	3,762	3,756	3,712	3,741
Total	**6,564**	**6,652**	**6,723**	**6,818**	**6,936**	**7,009**	**6,980**	**6,913**	**6,865**	**6,816**	**6,796**	**6,791**	**6,787**	**6,740**	**6,717**	**6,763**

[a] Includes N volatilized and leached/runoff during treatment, storage and transport before soil application.

Residue N Inputs: Residue N input is estimated as part of the DAYCENT simulation and is not an input to the model. The simulated total N inputs of above- and belowground residue N and fixed N that was not harvested are provided in Table A-1.

[45] Total Kjeldahl N is a measure of organically bound N and ammonia N in both solid and liquid wastes.

Other N Inputs: Other N inputs are estimated within the DAYCENT simulation, and thus input data are not required, including mineralization from decomposition of soil organic matter and asymbiotic fixation of N from the atmosphere. These additional inputs of N are addressed to include all emissions from managed lands, as recommended by IPCC (2006). The simulated total N inputs from other sources are provided in Table A-1.

Table A-1: Crop Residue N and Other N Inputs to Major Crops as Simulated by DAYCENT (Gg N)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Residue N[a]	4,696	4,348	4,897	5,005	4,622	4,882	4,855	4,912	4,675	5,162	4,791	4,579	4,396	4,940	4,511	4,817
Other N Inputs[b]	24,696	23,565	24,432	24,843	23,170	23,209	23,987	23,804	25,642	22,660	24,171	23,506	22,944	23,248	21,493	24,700

[a] Residue N inputs include unharvested fixed N from legumes as well as crop residue N.
[b] Other N inputs include mineralization from decomposition of soil organic matter as well as asymbiotic fixation of N from the atmosphere.

Crop Rotation and Land Management Information: Data for the 63 agricultural regions were obtained for specific timing and type of cultivation, timing of planting/harvest, and crop rotation schedules (Hurd 1930, 1929, Latta 1938, Iowa State College Staff Members 1946, Bogue 1963, Hurt 1994, USDA 2000b, 2000c, CTIC 1998, Piper et al. 1924, Hardies and Hume 1927, Holmes 1902, 1929, Spillman 1902, 1905, 1907, 1908, Chilcott 1910, Smith 1911, Kezer ca 1917, Hargreaves 1993, ERS 2002, Warren 1911, Larson et al. 1922, Russell et al. 1922, Elliott and Tapp 1928, Elliott 1933, Ellsworth 1929, Garey 1929, Holmes 1929, Hodges et al. 1930, Bonnen and Elliott 1931, Brenner et al. 2002, 2001, Smith et al. 2002). As with N fertilizer and manure additions, data were not complete, so regional averages were used to fill spatial gaps in the data sets and interpolation/extrapolation was used to fill temporal gaps.

Native Vegetation by County: Pre-agricultural land cover for each county was designated according to the potential native vegetation used in the Vegetation-Ecosystem Modeling and Analysis Project (VEMAP 1995) analysis, which was based on the Kuchler (1964) Potential Vegetation Map for the conterminous United States.

Daily Weather Data by County: Daily maximum/minimum temperature and precipitation data were obtained from the DAYMET model, which generates daily surface precipitation, temperature, and other meteorological data at 1 km^2 resolution driven by weather station observations and an elevation model (Thornton et al. 2000, 1997, Thornton and Running, 1999; DAYMET, no date). It is necessary to use computer-generated weather data because weather station data do not exist in each county and, moreover, even fewer are located in agricultural lands. Weather station data are for a point in space, and the DAYMET modeling process uses this information with interpolation algorithms to derive weather patterns for areas between these stations. DAYMET weather data are available for the United States at 1 km^2 resolution for 1980 through 2003. For each county, DAYMET weather data was selected from the 1 km^2 cell that occurred in agricultural lands according the National Land Cover Dataset (Vogelman et al. 2001). The grid cells formed the basis for county-scale PDFs based on the frequency of cells with same weather patterns. (Note: separate PDFs were developed for cropland, pasture/hay land, and rangeland.) A weather record was then randomly selected from the PDFs in each iteration of the Monte Carlo analysis to represent variation in precipitation and temperature at the county scale. Weather data from 2003 was assumed to represent 2004 and 2005 as well.

Soil Properties by County: Soil texture data required by DAYCENT were obtained from STATSGO (Soil Survey Staff, Natural Resources Conservation Service, 2005). Observed data for soil hydraulic properties needed for model inputs were not available, so they were calculated from STATSGO (Soil Survey Staff, Natural Resources Conservation Service, 2005) texture class and Saxton et al.'s (1986) hydraulic properties calculator. Similar to the weather data, soil types within the STATSGO map that occurred in agricultural lands according to the National Land Cover Dataset (Vogelman et al. 2001) were used to form a county-scale PDF. Specifically, the PDFs were an area-weighted proportion for the extent of overlap between map units and agricultural land. (Note: separate PDFs were developed for cropland, pasture/hay land, and rangeland.) Individual soil types were randomly selected from the PDFs during each iteration of the Monte Carlo analysis to represent variation in soil texture and depth at the county scale.

Crop Areas by County: County-level total crop area data were downloaded from the NASS web site for the years 1990 through 2004 (USDA 2005h), and these data formed the basis to scale emissions from individual crop types to an entire county.

Non-Major Crop Types: Tier 1 Method

The activity data required for calculating emissions from non-major crop types include: (1) the amount of mineral N in synthetic fertilizers that are applied annually, (2) the amount of N in non-manure organic commercial fertilizers, and (3) the amount of N in the above- and below-ground residue retained on and in soils of all non-major crops.

Application of Synthetic Commercial Fertilizers: A process-of-elimination approach was used to estimate N fertilizer additions to non-major crop types. N fertilizer additions to major crops, grasslands, forests, and settlements were summed, this sum was subtracted from total annual fertilizer sales, and the difference was assumed to be applied to non-major crop types. Total fertilizer application is found in Table A- 177.

Table A- 177: Synthetic Fertilizer N Added to Non-Major Crops (Gg N)

Fertilizer Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Fertilizer N	762	945	670	1,438	1,527	1,619	1,305	1,514	1,562	1,832	1,405	1,367	1,474	2,281	2,549	2,332

Other Commercial Organic Fertilizers:[46] Estimates of total national annual N additions of other organic fertilizers were derived from organic fertilizer statistics (TVA 1991, 1992a, 1993, 1994; AAPFCO 1995, 1996, 1997, 1998, 1999, 2000a, 2000b, 2002, 2003, 2004, 2005, 2006). The organic fertilizer data, which are recorded in mass units of fertilizer, had to be converted to mass units of N by multiplying the consumption values by the average organic fertilizer N contents provided in the annual fertilizer publications. These N contents are weighted average values, and vary from year-to-year (ranging from 2.3 percent to 3.9 percent over the period 1990 through 2005). The fertilizer consumption data are recorded in "fertilizer year" totals, (i.e., July to June), but were converted to calendar year totals. This was done by assuming that approximately 35 percent of fertilizer usage occurred from July to December and 65 percent from January to June (TVA 1992b). July to December values were not available for calendar year 2005, so a "least squares line" statistical extrapolation using the previous 14 years of data was used to arrive at an approximate value. Annual consumption of other organic fertilizers is presented in Table A- 178.

Table A- 178: Other Organic Commercial Fertilizer Consumption on Agricultural Lands (Gg N)

Fertilizer Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Other Commercial Organic Fertilizer N*	8	13	12	11	13	16	18	18	18	19	20	18	15	14	14	17

* Includes dried blood, dried manure, tankage, compost, other. Excludes manure and sewage sludge used as commercial fertilizer.

Retention of Crop Residue: Annual crop yield (metric tons per hectare) and area planted (hectare) statistics for non-major N-fixing crops, including bean and pulse crops, were taken from U.S. Department of Agriculture crop production reports (USDA 1994a, 1998b, 2003, 2005i, 2006b). Crop yield per hectare and area planted were multiplied to determine total crop yield for each crop, which was then converted to tons of dry matter product using the residue dry matter fractions shown in Table A- 179. Dry matter yield was then converted to tons of above- and below-ground biomass N. Above-ground biomass was calculated by using linear equations to estimate above-ground biomass given dry matter crop yields and below-ground biomass was calculated by multiplying above-ground biomass by the below-to-above-ground biomass ratio. N inputs were estimated by multiplying above- and below-ground biomass by respective N concentrations. All ratios and equations used to calculate residue N inputs (shown in Table A- 179) are from IPCC (2006) and Williams (2006). The resulting annual biomass N inputs are presented in Table A- 180.

Table A- 179: Key Assumptions for Production of Non-Major Crops and Retention of Crop Residues

Crop	Dry Matter Fraction of Harvested Product	Above-ground Residue		Ratio of Below-ground Residue to Above-ground Biomass	Residue N Fraction	
		Slope	Intercept		Above-ground	Below-ground
Peanuts for Nuts	0.94	1.07	1.54	0.20	0.016	0.014
Dry Edible Beans	0.90	0.36	0.68	0.19	0.010	0.010
Dry Edible Peas	0.91	1.13	0.85	0.19	0.008	0.008

[46] Other commercial organic fertilizers include dried blood, dried manure, tankage, compost, other, but exclude manure and sewage sludge used as commercial fertilizers.

Austrian Winter Peas	0.91	1.13	0.85	0.19	0.008	0.008
Lentils	0.91	1.13	0.85	0.19	0.008	0.008
Wrinkled Seed Peas	0.91	1.13	0.85	0.19	0.008	0.008
Barley	0.89	0.98	0.59	0.22	0.007	0.014
Oats	0.89	0.91	0.89	0.25	0.007	0.008
Rye	0.88	1.09	0.88	0.22	0.005	0.011
Millet	0.90	1.43	0.14	0.22	0.007	0.009
Rice	0.89	0.95	2.46	0.16	0.007	0.009

Table A- 180: Nitrogen in Crop Residues Retained on Soils Producing Non-Major Crops (Gg N)

Product Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Peanuts for Nuts	65	78	71	63	71	63	65	64	69	69	66	75	63	70	72	77
Dry Edible Beans	17	17	14	14	16	17	16	16	16	17	15	14	17	14	13	15
Dry Edible Peas	9	10	9	10	9	11	10	11	11	11	10	10	11	11	14	16
Austrian Winter Peas	8	8	8	8	8	8	8	8	8	8	8	8	8	8	8	8
Lentils	9	9	9	9	9	9	9	9	9	9	10	10	9	9	10	11
Wrinkled Seed Peas	9	9	8	9	8	9	8	8	8	8	8	8	8	8	9	8
Barley	122	130	128	121	107	102	110	104	101	80	95	78	75	85	85	69
Oats	91	66	77	65	58	53	43	48	47	45	46	44	45	47	43	43
Rye	17	16	17	16	17	16	15	15	17	17	15	15	14	15	15	16
Millet	7	7	7	7	7	7	7	7	7	7	4	8	3	5	7	6
Rice	82	84	90	82	97	88	87	91	92	100	94	103	102	98	110	107
Total	**434**	**433**	**439**	**405**	**409**	**383**	**379**	**382**	**387**	**372**	**371**	**373**	**356**	**372**	**386**	**376**

Step 1b: Activity Data for Direct Emissions from Drainage and Cultivation of Organic Cropland Soils

Tier 1 Method

Estimates of the areas of drained and cultivated organic cropland soils in 1982, 1992, and 1997 were obtained from the U.S. Department of Agriculture *1997 National Resources Inventory* (USDA 2000b, as extracted by Eve 2001, and revised by Ogle 2002).[47] These areas were grouped by broad climatic region[48] using temperature and precipitation estimates from Daly et al. (1994, 1998), and then further aggregated to derive total land in temperate and sub-tropical regions. These final areas were then linearly interpolated and extrapolated to obtain estimates not reported in the NRI between 1990 and 2005 (Table A- 181).

Table A- 181: Cultivated Organic Soil Area (Thousand Hectares)

Year	Temperate Area	Sub-Tropical Area
1990	432	192
1991	431	193
1992	429	194
1993	431	194
1994	433	195
1995	435	195
1996	437	196
1997	439	196
1998	441	197
1999	443	197
2000	445	197
2001	447	198
2002	449	198
2003	451	199
2004	453	199
2005	455	199

[47] These areas do not include Alaska, but Alaska's cropland area accounts for less than 0.1 percent of total U.S. cropland area, so this omission is not significant.

[48] These climatic regions were: (1) cold temperate, dry, (2) cold temperate, moist, (3) sub-tropical, dry, (4) sub-tropical, moist, (5) warm temperate, dry, and (6) warm temperate, moist.

Step 1c: Activity Data for Direct Emissions from Grassland Management

N_2O emissions from grasslands were computed using DAYCENT and the Tier 1 methodology. DAYCENT simulations addressed the influence of legume seeding, managed manure N amendments (i.e., not pasture, range, and paddock [PRP] manure), and synthetic fertilizer applications, in addition to the unmanaged manure N that was excreted by livestock and deposited directly onto soils (i.e., PRP manure). PRP manure N additions that were not included in the DAYCENT simulations and sewage sludge amendments to agricultural soils were addressed using the Tier 1 method.

Tier 3 DAYCENT Simulations

Activity data for DAYCENT simulations of grasslands (i.e., climate, soils, and N inputs) were based on the same sources as those used for major crop types described in Step 1a, except county-level area data on privately-owned pasture and rangeland areas (i.e., not federal) from the National Resources Inventory (USDA 2000b). A key source of N for grasslands is PRP manure N deposition. In the DAYCENT model, N additions to soils from grazing animal waste are simulated as a function of forage growth rates and nutrient contents, grazing intensity, soil texture, weather, and other factors that interact to control N flows among plants, animals, and soil. N fixation by legumes, and N residue inputs from senesced grass litter were included as sources of N to the soil, which are estimated as a function of vegetation type, weather, and soil properties. Similar to the methodology for major crops, "other N inputs" were simulated within the model framework in order to estimate soil N_2O emissions from managed lands (IPCC 2006), including mineralization from decomposition of soil organic matter and litter, as well as asymbiotic N fixation from the atmosphere. Decomposition rates are a function of litter quality and quantity, soil texture, water content and temperature, and other factors. Total annual amounts of PRP manure N, mineral N fertilizer application, manure N amendments, forage legume N and "other N inputs" can be found in Table A- 182.

Table A- 182: Synthetic Fertilizer N, PRP Manure N, Organic Manure N Amendment, Forage Legume N, and Other N Inputs Simulated with the DAYCENT Model (Gg N)

Fertilizer Type	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Fertilizer N	8,323	8,293	8,631	8,190	8,495	8,068	8,694	8,489	8,450	8,199	8,354	8,098	8,269	7,775	7,805	7,719
PRP Manure N	2,917	2,914	3,284	3,301	2,749	3,261	2,378	3,276	2,872	3,172	2,641	2,807	2,146	2,716	2,482	2,905
Managed Manure	1,146	1,172	1,165	1,188	1,192	1,199	1,195	1,206	1,232	1,229	1,260	1,272	1,244	1,230	1,267	1,270
Residue N[a]	4,696	4,348	4,897	5,005	4,622	4,882	4,855	4,912	4,675	5,162	4,791	4,579	4,396	4,940	4,511	4,817
Other N Inputs[b]	24,696	23,565	24,432	24,843	23,170	23,209	23,987	23,804	25,642	22,660	24,171	23,506	22,944	23,248	21,493	24,700

[a] Residue N inputs include unharvested fixed N from legumes as well as crop residue N.
[b] Other N inputs include mineralization from decomposition of soil organic matter and litter, as well as asymbiotic fixation of N from the atmosphere.

Tier 1 Method: Additional Direct Soil N_2O Emissions

The Tier 1 method was used to estimate emissions from PRP manure that were not simulated with DAYCENT, in addition to emissions due to sewage sludge amendments to agricultural soils.

PRP Manure: Manure N additions from grazing animals were modeled within DAYCENT, but the simulations only captured approximately 70 percent of total manure production in this category. It is reasonable that DAYCENT did not account for all PRP manure, because the NRI data does not include some grassland areas such as federal grasslands. Soil N_2O emissions from the remaining manure N were estimated using Tier 1 methods. Activity data for PRP manure N excretion from dairy cattle, beef cattle, swine, sheep, goats, poultry, and horses, were derived from multiple sources as described in the following text.

Information regarding dairy farm grazing was obtained from communications with personnel from state Natural Resource Conservation Service offices, state universities, and other experts (Poe et al. 1999, Deal 2000, Johnson 2000, Miller 2000, Stettler 2000, Sweeten 2000, Wright 2000). Because grazing operations are typically related to the number of animals on a farm, farm-size distribution data reported in the *1992, 1997, and 2002 Census of Agriculture* (USDA 2005g) were used in conjunction with the state data obtained from personal communications to determine the percentage of total dairy cattle that graze. An overall percent of dairy waste that is deposited in PRP was developed for geographic regions of the United States. These percentages were applied to the total annual dairy cow and heifer state population data for 1990 through 2005, which were obtained from NASS (USDA 1995a, 1999a, 2004a, 2005a-b, 2006a).

To determine the population of beef cattle that are on PRP, the following assumptions were made: (1) beef cows, bulls, and calves were not housed on feedlots; (2) a portion of heifers and steers were on feedlots; and (3) all beef cattle that were not housed on feedlots were located on PRP (i.e., total population minus population on feedlots equals population on PRP) (Milton 2000). Information regarding the percentage of heifers and steers on feedlots was obtained from USDA personnel (Milton 2000) and used in conjunction with NASS population data (USDA 1995a, 1999a, 2004a, 2005a, 2006a) to determine the population of steers and heifers on PRP.

Based on the assumption that smaller facilities are less likely to utilize manure management systems, farm-size distribution data reported in the *1992, 1997, and 2002 Census of Agriculture* (USDA 2005g) were used to determine the percentage of all swine whose manure is not managed (i.e., the percentage on PRP). These percentages were applied to the average of the quarterly NASS population data for swine (USDA 1994b, 1998c, 2004b, 2005c, 2006a) to determine the population of swine on PRP.

Total sheep and lamb population data were obtained from NASS (USDA 1994c, 1999c, 2004e, 2005f). To determine the number of sheep and lamb in managed systems, information on the number of sheep and lamb on feed were obtained from U.S. Department of Agriculture for 1990 through 1993 (USDA 1994c). However, population data for lamb and sheep on feed were not available after 1993, so the number of lamb and sheep on feed for 1994 through 2005 were estimated using the average of the percent of lamb and sheep on feed from 1990 through 1993. In addition, all of the sheep and lamb on feed were not necessarily managed on feedlots; they may have been unmanaged on pasture/crop residue supplemented by feed. To estimate the portion of on feed animals that are on PRP, data were obtained from U.S. Department of Agriculture for lambs only in 1993 (USDA 1994c). To calculate the number of sheep and lamb on feedlots for all years, it was assumed that the percentage of sheep and lamb on feedlots versus pasture/crop residue is the same as that for lambs in 1993 (Anderson 2000). It was assumed that all sheep and lamb manure not deposited on feedlots was deposited on PRP (Anderson 2000).

It was assumed that 92 percent of goat manure was deposited on PRP (Safley et al. 1992). Annual goat population data by state were available for only 1992, 1997, and 2002 (USDA 2005e). The data for 1992 were used for 1990 through 1992 and the data for 2002 were used for 2002 through 2005. Data for 1993 through 1996 and 1998 through 2001 were linearly interpolated using the 1992, 1997, and 2002 data. In addition, it was assumed that one percent of poultry manure for broilers and turkeys was deposited on PRP (Safley et al. 1992). Poultry population data were obtained from NASS (USDA 1995b, 1998a, 1999b, 2000a, 2004c-d, 2005d-e, 2006a). The annual population data for boilers and turkeys were adjusted for turnover (i.e., slaughter) rate (Lange 2000). Lastly, it was assumed that 92 percent of horse manure was deposited on PRP (Safley et al. 1992). Horse population data were obtained from the FAOSTAT database (FAO 2006).

For each animal type, the population of animals within PRP systems was multiplied by an average animal mass constant (USDA 1996, ASAE 1999; NRC 2000, ERG 2003, EPA 1992, Safley 2000) to derive total animal mass for each animal type. Total Kjeldahl N excreted per year was then calculated for each animal type using daily rates of N excretion per unit of animal mass (USDA 1996, ASAE 1999). Annual N excretion was then summed over all animal types to yield total N in PRP manure (Table A- 176).

Sewage Sludge Amendments: Sewage sludge is generated from the treatment of raw sewage in public or private wastewater treatment works and is typically used as a soil amendment or is sent to waste disposal facilities such as landfills. Estimates of the amounts of sewage sludge N applied to agricultural lands were derived from national data on sewage sludge generation, disposition, and N content. Total sewage sludge generation data for 1988, 1996, and 1998, and a projection for 2000, in dry mass units, were obtained from EPA reports (EPA 1993, 1999), and linearly interpolated to estimate values for the intervening years. Sewage sludge generation data are not available for 2001 through 2005 (Bastian 2002, 2003, 2005, 2006), so the 2000 projection was linearly extrapolated using the growth in national wastewater flow between 1996 and 2000 (EPA 1997, 2003). The total sludge generation estimates were then converted to units of N by applying an average N content of 3.3 percent (Metcalf and Eddy 1991), and disaggregated into use and disposal practices using historical data and projections in EPA (1993) and EPA (1999). The use and disposal practices were agricultural land application, other land application, surface disposal, incineration, landfilling, ocean dumping (ended in 1992), and other disposal. The resulting estimates of sewage sludge N applied to agricultural land were used here; the estimates of sewage sludge N applied to other land and surface-disposed were used in estimating N_2O fluxes from soils in *Settlements Remaining Settlements* (see section 7.5 of the Land Use, Land-Use Change, and Forestry chapter). Both of these data sets are presented in Table A- 183.

Table A- 183: Sewage Sludge Nitrogen by Disposal Practice (Gg N)

Disposal Practice	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Applied to Agricultural Soils	51	58	65	72	78	85	87	89	90	93	97	100	103	106	109	113
Other Land Application	27	30	34	37	41	44	45	46	47	49	51	52	54	55	57	59
Surface Disposal	16	15	14	12	10	9	9	9	9	8	8	8	8	8	8	8
Total	**94**	**103**	**112**	**121**	**130**	**138**	**141**	**144**	**146**	**150**	**155**	**160**	**164**	**169**	**174**	**179**

Step 1d: Activity Data for Indirect N_2O Emissions from Managed Soils of all Land-Use Types and Managed Manure Systems

Volatilization leads to emissions of NH_3 and NO_x to the atmosphere from N that was applied or deposited as synthetic fertilizer, livestock manure, sewage sludge, and other organic amendments. In turn, this N is returned to soils through atmospheric deposition, thereby increasing mineral N availability and enhancing N_2O production. Additional N is lost from soils through leaching as water percolates through a soil profile and through runoff with overland water flow. These losses of N enter groundwater and waterways, from which a portion is emitted as N_2O. The activity data for synthetic fertilizer, livestock manure, other organic amendments, residue N inputs, sewage sludge N, and other N inputs are the same as those used in the calculation of direct emissions from agricultural mineral soils, and may be found in Table A- 175 through Table A- 178, Table A- 180, and Table A- 183. Volatilization and leaching/runoff of N from manure during storage, treatment, and transport was also considered (Table A- 176). The activity data for computing direct N_2O emissions from settlements and forest lands are described in the Land Use, Land-Use Change, and Forestry Chapter (Chapter 7).

Using the DAYCENT model, volatilization and leaching/surface run-off of N from soils was computed internally for major crop types and grasslands. DAYCENT simulates the processes leading to these losses of N based on environmental conditions (i.e., weather patterns and soil characteristics), management impacts (plowing, irrigation, harvest, etc.), and soil N availability. Note that the DAYCENT method accounts for losses of N from all anthropogenic activity, not just the inputs of N from mineral fertilization and organic amendments, which are addressed in the Tier 1 methodology. Similarly, the N available for producing indirect emissions resulting from grassland management as well as deposited PRP manure was also calculated by DAYCENT. Volatilized losses of N were summed for each day in the annual cycle to provide an estimate of the amount of N subject to indirect N_2O emissions. In addition, the daily losses of N through leaching and runoff in overland flow were summed for the annual cycle.

The Tier 1 method was used to estimate N losses from mineral soils due to volatilization and leaching/runoff for non-major crop types, PRP manure not accounted for by DAYCENT, forest lands, and settlements. To estimate volatilized losses, synthetic fertilizers, manure, sewage sludge, and other organic N inputs were multiplied by the fraction subject to gaseous losses using the respective 1996 IPCC default values (0.1 kg N/kg N added as mineral fertilizers, and 0.2 kg N/kg N added as manure; IPCC/UNEP/OECD/IEA 1997). Leaching/runoff losses of N were estimated by summing the N additions from synthetic and other organic fertilizers, manure, sewage sludge, and above- and belowground crop residues, and then multiplying by the 1996 IPCC default fraction subject to leaching/runoff losses (0.3 kg N/kg N applied; IPCC/UNEP/OECD/IEA 1997). Furthermore, the estimates of volatilized N losses included some N from managed manure not applied to crops (or pastures), which was assumed to volatilize during storage, treatment and transport and later be added to soils through atmospheric deposition. Managed manure N lost during storage, treatment, and transport could be volatilized or lost through leaching and runoff, but there is insufficient information to determine the amount of N loss through each pathway. Consequently, losses are assumed to occur through volatilization only as a conservative estimate of indirect N_2O emissions, because the emission factor for volatilization is slightly higher than for leaching/runoff (Brumme et al. 1999, Butterbach-Bahl et al. 1997, Cloughet al. 2006, Corre et al. 1999, Denier van der Gon and Bleeker 2005, Dong et al. 2004, Hiscock et al. 2002, 2003, IPCC/UNEP/OECD/IEA 1997, Reay et al. 2004, 2005, Sawamoto et al. 2005, Williams-Jacobse 2002, IPCC 2006).

Volatilized N from major crop types, minor crop types, grasslands, settlements, and forest lands were added to the N volatilized during manure storage and handling to obtain the total annual losses for this pathway. Similarly, the annual amounts of N lost from soil profiles through leaching and surface runoff were summed to obtain the total losses for this pathway.

Step 2: Estimate Direct N_2O Emissions from Cropland Soils

In this step, N_2O emissions were calculated for direct N_2O emissions due to the N additions and cultivation of major crop types, N additions to non-major crop types, and direct N_2O emissions due to drainage and cultivation of organic soils.

Step 2a: Direct N_2O Emissions from Cropland Mineral Soils

Two methods were used to estimate direct N_2O emissions from N additions and crop production on mineral soils. The DAYCENT ecosystem model was used to estimate emissions from major crop types, while the Tier 1 methodology was used to estimate emissions from crops considered non-major types, which are grown on a considerably smaller portion of land than the major types.

Major Crops: Tier 3 DAYCENT Simulations

Simulations were performed over three major time periods for each county in the United States using the DAYCENT model. The first time period was used for simulation of native vegetation up to date of cultivation in the county (1 A.D. to plow out). Plow out was assumed to occur between 1600 and 1850, depending on the state in which the county lies. Simulation of at least 1600 years of native vegetation was needed to initialize soil organic matter (SOM) pools in the model. The second time period of the simulation started at plowout and represents historical agricultural practices up to the modern period (plow out to 1970). Simulation of the historical cropping period was needed to establish modern day SOM levels, which is important because N_2O emissions are sensitive to the amount of SOM. Lastly, simulations were performed for modern agricultural period (1971 through 2005).

Corn, soybeans, wheat, alfalfa hay, other hay, sorghum, and cotton are defined as major crops and were simulated in every county where they were grown. These crops represent approximately 90 percent of total principal cropland in the United States. Principal crop types, as defined by NASS (USDA 2003), include all grain, hay and row crops as well as vegetables for processing, but not commercial vegetable crops or orchards. For rotations that include a cycle that repeats every two or more years (e.g., corn/soybeans, wheat/corn/fallow) different simulations were performed where each phase of the rotation was simulated every year. For example, in regions where wheat/corn/fallow cropping is used, 3 rotations were simulated: one with wheat grown the first year, a second with corn the first year and a third with fallow the first year. This ensured that each crop was represented during each year in one of the three simulations. In cases where the same crop was grown in the same year in two or more distinct rotations for a region, N_2O emissions were averaged across the different rotations to obtain a value for that crop. Emissions from cultivated fallow land were also included. Fallow area was assumed to be equal to winter wheat area in regions where winter wheat/fallow rotations are the dominant land management for winter wheat.

The simulations reported here assumed conventional tillage cultivation, gradual improvement of cultivars, and gradual increases in fertilizer application until 1989. We accounted for improvements of cultivars (cultivated varieties), because it is unrealistic to assume that modern corn is identical to corn grown in 1900 in terms of yield potential, N demand, etc. Realistic simulations of historical land management and vegetation type are important because they influence present day soil C and N levels, which influence present day N cycling and associated N_2O emissions. These simulations included approximately 90 percent of principal cropland area and approximately 86 percent of total cropped area. Total cropped area includes principal crops, perennial crops (e.g., fruit and nut trees), and commercial vegetables.

Uncertainty estimation was an integral part of this analysis; uncertainty with the input data for the county-scale simulations and structural uncertainty associated with the DAYCENT model were both addressed. In the first step of the uncertainty assessment, a Monte Carlo Analysis was used to propagate input data uncertainty through the modeling process. Thus, input data were randomly selected from PDFs for weather records, soil type, mineral N fertilization rate, and organic amendments. See Step 1A for additional discussion about the PDFs. After selecting a set of inputs for a county, DAYCENT was used to simulate each crop and then the process was repeated until 100 iterations were completed. Due to the computationally intensive requirements for DAYCENT, it was not possible to simulate every county with the Monte Carlo Analysis. Two counties were selected from each of the 63 agricultural regions and additional counties were added based on the variance in N_2O emissions across regions from the past year's inventory, using a Neyman allocation (Cochran 1977). A Neyman allocation is based on the variance in N_2O emissions across the 63 regions; regions with larger variances are allocated a larger number of counties for the Monte Carlo Analysis. A total of 300 counties were included in the Monte Carlo Analysis, which is approximately

10 percent of all counties in the analysis. In addition, all counties were simulated once based on the dominant conditions from the PDFs (i.e., most common soil type, weather condition, manure amendment, and mineral fertilizer rate).

In the second step of the uncertainty analysis, the uncertainty inherent in model formulation and parameterization was determined. An empirically-based procedure was used to quantify those uncertainties as described by Ogle et al. (2007), in which an uncertainty estimator was developed based on the relationship between modeled results and field measurements. DAYCENT was used to simulate 8 agricultural experiments with 50 treatments. A linear-mixed effect model was developed in which measurements were statistically modeled as a function of modeled emissions. DAYCENT results were a highly significant predictor of the measurements, with a p-value of <0.01. Several other variables were included in the model to determine if DAYCENT results were more accurate under certain conditions, such as climate or soil types, but no significant relationships were found at an alpha level of 0.05. Random effects were included in the model to capture the dependence in time series and data collected from the same site, which were needed to estimate appropriate standard deviations for parameter coefficients.

The structural uncertainty estimator accounted for bias and prediction error in the DAYCENT model results, as well as random error associated with fine-scale emission predictions in counties over a time series from 1990 to 2005. To apply the uncertainty estimator, the DAYCENT results were used as an input to the linear mixed effect model after randomly selecting parameter coefficients from their joint probability distribution, in addition to random draws from PDFs representing the uncertainty due to site and site by year random effects.

In DAYCENT, once N enters the plant/soil system, the model cannot distinguish among the original sources of the N to determine which management activity led to specific N_2O emissions. This means, for example, that N_2O emissions from applied synthetic fertilizer cannot be separated from emissions due to N inputs from crop residue strictly using model outputs. It is desirable, however to report emissions associated with specific practices to the extent possible. Thus, for each crop in a county, the N inputs during the simulation were determined for anthropogenic practices discussed in IPCC/UNEP/OECD/IEA (1997), including synthetic mineral N fertilization, organic amendments, and crop residue N added to soils (including N-fixing crops). The percentage of N input for anthropogenic practices was divided by the total N input, and this proportion was used to determine the amount of N_2O emissions assigned to each of the practices.[49] For example, if 70 percent of the mineral N made available in the soil was due to mineral fertilization, then 70 percent of the N_2O emissions were assigned to this practice. The remainder of the soil N_2O emissions is reported under "other N inputs," which includes mineralization due to decomposition of soil organic matter and litter, as well as asymbiotic fixation of mineral N in soils from the atmosphere. Asymbiotic N fixation by soil bacteria is a minor source of N, typically not exceeding 10 percent of total N inputs to agroecosystems. Mineralization of soil organic matter is a more significant source of N, but is still typically less than half of the amount of N made available in the soil compared to fertilization, manure amendments, and symbiotic fixation. Accounting for the influence of "other N inputs" was necessary in order to meet the recommendation of reporting all emissions from managed lands (IPCC 2006).

In order to obtain a final estimate, the emissions were summed for each practice as simulated for the dominant conditions. In a second step, the difference was computed on a county basis between the mean emissions from the Monte Carlo and dominant condition simulation. This difference was used to adjust the result from simulations based on the dominant condition (i.e., if the Monte Carlo mean was slightly higher than the dominant condition, the total would be raised by the difference) (Del Grosso et al., In prep). Values were aggregated to the 63 region scale, and then summed to the entire nation. The uncertainty was based on the variance in simulated N_2O emissions for the iterations in the Monte Carlo Analysis and the variance associated with difference between the means from the Monte Carlo Analysis and the simulated N_2O emissions for the dominant condition, expressed as a 95 percent confidence interval.

[49] This method is a simplification of reality to allow partitioning of N_2O emissions. It is assumed that all N inputs have an identical chance of being converted to N_2O. This is unlikely to be the case, but DAYCENT does not track mineral N by source category so this approximation is the only approach that can be used for approximating the portion of N_2O emissions by source of N input.

Non-Major Crops: Tier 1 Method

To estimate direct N_2O emissions from N additions to non-major crops, the amount of applied synthetic and other organic fertilizer (i.e., dried blood, dried manure, tankage, compost, and other) N in each year was added to N inputs from crop residues, and the total N was multiplied by the IPCC default emission factor of 0.01 kg N_2O-N/kg N (IPCC 2006, Bouwman et al. 2002a, 2002b, Novoa and Tejeda 2006, Stehfest and Bouwman 2006). The uncertainty was determined based on simple error propagation methods (IPCC 2006). The uncertainty in the default emission factor is 0.3-3.0 kg N_2O-N/kg N (IPCC 2006, Bouwman et al. 2002a, 2002b, Novoa and Tejeda 2006, Stehfest and Bouwman 2006). Uncertainty in activity data is ± 20 percent for fertilizer additions (Mosier 2004).[50] Uncertainties in the emission factor and fertilizer additions were combined with uncertainty in the equations used to calculate residue N additions from above- and below-ground biomass dry matter and N concentration to derive overall uncertainty.

Step 2b: Direct N_2O Emissions Due to Drainage and Cultivation of Organic Cropland Soils

To estimate annual N_2O emissions from drainage and cultivation of organic cropland soils, the area of cultivated organic soils in temperate regions was multiplied by the IPCC default emission factor for temperate soils (8 kg N_2O-N/ha cultivated (IPCC 2006, Klemedtsson et al. 1999, IPCC 2000)) and the corresponding area in the sub-tropical regions was multiplied by the average (12 kg N_2O-N/ha cultivated) of the temperate and tropical (16 kg N_2O-N/ha cultivated (IPCC 2006, Klemedtsson et al. 1999, IPCC 2000)) IPCC default emission factors. The uncertainty was determined based on simple error propagation methods (IPCC 2006). Uncertainty in the default emission factor is 2-24 kg N_2O-N/ha (IPCC 2006, Klemedtsson et al. 1999, IPCC 2000).

Step 2c: Estimate Total Direct N_2O Emissions from Cropland Soils

In this step, total direct N_2O emissions from cropland soils are calculated by summing direct emissions due to anthropogenic activity on mineral soils with emissions resulting from the drainage and cultivation of organic cropland soils (Table A- 184). Uncertainties were combined using the simple error propagation method (IPCC 2006).

Table A- 184: Direct N_2O Emissions from Cropland Soils (Tg CO_2 Eq.)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Mineral Soils	**219.3**	**217.2**	**217.5**	**233.9**	**223.5**	**211.4**	**252.6**	**242.8**	**262.7**	**228.6**	**247.6**	**249.7**	**231.1**	**223.5**	**217.9**	**231.2**
Major Crops	**213.4**	**210.4**	**212.1**	**224.8**	**214.0**	**201.6**	**244.4**	**233.5**	**253.1**	**217.7**	**238.8**	**241.2**	**222.1**	**210.5**	**203.6**	**217.9**
Synthetic Fertilizer	79.9	84.4	82.9	82.9	84.7	77.2	94.0	88.9	89.9	80.5	85.0	87.5	83.0	73.5	76.1	75.5
Managed Manure	10.3	11.2	10.5	11.3	11.2	10.9	12.1	12.0	12.3	11.4	12.0	12.8	11.9	11.2	11.6	11.6
Residue N[a]	12.9	11.7	15.1	15.4	13.0	13.9	17.0	16.4	14.4	16.6	16.7	14.8	13.4	16.5	12.8	14.1
Other N Inputs[b]	110.3	103.1	103.6	115.2	105.2	99.6	121.3	116.2	136.5	109.2	125.1	126.0	113.8	109.4	103.1	116.6
Non-Major Crops	**5.9**	**6.8**	**5.5**	**9.0**	**9.5**	**9.8**	**8.3**	**9.3**	**9.6**	**10.8**	**8.7**	**8.6**	**9.0**	**13.0**	**14.4**	**13.3**
Synthetic Fertilizer	3.7	4.6	3.3	7.0	7.4	7.9	6.4	7.4	7.6	8.9	6.8	6.7	7.2	11.1	12.4	11.4
Other Organic Commercial Fertilizer	+	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Residue N[a]	2.1	2.1	2.1	2.0	2.0	1.9	1.8	1.9	1.9	1.8	1.8	1.8	1.7	1.8	1.9	1.8
Organic Soils	**2.8**	**2.8**	**2.8**	**2.8**	**2.8**	**2.8**	**2.8**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**
Total	**222.1**	**220.0**	**220.3**	**236.7**	**226.4**	**214.2**	**255.5**	**245.7**	**265.6**	**231.4**	**250.5**	**252.6**	**234.0**	**226.4**	**220.9**	**234.2**

+ Less than 0.05 Tg CO_2 Eq.
[a] Residue N inputs include unharvested fixed N from legumes as well as crop residue N.
[b] Other N inputs include mineralization from decomposition of soil organic matter as well as asymbiotic fixation.

[50] Note that due to lack of data, uncertainties in managed manure N production, PRP manure N production, other organic fertilizer amendments, indirect losses of N in the DAYCENT simulations, and sewage sludge amendments to soils are currently treated as certain; every attempt will be made to include these sources of uncertainty in future Inventories.

Step 3: Estimate Direct N_2O Emissions from Grasslands

A combination of DAYCENT and the Tier 1 method was also used to estimate direct N_2O emissions from soils in grasslands (pastures and rangeland). Managed pastures were simulated with DAYCENT by assuming that the vegetation mix includes forage legumes and grasses and that grazing intensity was moderate to heavy. Rangelands were simulated without forage legumes and grazing intensity was assumed to be light to moderate. The methodology used to conduct the DAYCENT simulations of grasslands was similar to that for major crop types described above in Step 2a, including the analysis addressing uncertainty in the model inputs. The one exception is that the structural uncertainty associated with DAYCENT model predictions was not addressed because of limited measurement data from grasslands. Addressing this uncertainty is a planned improvement as more measurement data become available for grassland soil N_2O emissions.

The N excreted by livestock not accounted for by DAYCENT simulations of pasture and rangeland (~25 percent of total PRP manure) was multiplied by the IPCC (2006), IPCC/UNEP/OECD/IEA (1997) default emission factor (0.02 kg N_2O-N/kg N excreted). DAYCENT simulated emissions were added to the emissions estimated using the IPCC Tier 1 method to provide the national total for direct N_2O losses from grasslands (see Table A- 185). The uncertainty was determined based on the Tier 1 error propagation methods provided by the IPCC (2006). Uncertainty in the default emission factor is 0.007-0.06 kg N_2O-N/kg N (IPCC 2006).

Table A- 185: Direct N_2O Emissions from Grasslands (Tg CO_2 Eq.)

	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
DAYCENT	**78.3**	**82.3**	**76.1**	**78.7**	**71.0**	**68.5**	**78.7**	**68.6**	**69.1**	**61.5**	**62.0**	**64.8**	**63.9**	**60.4**	**58.7**	**67.7**
Synthetic Fertilizer	2.0	2.2	1.9	2.0	1.9	1.7	2.1	1.7	1.7	1.6	1.6	1.7	1.8	1.6	1.7	1.7
PRP Manure	7.0	7.6	7.4	7.8	6.3	7.0	6.1	6.9	6.2	6.3	5.3	5.8	4.9	5.3	5.2	6.1
Managed Manure	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
Residue N[a]	34.4	34.5	33.1	33.9	31.4	29.9	34.7	30.4	30.6	28.0	28.1	29.9	28.0	27.9	26.4	29.8
Other N Inputs[b]	34.5	37.6	33.3	34.6	31.1	29.6	35.3	29.1	30.2	25.2	26.5	27.0	28.8	25.2	25.0	29.7
Tier 1	**9.6**	**9.8**	**7.1**	**7.3**	**13.7**	**9.3**	**17.8**	**7.8**	**10.8**	**7.4**	**11.9**	**10.0**	**16.2**	**10.6**	**12.5**	**8.7**
PRP Manure	9.4	9.6	6.8	7.0	13.3	8.8	17.3	7.3	10.4	7.0	11.5	9.5	15.7	10.1	12.0	8.1
Sewage Sludge	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**+**	**0.5**	**0.5**	**0.5**	**0.5**
Total	**88.0**	**92.1**	**83.2**	**86.1**	**84.7**	**77.8**	**96.5**	**76.3**	**79.9**	**68.9**	**73.9**	**74.8**	**80.1**	**71.0**	**71.3**	**76.4**

+ Less than 0.05 Tg CO_2 Eq.

[a] Residue N inputs include unharvested fixed N from legumes as well as crop residue N.

[b] Other N inputs include mineralization from decomposition of soil organic matter and litter, as well as asymbiotic fixation

Step 4: Estimate Indirect N_2O Emissions for All Land-Use Types and Managed Manure Systems

In this step, N_2O emissions were calculated for the two indirect emission pathways (N_2O emissions due to volatilization, and N_2O emissions due to leaching and runoff of N), which were then summed to yield total indirect N_2O emissions from croplands, grasslands, forest lands, and settlements. Also included were indirect emissions associated with loss of N from managed manure systems.

Step 4a: Indirect Emissions Due to Volatilization

Indirect emissions from volatilization were calculated according to the amount of mineral N that was transported in gaseous forms from (1) the soil profile, and (2) managed manure N lost during storage, treatment, and transport, and later emitted as soil N_2O following atmospheric deposition.[51] See Step 1D for additional information about the methods used to compute N losses due to volatilization. The estimated N volatilized for all land-use and livestock activities was multiplied by the IPCC default emission factor of 0.01 kg N_2O-N/kg N (IPCC 2006, Brumme et al. 1999, Butterbach-Bahl et al. 1997, Corre et al. 1999, Denier van der Gon and Bleeker 2005,

[51] Although some managed manure N is likely to be lost through leaching and in runoff during transport, treatment and storage, all N lost before soil application is considered volatilized because the emission factor for volatilization is slightly higher than for leached/runoff N.

IPCC/UNEP/OECD/IEA 1997) to compute total N_2O emissions from volatilization. The resulting estimates are provided in Table A- 186. The uncertainty was determined using simple error propagation methods (IPCC 2006). Uncertainty in the default emission factor is 0.002-0.05 kg N_2O-N/kg N (IPCC 2006, Brumme et al. 1999, Butterbach-Bahl et al. 1997, Corre et al. 1999, Denier van der Gon and Bleeker 2005, IPCC/UNEP/OECD/IEA 1997). Uncertainty in the default factor for volatilization from synthetic fertilizer is 0.03-0.3 (kg NH_3-N+NOx-N (/kg N (IPCC 2006, IPCC/UNEP/OECD/IEA 1997). Uncertainty in the default factor for volatilization from manure is 0.05-0.5 (kg NH_3-N+NOx-N)/kg N (IPCC 2006, IPCC/UNEP/OECD/IEA 1997).

Table A- 186: Indirect N_2O Emissions (Tg CO_2 Eq.)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Volatilization and Atm. Deposition	**23.4**	**23.3**	**23.5**	**23.7**	**23.7**	**23.7**	**23.6**	**23.9**	**24.3**	**23.7**	**23.6**	**23.4**	**23.6**	**23.9**	**23.5**	**24.6**
Croplands	4.6	4.4	4.6	4.8	4.8	4.9	4.9	5.1	5.3	5.2	5.3	4.9	5.0	5.4	5.3	5.4
Settlements	0.5	0.5	0.6	0.6	0.6	0.6	0.6	0.6	0.6	0.6	0.6	0.6	0.6	0.6	0.6	0.6
Forest Land	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
Grasslands	10.7	10.6	10.7	10.5	10.3	10.3	10.2	10.1	10.1	9.6	9.3	9.4	9.3	9.4	9.1	9.9
Managed Manure[a]	7.5	7.7	7.7	7.9	7.9	8.0	7.9	8.2	8.2	8.3	8.4	8.5	8.7	8.5	8.5	8.5
Surface Leaching & Run-off	**33.4**	**31.9**	**31.8**	**50.2**	**24.4**	**37.7**	**37.5**	**31.7**	**35.2**	**25.9**	**28.8**	**38.2**	**28.3**	**28.8**	**23.1**	**30.0**
Croplands	22.6	18.7	19.3	27.3	15.2	22.3	21.9	18.4	22.6	17.2	19.7	21.2	17.5	20.3	14.8	20.7
Settlements	1.2	1.2	1.2	1.2	1.3	1.2	1.3	1.3	1.3	1.3	1.3	1.2	1.3	1.3	1.3	1.3
Forest Land	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
Grasslands	9.6	12.0	11.3	21.6	7.9	14.0	14.3	11.9	11.2	7.4	7.8	15.7	9.5	7.1	6.9	7.9
Total	**56.8**	**55.1**	**55.3**	**73.8**	**48.1**	**61.4**	**61.1**	**55.6**	**59.5**	**49.7**	**52.4**	**61.6**	**52.0**	**52.8**	**46.6**	**54.6**

+ Less than 0.05 Tg CO_2 Eq.

[a] Accounts for loss of manure N during transport, treatment and storage.

Step 4b: Indirect Emissions Due to Leaching and Runoff

Indirect emissions from leaching of mineral N from soils and losses in overland flow of runoff waters were calculated according to the amount of mineral N that was transported from soil profiles in aqueous forms after originating from anthropogenic activity. See Step 1D for additional information about the methods used to compute N losses from soils due to leaching and runoff in overland water flows.

The total amount of N transported from soil profiles in aqueous forms was multiplied by the IPCC default emission factor of 0.0075 kg N_2O-N/kg N (IPCC 2006, Cloughet al. 2006, Dong et al. 2004, Hiscock et al. 2002, 2003, Reay et al. 2004, 2005, Sawamoto et al. 2005, Williams-Jacobse 2002) to provide the emission estimate from this source. The resulting emission estimates are provided in Table A- 187. The uncertainty was determined based on simple error propagation methods (IPCC 2006). Uncertainty in the default emission factor is 0.0005-0.025 kg N_2O-N/kg N (IPCC 2006, Cloughet al. 2006, Dong et al. 2004, Hiscock et al. 2002, 2003, Reay et al. 2004, 2005, Sawamoto et al. 2005, Williams-Jacobse 2002). Uncertainty in the default factor for leaching/runoff from N inputs is 0.1-0.8 kg NO_3-N/kg N (IPCC 2006, IPCC/UNEP/OECD/IEA 1997).

Step 5: Estimate Total N_2O Emissions for U.S. Soils

Total emissions were estimated by adding total direct emissions (from major crop types and non-major crop types on mineral cropland soils, drainage and cultivation of organic soils, and grassland management) to indirect emissions for all land use and management activities. U.S. national estimates for this source category are provided in Table A- 187. Uncertainties in the final estimate were combined using simple error propagation methods (IPCC 2006), and expressed as a 95 percent confidence interval.

Table A- 187: Total N_2O Emissions from Agricultural Soil Management (Tg CO_2 Eq.)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Total Direct	**310.1**	**312.1**	**303.5**	**322.7**	**311.1**	**292.0**	**352.0**	**322.0**	**345.5**	**300.3**	**324.4**	**327.4**	**314.1**	**297.4**	**292.1**	**310.5**
Direct Emissions from Mineral Cropland Soils	**219.3**	**217.2**	**217.5**	**233.9**	**223.5**	**211.4**	**252.6**	**242.8**	**262.7**	**228.6**	**247.6**	**249.7**	**231.1**	**223.5**	**217.9**	**231.2**
Synthetic Fertilizer	83.6	89.0	86.1	89.9	92.1	85.1	100.3	96.3	97.5	89.4	91.9	94.2	90.2	84.6	88.5	86.9
Organic Amendment[a]	10.3	11.3	10.5	11.4	11.3	10.9	12.2	12.1	12.4	11.5	12.1	12.9	12.0	11.2	11.6	11.7
Residue N[b]	15.0	13.8	17.2	17.4	15.0	15.8	18.8	18.2	16.3	18.4	18.5	16.6	15.1	18.3	14.7	16.0
Other N Inputs[c]	110.3	103.1	103.6	115.2	105.2	99.6	121.3	116.2	136.5	109.2	125.1	126.0	113.8	109.4	103.1	116.6
Direct Emissions from	**2.8**	**2.8**	**2.8**	**2.8**	**2.8**	**2.8**	**2.8**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**	**2.9**

Drained Organic Cropland Soils																
Direct Emissions from Grasslands	**88.0**	**92.1**	**83.2**	**86.1**	**84.7**	**77.8**	**96.5**	**76.3**	**79.9**	**68.9**	**73.9**	**74.8**	**80.1**	**71.0**	**71.3**	**76.4**
Synthetic Mineral Fertilizer	2.0	2.2	1.9	2.0	1.9	1.7	2.1	1.7	1.7	1.6	1.6	1.7	1.8	1.6	1.7	1.7
PRP Manure	16.4	17.1	14.2	14.8	19.6	15.8	23.4	14.3	16.6	13.3	16.8	15.3	20.6	15.5	17.2	14.3
Managed Manure	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+	+
Sewage Sludge	+	+	+	+	+	+	+	+	+	+	+	+	0.5	0.5	0.5	0.5
Residue[b]	34.4	34.5	33.1	33.9	31.4	29.9	34.7	30.4	30.6	28.0	28.1	29.9	28.0	27.9	26.4	29.8
Other N Inputs[c]	34.5	37.6	33.3	34.6	31.1	29.6	35.3	29.1	30.2	25.2	26.5	27.0	28.8	25.2	25.0	29.7
Total Indirect	**56.8**	**55.1**	**55.3**	**73.8**	**48.1**	**61.4**	**61.1**	**55.6**	**59.5**	**49.7**	**52.4**	**61.6**	**52.0**	**52.8**	**46.6**	**54.6**
Volatilization	23.4	23.3	23.5	23.7	23.7	23.7	23.6	23.9	24.3	23.7	23.6	23.4	23.6	23.9	23.5	24.6
Leaching/Runoff	33.4	31.9	31.8	50.2	24.4	37.7	37.5	31.7	35.2	25.9	28.8	38.2	28.3	28.8	23.1	30.0
Total Emissions	**366.9**	**367.2**	**358.8**	**396.5**	**359.2**	**353.4**	**413.1**	**377.6**	**405.0**	**350.0**	**376.8**	**389.0**	**366.1**	**350.2**	**338.8**	**365.1**

+ Less than 0.05 Tg CO_2 Eq.
[a] Organic amendment inputs include managed manure amendments and other commercial organic fertilizer (i.e., dried blood, dried manure, tankage, compost, and other).
[b] Residue N inputs include unharvested fixed N from legumes as well as crop residue N.
[c] Other N inputs include mineralization from decomposition of soil organic matter, as well as asymbiotic fixation.

Direct and indirect emissions of soil N_2O vary regionally in both croplands and grasslands as a function of N inputs, weather, and soil type. Direct emissions from cropped soils are high in the Corn Belt, where N inputs associated with corn and soybean cropping are high (Figure A- 5). Direct N_2O is also high in some northern counties of the Great Lake states and New England. Although a small portion of the land in these counties is cropped, emissions on a per unit area basis are high because subsurface soil layers are likely to remain frozen when surface soil layers thaw in spring. This creates saturated conditions near the surface that facilitate denitrification and N_2O emissions. Emissions are also high in the lower Mississippi Valley and some counties in eastern Texas where high N inputs and fine-textured soils lead to high denitrification rates. High emissions in some Western counties are associated with intensive, irrigated cropping. Indirect emissions tend to be high in the southeast United States (Figure A- 6), where sandy soils facilitate leaching, which is an important contributor to indirect emissions. Emissions are moderate to high in the Corn Belt, some counties in eastern Texas, and some counties in the West. In most of these counties, emissions are associated with N inputs but in some counties, particularly the arid West, high indirect emissions can result from rare large rainfall events that leach large amounts of NO_3 below the rooting zone and into the subsoil.

Direct and indirect emissions from grasslands are typically lower than those from croplands (Figure A- 7 and Figure A- 8) because N inputs, particularly from synthetic fertilizer, tend to be lower. As with cropped soils, emissions from grasslands on a per unit area basis are high in the Northeast and the Great Lake states (Figure A- 7) because spring thawing of surface soil layers while deeper layers remain frozen facilitates the denitrification process. In general, emissions are lower in the western United States because grasslands in the East are more intensively managed (legume seeding, fertilization) while Western rangelands receive little, if any, N inputs. Indirect emissions from grasslands are high in the Southeast (Figure A- 8), mainly because large amounts of NO_3 are leached from coarse-textured soils in this region. As with cropped soils, high indirect emissions from grasslands in the arid West can result from rare but large rainfall events that leach large amounts of NO_3 below the rooting zone and into the subsoil.

Figure A- 5: Major Crops, Direct N_2O Emissions Estimated Using the DAYCENT Model, 1990-2005 (kg N/ha/yr)

Figure A- 6: Major Crops, N Losses Leading to Indirect N_2O Emissions Estimated Using the DAYCENT Model, 1990-2005 (kg N/ha/yr)

Figure A- 7: Grasslands, Direct N_2O Emissions Estimated Using the DAYCENT Model, 1990-2005 (kg N/ha/yr)

Figure A- 8: Grasslands, N Losses Leading to Indirect N2O Emissions Estimated Using the DAYCENT Model, 1990-2005 (kg N/ha/yr)

3.12. Methodology for Estimating Net Carbon Stock Changes in Forest Lands Remaining Forest Lands

This sub-annex expands on the methodology used to calculate net changes in carbon (C) stocks in forest ecosystems and in harvested wood products. Some of the details of C conversion factors and procedures for calculating net CO_2 flux for forests are provided below; full details of selected topics may be found in the cited references.

Carbon Stocks and Net Changes in Forest Ecosystem Carbon Stocks

Two forest inventories exist for most forest land in the 48 states of the conterminous United States. C stocks are estimated for each inventory, at the level of permanent inventory plots. C per hectare (for a sample location) is multiplied by the total number of hectares that the plot represents, and then totals are summed for an area of interest, such as the state of Arizona. Net annual C stock changes are calculated by taking the difference between the inventories and dividing by the number of years between the inventories for a selected state or sub-state area.

Forest inventory data

The estimates of forest C stocks are based on data derived from forest inventory surveys. Forest inventory data were obtained from the USDA Forest Service, Forest Inventory and Analysis (FIA) program (Frayer and Furnival 1999). FIA data include remote sensing information to determine forest areas, and collection of measurements in the field at sample locations called plots. Tree measurements include diameter and species. On a subset of plots, additional measurements or samples are taken of down dead wood, litter, and soil C; however, these are not yet available for C estimation. The field protocols are thoroughly documented and available for download from the USDA Forest Service (2006a). The inventory was designed for timber volume estimation rather than C stock estimation, so most C pools are not measured or sampled directly. Bechtold and Patterson (2005) provide the estimation procedures for standard forest inventory results. The data are freely available for download at USDA Forest Service (2006b) as the Forest Inventory and Analysis Database (FIADB) Version 2.1. The data identified as "snapshot" files and cited as FISDB 2.1, are the primary sources of inventory data used to estimate forest C stocks.

Forest inventory data is limited or non-existent in Alaska, Hawaii, and U.S. territories, and, thus, this inventory assumes that these areas account for a net C change of zero. Some territories may have fairly good data; however, that has not yet been organized and used to estimate C. Hawaii and the territories have relatively small areas of forest land and, thus, inclusion in the inventory would probably not affect the overall C budget to a great degree. Alaska has over 50 million hectares of forest land, however, and could have a significant effect in terms of C emissions and sinks. A review of scientific literature indicates that accounting for the flux of C in Alaskan forests would change U.S. national forest C flux estimates by –5 percent to 10 percent (not including harvested wood). A planned improvement for this section is, thus, to develop better estimates of C flux in Alaska.

Agroforestry systems are also not currently accounted for in the U.S. inventory, since they are not explicitly inventoried by either of the two primary national natural resource inventory programs: the Forest Inventory and Analysis (FIA) program of the USDA Forest Service and the National Resources Inventory (NRI) of the USDA Natural Resources Conservation Service (Perry et al. 2005). The majority of these tree-based practices do not meet the size and definitions for forests within each of these resource inventories. The size characteristics that exclude them from inventories also allow these systems to provide their many services without taking the land out of agricultural production, making them an appealing C sequestration option. Agroforestry in the United States has been defined as "intensive land-use management that optimizes the benefits (physical, biological, ecological, economic, social) from bio-physical interactions created when trees and/or shrubs are deliberately combined with crops and/or livestock."(Gold et al. 2000). In the United States, there are six categories of agroforestry practices:

riparian forest buffers, windbreaks, alley cropping, silvopasture, forest farming and special applications.[52] These practices are used to address many issues facing agricultural lands, such as economic diversification, habitat fragmentation, and water quality. While providing these services and regardless of intent, these tree-based plantings will also reduce atmospheric CO_2. This occurs directly through CO_2 sequestration into woody biomass, and indirectly through enhancement of agricultural production, trapping wind-blown and surface runoff sediments, and/or reducing CO_2 emissions through fuel-use savings (Brandle et al. 1992). These practices are not worth reporting individually, but can potentially be quite large when taken into account within a whole-farm or within an aggregating larger entity (i.e., state-level) (Brandle et al. 1992, Schoeneberger 2006). The sequestration potential through agroforestry practices in the United States has been estimated to be approximately 90.3 Mt C/year by 2025 (Nair and Nair 2003).

Summing state-level C stocks to calculate United States net C flux in forest ecosystems

The overall approach for determining forest C stocks and stock change is to estimate forest C stocks based on data from two or more forest surveys conducted several years apart for each state or sub-state. There are generally 2 to 4 surveys per state available electronically, beginning with a pre-1990 survey. C stocks are calculated separately for each state based on available inventories conducted since 1990 and for the inventory closest to, but prior to, 1990 (the base year). This approach ensures that the period 1990 to present can be adequately represented. Surveys conducted prior to and in the early to mid 1990s focused on land capable of supporting timber production (timberland).[53] As a result, information on less productive forest land or lands reserved from harvest was limited. Inventory field crews periodically measured all the plots in a state at a frequency of every 5 to 14 years. Generally, forests in states with fast-growing (and therefore rapidly changing) forests tended to be surveyed more often than states with slower-growing (and therefore slowly changing) forests. Older surveys for some states, particularly in the West, also have National Forest System lands and/or reserved lands surveyed at different times than productive, privately-owned forest land in the state. Periodic data for each state, thus, became available at irregular intervals and determining the year of data collection associated with the survey can sometimes be difficult. Table A-188 provides a list of the specific surveys used here and Smith et al. (in prep.) provides further details.

Table A-188. Source of Forest Inventory and Average Year of Field Survey Used to Estimate Statewide Carbon Stocks.

State/Substate[a]	Source of Inventory Data, Report/Inventory Year[b]	Average Year Assigned to Inventory[c]
Alabama	Southern, 1982	1982
	FISDB 2.1, 1990	1990
	FISDB 2.1, 2000	1999
	FISDB 2.1, 2004	2003
Arizona, NF non-woodlands	1987 RPA	1985
	FISDB 2.1, 1999	1996
	FISDB 2.1, 2005	2003
Arizona, NF woodlands	1987 RPA	1984
	FISDB 2.1, 1999	1996
	FISDB 2.1, 2005	2003

[52] More information on agroforestry practices can be found online at <http://www.unl.edu/nac>.

[53] Forest land in the United States includes land that is at least 10 percent stocked with trees of any size. Timberland is the most productive type of forest land, which is on unreserved land and is producing or capable of producing crops of industrial wood. Productivity is at a minimum rate of 20 cubic feet of industrial wood per acre per year. The remaining portion of forest land is classified as either reserved forest land, which is forest land withdrawn from timber use by statute or regulation, or other forest land, which includes less productive forests on which timber is growing at a rate less than 20 cubic feet per acre per year. Forest land includes woodlands, which describes forest types consisting primarily of species that have their diameter measured at root collar, and for which there is no site index equations, nor stocking guides (Woudenberg 2006). These may include areas with a tree cover of between 5-9.9%; however, it is not known how much woodland is below 10% tree cover. In 2002, there were about 199 million hectares of timberland in the conterminous United States, which represented 79 percent of all forest lands over the same area (Smith et al. 2004b).

Arizona, non-NF non-woodlands	FISDB 2.1, 1985	1986
	FISDB 2.1, 1999	1995
	FISDB 2.1, 2005	2003
Arizona, non-NF woodlands	FISDB 2.1, 1999	1989
	FISDB 2.1, 2003	2003
Arkansas	Eastwide, 1988	1988
	FISDB 2.1, 1995	1996
	FISDB 2.1, 2005	2003
California, NF non-chaparral	1987 RPA	1980
	1997 RPA	1991
	FISDB 2.1, 2005	2003
California, non-NF non-chaparral	1987 RPA	1982
	Westwide, 1994	1994
	FISDB 2.1, 2005	2004
Colorado, NF non-woodlands	1997 RPA	1981
	FISDB 2.1, 2005	2004
Colorado, NF woodlands	1997 RPA	1975
	FISDB 2.1, 1984	1997
	FISDB 2.1, 2005	2004
Colorado, non-NF non-woodlands	Westwide, 1983	1980
	FISDB 2.1, 2005	2004
Colorado, non-NF woodlands	Westwide, 1983	1983
	FISDB 2.1, 2005	2004
Connecticut	FISDB 2.1, 1985	1985
	FISDB 2.1, 1998	1998
	FISDB 2.1, 2004	2004
Delaware	FISDB 2.1, 1986	1986
	FISDB 2.1, 1999	1999
	FISDB 2.1, 2004	2005
Florida	FISDB 2.1, 1987	1987
	FISDB 2.1, 1995	1995
	FISDB 2.1, 2005	2004
Georgia	FISDB 2.1, 1989	1989
	FISDB 2.1, 1997	1997
	FISDB 2.1, 2004	2001
Idaho, Caribou-Targhee NF	1987 RPA	1976
	Westwide, 1991	1992
	FISDB 2.1, 2005	2005
Idaho, Kootenai NF	1987 RPA	1988
	FISDB 2.1, 1991	1995
	FISDB 2.1, 2005	2006
Idaho, Payette NF	1987 RPA	1982
	FISDB 2.1, 2005	2005
Idaho, Salmon-Challis NF	1987 RPA	1978
	FISDB 2.1, 2005	2005
Idaho, Sawtooth NF	Westwide, 1991	1983
	FISDB 2.1, 1991	1996
	FISDB 2.1, 2005	2005
Idaho, all other NF	Westwide, 1991	1988
	FISDB 2.1, 1991	2000
	FISDB 2.1, 2005	2005
Idaho, non-NF non-woodlands	1987 RPA	1982
	FISDB 2.1, 1991	1990
	FISDB 2.1, 2005	2005
Idaho, non-NF woodlands	FISDB 2.1, 1991	1982
	FISDB 2.1, 2005	2005
Illinois	FISDB 2.1, 1985	1985
	FISDB 2.1, 1998	1998
	FISDB 2.1, 2004	2003
Indiana	FISDB 2.1, 1986	1986
	FISDB 2.1, 1998	1998
	FISDB 2.1, 2003	2001

	FISDB 2.1, 2005	2003
Iowa	FISDB 2.1, 1990	1990
	FISDB 2.1, 2003	2002
	FISDB 2.1, 2004	2003
Kansas	FISDB 2.1, 1981	1981
	FISDB 2.1, 1994	1994
	FISDB 2.1, 2004	2003
Kentucky	FISDB 2.1, 1988	1987
	FISDB 2.1, 2004	2002
Louisiana	Southern, 1984	1984
	FISDB 2.1, 1991	1991
	FISDB 2.1, 2003	2003
Maine	Eastwide, 1982	1983
	FISDB 2.1, 1995	1995
	FISDB 2.1, 2003	2002
Maryland	FISDB 2.1, 1986	1986
	FISDB 2.1, 1999	2000
	FISDB 2.1, 2004	2005
Massachusetts	FISDB 2.1, 1985	1985
	FISDB 2.1, 1998	1998
	FISDB 2.1, 2004	2004
Michigan	FISDB 2.1, 1980	1980
	FISDB 2.1, 1993	1993
	FISDB 2.1, 2004	2003
Minnesota	FISDB 2.1, 1977	1977
	FISDB 2.1, 1990	1989
	FISDB 2.1, 2003	2001
	FISDB 2.1, 2004	2002
Mississippi	Eastwide, 1987	1987
	FISDB 2.1, 1994	1994
Missouri	FISDB 2.1, 1989	1988
	FISDB 2.1, 2003	2002
	FISDB 2.1, 2004	2003
Montana, NF	1987 RPA	1988
	FISDB 2.1, 1989	1996
	FISDB 2.1, 2005	2005
Montana, non-NF non-reserved	FISDB 2.1, 1989	1989
	FISDB 2.1, 2005	2005
Montana, non-NF reserved	1997 RPA	1990
	FISDB 2.1, 2005	2005
Nebraska	FISDB 2.1, 1983	1983
	FISDB 2.1, 1994	1995
	FISDB 2.1, 2004	2003
Nevada, NF non-woodlands	1987 RPA	1974
	FISDB 2.1, 1989	1997
	FISDB 2.1, 2005	2005
Nevada, NF woodlands	1987 RPA	1978
	FISDB 2.1, 1989	1997
	FISDB 2.1, 2005	2005
Nevada, non-NF non-woodlands	1997 RPA	1985
	FISDB 2.1, 2005	2005
Nevada, non-NF woodlands	FISDB 2.1, 1989	1980
	FISDB 2.1, 2005	2005
New Hampshire	FISDB 2.1, 1983	1983
	FISDB 2.1, 1997	1997
	FISDB 2.1, 2003	2004
New Jersey	FISDB 2.1, 1987	1987
	FISDB 2.1, 1999	1999
	FISDB 2.1, 2004	2005
New Mexico, NF non-woodlands	1987 RPA	1986
	FISDB 2.1, 1999	1997
New Mexico, NF woodlands	1987 RPA	1986

	FISDB 2.1, 1999	1997
New Mexico, non-NF non-woodlands	FISDB 2.1, 1987	1987
	FISDB 2.1, 1999	1999
New Mexico, non- NF woodlands	FISDB 2.1, 1999	1989
New York, non-reserved	Eastwide, 1980	1981
	FISDB 2.1, 1993	1993
	FISDB 2.1, 2004	2003
New York, reserved	1987 RPA	1988
	FISDB 2.1, 2004	2003
North Carolina	FISDB 2.1, 1984	1984
	FISDB 2.1, 1990	1990
	FISDB 2.1, 2002	2001
North Dakota	FISDB 2.1, 1980	1979
	FISDB 2.1, 1995	1995
	FISDB 2.1, 2005	2003
Ohio	1987 RPA	1988
	FISDB 2.1, 1991	1991
	FISDB 2.1, 2004	2003
Oklahoma, Central & West	FISDB 2.1, 1993	1989
Oklahoma, East	Southern, 1986	1986
	FISDB 2.1, 1993	1993
Oregon, NF East	1987 RPA	1988
	Westwide, 1992	1995
	FISDB 2.1, 2005	2004
Oregon, non-NF East	1987 RPA	1977
	FISDB 2.1, 1992	1990
	FISDB 2.1, 1999	1999
	FISDB 2.1, 2005	2003
Oregon, NF West	1987 RPA	1987
	Westwide, 1992	1995
	FISDB 2.1, 2005	2004
Oregon, non-NF West	Westwide, 1992	1986
	FISDB 2.1, 2005	2004
Pennsylvania	FISDB 2.1, 1989	1990
	FISDB 2.1, 2004	2003
Rhode Island	FISDB 2.1, 1985	1985
	FISDB 2.1, 1998	1999
	FISDB 2.1, 2004	2004
South Carolina	FISDB 2.1, 1986	1986
	FISDB 2.1, 1993	1993
	FISDB 2.1, 2001	2001
	FISDB 2.1, 2005	2004
South Dakota, NF	1997 RPA	1986
	FISDB 2.1, 1995	1999
	FISDB 2.1, 2004	2003
South Dakota, non-NF	1987 RPA	1986
	FISDB 2.1, 1995	1995
	FISDB 2.1, 2004	2003
Tennessee	FISDB 2.1, 1989	1989
	FISDB 2.1, 1999	1998
	FISDB 2.1, 2003	2002
Texas	Southern, 1986	1986
	FISDB 2.1, 1992	1992
	FISDB 2.1, 2003	2003
Utah, non-woodlands	1987 RPA	1978
	FISDB 2.1, 1993	1993
	FISDB 2.1, 2005	2003
Utah, woodlands	1987 RPA	1979
	FISDB 2.1, 1993	1994
	FISDB 2.1, 2005	2003
Vermont	FISDB 2.1, 1983	1983
	FISDB 2.1, 1997	1997

	FISDB 2.1, 2004	2004
Virginia	FISDB 2.1, 1984	1985
	FISDB 2.1, 1992	1991
	FISDB 2.1, 2001	2000
	FISDB 2.1, 2003	2001
Washington, NF East	1987 RPA	1988
	Westwide, 1991	1995
	FISDB 2.1, 2005	2004
Washington, non-NF East	1987 RPA	1981
	FISDB 2.1, 1991	1992
	FISDB 2.1, 2005	2004
Washington, NF West	1987 RPA	1988
	Westwide, 1991	1995
	FISDB 2.1, 2005	2004
Washington, non-NF West	1987 RPA	1979
	FISDB 2.1, 1991	1991
	FISDB 2.1, 2005	2004
West Virginia	FISDB 2.1, 1989	1988
	FISDB 2.1, 2000	2001
	FISDB 2.1, 2004	2005
Wisconsin	FISDB 2.1, 1983	1982
	FISDB 2.1, 1996	1995
	FISDB 2.1, 2004	2002
Wyoming, NF	1997 RPA	1982
	2002 RPA	1997
	FISDB 2.1, 2000	2000
Wyoming, non-NF non-reserved woodland	FISDB 2.1, 1984	1984
	FISDB 2.1, 2000	2002
Wyoming, non-NF reserved	1997 RPA	1985
	FISDB 2.1, 2000	2000
Wyoming, non-NF non-reserved non-woodland	FISDB 2.1, 1984	1984
	FISDB 2.1, 2000	2002

[a] Substate areas include National Forests (NF), all forest ownerships except National Forest (non-NF), woodlands (Forest land dominated by woodland species, such as pinyon and juniper, where stocking cannot be determined, condition has at least 5% crown cover by trees of any size or has had at least 5% cover in the past[USDA Forest Service, National Field Guide, 2006a]), non-woodlands (used for clarity to emphasize that woodlands are not included), non-chaparral (used for clarity to emphasize that no chaparral is included), reserved (Forest land withdrawn from timber utilization through statute, administrative regulation, or designation, Smith et al. (2004b)), and non-reserved (forest land that is not reserved, used for clarity). Some National Forests are listed individually by name, e.g., Payette NF. Oregon and Washington were divided into eastern and western forests (east or west of the crest of the Cascade Mountains). In Oklahoma, east refers to the eastern 18 counties. Other counties were traditionally considered unproductive in terms of forest products. Central & West refer to all other counties, but only 1 inventory was conducted. Thus, we are including this substate in stock estimation, not in flux.

[b] FISDB 2.1 is the snapshot version of FIADB 2.1 as available on Internet September 8, 2006 (USDA Forest Service 2006c). Eastwide (Hansen et al. 1992) and Westwide (Woudenberg and Farrenkopf 1995) inventory data are formats that predate the FIADB data. Southern inventories are from older surveys compiled for the Southern region (CD of Southern data from Linda Heatherly, FIA Southern Region, July 7, 2004). RPA data are periodic national summaries. The year is the nominal, or reporting, year associated with each dataset.

[c] Average year is based on average measurement year of forest land survey plots and rounded to the nearest integer year.

A new national plot design and annual sampling (USDA Forest Service 2006a) was introduced by FIA about ten years ago. Most states have only recently been brought into this system, though. Annual sampling means that a portion of plots throughout each state is sampled each year, with the goal of measuring all plots once every 5 years. Sampling is designed such that partial inventory cycles provide usable, unbiased samples of forest inventory, but with higher standard errors than the full cycle. Thus, many states have relatively recent partial inventories; that is, they are part-way through an annual inventory cycle. All annual surveys initiated since 1998 have followed the new national plot design for all forest land, including reserved and less productive lands.

For each pool in each state in each year, C stocks are estimated by linear interpolation between survey years. Similarly, fluxes, or net stock changes, are estimated for each pool in each state by dividing the difference between two successive stocks by the number of intervening years between surveys. Thus, the number of separate stock change estimates for each state or sub-state is one less than the number of available inventories. Stocks and fluxes since the most recent survey are based on extrapolating estimates from the last two surveys. C stock and flux estimates for each pool are summed over all states to form

estimates for the conterminous United States. Summed fluxes and stocks are presented in Table A- 189 and Table A- 190, respectively. Table A- 190 also provides an estimate of forest area based on the interpolation procedure described above.

Table A-189. Net Annual Changes in Carbon Stocks (Tg C yr^{-1}) in Forest and Harvested Wood Pools, 1990-2005

Carbon Pool	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Forest	**(127.2)**	**(151.4)**	**(151.3)**	**(144.0)**	**(162.5)**	**(164.2)**	**(167.0)**	**(173.0)**	**(159.3)**	**(150.5)**	**(144.4)**	**(151.5)**	**(162.4)**	**(162.4)**	**(162.4)**	**(162.4)**
Live, aboveground	(68.7)	(83.3)	(85.2)	(80.9)	(83.1)	(90.3)	(93.2)	(97.6)	(91.8)	(92.9)	(94.7)	(98.3)	(102.7)	(102.7)	(102.7)	(102.7)
Live, belowground	(17.4)	(19.1)	(18.8)	(17.9)	(18.2)	(19.0)	(18.9)	(20.3)	(19.5)	(19.8)	(20.1)	(20.8)	(21.7)	(21.7)	(21.7)	(21.7)
Dead Wood	(10.0)	(12.6)	(12.6)	(12.7)	(16.0)	(16.6)	(15.9)	(17.0)	(17.1)	(14.7)	(13.1)	(13.6)	(14.3)	(14.3)	(14.3)	(14.3)
Litter	(17.9)	(18.5)	(17.6)	(17.9)	(20.8)	(13.5)	(13.3)	(13.3)	(10.7)	(11.3)	(9.8)	(12.9)	(14.2)	(14.2)	(14.2)	(14.2)
Soil Organic Carbon	(13.2)	(18.0)	(17.1)	(14.6)	(24.3)	(24.8)	(25.6)	(24.7)	(20.2)	(11.8)	(6.7)	(5.9)	(9.5)	(9.5)	(9.5)	(9.5)
Harvested Wood	**(36.0)**	**(33.6)**	**(33.4)**	**(32.5)**	**(32.8)**	**(31.5)**	**(29.8)**	**(31.1)**	**(30.3)**	**(31.6)**	**(29.8)**	**(24.6)**	**(25.3)**	**(25.0)**	**(27.8)**	**(28.2)**
Products in Use	(17.2)	(14.5)	(15.9)	(14.6)	(15.5)	(14.6)	(13.7)	(14.4)	(13.2)	(14.0)	(12.6)	(8.5)	(9.3)	(9.1)	(11.8)	(12.1)
SWDS	(18.8)	(19.1)	(17.5)	(17.9)	(17.3)	(16.9)	(16.1)	(16.7)	(17.1)	(17.6)	(17.2)	(16.1)	(16.0)	(15.9)	(16.0)	(16.1)
Total Net Flux	**(163.2)**	**(185.0)**	**(184.7)**	**(176.5)**	**(195.3)**	**(195.7)**	**(196.8)**	**(204.1)**	**(189.6)**	**(182.1)**	**(174.2)**	**(176.1)**	**(187.7)**	**(187.4)**	**(190.2)**	**(190.6)**

Table A-190. Carbon Stocks (Tg C) in Forest and Harvested Wood Pools, 1990-2006

	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005	2006
Forest Area (1000 ha)	**242,300**	**242,925**	**243,629**	**244,329**	**245,057**	**245,946**	**246,797**	**247,667**	**248,577**	**249,451**	**250,275**	**251,110**	**251,977**	**252,879**	**253,782**	**254,684**	**255,587**
Carbon Pool																	
Forest	**39,026**	**39,153**	**39,305**	**39,456**	**39,600**	**39,762**	**39,927**	**40,094**	**40,267**	**40,426**	**40,576**	**40,721**	**40,872**	**41,035**	**41,197**	**41,359**	**41,522**
Live, aboveground	14,164	14,232	14,316	14,401	14,482	14,565	14,655	14,748	14,846	14,938	15,031	15,125	15,224	15,326	15,429	15,532	15,634
Live, belowground	2,794	2,811	2,830	2,849	2,867	2,885	2,904	2,923	2,943	2,963	2,983	3,003	3,024	3,045	3,067	3,089	3,110
Dead Wood	2,354	2,364	2,376	2,389	2,402	2,418	2,434	2,450	2,467	2,485	2,499	2,512	2,526	2,540	2,555	2,569	2,583
Litter	4,404	4,422	4,441	4,458	4,476	4,497	4,511	4,524	4,537	4,548	4,559	4,569	4,582	4,596	4,610	4,625	4,639
Soil Organic Carbon	15,310	15,323	15,341	15,359	15,373	15,398	15,422	15,448	15,473	15,493	15,505	15,511	15,517	15,527	15,536	15,546	15,555
Harvested Wood	1,888	1,927	1,965	2,000	2,033	2,067	2,100	2,132	2,164	2,194	2,225	2,255	2,287	2,317	2,341	2,367	2,395
Products in Use	1,184	1,203	1,221	1,238	1,252	1,268	1,283	1,298	1,313	1,327	1,341	1,354	1,368	1,381	1,389	1,399	1,411
SWDS	704	724	744	762	781	799	817	834	851	867	884	901	919	936	952	968	984
Total Carbon Stock	**40,914**	**41,080**	**41,270**	**41,456**	**41,633**	**41,829**	**42,027**	**42,226**	**42,431**	**42,620**	**42,801**	**42,976**	**43,159**	**43,352**	**43,538**	**43,726**	**43,917**

Table A- 191 shows average C density values for forest ecosystem C pools according to region and forest types. These values were calculated by applying plot-level C estimation procedures as described above to the most recent inventory per state (Table A-188). C density values reflect the most recent survey for each state as available in the FIADB, not potential maximum C storage. C densities are affected by the distribution of stand sizes within a forest type, which can range from regenerating to mature stands. A large proportion of young stands in a particular forest type is likely to reduce the regional average for C density.

Table A- 191. Average carbon density (Mg C/ha) by carbon pool and forest area (1000 ha) according to region and forest type, based on the most recent inventory survey available for each state from FIA (see Table A-188), corresponding to an average year of 2002

Region (States) Forest Types	Above-ground Biomass	Below-ground Biomass	Dead Wood	Litter	Soil Organic Carbon	Forest Area
	Carbon Density (Mg/ha)					(1,000 ha)
Northeast						
(CT,DE,MA,MD,ME,NH,NJ,NY,OH,PA,RI,VT,WV)						
White/Red/Jack Pine	93.0	19.2	11.3	13.6	78.1	1,949
Spruce/Fir	51.2	10.8	11.8	30.6	98.0	2,984
Oak/Pine	74.8	14.7	9.0	27.1	66.9	1,334
Oak/Hickory	79.7	15.1	10.3	8.0	53.1	11,204
Elm/Ash/Cottonwood	50.6	9.6	8.0	23.8	111.7	1,080
Maple/Beech/Birch	75.1	14.4	12.4	26.4	69.6	15,643
Aspen/Birch	46.1	9.1	7.7	8.5	87.4	1,659
Minor Types and Nonstocked	50.4	10.2	6.9	13.9	82.4	1,302
All	72.6	14.1	11.0	19.2	69.7	37,155
Northern Lake States						
(MI,MN,WI)						
White/Red/Jack Pine	54.1	11.3	8.0	12.3	120.8	1,822
Spruce/Fir	40.8	8.6	8.3	32.5	261.8	3,075
Oak/Hickory	70.6	13.4	10.3	7.8	97.1	2,909
Elm/Ash/Cottonwood	50.1	9.6	8.6	25.4	179.9	1,682
Maple/Beech/Birch	71.0	13.6	10.9	26.4	134.3	5,139
Aspen/Birch	41.7	8.1	8.4	8.3	146.1	5,341
Minor Types and Nonstocked	37.0	7.3	6.1	11.0	126.6	892
All	54.4	10.6	9.1	18.1	153.1	20,859
Northern Prairie States						
(IA,IL,IN,KS,MO,ND,NE,SD)						
Ponderosa Pine	42.1	8.9	6.9	14.3	48.5	576
Oak/Pine	51.8	10.1	7.2	25.1	39.8	563
Oak/Hickory	69.2	13.1	9.3	7.6	49.0	8,209
Elm/Ash/Cottonwood	71.9	13.5	11.1	23.6	83.1	1,737
Maple/Beech/Birch	63.5	12.0	8.8	24.9	70.9	1,111
Minor Types and Nonstocked	37.2	7.4	5.9	12.5	57.6	876
All	65.0	12.3	9.1	12.5	55.5	13,072
South Central						
(AL,AR,KY,LA,MS,OK,TN,TX)						
Longleaf-slash pine	37.0	7.5	3.8	10.6	55.5	1,316
Loblolly-shortleaf pine	43.5	8.9	4.8	9.6	41.9	12,469
Oak-pine	50.4	9.8	6.3	9.6	41.7	6,134
Oak-hickory	61.5	11.6	7.1	6.3	38.6	19,866
Oak-gum-cypress	72.2	13.7	8.8	6.3	52.8	5,178
Elm-ash-cottonwood	56.8	10.7	7.7	5.8	49.9	2,456
Minor types and nonstocked	48.5	9.3	6.6	7.8	47.1	1,299
All	55.4	10.7	6.5	7.7	42.6	48,718
Southeast						
(FL,GA,NC,SC,VA)						
Longleaf-slash pine	34.1	6.9	3.7	9.7	110.0	4,067
Loblolly-shortleaf pine	46.9	9.6	5.8	9.3	72.9	8,748
Oak-pine	49.8	9.7	5.5	9.1	61.4	4,534
Oak-hickory	70.5	13.3	8.1	6.3	45.3	12,201
Oak/Gum/Cypress	72.5	14.0	9.0	6.3	158.0	4,204

Elm/Ash/Cottonwood	65.1	12.3	10.1	6.0	95.7	675
Minor Types and Nonstocked	48.1	9.3	6.4	6.4	96.5	1,192
All	57.3	11.1	6.8	7.8	77.5	35,621
Pacific Northwest, Westside						
(Western OR and WA)						
Douglas-fir	143.8	30.2	31.4	31.6	94.8	5,583
Fir/Spruce/Mt. Hemlock	149.0	31.5	38.3	37.9	62.1	1,286
Hemlock/Sitka Spruce	176.4	37.2	45.5	37.3	116.3	1,676
Alder/Maple	82.6	16.2	21.0	7.4	115.2	1,280
Minor Types and Nonstocked	69.5	13.8	12.0	13.4	84.8	1,302
All	133.6	27.9	30.8	28.3	95.4	11,127
Pacific Northwest, Eastside						
(Eastern OR and WA)						
Pinyon/Juniper	13.2	2.6	2.4	21.1	46.9	810
Douglas-fir	78.3	16.4	18.5	36.2	94.8	1,987
Ponderosa Pine	49.8	10.4	10.1	22.6	50.7	2,956
Fir/Spruce/Mt. Hemlock	93.2	19.7	26.4	37.9	62.1	1,627
Lodgepole Pine	41.4	8.7	9.8	21.1	52.0	1,039
Western Larch	66.6	13.9	18.0	36.0	45.1	286
Minor Types and Nonstocked	29.4	5.8	12.9	22.4	79.5	1,451
All	56.1	11.7	14.3	27.8	64.9	10,157
Pacific Southwest						
(CA)						
Pinyon/Juniper	26.0	5.1	2.0	21.1	26.3	924
Douglas-fir	165.0	34.3	34.3	35.4	40.1	433
Ponderosa Pine	54.9	11.4	10.6	35.0	41.3	396
Fir/Spruce/Mt. Hemlock	160.6	33.9	44.1	38.3	51.9	800
Lodgepole Pine	95.1	20.1	19.7	39.1	35.2	389
Redwood	200.0	41.7	42.1	60.4	53.8	261
California Mixed Conifer	116.6	24.4	28.7	37.6	49.8	3,763
Western Oak	67.3	12.8	7.6	29.2	27.6	3,792
Tanoak/Laurel	129.2	25.4	19.0	27.5	27.6	860
Minor Types and Nonstocked	34.5	6.7	8.8	23.6	37.7	1,834
All	89.4	18.2	18.1	31.9	38.1	13,451
Rocky Mountain, North						
(ID,MT)						
Douglas-fir	73.0	15.4	13.8	37.1	38.8	5,774
Ponderosa Pine	41.1	8.6	7.5	23.1	34.3	1,926
Fir/Spruce/Mt. Hemlock	68.1	14.4	21.4	37.4	44.1	4,475
Lodgepole Pine	52.6	11.2	9.9	23.2	37.2	2,714
Western Larch	60.8	12.9	14.1	36.0	34.2	463
Minor Types and Nonstocked	27.4	5.5	9.7	24.9	42.9	4,057
All	56.0	11.8	13.5	31.2	40.1	19,409
Rocky Mountain, South						
(AZ,CO,NM,NV,UT,WY)						
Pinyon/Juniper	22.3	4.5	0.9	21.1	19.7	19,800
Douglas-fir	73.8	15.6	16.6	38.0	30.9	1,808
Ponderosa Pine	48.1	10.2	8.2	23.6	24.1	3,502
Fir/Spruce/Mt. Hemlock	80.3	17.0	22.8	38.8	31.5	4,156
Lodgepole Pine	53.8	11.4	13.0	24.0	27.0	2,108
Aspen/Birch	58.7	11.3	12.1	28.5	58.8	2,497
Western Oak	20.1	3.8	2.3	27.1	38.0	2,910
Minor Types and Nonstocked	15.9	2.9	4.1	23.3	25.2	5,207
All	35.2	7.2	6.1	25.1	26.4	41,989
United States (lower 48 states)	**60.6**	**12.0**	**10.1**	**18.3**	**61.5**	**251,558**

The FIADB snapshot dataset includes some of the older periodic data from many states and, where necessary for some states, draws older survey data from other FIA-based datasets. The Resources Planning Act Assessment (RPA) database, which includes periodic summaries of state inventories, is one example. The basic difference between the RPA database and the FIADB is that the FIADB includes some informative additional details such as individual-tree data. Having only plot-level information (such as volume per hectare), limits the conversion

to biomass. This does not constitute a substantial difference for the overall state-wide estimates, but it does affect plot-level precision (Smith et al. 2004a). In the past, FIA made their data available in tree-level Eastwide (Hansen et al. 1992) or Westwide (Woudenberg and Farrenkopf 1995) formats, which included inventories for Eastern and Western states, respectively. The current Inventory estimates rely in part on older tree-level data that are not available on the current FIADB site and older tree-level inventories from the Southern FIA unit (Heatherly 2006). All FIADB snapshot data used for C stock estimates were obtained from USDA Forest Service (2006b). More complete information about these data is available on the internet at the Forest Inventory and Analysis Datacenter (USDA Forest Service 2006b), and the results of these databases are found in Waddell et al. (1989), Smith et al. (2001), and Smith et al. (2004b).

An historical focus of the FIA program was to provide information on timber resources of the United States. For this reason, prior to 1998, some forest land, which were less productive or reserved (i.e., land where harvesting was prohibited by law), were less intensively surveyed. This generally meant that on these less productive lands, forest type and area were identified but data were not collected on individual tree measurements. The practical effect that this evolution in inventories has had on estimating forest C stocks from 1990 through the present is that some older surveys of lands do not have the stand-level values for merchantable volume of wood or stand age. Any data gaps identified in the surveys taken before 1998 were filled by assigning average C densities calculated from the more complete, later inventories from the respective states. The overall effect of this necessary approach to generate estimates for C stock is that no net change in C density occurs on those lands with gaps in past surveys.

Estimating C stocks from forest inventory data

For each inventory summary in each state, data are converted to C units or augmented by other ecological data. This collection of conversion factors and models is referred to as FORCARB2 (Birdsey and Heath 1995, Birdsey and Heath 2001, Heath et al. 2003, Smith et al. 2004a). The conversion factors and model coefficients are usually categorized by region, and forest type. Classifications for both region and forest type are subject to change depending on the particular coefficient set. Thus, region and type are specifically defined for each set of estimates. Factors are applied to the survey data at the scale of FIA inventory plots. The results are estimates of C density (Mg per hectare) for the various forest pools. C density for live trees, standing dead trees, understory vegetation, down dead wood, litter, and soil organic matter are estimated. All non-soil pools except litter can be separated into aboveground and belowground components. The live tree and understory C pools are pooled as biomass in this inventory. Similarly, standing dead trees and down dead wood are pooled as dead wood in this inventory. C stocks and fluxes for *Forest Land Remaining Forest Land* are reported in pools following IPCC (2003).

Live tree C pools

The tree C pools include aboveground and belowground (coarse root) C mass of live trees. Separate estimates are made for full-tree and aboveground-only biomass in order to estimate the belowground component. Most tree C estimates are based on Jenkins et al. (2003) and are functions of species groups and diameter. For example, the equation for estimating aboveground biomass for a live tree of a species in the aspen/alder/cottonwood/willow group is:

$$\text{Biomass (kg dry weight)} = e^{(-2.2094 + 2.3867 \times \ln(\text{diameter}))}$$

Diameter is cm at diameter breast height (d.b.h.), which is measured at 1.37 m above the forest floor. C is calculated by multiplying biomass by 0.5 because biomass is 50 percent of dry weight (IPCC/UNEP/OECD/IEA 1997). A full set of coefficients can be found in Jenkins et al. (2003; Table 4). Belowground root biomass is estimated as a ratio of roots to total aboveground biomass. The equation for ratio of root biomass of a live tree in the aspen/alder/cottonwood/willow group is:

$$\text{Ratio} = e^{(-1.6911 + 0.8160/\text{diameter})}$$

Belowground biomass is calculated by multiplying the ratio by total aboveground biomass. A full set of coefficients can be found in Jenkins et al. (2003; Table 6). The C per tree is summed for each plot, and multiplied by the appropriate expansion factors to obtain a C stock estimate for the plot.

Some inventory data do not provide measurements of individual trees; tree C in these plots is estimated from plot-level growing stock volume of live trees and equations given in Table A- 192 and Table A- 193. These equations are updates of those in Smith et al. (2003), modified to reduce error and correspond to common forest types defined by inventories. Separate estimates are made for whole-tree and aboveground-only biomass based on forest type group and region. The belowground portion is determined as the difference between the two estimates. C density is estimated based on the growing stock volume of the plot, where growing stock includes live trees of commercial species meeting specified stands of quality of vigor. Only trees 5.0 inches d.b.h. and larger are included in growing stock volume (Smith et al. 2004b). The full sets of coefficients are in Table A- 192 and Table A- 193. For example, the total C in tree biomass per hectare of aspen-birch in the North averages 8.1 Mg C/ha if growing-stock volume is zero. If growing-stock volume is greater than zero, the estimate is in two parts. Average C density of non-growing-stock trees (sapling and cull trees) is 14.3 Mg C/ha, and the equation for C in growing-stock trees is:

$$\text{Growing-stock trees (Mg C/ha)} = e^{(-0.337 + \ln(\text{volume}) \times 0.933)}$$

Units for volume are m^3/ha.

Table A- 192. Coefficients for estimating carbon density of live trees (above- and below-ground, MgC/ha) by region and type for plot-level data such as RPA data.[a]

Region[b]	Forest type group[c]	Carbon density, if Growing Stock Volume (GSV)=0	C density for non-Growing Stock (GS), if GSV > 0	Coefficient A	Coefficient B
North	Aspen/Birch	8.138	14.335	-0.337	0.933
	Elm/Ash/Cottonwood	16.187	18.707	-0.206	0.920
	Maple/Beech/Birch	6.938	17.054	-0.170	0.925
	Oak/Hickory	13.083	15.914	-0.079	0.932
	Hardwood minor types	10.376	14.127	0.002	0.890
	Oak/Pine	4.079	15.473	-0.146	0.908
	Ponderosa Pine & Exotic Softwood	2.595	6.895	-0.074	0.886
	Loblolly/Shortleaf Pine	6.277	9.766	-0.415	0.943
	Spruce/Fir	6.424	16.903	-0.487	0.947
	White/Red/Jack Pine	3.908	12.117	-0.349	0.924
	Softwood minor types	6.277	17.234	-0.380	0.970
	Non-stocked	1.054	1.238	-0.174	0.866
Pacific Coast	Alder/Maple	8.425	4.444	0.056	0.828
	Other Western Hardwoods	8.425	10.483	0.041	0.864
	Tanoak/Laurel	8.425	10.203	-0.167	0.917
	Western Oak	8.425	7.400	0.344	0.850
	Hardwood minor types	8.425	4.802	0.333	0.770
	California Mixed Conifer	10.102	4.727	0.137	0.843
	Douglas-fir	2.752	4.961	0.180	0.834
	Fir/Spruce/Mt. Hemlock	10.102	6.462	0.171	0.834
	Hemlock/Sitka Spruce	10.102	8.034	0.085	0.830
	Lodgepole Pine	10.102	5.733	-0.129	0.857
	Pinyon/Juniper	22.552	5.065	-0.070	0.842
	Ponderosa Pine	10.102	2.262	0.145	0.813
	Western Larch	10.102	5.254	-0.264	0.853
	Softwood minor types	10.102	6.771	0.466	0.783
	RPA Western Hardwoods	8.425	7.460	0.302	0.831
	Non-stocked	0.880	0.300	0.049	0.806
Rocky Mountain	Aspen/Birch	4.594	9.516	0.324	0.792
	Harwood minor types	4.866	11.844	0.266	0.814
	Douglas-fir	1.987	5.363	0.331	0.825
	Fir/Spruce/Mt. Hemlock	1.987	6.693	0.065	0.825
	Lodgepole Pine	1.080	8.051	0.003	0.804
	Other Western Softwoods	1.987	12.217	0.361	0.796
	Ponderosa Pine	1.987	5.574	0.382	0.771
	Softwood minor types	1.987	5.496	-0.152	0.836
	RPA Western Hardwood	13.714	11.678	0.246	0.807
	Pinyon/Juniper	22.927	23.301	0.254	0.794
	West. Oak/Other West. Hardwoods	14.441	18.544	0.215	0.796

	Non-stocked	1.111	0.568	0.257	0.732
South	Elm/Ash/Cottonwood	12.841	21.633	-0.144	0.896
	Oak/Gum/Cypress	7.176	23.919	-0.216	0.907
	Oak/Hickory	14.594	20.007	-0.031	0.886
	Hardwood minor types	47.316	40.194	-0.442	0.960
	Oak/Pine	4.106	17.933	-0.086	0.858
South	Loblolly/Shortleaf Pine	3.892	12.466	0.206	0.773
	Longleaf/Slash Pine	4.441	8.694	0.110	0.772
	Softwood minor types	7.161	20.189	-0.085	0.868
	Non-stocked	0.467	0.943	0.019	0.734

[a] Prediction of C in growing-stock trees is based on exp(A + B*ln(growing stock volume)).
[b] Regions are North (CT, DE, IA, IL, IN, KS, MA, MD, ME, MI, MN, MO, ND, NE, NH, NJ, NY, OH, PA, RI, SD, VT, WI, WV); Pacific Coast (CA, OR, WA); Rocky Mountain (AZ, CO, ID, MT, NM, NV, UT, WY); and South (AL, AR, FL, GA, KY, LA, MS, NC, OK, SC, TN, TX, VA).
[c] Forest type groups are identified in appendix D of the FISDB Users Guide (USDA Forest Service, 2006c).

Table A- 193. Coefficients for estimating carbon density of live trees (aboveground only, MgC/ha) by region and type for plot-level data such as RPA data.[a] Density of live trees (aboveground)[a]

Region[b]	Forest type group[c]	Carbon density, if Growing Stock Volume (GSV)=0	C density for non-GS, if GSV > 0	Coefficient A	Coefficient B
North	Aspen/Birch	6.697	11.880	-0.521	0.934
	Elm/Ash/Cottonwood	13.585	15.653	-0.387	0.922
	Maple/Beech/Birch	5.762	14.219	-0.352	0.926
	Oak/Hickory	10.960	13.306	-0.260	0.933
	Hardwood minor types	8.647	11.796	-0.166	0.888
	Oak/Pine	3.368	12.881	-0.335	0.909
	Ponderosa Pine & Exotic Softwood	2.116	5.671	-0.269	0.886
	Loblolly/Shortleaf Pine	5.098	8.070	-0.620	0.946
	Spruce/Fir	5.206	13.833	-0.684	0.948
	White/Red/Jack Pine	3.174	10.010	-0.548	0.926
	Softwood minor types	5.098	14.246	-0.570	0.971
	Non-stocked	0.880	1.032	-0.357	0.866
Pacific Coast	Alder/Maple	7.006	3.676	-0.138	0.830
	Other Western Hardwoods	7.006	8.709	-0.154	0.867
	Tanoak/Laurel	7.006	8.469	-0.355	0.918
	Western Oak	7.006	6.163	0.167	0.850
	Hardwood minor types	7.006	3.974	0.136	0.773
	California Mixed Conifer	8.309	3.883	-0.061	0.844
	Douglas-fir	2.235	4.072	-0.017	0.835
	Fir/Spruce/Mt. Hemlock	8.309	5.285	-0.027	0.835
	Hemlock/Sitka Spruce	8.309	6.586	-0.113	0.831
	Lodgepole Pine	8.309	4.674	-0.327	0.858
	Pinyon/Juniper	18.583	4.170	-0.263	0.842
	Ponderosa Pine	8.309	1.849	-0.053	0.814
	Western Larch	8.309	4.282	-0.461	0.853
	Softwood minor types	8.309	5.563	0.267	0.784
	RPA Western Hardwoods	7.006	6.202	0.119	0.831
	Non-stocked	0.724	0.247	-0.146	0.808
Rocky Mountain	Aspen/Birch	3.798	7.914	0.139	0.793
	Harwood minor types	4.027	9.936	0.084	0.815
	Douglas-fir	1.616	4.388	0.134	0.826
	Fir/Spruce/Mt. Hemlock	1.616	5.466	-0.133	0.826
	Lodgepole Pine	0.871	6.571	-0.195	0.805
	Other Western Softwoods	1.616	10.031	0.165	0.797
	Ponderosa Pine	1.616	4.569	0.185	0.772
	Softwood minor types	1.616	4.473	-0.350	0.837
	RPA Western Hardwood	11.341	9.704	0.054	0.809
	Pinyon/Juniper	18.867	19.173	0.059	0.794
	West. Oak/Other West. Hardwoods	11.942	15.353	0.021	0.796
	Non-stocked	0.916	0.466	0.061	0.733
South	Elm/Ash/Cottonwood	10.749	18.129	-0.323	0.897

Oak/Gum/Cypress	5.987	20.004	-0.400	0.909
Oak/Hickory	12.223	16.731	-0.215	0.888
Hardwood minor types	39.737	33.739	-0.631	0.964
Oak/Pine	3.394	14.923	-0.277	0.859
Loblolly/Shortleaf Pine	3.172	10.288	0.012	0.773
Longleaf/Slash Pine	3.634	7.176	-0.088	0.773
Softwood minor types	5.893	16.751	-0.280	0.869
Non-stocked	0.388	0.788	-0.171	0.735

[a] Prediction of aboveground C in growing-stock trees is based on exp(A + B*ln(growing stock volume).
[b] Regions are North (CT, DE, IA, IL, IN, KS, MA, MD, ME, MI, MN, MO, ND, NE, NH, NJ, NY, OH, PA, RI, SD, VT, WI, WV); Pacific Coast (CA, OR, WA); Rocky Mountain (AZ, CO, ID, MT, NM, NV, UT, WY); and South (AL, AR, FL, GA, KY, LA, MS, NC, OK, SC, TN, TX, VA).
[c] Forest type groups are identified in appendix D of the FISDB Users Guide (USDA Forest Service, 2006c).

Understory vegetation

Understory vegetation is a minor component of biomass. Understory vegetation is defined as all biomass of undergrowth plants in a forest, including woody shrubs and trees less than one-inch d.b.h. In this inventory, it is assumed that 10 percent of understory C mass is belowground. This general root-to-shoot ratio (0.11) is near the lower range of temperate forest values provided in IPCC (2003) and was selected based on two general assumptions: ratios are likely to be lower for light-limited understory vegetation as compared with larger trees, and a greater proportion of all root mass will be less than 2 mm diameter.

Estimates of C density are based on information in Birdsey (1996), which was applied to FIA permanent plots. These were fit to the equation:

$$\text{Ratio} = e^{(A - B \times \ln(\text{live tree C density})}$$

In this equation, "ratio" is the ratio of understory C density (Mg C/ha) to live tree C density (above- and below-ground) in Mg C/ha. An additional coefficient is provided as a maximum ratio; that is, any estimate predicted from the equation that is greater than the maximum ratio is set equal to the maximum ratio. A full set of coefficients is in Table A- 194. Regions and forest types are the same classifications described in Smith et al. (2003). As an example, the basic calculation for understory C in aspen-birch forests in the Northeast is:

$$\text{Understory (Mg C/ha)} = (\text{live tree C density}) \times e^{(0.855 - 1.03 \times \ln(\text{tree C density})}$$

This calculation is followed by three possible modifications. First, the maximum value for the ratio is set to 2.02 (see value in column "maximum ratio"); this also applies to stands with zero tree C, which is undefined in the above equation. Second, the minimum ratio is set to 0.005 (Birdsey 1996). Third, nonstocked and pinyon/juniper stands are set to constant ratios defined by coefficient A.

Table A- 194. Coefficients for estimating the ratio of carbon density of understory vegetation (above- and belowground, MgC/ha)[a] by region and forest type. The ratio is multiplied by tree carbon density on each plot to produce understory vegetation.

Region[b]	Forest Type[b]	A	B	Maximum ratio[c]
NE	Aspen-Birch	0.855	1.032	2.023
	MBB/Other Hardwood	0.892	1.079	2.076
	Oak-Hickory	0.842	1.053	2.057
	Oak-Pine	1.960	1.235	4.203
	Other Pine	2.149	1.268	4.191
	Spruce-Fir	0.825	1.121	2.140
	White-Red-Jack Pine	1.000	1.116	2.098
	Nonstocked	2.020	2.020	2.060
NLS	Aspen-Birch	0.777	1.018	2.023
	Lowland Hardwood	0.650	0.997	2.037
	Maple-Beech-Birch	0.863	1.120	2.129
	Oak-Hickory	0.965	1.091	2.072
	Pine	0.740	1.014	2.046
	Spruce-Fir	1.656	1.318	2.136
	Nonstocked	1.928	1.928	2.117
NPS	Conifer	1.189	1.190	2.114

	Lowland Hardwood	1.370	1.177	2.055
	Maple-Beech-Birch	1.126	1.201	2.130
	Oak-Hickory	1.139	1.138	2.072
	Oak-Pine	2.014	1.215	4.185
	Nonstocked	2.052	2.052	2.072
	Douglas-fir	2.084	1.201	4.626
	Fir-Spruce	1.983	1.268	4.806
	Hardwoods	1.571	1.038	4.745
PSW	Other Conifer	4.032	1.785	4.768
	Pinyon-Juniper	4.430	4.430	4.820
	Redwood	2.513	1.312	4.698
	Nonstocked	4.431	4.431	4.626
	Douglas-fir	1.544	1.064	4.626
	Fir-Spruce	1.583	1.156	4.806
	Hardwoods	1.900	1.133	4.745
PWE	Lodgepole Pine	1.790	1.257	4.823
	Pinyon-Juniper	2.708	2.708	4.820
	Ponderosa Pine	1.768	1.213	4.768
	Nonstocked	4.315	4.315	4.626
	Douglas-fir	1.727	1.108	4.609
	Fir-Spruce	1.770	1.164	4.807
	Other Conifer	2.874	1.534	4.768
PWW	Other Hardwoods	2.157	1.220	4.745
	Red Alder	2.094	1.230	4.745
	Western Hemlock	2.081	1.218	4.693
	Nonstocked	4.401	4.401	4.589
	Douglas-fir	2.342	1.360	4.731
	Fir-Spruce	2.129	1.315	4.749
	Hardwoods	1.860	1.110	4.745
RMN	Lodgepole Pine	2.571	1.500	4.773
	Other Conifer	2.614	1.518	4.821
	Pinyon-Juniper	2.708	2.708	4.820
	Ponderosa Pine	2.099	1.344	4.776
	Nonstocked	4.430	4.430	4.773
	Douglas-fir	5.145	2.232	4.829
	Fir-Spruce	2.861	1.568	4.822
	Hardwoods	1.858	1.110	4.745
RMS	Lodgepole Pine	3.305	1.737	4.797
	Other Conifer	2.134	1.382	4.821
	Pinyon-Juniper	2.757	2.757	4.820
	Ponderosa Pine	3.214	1.732	4.820
	Nonstocked	4.243	4.243	4.797
	Bottomland Hardwood	0.917	1.109	1.842
	Misc. Conifer	1.601	1.129	4.191
	Natural Pine	2.166	1.260	4.161
SC	Oak-Pine	1.903	1.190	4.173
	Planted Pine	1.489	1.037	4.124
	Upland Hardwood	2.089	1.235	4.170
	Nonstocked	4.044	4.044	4.170
	Bottomland Hardwood	0.834	1.089	1.842
	Misc. Conifer	1.601	1.129	4.191
	Natural Pine	1.752	1.155	4.178
SE	Oak-Pine	1.642	1.117	4.195
	Planted Pine	1.470	1.036	4.141
	Upland Hardwood	1.903	1.191	4.182
	Nonstocked	4.033	4.033	4.182

[a]Prediction of ratio of understory C to live tree C is based on the equation: Ratio=exp(A-B*ln(tree_carbon_tph)), where "ratio" is the ratio of understory C density to live tree (above-and below- ground) C density, and "tree_carbon_density" is live tree (above-and below- ground) C density in Mg C/ha.

[b] Regions and types as defined in Smith et al. (2003)

[c]Maximum ratio: any estimate predicted from the equation that is greater than the maximum ratio is set equal to the maximum ratio.

Dead Wood

The standing dead tree C pools include aboveground and belowground (coarse root) mass. Estimates for standing dead tree C are not based on FIA standing dead tree data because of yet unresolved problems with consistency among the state inventories. Instead, the estimates are based a ratio of growing stock volume of live trees by region and forest type groups, applied at the FIA plot-level. The standing dead tree equations estimate mass; they are converted to C mass by multiplying by 0.5. An example calculation for standing dead tree C in aspen-birch forests in the Northeast is:

$$\text{Dry weight (Mg/ha)} = 1.0 \times (\text{growing stock volume})^{0.499}$$

It is multiplied by 0.5 to obtain Mg C/ha. All coefficients are provided in Table A- 195. Note that nonstocked stands are assigned a constant C density (the value of Coefficient A).

Down dead wood is defined as pieces of dead wood greater than 7.5 cm diameter, at transect intersection, that are not attached to live or standing dead trees. Down dead wood includes stumps and roots of harvested trees. Ratio estimates of down dead wood to live tree biomass were developed using FORCARB2 simulations and applied at the plot level (Smith et al. 2004a). Estimates for down dead wood are according to the region and forest type classifications described in Smith et al. (2003). A full set of ratios is provided in Table A-196. An example calculation for down dead wood in aspen-birch forests in the Northeast is:

$$\text{C density (Mg C/ha)} = (\text{live tree C density, above- and below- ground}) \times (0.078) = 7.8\% \text{ of live tree C}$$

Conversion to C mass is not necessary because the live tree value is already in terms of C.

Table A- 195. Coefficients for estimating standing dead tree carbon (MgC/ha) by region and forest type group.[a]

Region[b]	Forest type group[c]	Coefficient A	Coefficient B
MTN	Douglas-fir	3.935	0.312
	Fir/Spruce/Mt. Hemlock	4.550	0.358
	Hemlock/Sitka Spruce	1.000	0.569
	Lodgepole Pine	1.177	0.501
	Ponderosa Pine	1.000	0.455
	Nonstocked	12.855	--
MTS	Douglas-fir	2.200	0.460
	Fir/Spruce/Mt. Hemlock	6.923	0.293
	Lodgepole Pine	1.177	0.501
	Ponderosa Pine	1.944	0.292
	Nonstocked	4.232	--
NC	Aspen/Birch	1.962	0.400
	Elm/Ash/Cottonwood	3.755	0.253
	Maple/Beech/Birch	3.442	0.219
	Planted Pine	1.000	0.298
	Oak/Hickory	2.949	0.236
	Oak/Pine	1.364	0.394
	Spruce/Fir	1.320	0.472
	White/Red/Jack Pine	2.844	0.266
	Nonstocked	2.634	--
NE	Aspen/Birch	1.000	0.499
	Elm/Ash/Cottonwood	4.992	0.134
	Maple/Beech/Birch	3.041	0.306
	Oak/Hickory	3.332	0.191
	Oak/Pine	1.725	0.311
	Spruce/Fir	5.893	0.190
	White/Red/Jack Pine	2.841	0.254
	Nonstocked	2.876	--
PC	California Mixed Conifer	1.000	0.608
	Douglas-fir	1.237	0.559
	Douglas-fir Planted	10.145	0.112
	Fir/Spruce/Mt. Hemlock	4.235	0.415
	Hemlock/Sitka Spruce	1.546	0.562
	Redwood	5.385	0.287

	Nonstocked	7.377	--
	Elm/Ash/Cottonwood	2.393	0.284
	Loblolly/Shortleaf Pine	1.203	0.271
	Loblolly/Shortleaf Planted	1.000	0.138
SC	Oak/Gum/Cypress	4.234	0.121
	Oak/Hickory	2.396	0.186
	Oak/Pine	1.133	0.337
	Nonstocked	0.286	--
	Elm/Ash/Cottonwood	1.358	0.476
	Loblolly/Shortleaf Pine	1.000	0.324
	Loblolly/Shortleaf Planted	1.000	0.265
SE	Oak/Gum/Cypress	1.770	0.329
	Oak/Hickory	2.256	0.257
	Oak/Pine	1.000	0.351
	Nonstocked	0.349	--
South	Longleaf/Slash Pine	1.000	0.184
	Longleaf/Slash Planted	1.000	0.106
	Alder/Maple	2.190	0.466
	Aspen/Birch	3.062	0.376
	Pinyon/Juniper	3.163	0.100
West	Tanoak/Laurel	1.000	0.593
	Western Hardwood/Woodlands	5.595	0.181
	Western Larch	2.049	0.449
	Western Oak	1.996	0.348

[a]Standing dead tree C is based on the equation: mass (Mg/ha) = A* (live-tree growing stock volume)^B. Note that nonstocked stands are assigned a constant C density (the value listed under coefficient A). Note that the standing dead tree equations are for biomass. To convert to C mass, multiply by 0.5.
[b]Regions are PC (CA,OR-West,WA-West), MTN (OR-East,WA-East,ID,MT), MTS (AZ,CO,NM,NV,UT,WY), West (regions PC, MTN, and MTS), NC (IA, IL, IN, KS, MI, MN, MO, ND, NE, SD, WI), NE (CT, DE, MA, MD, ME, NH, NJ, NY, OH, PA, RI, VT, WV), SC (AL, AR, KY, LA, MS, OK, TN, TX), SE (FL, GA, NC, SC, VA), and South (regions SC and SE).
[c]Forest types are described in appendix D of the FISDB users guide (USDA Forest Service, 2006c). Minor forest types within a region that are not explicitly defined/listed in the table of coefficients are assigned to a similar hardwood or softwood forest type.

Table A-196. Ratio for estimating down dead wood by region and forest type. The ratio is multiplied by the live tree carbon density on a plot to produce down dead wood carbon density (MgC/ha).

Region[a]	Forest type[a]	Ratio	Region (cont'd)	Forest type (cont'd)	Ratio (cont'd)
	Aspen-Birch	0.078		Douglas-fir	0.100
	MBB/Other Hardwood	0.071		Fir-Spruce	0.090
	Oak-Hickory	0.068		Other Conifer	0.073
NE	Oak-Pine	0.061	PWW	Other Hardwoods	0.062
	Other Pine	0.065		Red Alder	0.095
	Spruce-Fir	0.092		Western Hemlock	0.099
	White-Red-Jack Pine	0.055		Nonstocked	0.020
	Nonstocked	0.019		Douglas-fir	0.062
	Aspen-Birch	0.081		Fir-Spruce	0.100
	Lowland Hardwood	0.061		Hardwoods	0.112
	Maple-Beech-Birch	0.076	RMN	Lodgepole Pine	0.058
NLS	Oak-Hickory	0.077		Other Conifer	0.060
	Pine	0.072		Pinyon-Juniper	0.030
	Spruce-Fir	0.087		Ponderosa Pine	0.087
	Nonstocked	0.027		Nonstocked	0.018
	Conifer	0.073		Douglas-fir	0.077
	Lowland Hardwood	0.069		Fir-Spruce	0.079
	Maple-Beech-Birch	0.063		Hardwoods	0.064
NPS	Oak-Hickory	0.068	RMS	Lodgepole Pine	0.098
	Oak-Pine	0.069		Other Conifer	0.060
	Nonstocked	0.026		Pinyon-Juniper	0.030
PSW	Douglas-fir	0.091		Ponderosa Pine	0.082
	Fir-Spruce	0.109		Nonstocked	0.020
	Hardwoods	0.042	SC	Bottomland Hardwood	0.063
	Other Conifer	0.100		Misc. Conifer	0.068

	Pinyon-Juniper	0.031
	Redwood	0.108
	Nonstocked	0.022
PWE	Douglas-fir	0.103
	Fir-Spruce	0.106
	Hardwoods	0.027
	Lodgepole Pine	0.093
	Pinyon-Juniper	0.032
	Ponderosa Pine	0.103
	Nonstocked	0.024

	Natural Pine	0.068
	Oak-Pine	0.072
	Planted Pine	0.077
	Upland Hardwood	0.067
	Nonstocked	0.013
SE	Bottomland Hardwood	0.064
	Misc. Conifer	0.081
	Natural Pine	0.081
	Oak-Pine	0.063
	Planted Pine	0.075
	Upland Hardwood	0.059
	Nonstocked	0.012

[a] Regions and types as defined in Smith et al. (2003).

Litter carbon

C of the litter layer is sampled on a subset of the FIA plots. However, the data are not yet available. Litter C is the pool of organic C (including material known as duff, humus, and fine woody debris) above the mineral soil and includes woody fragments with diameters of up to 7.5 cm. Estimates are based on equations of Smith and Heath (2002) and applied at the plot level. The equations describe processes for decay or loss of forest floor following harvest and the net accumulation of new forest floor material following stand growth. For example, total forest floor C at a given number of years after a clearcut harvest for aspen-birch forests in the North is:

$$\text{Total forest floor C (Mg C/ha)} = (18.4\times\text{years})/(53.7+\text{years}) + 10.2\times e^{(-\text{years}/9.2)}$$

See Table 4 of Smith and Heath (2002) for the complete set of coefficients. Note that these are direct estimates of C density, the 0.5 conversion does not apply to litter.

Soil organic carbon

Soil organic carbon (SOC) is currently sampled to a 20 cm depth on subsets of FIA plots, however, these data are not available United States-wide. Thus, estimates of SOC are based on the national STATSGO spatial database (USDA 1991), and the general approach described by Amichev and Galbraith (2004). In their procedure, SOC was calculated for the conterminous United States using the STATSGO database, and data gaps were filled by representative values from similar soils. Links to region and forest type groups were developed with the assistance of the USDA Forest Service FIA Geospatial Service Center by overlaying FIA forest inventory plots on the soil C map. The average SOC densities are provided in Table A- 191.

Emissions from fires

CO_2

As stated in other sections, the forest inventory approach implicitly accounts for emissions due to disturbances because only C remaining in the forest is estimated. Net C stock change is estimated by subtracting consecutive C stock estimates. A disturbance removes C from the forest. The inventory data, on which net C stock estimates are based, already reflects the C loss. Therefore, estimating the CO_2 emissions from a disturbance such as fire, and adding those to the net CO_2 change in forests is eventually double-counting the loss from fire because the inventory data already reflect the loss.

There is interest, however, in the size of the CO_2 emissions from disturbances such as fire. The IPCC (2003) methodology was employed. Wildfire area statistics are available, but they include non-forest land, such as shrublands and grasslands. It was thus necessary to develop a rudimentary estimate of the percent of area burned in forest by multiplying the reported area burned by the ratio of area of forest land considered to be under protection from fire as compared to the total area considered to be under protection from fire. The average C density in the lower 48 states for aboveground biomass C, dead wood C, and litter layer is 89.0 Mg/ha. A default value of 0.40 from IPCC (2003) was assumed for the amount of biomass burned by wildfire (combustion factor value). Thus, approximately 35.6 Mg/ha C is estimated to be emitted by wildfire. For Alaska, an average C density of 70 Mg/ha was estimated from Heath et al. (2003); this translates into 28 Mg/ha emitted. These were multiplied by estimates of

forest area burned by year and displayed in Table A-197. C estimates were multiplied by 3.667 to yield CO_2 units. Some combusted wood may continue to decay through time; however, as noted previously this is a rudimentary estimate which is already accounted for in the net C change estimates. Total CO_2 emissions for the 48 states and Alaska in 2005 were estimated to be 126.9 Tg/yr.

Table A-197. Areas (hectares) from wildfire statistics and corresponding estimates of carbon and CO_2 (Tg/yr) emissions for the lower 48 states and Alaska[1]

	"Lower 48" states				Alaska			
Year	Area reported[2] (ha)	Forest area burned[3] (ha)	Carbon emitted (Tg/yr)	CO_2 emitted (Tg/yr)	Area reported[4] (ha)	Forest area burned[3] (ha)	Carbon emitted (Tg/yr)	CO_2 emitted (Tg/yr)
1990	915,985	327,109	11.6	42.7	1,290,717	336,159	9.4	34.5
1991	197,107	70,389	2.5	9.2	708,465	184,515	5.2	18.9
1992	939,804	335,615	11.9	43.8	54,778	14,267	0.4	1.5
1993	646,406	230,839	8.2	30.1	288,588	75,161	2.1	7.7
1994	1,804,209	644,304	22.9	84.1	107,534	28,007	0.8	2.9
1995	919,359	328,314	11.7	42.9	17,784	4,632	0.1	0.5
1996	2,469,444	881,867	31.4	115.1	242,515	63,162	1.8	6.5
1997	665,999	237,836	8.5	31.0	820,258	213,631	6.0	21.9
1998	893,934	319,234	11.4	41.7	48,867	12,727	0.4	1.3
1999	1,884,440	672,955	24.0	87.8	406,883	105,970	3.0	10.9
2000	3,102,299	1,107,867	39.4	144.6	306,062	79,712	2.2	8.2
2001	1,350,447	482,260	17.2	63.0	88,268	22,989	0.6	2.4
2002	1,923,967	687,071	24.5	89.7	883,576	230,122	6.4	23.6
2003	1,746,370	623,649	22.2	81.4	243,911	63,525	1.8	6.5
2004	108,258	38,660	1.4	5.0	2,639,840	687,530	19.3	70.6
2005	1,628,000	581,378	20.7	75. 9	1,887,407	491,564	13.8	50.5

[1] Note that these emissions have already been accounted for in the net C sequestration estimates (i.e., net flux already accounts for the amount sequestered minus any emissions)
[2] National Interagency Coordination Center (2006a).
[3] Ratios calculated using estimates in Smith et al. (2004b) and USDA Forest Service (1992).
[4] Alaska Department of Natural Resources (2006).

Non-CO_2

The IPCC (2003) Tier 2 default methodology was used to calculate non-CO_2 emissions. Emissions estimates for CH_4 are calculated by multiplying the total estimated C emitted from forest burned by gas-specific emissions ratios and conversion factors. N_2O emissions are calculated in the same manner, but are also multiplied by a N-C ratio of 0.01 as recommended by IPCC (2003). The equations used are:

$$CH_4 \text{ Emissions} = (\text{C released}) \times (\text{emission ratio}) \times 16/12$$

$$N_2O \text{ Emissions} = (\text{C released}) \times (\text{N/C ratio}) \times (\text{emission ratio}) \times 44/28$$

The resulting estimates are presented in Table A-198.

Table A-198. Estimated carbon released and estimates of non-CO_2 emissions (Tg/yr) for U.S. forests[1]

Year	C emitted (Tg/yr)	CH_4 emitted (Tg/yr)	N_2O (Tg/yr)
1990	21.058	0.337	0.002
1991	7.672	0.123	0.001
1992	12.347	0.198	0.001
1993	10.322	0.165	0.001
1994	23.721	0.380	0.003
1995	11.818	0.189	0.001
1996	33.163	0.531	0.004
1997	14.449	0.231	0.002
1998	11.721	0.188	0.001
1999	26.924	0.431	0.003
2000	41.672	0.667	0.005
2001	17.812	0.285	0.002
2002	30.903	0.494	0.003

2003	23.981	0.384	0.003
2004	20.627	0.330	0.002
2005	34.461	0.551	0.004

[1] Calculated based on C emission estimates in Table A-197 and default factors in IPCC (2003)

Carbon in Harvested Wood Products

Estimates of the harvested wood product (HWP) contribution to forest C sinks and emissions (hereafter called "HWP Contribution") are based on methods described in Skog (in preparation) using the WOODCARB II model. These are based on the methods suggested in IPCC (2006) for estimating HWP C. Key changes from estimates provided in previous years are indicated in the Recalculations Discussion section. The United States uses the production accounting approach (as in previous years) to report HWP contribution (Table A-199). Estimates resulting from use of the two alternative approaches, the stock change and atmospheric flow approaches, are also presented for comparison (Table A-200).

Estimates of 5 HWP variables that can be used to calculate HWP contribution for the stock change and atmospheric flow approaches for imports and exports are provided in Table A-199. The HWP variables estimated are:

1A) annual change of C in wood and paper products in use in the United States,

1B) annual change of C in wood and paper products in solid waste disposal sites (SWDS) in the United States,

2A) annual change of C in wood and paper product in use in the United States and other countries where the wood came from trees harvested in the United States,

2B) annual change of C in wood and paper products in SWDS in the United States and other countries where the wood came from trees harvested in the United States,

3) C in imports of wood, pulp, and paper to the United States,

4) C in exports of wood, pulp and paper from the United States, and

5) C in annual harvest of wood from forests in the United States.

Table A-199. Harvested wood products from wood harvested in United States – Annual additions carbon to stocks and total stocks

	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Net carbon additions per year (Tg C per year)																
Total Harvested wood carbon	36.0	33.5	33.4	32.5	32.8	31.5	29.8	31.1	30.3	31.6	29.8	24.6	25.3	25.0	27.8	28.2
Products in use	17.2	14.5	15.9	14.6	15.5	14.6	13.7	14.4	13.2	14.0	12.6	8.5	9.3	9.1	11.8	12.1
Solid wood products	14.1	11.6	12.4	12.0	11.9	11.1	11.3	11.6	11.3	12.0	11.7	10.0	10.6	9.8	11.3	11.8
Paper products	3.1	2.8	3.4	2.6	3.6	3.5	2.4	2.8	1.9	1.9	0.8	-1.5	-1.3	-0.7	0.5	0.3
Products in SWDS	18.8	19.1	17.5	17.9	17.3	16.9	16.1	16.7	17.1	17.6	17.2	16.1	16.0	15.9	16.0	16.1
Solid wood products	10.8	11.9	10.3	10.3	10.4	11.3	11.1	10.8	10.7	10.9	11.0	11.0	11.2	11.3	11.5	11.7
Paper products	7.9	7.2	7.2	7.5	6.8	5.6	5.0	5.9	6.4	6.7	6.2	5.1	4.8	4.6	4.5	4.4
Total Carbon stocks (Tg C)																
Total Harvested wood carbon	1,888	1,926	1,964	2,000	2,034	2,067	2,100	2,133	2,164	2,194	2,225	2,255	2,287	2,317	2,341	2,367
Products in use	1,184	1,203	1,221	1,238	1,252	1,268	1,283	1,298	1,313	1,327	1,341	1,354	1,368	1,381	1,389	1,399
Products in SWDS	704	724	744	762	781	799	817	834	851	867	884	901	919	936	952	968

Table A-200: Comparison of Net Annual Change in Harvested Wood Products Carbon Stocks Using Alternative Accounting Approaches

	HWP Contribution to LULUCF Emissions/ removals (Tg CO_2 Eq.)		
Inventory Year	**Stock Change Approach**	**Atmospheric Flow Approach**	**Production Approach**
1990	(145.9)	(154.7)	(131.9)
1991	(131.6)	(146.7)	(123.0)
1992	(134.2)	(145.8)	(122.3)
1993	(140.1)	(141.1)	(119.2)
1994	(142.4)	(142.3)	(120.2)
1995	(137.7)	(139.8)	(115.4)
1996	(133.4)	(133.5)	(109.4)
1997	(141.8)	(137.8)	(114.1)
1998	(149.6)	(132.5)	(111.2)
1999	(158.7)	(136.6)	(115.9)
2000	(152.0)	(129.0)	(109.1)
2001	(136.9)	(108.9)	(90.2)
2002	(142.1)	(109.6)	(92.6)
2003	(143.3)	(107.8)	(91.4)
2004	(171.8)	(117.8)	(102.1)
2005	(145.4)	(93.9)	(103.4)

Note: Parentheses indicate net C sequestration (i.e., a net removal of C from the atmosphere).

Table A-201. Harvested Wood Products Sectoral Background Data for LULUCF - United States (production approach)

Inventory year	1A Annual Change in stock of HWP in use from consumption	1B Annual Change in stock of HWP in SWDS from consumption	2A Annual Change in stock of HWP in use produced from domestic harvest	2B Annual Change in stock of HWP in SWDS produced from domestic harvest	3 Annual Imports of wood, and paper products + wood fuel, pulp, recovered paper, roundwood/ chips	4 Annual Exports of wood, and paper products + wood fuel, pulp, recovered paper, roundwood/ chips	5 Annual Domestic Harvest	6 Annual release of carbon to the atmosphere from HWP consumption (from fuelwood & products in use and products in SWDS)	7 Annual release of carbon to the atmosphere from HWP (including firewood) where wood came from domestic harvest (from products in use and products in SWDS)	8 HWP Contribution to LULUCF CO_2 emissions/ removals
	$\Delta C_{HWP\ IU\ DC}$	$\Delta C_{HWP\ SWDS\ DC}$	$\Delta C_{HWP\ IU\ DH}$	$\Delta C_{HWP\ SWDS\ DH}$	P_{IM}	P_{EX}	H	$\uparrow C_{HWP\ DC}$	$\uparrow C_{HWP\ DH}$	
									Gg C /yr	Gg CO_2 /yr
1990	17,000	22,700	17,200	18,800	12,700	15,100	142,300	100,100	106,300	(131,900)
1991	13,100	22,800	14,500	19,100	11,600	15,700	144,400	104,400	110,900	(123,000)
1992	15,700	20,900	15,900	17,500	12,900	16,000	139,400	99,600	106,000	(122,300)
1993	17,000	21,300	14,600	17,900	14,500	14,800	134,600	96,100	102,100	(119,200)
1994	18,200	20,600	15,500	17,300	15,700	15,700	134,800	95,900	102,000	(120,200)
1995	17,300	20,300	14,600	16,900	16,700	17,300	137,000	98,900	105,600	(115,400)
1996	17,000	19,400	13,700	16,100	16,700	16,700	134,500	98,100	104,600	(109,400)
1997	18,800	19,900	14,400	16,700	18,000	16,900	135,400	97,900	104,300	(114,100)
1998	20,300	20,500	13,200	17,100	19,700	15,000	135,000	98,900	104,700	(111,200)
1999	22,000	21,200	14,000	17,600	21,300	15,200	134,900	97,700	103,300	(115,900)
2000	20,500	21,000	12,600	17,200	22,400	16,200	134,500	99,300	104,700	(109,100)
2001	17,300	20,000	8,500	16,100	23,000	15,300	128,600	98,900	104,000	(90,200)
2002	18,600	20,100	9,300	16,000	24,600	15,700	127,600	97,700	102,300	(92,600)
2003	18,900	20,200	9,100	15,900	26,000	16,300	125,000	95,600	100,000	(91,400)
2004	26,100	20,700	11,800	16,000	31,600	16,900	130,500	98,300	102,600	(102,100)
2005p	19,200	20,400	12,100	16,100	31,500	17,400	132,900	107,300	104,700	(103,400)

Note: $\uparrow C_{HWP\ DC} = H + P_{IM} - P_{EX} - \Delta C_{HWP\ IU\ DC} - \Delta C_{HWP\ SWDS\ DC}$ AND $\uparrow C_{HWP\ DH} = H - \Delta C_{HWP\ IU\ DH} - \Delta C_{HWP\ SWDS\ DH}$. Parentheses indicate net C sequestration (i.e., a net removal of C from the atmosphere).
The letter "p" next to an inventory year indicates a preliminary estimate.

Annual estimates of variables 1A, 1B, 2A and 2B were calculated by tracking the additions to and removals from the pool of products held in end uses, that is products in use such as housing or publications, and the pool of products held in SWDS. In the case of variables 2A and 2B, the pools include products exported and held in other countries and the pools in the United States exclude products made from wood harvested in other countries. Solidwood products added to pools include lumber and panels. End-use categories for solidwood include single and multifamily housing, alteration and repair of housing, and other end uses. There is one product category and one end-use category for paper. Additions to and removals from pools are tracked beginning in 1900, with the exception that additions of softwood lumber to housing begins in 1800. Solidwood and paper product production and trade data are from USDA Forest Service and other sources (Hair and Ulrich 1963; Hair 1958; USDC Bureau of Census 1976; Ulrich, 1985, 1989; Steer 1948; AF&PA 2006a, 2006b; Howard 2003 & forthcoming).

The rate of removals from products in use and the rate of decay of products in SWDS is specified by first order (exponential) decay curves with given half-lives (time at which half of amount placed in use will have been discarded from use). Half-lives for products in use, determined after calibration of the model to meet two validation criteria, are shown in Table A-202. The first validation criteria is that the WOODCARB II model estimate of C in houses standing in 2001 needed to match an independent estimate of C in housing based on U.S. Census and USDA Forest Service survey data. The second criteria is that the WOODCARB II model estimate of wood and paper being discarded to SWDS needed to match EPA estimates of discards over the period 1990 to 2000. This calibration strongly influences the estimate of variable 1A, and to a lesser extent variable 2A. The calibration also determines the amounts going to SWDS.

Decay parameters for products in SWDS are shown in Table A-203. Estimates of 1B and 2B also reflect the change over time in the fraction of products discarded to SWDS (versus, burning or recycling) and the fraction of SWDS that are sanitary landfills versus dumps.

Variables 2A and 2B are used to estimate HWP contribution under the production accounting approach. A key assumption for estimating these variables is that products exported from the United States and held in pools in other countries have the same half lives for products in use, the same percentage of discarded products going to SWDS, and the same decay rates in SWDS. Summaries of net fluxes and stocks for harvested wood in products and SWDS are in Table A- 189 and Table A- 190.

Table A-202. Half-life of solidwood and paper products in end uses

Parameter	Value	Units
Half life of wood in single family housing 1920 and before	78.0	Years
Half life of wood in single family housing 1920 - 1939	78.0	Years
Half life of wood in single family housing 1940 - 1959	80.0	Years
Half life of wood in single family housing 1960 - 1979	81.9	Years
Half life of wood in single family housing 1980 +	83.9	Years
Ratio of multifamily half live to single family half life	0.61	
Ratio of repair and alterations half life to single family half life	0.30	
Half life for other solidwood product in end uses	38.0	Years
Half life of paper in end uses	2.54	Years

Table A-203. Parameters determining decay of wood and paper in SWDS

Parameter	Value	Units
Percentage of wood and paper in dumps that is subject to decay	100%	
Percentage of wood in landfills that is subject to decay	23%	
Percentage of paper in landfills that is subject to decay	56%	
Half life of wood in landfills / dumps (portion subject to decay)	29	Years
Half life of paper in landfills/ dumps (portion subject to decay)	14.5	Years

Uncertainty Analysis

The uncertainty analyses for total net flux of forest C (see uncertainty table in LULUCF chapter) are consistent with the IPCC-recommended Tier 2 methodology (IPCC 2003). Separate analyses are produced for forest ecosystem and HWP flux. The uncertainty estimates are from Monte Carlo simulations of the respective models and input data. Methods generally follow those described in Heath and Smith (2000b), Smith and Heath (2000), and Skog et al. (2004). Briefly, uncertainties surrounding input data or model processes are quantified as probability

densities (PDFs) so that a series of sample values can be randomly selected from the distributions. Model simulations are repeated a large number of times to numerically simulate the effect of the random PDF selections on estimated total C flux. The separate results from the ecosystem and HWP simulations are pooled for total uncertainty (see uncertainty table in LULUCF chapter).

Uncertainty surrounding current net C flux in forest ecosystems is based on uncertainty in the two most recent state or sub-state C stocks, which are summed to the national total. C stocks are based on plot-level estimates, and, therefore, uncertainty analysis starts at the plot level. Uncertainty surrounding C density (Mg/ha) is defined for each of six C pools for each inventory plot. Live trees are generally assigned normal PDFs, which are defined according to variability information in Jenkins et al. (2003) and the species and number of trees measured on each FIA plot. Plot-level live tree C estimates from RPA data are based on volume; these PDFs also include an additional level of uncertainty based on their respective regression equations. Similarly, the normally-distributed PDFs for standing dead trees are based on both volume regression and the individual-tree uncertainties related to the Jenkins et al. (2003) based estimates. Definitions of these normal distributions, which centered on expected values, depend on region, type, and specific plot information. Where data gaps—tree or volume—are identified for older inventory data, corresponding averages from later data are applied for live and standing dead tree C densities. Uniform PDFs with a range of ±90 percent of the average are used for these plots.

Distributions for the remaining C pools and the area expansion are triangular or uniform, which partly reflects the lower level of information available about these estimates. Down dead wood, understory, and litter are assigned triangular distributions with the mean at the expected value for each plot and the minimum and mode at 10 percent of the expected value. The use of these PDFs skewed to the right reflects the assumption that a small proportion of plots will have relatively high C densities. Joint sampling of PDFs is specified for two pairs of samples: understory PDF sampling is slightly negatively correlated with live tree PDF sampling, and down dead wood sampling is slightly positively correlated with live tree sampling. This also reflects the structure of the estimates, which are dependent on live tree C. Soil organic C is defined as a uniform PDF at ±50 percent of the mean. Uncertainty surrounding plot area is assigned a symmetric triangular distribution and defined following the discussion of accuracy standards for the inventory data in USDA Forest Service (2006c). The cumulative sum of random samples from the 7 PDFs (6 C and 1 expansion) defined a PDF for each plot. These were summed to the state or sub-state total stocks.

The three most significant changes in forest ecosystem simulation methods as compared with the previous inventory probably had very little effect on the overall analysis. The first of these is the greater uncertainty (and uniform PDF) assigned to the average tree C densities assigned in some of the older data. The second change is in the triangular PDFs for understory, down dead wood, and litter, which moved the minimum value from 0 to 10 percent of the mean. Finally, true Monte Carlo simulation is used in contrast to the Latin hypercube sampling used for the previous inventory. This method is employed to facilitate the requirement of 10,000 iterations in the simulation. A part of the QA/QC process included verifying that the PDFs were adequately sampled.

Uncertainty about net C flux in HWP is based on Skog et al. (2004). However, the uncertainty analysis simulation has been revised in conjunction with overall revisions in the HWP model (Skog in preparation). Latin hypercube sampling is the basis for the Monte Carlo simulation. Estimates of the HWP variables and HWP Contribution under the production approach are subject to many sources of uncertainty. A previous estimate of uncertainty in HWP Contribution under the production approach in this report was almost ±40 percent. This percentage error estimate turns out to be close to the actual adjustment of the estimates made in this report. A preliminary estimate of uncertainty is provided that evaluated the effect of uncertainty in 13 sources, including production and trade data and parameters used to make the estimate. Uncertain data and parameters include data on production and trade and factors to convert them to C, the Census-based estimate of C in housing in 2001, the EPA estimate of wood and paper discarded to SWDS for 1990 to 2000, the limits on decay of wood and paper in SWDS, the decay rate (half-life) of wood and paper in SWDS, the proportion of products produced in the United States made with wood harvested in the United States, and the rate of storage of wood and paper C in other countries that came from United States harvest, compared to storage in the United States.

3.13. Methodology for Estimating Net Changes in Carbon Stocks in Mineral and Organic Soils on Croplands and Grasslands

This sub-annex describes the methodologies used to calculate annual carbon (C) stock changes from mineral and organic soils under agricultural management, including *Cropland Remaining Cropland*, *Land Converted to Cropland*, *Grassland Remaining Grassland*, and *Land Converted to Grassland*. Three types of methodologies were applied: 1) a Tier 3 approach, employing the Century simulation model, 2) Tier 2 methods with country-specific stock change and emission factors; and 3) Tier 2 methods for estimating additional changes in mineral soil C stocks due to sewage sludge additions to soils and enrollment changes in the Conservation Reserve Program after 1997.

The soil C inventory uses a Tier 3 Approach for the majority of agricultural lands. This approach has several advantages over the IPCC Tier 1 or 2 approaches:

- it utilizes actual weather data at county scales, rather than a broad climate region classification, enabling quantification of inter-annual variability in C fluxes at finer spatial scales;
- the model uses a more detailed characterization of spatially-mapped soil properties that influence soil C dynamics, as opposed to the broad soil taxonomic classifications of the IPCC methodology;
- the simulation approach provides a more detailed representation of management influences and their interactions than are represented by a discrete factor-based approach in the Tier 1 and 2 methods; and
- soil C changes are estimated on a more continuous basis (monthly) as a function of the interaction of climate, soil, and land management, compared with the linear change between start and end of the inventory that is used with the Tier 1 and 2 methods.

The Century model was chosen as an appropriate tool for a Tier 3 application based on several criteria:

- The model was developed in the United States and has been extensively tested and verified for U.S. conditions. In addition, the model has been widely used by researchers and agencies in many other parts of the world for simulating soil C dynamics at local, regional and national scales (e.g., Brazil, Canada, India, Jordan, Kenya, Mexico).
- The model is capable of simulating cropland, grassland, forest, and savanna ecosystems and land-use transitions between these different land uses. It is, thus, well suited to model land-use change effects.
- The model was designed to simulate all major types of management practices that influence soil C dynamics, with the exception of cultivated organic soils and a few crops that have not been parameterized for Century simulations (e.g., rice, perennial/horticultural crops, and tobacco). For these latter cases, an IPCC Tier 2 method has been used.
- The model has input data requirements that were obtainable from existing national databases in the United States. The exceptions are CRP enrollment after 1997 and sewage sludge amendments to soils, which are not known at a sufficient resolution to use the Tier 3 model. Soil C stock changes associated with these practices are addressed with a Tier 2 method.

Century Model Description

The Century model simulates C (and also N, P, and S) dynamics, soil temperature, and water dynamics for cropland, grassland, forest, and savanna (mixed forest-grassland) systems. For the U.S. inventory application, only C and N dynamics have been included for several reasons: to simplify the analysis and reduce data requirements; and because P and S interactions are less important as determinants of land-use and management-induced changes in soil C stocks for U.S. agricultural systems.

The model consists of four main components: 1) soil organic matter and nutrient dynamics; 2) plant growth processes; 3) water and temperature dynamics; and 4) management practices. The model was designed to work with readily available input data: monthly weather data (e.g., temperature and precipitation); soil physical properties (e.g., soil texture, drainage condition, rooting depth); and information about land use/land cover (e.g., vegetation attributes) and management activities (see below). The model operates on a monthly time step (with weekly time steps used for soil water dynamics).

Dynamics of organic C and N (Figure A-9) are simulated for the surface and subsurface litter pools, and the top 20 cm of the soil profile; mineral N dynamics are simulated through the whole soil profile. Organic C and N stocks are represented by two plant litter pools (termed metabolic and structural) and three soil organic matter (SOM) pools (termed active, slow, and passive). The metabolic litter pool represents the easily decomposable constituents of plant residues, while the structural litter pool is composed of more recalcitrant, ligno-cellulose plant materials. The three SOM pools represent a gradient in decomposability, from active SOM (representing microbial biomass and associated metabolites) having a rapid turnover (months to years), to passive SOM (representing highly processed, humified, condensed decomposition products), which is highly recalcitrant, with mean residence times on the order of several hundred years. The slow pool represents decomposition products of intermediate stability, having a mean residence time on the order of decades and is the fraction that tends to change the most in terms of C content in response to changes in land use and management. Soil texture influences turnover rates of the slow and passive pools, whereby the clay and silt-sized mineral fraction of the soil provides physical protection from microbial attack, leading to slower decomposition and greater SOM stabilization in finely textured soils. Soil temperature and moisture, tillage disturbance, aeration, and other factors influence the decomposition and loss of C from the soil organic matter pools.

Figure A-9 . Flow diagram of Carbon submodel (A) and Nitrogen submodel (B)

The plant growth submodel simulates C assimilation through photosynthesis, N uptake, dry matter production, partitioning of C within the crop or forage, senescence, and mortality. The primary function of the growth submodel is to estimate the amount, type, and timing of organic matter inputs to soil and to represent the influence of the plant on soil water, temperature, and N balance. Yield and removal of harvested biomass are also simulated. Separate submodels are designed to simulate herbaceous plants (i.e., agricultural crops and grasses) and woody vegetation (i.e., trees and scrubs). Only the herbaceous plant submodel is currently used in the agricultural inventory applications. Maximum monthly net primary production (NPP) rate (i.e., a crop and forage species/variety parameter specified in the model input files) is modified by air temperature and available water to estimate a potential monthly NPP, which is then further subject to nutrient limitations in order to estimate actual NPP and biomass allocation.

The soil water balance submodel calculates water balance components and changes in soil water availability, which influences both plant growth and decomposition/nutrient cycling processes. The moisture content of soils are simulated through a multi-layer profile based on precipitation, snow accumulation and melting, interception, soil and canopy evaporation, transpiration, soil water movement, runoff, and drainage.

The final main component of the model is the management submodel, which includes options for specifying crop type, crop sequence (e.g., rotation), tillage, fertilization, organic matter addition (e.g., manure amendments), harvest (with variable residue removal), drainage, irrigation, burning, and grazing intensity. An input "schedule" file is used to simulate the timing of management activities and temporal trends; schedules can be organized into discrete time blocks to define a repeated sequence of events (e.g., a crop rotation or a frequency of disturbance such as a burning cycle for perennial grassland). Management options can be specified for any month of a year within a scheduling block, where management codes point to operation-specific parameter files (referred to as *.100 files), which contain the information used to simulate management effects within the model process algorithms. User-specified management activities can be defined by adding to or editing the contents of the *.100 files. Additional details of the model formulation are given in Parton et al. (1987, 1988, 1994) and Metherell et al. (1993), and archived copies of the model source code are available.

IPCC Tier 2 Method Description

The IPCC method has been developed to estimate C stock changes and CO_2 fluxes between soils and the atmosphere based on land-use and management activity (IPCC/UNEP/OECD/IEA 1997, Ogle et al. 2003). For mineral soils (i.e., all soil orders from the USDA taxonomic classification except Histosols), the IPCC inventory method uses reference C values to establish baseline C stocks that are modified based on agricultural activities using land-use change, tillage, and input factors. For this inventory, the standard IPCC approach was modified to use agricultural SOC stocks as the reference condition, rather than uncultivated soils under native vegetation. This modification was needed because soil measurements under agricultural management are much more common and

easily identified in the National Soil Survey Characterization Database (NRCS 1997), and, thus, these measurements formed the basis to estimate reference C stocks. Measurements of soils under native vegetation are uncommon in the major agricultural regions of the United States because most of the area has been converted into cropland. In addition, country-specific factors were derived for land-use change, tillage and input factors.

Organic soils used for agricultural production are treated in a separate calculation. These soils are made up of deep (greater than 30 cm) layers of organic material that can decompose at a steady rate over several decades following drainage for crop production or grazing (IPCC/UNEP/OECD/IEA 1997). The IPCC approach uses an emission factor to estimate annual losses of CO_2 from cultivated organic soils, rather than an explicit stock change approach.

Methodological Steps for Derivation of Soil Organic C Stock Change Estimates

The inventory of soil C stock changes in U.S. agricultural land combines Tier 2 and 3 approaches. A simulation-based Tier 3 approach was used to estimate soil C changes for most agricultural land (approximately 90 percent of total cropland and grassland) comprising the dominant cropping and grazing systems in the United States, for which the model has been well-tested. Estimates for the remaining area, comprised of less common crop systems (e.g., horticultural, vegetable, tobacco, rice) and all agricultural land occurring on drained organic soils, were developed using the Tier 2 approach. Tier 2 methods were used to estimate additional changes in mineral soil C stocks due to sewage sludge additions to soils, and enrollment changes in the Conservation Reserve Program after 1997. Most of the activity data sources were common to the Tier 2 and Tier 3 approaches, and, hence, they are described in an integrated fashion below. Additional activity data required for the methods are described in adjoining sections, followed by the computation steps.

Step 1: Derive Activity Data

Activity data were compiled for the Tier 3 Century model and Tier 2 IPCC methods, including climate data, soil characteristics, and land-use/management activity data. The first step was to obtain land-use/management activity data, and determine the land base for areas under agricultural management. The land base was subdivided into *Cropland Remaining Cropland*, *Land Converted to Cropland*, *Grassland Remaining Grassland*, and *Land Converted to Grassland*. Land parcels classified as *Cropland Remaining Cropland* and *Grassland Remaining Grassland* had been in these uses since 1982. Otherwise, the land parcel was classified as *Land Converted to Cropland* or *Land Converted to Grassland* based on the most recent use in the inventory time period (further elaboration provided in Step 1a). The areas modeled with Century and those estimated with the Tier 2 IPCC method were also subdivided. Parcels of land were allocated to the Tier 2 approach if: they occurred on organic soils; had a non-agricultural use such as forest and federal lands;[54] contained mineral soils that are very gravelly, cobbly, or shaley (i.e., classified as soils that have greater than 35 percent of soil volume comprised of gravel, cobbles or shale); or produced vegetables, perennial/horticultural crops, tobacco or rice, which was either grown continuously or in rotation with other crops. Finally, additional data were collected specific to each method, describing other key management activities (e.g., tillage management, fertilizer and manure addition rates) and environmental conditions (e.g., climate and soil characteristics).

Step 1a: Determine the Land Base and Classify Management Systems

Land Base—The National Resources Inventory (NRI) provided the basis for identifying the U.S. agricultural land base, and classifying parcels into *Cropland Remaining Cropland*, *Land Converted to Cropland*, *Grassland Remaining Grassland*, and *Land Converted to Grassland* (USDA-NRCS 2000). The NRI has a stratified multi-stage sampling design, where primary sample units are stratified on the basis of county and township boundaries defined by the U.S. Public Land Survey (Nusser and Goebel 1997). Within a primary sample unit, typically a 160-acre (64.75 ha) square quarter-section, three sample points are selected according to a restricted randomization procedure. Each point in the survey is assigned an area weight (expansion factor) based on other known areas and land-use information (Nusser and Goebel 1997). An extensive amount of soils, land-use, and land

[54] Federal land is treated as forest or nominal grassland for purposes of these calculations, although the specific use is not identified in the NRI survey (USDA-NRCS 2000).

management data are collected during each survey, which occurs every five years (Nusser et al. 1998).[55] Primary sources for data include aerial photography and remote sensing materials as well as field visits and county office records.

NRI points were included in the land base if they were identified as cropland or grassland[56] in the 1992 or 1997 surveys (Table A-204). Overall, more than 400,000 NRI points were included in the inventory calculations. Each point represents a specific land parcel based upon the weighted expansion factors. To subdivide land parcels into *Cropland Remaining Cropland*, *Land Converted to Cropland*, *Grassland Remaining Grassland*, and *Land Converted to Grassland*, the time series was divided into four inventory time periods: 1980-84, 1985-89, 1990-94 and 1995-2000, using NRI data from 1982, 1987, 1992, and 1997, respectively.[57] At the end of each inventory time period, lands occurring on mineral and organic soils were classified into the four land use/land-use change categories based on whether the parcel was in the same use during the previous inventory time periods or had been converted from another land use. Lands are retained in the land-use change categories (i.e., *Land Converted to Cropland* and *Land Converted to Grassland*) for 20 years as recommended by the IPCC guidelines (IPCC 2006). The total agricultural land base was 386 million hectares.

Table A-204: Total Land Areas for the Agricultural Soil C Inventory, Subdivided by Land Use and Land Use Change Categories (Million Hectares)

	Land Areas (10^6 ha)					
	1990-95			1995-2000		
Category	Century	IPCC	Total	Century	IPCC	Total
Mineral	**339.03**	**46.01**	**385.04**	**339.03**	**46.01**	**385.04**
Cropland Remaining Cropland	127.50	29.20	156.70	123.73	28.00	151.73
Land Converted to Cropland	7.08	2.33	9.41	11.40	1.25	12.65
Grassland Remaining Grassland	187.28	8.00	195.28	184.37	6.25	190.62
Land Converted to Grassland	17.17	4.02	21.19	19.53	2.44	21.97
Non-Agricultural Uses[1]	0.00	2.46	2.46	0.00	8.08	8.08
Organic	**0.00**	**1.34**	**1.34**	**0.00**	**1.34**	**1.34**
Cropland Remaining Cropland	0.00	0.70	0.70	0.00	0.72	0.72
Land Converted to Cropland	0.00	0.07	0.07	0.00	0.03	0.03
Grassland Remaining Grassland	0.00	0.50	0.50	0.00	0.48	0.48
Land Converted to Grassland	0.00	0.04	0.04	0.00	0.07	0.07
Non-Agricultural Uses[1]	0.00	0.04	0.04	0.00	0.05	0.05
Total	**339.03**	**47.35**	**386.38**	**339.03**	**47.35**	**386.38**

[1] These non-agricultural uses were converted to or from cropland or grassland during the early or late 1990s (i.e., were cropland or grassland in either 1992 or 1997 according to the NRI).

Management System Classification—NRI points were classified into specific crop rotations, continuous pasture/rangeland, and other non-agricultural uses based on the survey data (Table A-2). For model calculations, land-use and management activities were grouped into inventory time periods (i.e., time "blocks") for 1980-84, 1985-89, 1990-94 and 1995-2000, using NRI data from 1982, 1987, 1992, and 1997, respectively.

Crops were reported for the NRI inventory year and the three previous years for each NRI point location; hence, the full crop sequence is lacking one year in five. The most recent national-level data available for NRI were for 1997; crop rotations existing in 1997 were, thus, extended to 2005 in order to cover the last inventory time period. In addition, NRI differentiates between improved and unimproved grassland, where improvements include irrigation and interseeding of legumes.

As discussed above, Century was used to model NRI points on most mineral soils for most crops. The exceptions were for land that had one or more years of vegetable crops, tobacco, perennial/horticultural crops, and/or rice, in addition to soils that were classified as very or extremely gravelly, cobbly, or shaley (i.e., where

[55] In the current inventory, NRI data only provide land-use and management statistics through 1997, but the inventory is now transitioning into annual data collection. It is anticipated that new statistics will be released in the coming year for 2000 through 2003.

[56] Includes non-federal lands only, because federal lands are not classified into land uses as part of the NRI survey (i.e, they are only designated as federal lands).

[57] The first two inventory time periods occur before the 1990 baseline year of the reporting period and, therefore, are not included in this report.

greater 35 percent of the soil volume is composed of gravel, cobbles or shale), regardless of the crops present. Change in soil organic C for these NRI point locations were estimated using the IPCC Tier 2 method. Century was also used to estimate stock changes in grassland, except for soils that were classified as very or extremely gravelly, cobbly or shaley (Table A-205). Century has not been fully tested for non-major crops, horticultural or perennial crops, rice and agricultural use of organic soils. In addition, Century has not been adequately tested for soils with a high gravel, cobble or shale content. The IPCC Tier 2 method was used to estimate stock changes for this portion of the land base, as well as land converted from non-agricultural uses to crop or grassland during the reporting period. Again, Century has not been fully tested for these types of transitions.

Table A-205: Total Land Areas by Land-Use and Management System for the Entire U.S. Agricultural Land Base (Million Hectares)

	Land Areas (10^6 ha)					
	1990-95			1995-2000		
Land-Use/Management System	**Century**	**IPCC**	**Total**	**Century**	**IPCC**	**Total**
Cropland Systems	**134.58**	**31.53**	**166.11**	**135.13**	**29.25**	**164.38**
Irrigated Crops	9.55	7.27	16.82	9.58	6.91	16.49
Continuous Row Crops	36.90	4.12	41.02	39.68	3.63	43.31
Continuous Small Grains	11.76	1.25	13.01	13.51	1.04	14.55
Continuous Row Crops and Small Grains	14.48	2.30	16.78	12.66	1.95	14.61
Row Crops in Rotation with Hay and/or Pasture	3.37	0.30	3.67	3.41	0.23	3.65
Small Grains in Rotation with Hay and/or Pasture	0.85	0.06	0.91	0.91	0.06	0.96
Row Crops and Small Grains in Rotation with Hay and/or Pasture	0.31	0.03	0.34	0.31	0.04	0.35
Vegetable Crops	0.00	2.90	2.90	0.00	3.16	3.16
Low Residue Annual Crops (e.g., Tobacco or Cotton)	4.42	0.87	5.29	4.46	1.03	5.49
Small Grains with Fallow	17.60	2.01	19.61	14.34	1.31	15.65
Row Crops and Small Grains with Fallow	0.69	1.72	2.41	0.56	1.80	2.36
Row Crops with Fallow	2.09	0.52	2.60	1.66	0.34	2.01
Miscellaneous Crop Rotations	1.58	0.54	2.12	1.47	0.43	1.90
Continuous Rice	0.00	0.34	0.34	0.00	0.31	0.31
Rice in Rotation with other crops	0.00	1.78	1.78	0.00	1.91	1.91
Continuous Perennial or Horticultural Crops	0.00	2.57	2.57	0.00	2.50	2.50
Continuous Hay	5.56	0.59	6.15	6.70	0.50	7.21
Continuous Hay with Legumes or Irrigation	13.59	1.31	14.91	13.68	1.12	14.80
CRP	11.84	1.03	12.86	12.19	0.96	13.16
Aquaculture	0.00	0.01	0.01	0.00	0.01	0.01
Grassland Systems	**204.45**	**12.02**	**216.46**	**203.9**	**8.68**	**212.58**
Rangeland	158.66	5.98	164.64	158.94	5.16	164.10
Continuous Pasture	31.24	3.76	35.00	32.03	2.49	34.52
Continuous Pasture with Legumes or Irrigation (i.e., improved)	13.69	2.25	15.94	12.88	1.03	13.91
CRP	0.86	0.02	0.88	0.05	0.00	0.05
Non-Agricultural Systems	**0.00**	**2.46**	**2.46**	**0.00**	**8.08**	**8.08**
Forest	0.00	1.53	1.53	0.00	3.95	3.95
Federal	0.00	0.01	0.01	0.00	0.05	0.05
Water	0.00	0.11	0.11	0.00	0.25	0.25
Settlements	0.00	0.04	0.04	0.00	2.46	2.46
Miscellaneous	0.00	0.77	0.77	0.00	1.37	1.37
Total	**339.03**	**47.35**	**386.38**	**339.03**	**47.35**	**386.38**

Organic soils are categorized into land-use systems based on drainage for purposes of estimating C losses (IPCC/UNEP/OECD/IEA 1997). Undrained soils are treated as having no loss of organic C for purposes of the inventory. Drained soils are subdivided into those used for cultivated cropland, which are assumed to have high drainage and greater losses of C, and those used for managed pasture, which are assumed to have less drainage and smaller losses of C. Overall, the area of organic soils drained for cropland has remained relatively stable since 1982, but the area of organic soils managed as pasture has increased slightly (see Table A-206).

Table A-206: Total Land Areas for Drained Organic Soils By Land Management Category and Climate Region (Million Hectares)

	Land Areas (10^6 ha)					
IPCC Land-Use Category for Organic Soils	**Cold Temperate**		**Warm Temperate**		**Tropical**	
	1992	**1997**	**1992**	**1997**	**1992**	**1997**

Undrained	0.07	0.06	0.0020	0.0017	0.12	0.09
Managed Pasture (Low Drainage)	0.42	0.42	0.0136	0.0119	0.07	0.08
Cultivated Cropland (High Drainage)	0.33	0.34	0.0971	0.0974	0.19	0.20
Other Land Uses[2]	0.02	0.01	0.0002	0.0017	0.00	0.02
Totals	**0.84**	**0.84**	**0.11**	**0.11**	**0.39**	**0.39**

[1] Urban, water, and miscellaneous non-cropland, which are part of the agricultural land base, because these areas were converted from or into agricultural land uses during the 1990s.

Step 1b: Obtain Additional Management Activity Data for the Tier 3 Century Model

Management System Classification—Based on the classification of NRI data described in Step 1a, uncertainty in the areas associated with each management system was determined from the estimated sampling variance from the NRI survey (Nusser and Goebel 1997). See Step 2b for additional discussion.

Tillage practices—Tillage practices were estimated for each cropping system based on data compiled by the Conservation Technology Information Center (CTIC 1998). CTIC compiles data on cropland area under five tillage classes by major crop species and year for each county. Because the surveys involve county-level aggregate area, they do not fully characterize tillage practices as they are applied within a management sequence (e.g., crop rotation). This is particularly true for area estimates of cropland under no-till, which include a relatively high proportion of "intermittent" no-till, where no-till in one year may be followed by tillage in a subsequent year. For example, a common practice in maize-soybean rotations is to use tillage in the maize crop while no-till is used for soybean, such that no-till practices are not continuous in time. Due to the effects on soil C of the disturbance associated with periodic tillage, estimates of the area under continuous no-till were provided by experts at CTIC (Towery 2001).

Tillage practices were grouped into 3 categories: intensive, moderate, and no-tillage. Intensive tillage was defined as multiple tillage operations every year, including significant soil inversion (e.g., plowing, deep disking) and low surface residue coverage. This definition corresponds to the intensive tillage and "reduced" tillage systems as defined by CTIC (1998). No tillage was defined as not disturbing the soil except through the use of fertilizer and seed drills and where no-till is applied to all crops in the rotation. Moderate tillage made up the remainder of the cultivated area, including mulch tillage and ridge tillage as defined by CTIC and intermittent no-till. The specific tillage implements and applications used for different crops, rotations, and regions to represent the three tillage classes were derived from the 1995 Cropping Practices Survey by the Economic Research Service (ERS 1997).

Tillage data were further processed to construct probability distribution functions (PDFs) using CTIC tillage data. Transitions between tillage systems were based on observed county-level changes in the frequency distribution of the area under intensive, reduced and no-till from the 1980s through 1990s. Generally, the fraction of intensive tillage decreased during this time span, with concomitant increases in reduced till and no-till management. Transitions that were modeled and applied to NRI points occurring within a county were intensive tillage to reduced and no-till, and reduced tillage to no-till. The remaining amount of cropland was assumed to have no tillage change (e.g., intensive tillage remained in intensive tillage, etc.). Transition matrices were constructed from CTIC data to represent tillage changes for two time periods, combining the first two and the second two inventory time periods as single management blocks (i.e., 1980-1989, 1990-2000). Areas in each of the three tillage classes (conventional till (CT), reduced till (RT), no till (NT)) in 1989 (the first year the CTIC data were available) were used for the first time period, and data from 1997 were used for the second time period. Percentage areas of cropland in each county were calculated for each possible transition (e.g., CT→CT, CT→RT, CT→NT, RT→CT, RT→RT, RT→NT) to obtain a probability for each tillage transition at an NRI point. Since continuous NT constituted < 1 percent of total cropland prior to 1990, there were no transitions for NT→CT or NT→NT. Uniform probability distributions were established for each tillage scenario in the county. For example, a particular crop rotation had 80 percent chance of remaining in intensive tillage over the two decades, a 15 percent chance of a transition from intensive to reduced tillage and a 5 percent chance of a transition from intensive to no-till. The uniform distribution was subdivided into three segments with random draws in the Monte Carlo simulation (discussed in Step 2b) leading to intensive tillage over the entire time period if the value was greater than or equal to 0 and less than 80, a transition from intensive to reduced till if the random draw was equal to or greater than 80 and less than 95, or a transition from intensive to no-till if the draw was greater than or equal to 95. See step 2b for additional discussion of the uncertainty analysis.

Mineral Fertilizer application—Data on nitrogen fertilizer rates were obtained primarily from USDA's Economic Research Service's 1995 Cropping Practices Survey (ERS 1997). In this survey, data on inorganic nitrogen fertilization rates were collected for major crops (corn, cotton, soybeans, potatoes, winter wheat, durum wheat, and other spring wheat) in the major producing states. Note that all wheat data were combined into one category and assumed to represent all small grains. Estimates for sorghum fertilizer rates were derived from corn rates using a ratio of national average corn fertilizer rates to national average sorghum fertilizer rates derived from additional publications (NASS 2004, 1999, 1992; ERS 1988; Grant and Krenz 1985; USDA 1966, 1957, 1954).

The ERS survey parameter "TOT N" (total amount of N applied per acre), with a small number of records deleted as outliers, was used in determining the fraction of crop acres receiving fertilizer and the average fertilizer rates for a region. Mean fertilizer rates and standard deviations for irrigated and rainfed crops were produced for each state at the finest resolution available. State-level data were produced for surveyed states if a minimum of 15 data points existed for each of the two categories (irrigated and rainfed). If a state was not surveyed for a particular crop or if fewer than 15 data points existed for one of the categories, then data at the Farm Production Region level were substituted. If Farm Production Region data were not available, then U.S. level estimates (all major states surveyed) were used in the simulation for that particular crop in the state lacking sufficient data. Note that standard deviations for fertilizer rates on log scale were used to construct PDFs on a log-normal scale, in order to address uncertainties in application rates (see Step 2b for discussion of uncertainty methods).

Manure application—County-level manure nitrogen addition estimates were obtained from the Natural Resources Conservation Service (Edmonds et al. 2003). Working with the farm-level crop and animal data from the 1997 Census of Agriculture, NRCS has coupled estimates of manure nitrogen produced with estimates of manure nitrogen recoverability by animal waste management system to produce county-level estimates of manure nitrogen applied to cropland and pasture. Edmonds et al. (2003) defined a hierarchy of land use systems to which manure is applied, that included 24 crops, cropland used as pasture, and permanent pasture. They estimated the area amended with manure and manure nitrogen application rates in 1997 for both manure-producing farms and manure-receiving farms within a county, for two scenarios—before implementation of Comprehensive Nutrient Management Plans (baseline) and after implementation.

In order to derive estimates of manure application rates over time, managed manure N production estimates (which are available annually) were used to adjust the amount of area amended with manure on a county scale (Note: Edmonds et al. (2003) only provide information on application rates for 1997). Specifically, the managed manure N production in another year was divided by the managed manure N production in 1997. The amendment area in a county for 1997 was then multiplied by the ratio to reflect the probability of manure amendments based on the variation in managed manure N production across time. If more managed manure N was produced in a given year across a county relative to the amount produced in 1997 (ratio > 1), it was assumed that there was a higher probability of a manure amendment. In contrast, if less managed manure N was produced (ratio < 1), the probability of an amendment declined in comparison to 1997. A detailed description of the derivation of the managed manure N production data is provided in Annex 3.10. Managed manure N production in the 1980s was based on USDA estimates (Kellogg et al. 2000) after adjusting for relative differences in manure N production between the USDA dataset and estimates derived from the method described in Annex 3.10. Unmanaged manure classified as pasture/range/paddock manure was assumed to have negligible impacts on soil C stocks because of the tradeoff between reduced litterfall C versus C ingested by livestock and deposited on soils in manure.

For Century simulations, the amended areas were averaged for two time blocks (1980-1989, 1990-2000) similar to the tillage transitions. Rates for manure-producing farms and manure-receiving farms have been area-weighted and combined to produce a manure nitrogen application rate for each crop in a county. Several of the crops in Edmonds et al. (2003) have been area-weighted and combined into broader crop categories. For example, all small grain crops have been combined into one category. In order to address uncertainty, uniform probability distributions were constructed based on the proportion of land receiving manure versus the amount not receiving manure for each crop type and pasture. For example, if the 20 percent of land producing corn in a county was amended with manure, randomly drawing a value equal to or greater than 0 and less than 20 would lead to simulation with a manure amendment, while drawing a value greater than or equal to 20 and less than 100 would lead to no amendment in the simulation (see Step 2b for further discussion of uncertainty methods).

To estimate the C inputs associated with the manure N application rates (from Edmonds et al. 2003), C:N ratios for various manure types (based on animal species and manure management system) were estimated from data in the *Agricultural Waste Management Field Handbook* (USDA 1996) and the *On-Farm Composting Handbook*

(NRAES 1992). Weighted county-average C:N ratios for total manure applied were then calculated based on the C:N ratio and the manure N production rate for each manure type reported in the county. Manure C addition rates were then calculated by multiplying the county-average manure C:N ratio by the manure N application rates.

To account for the common practice of reducing inorganic nitrogen fertilizer inputs when manure is added to a cropland soil, a set of crop-specific reduction factors were derived from mineral fertilization data for land amended with manure versus land not amended with manure in the ERS 1995 Cropping Practices Survey (ERS 1997). In the simulations, mineral N fertilization rates were reduced for crops receiving manure nitrogen based on a fraction of the amount of manure nitrogen applied, depending on the crop and whether it was irrigated or a rainfed system. The reduction factors were selected from PDFs with normal densities in order to address uncertainties in this dependence between manure amendments and mineral fertilizer application.

Irrigation—NRI differentiates between irrigated and non-irrigated land but does not provide more detailed information on the type and intensity of irrigation. Hence, irrigation was modeled by assuming that applied water was sufficient to meet full crop demand (i.e., irrigation plus precipitation equaled potential evapotranspiration during the growing season).

Step 1c—Obtain Additional Management Activity Data for Tier 2 IPCC Method

Management System Classification—NRI points were assigned to IPCC input categories (low, medium, high and high with organic amendments) according to the classification provided in the IPCC guidelines (IPCC 2006). In order to estimate uncertainties, PDFs for the NRI land-use data were constructed as multivariate normal based on the total area estimates for each land use/management category and associated covariance matrix. Through this approach, dependencies in land use were taken into account resulting from the likelihood that current use is correlated with past use.

Tillage Practices—PDFs were also constructed for the CTIC tillage data, as bivariate normal on a log-ratio scale to reflect negative dependence among tillage classes. This structure ensured that simulated tillage percentages were non-negative and summed to 100 percent. CTIC data do not differentiate between continuous and intermittent use of no-tillage, which is important for estimating SOC storage. Thus, regional-based estimates for continuous no-tillage (defined as 5 or more years of continuous use) were modified based on consultation with CTIC experts, as discussed in Step 1a (downward adjustment of total no-tillage acres reported, Towery 2001).

Manure Amendments—Manure management is also a key practice in agricultural lands, with organic amendments leading to significant increases in SOC storage. USDA provides information on the amount of land amended with manure for 1997 based on manure production data and field-scale surveys detailing application rates that had been collected in the Census of Agriculture (Edmonds et al. 2003). Similar to the Century model discussion in Step1b, the amount of land receiving manure was based on the estimates provided by Edmonds et al. (2003), as a proportion of crop and grassland amended with manure within individual climate regions. The resulting proportions were used to re-classify a portion of crop and grassland into a new management category. Specifically, a portion of medium input cropping systems was re-classified as high input, and a portion of the high input systems was re-classified as high input with amendment. In grassland systems, the estimated proportions for land amended with manure were used to re-classify a portion of nominally-managed grassland as improved, and a portion of improved grassland as improved with high input. These classification approaches are consistent with the IPCC inventory methodology (IPCC/UNEP/OECD/IEA 1997, IPCC 2003). Uncertainties in the amount of land amended with manure were based on the sample variance at the climate region scale, assuming normal density PDFs (i.e., variance of the climate region estimates, which were derived from county-scale proportions).

Wetland Reserve—Wetlands enrolled in the Conservation Reserve Program have been restored in the Northern Prairie Pothole Region through the Partners for Wildlife Program funded by the U.S. Fish and Wildlife Service. The amount of restored wetlands was estimated from contract agreements (Euliss and Gleason 2002). While the contracts provide reasonable estimates of the amount of land restored in the region, they do not provide the information necessary to estimate uncertainty. Consequently, a ±50 percent range was used to construct the probability density functions for the uncertainty analysis.

Step 1d—Obtain Management Activity Data to Compute Additional Changes in Soil Organic C Stocks in Mineral Soils Due to Sewage Sludge Applications and CRP Enrollment after 1997

Two additional influences on soil organic C stocks in mineral soils were estimated using a Tier 2 method, including: sewage sludge additions to agricultural soils and changes in enrollment for the Conservation Reserve Program after 1997.

Total sewage sludge generation data for 1988, 1996, and 1998, and a projection for 2000, in dry mass units, were obtained from EPA reports (EPA 1993, 1999), and linearly interpolated to estimate values for the intervening years. Sewage sludge generation data are not available for 2001 through 2005 (Bastian 2002, 2003, 2005), so the 2000 projection was linearly extrapolated using the growth in national wastewater flow between 1996 and 2000 (EPA 1997, 2003). The total sludge generation estimates were then converted to units of nitrogen by applying an average N content of 3.3 percent (Metcalf and Eddy 1991), and disaggregated into use and disposal practices using historical data and projections in EPA (1993) and EPA (1999). The use and disposal practices were agricultural land application, other land application, surface disposal, landfilling, ocean dumping (ended in 1992), and other disposal. Sewage sludge N was assumed to be applied at the assimilative capacity provided in Kellogg et al. (2000), which is the amount of nutrients taken up by a crop and removed at harvest, representing the recommended application rate for manure amendments. This capacity varies from year to year, because it is based on specific crop yields during the respective year (Kellogg et al. 2000). Total sewage sludge N available for application was divided by the assimilative capacity to estimate the total land area over which sewage sludge had been applied. The resulting estimates were used for the estimation of soil C stock change.

The change in enrollment for the Conservation Reserve Program after 1997 was based on the amount of land under active contracts from 1998 through 2004 relative to 1997 (Barbarika 2006).

Step 1e: Obtain Climate and Soils Data

Tier 3 Century Model—Monthly weather data (temperature and precipitation) from the PRISM database (Parameter-elevation Regressions on Independent Slopes Model) (Daly et al. 1994) were used as an input to the Century model simulations for the period 1895-2005. PRISM is based on observed weather data from the National Weather Service network database and statistical models for interpolation and orographic corrections. The primary database consists of approximately 4×4 km grid cells. These data were averaged (weighted by area) for each county in the United States, so that counties are the finest spatial scale represented in the Century simulations.

Soil texture and natural drainage capacity (i.e., hydric vs. non-hydric soil characterization) were the main soil variables used as input to the Century model. Other soil characteristics needed in the simulation, such as field capacity and wilting point water contents, were estimated from soil texture data using pedo-transfer functions available in the model. Soil input data are derived from the NRI database, which contain descriptions for the soil type at each NRI point (used to specify land-use and management time series-see below). The data are based on field measurements collected as part of soil survey and mapping. Soils are classified according to "soil-series," which is the most detailed taxonomic level used for soil mapping in the United States. Surface soil texture and hydric condition were obtained from the soil attribute table in the NRI database. Texture is one of the main controls on soil C turnover and stabilization in the Century model, which uses particle size fractions of sand (50-2000 μm), silt (2-50 μm), and clay (< 2 μm) as inputs. NRI points were assigned to one of twelve texture classes for the simulations. Hydric condition specifies whether soils are poorly-drained, and hence prone to water-logging, or moderately to well-drained (non-hydric), in their native (pre-cultivation) condition.[58] Poorly drained soils can be subject to anaerobic (lack of oxygen) conditions if water inputs (precipitation and irrigation) exceed water losses from drainage and evapotranspiration. Depending on moisture conditions, hydric soils can range from being fully aerobic to completely anaerobic, varying over the year. Decomposition rates are modified according to a linear function that varies from 0.3 under completely anaerobic conditions to 1.0 under fully aerobic conditions (default parameters in Century).[59]

[58] Artificial drainage (e.g., ditch- or tile-drainage) is simulated as a management variable.

[59] Hydric soils are primarily subject to anaerobic conditions outside the plant growing season (i.e., in the absence of active plant water uptake). Soils that are water-logged during much of the year are typically classified as organic soils (e.g., peat), which are not simulated with the Century model.

IPCC Tier 2 Method—The IPCC inventory methodology for agricultural soils divides climate into eight distinct zones based upon average annual temperature, average annual precipitation, and the length of the dry season (IPCC/UNEP/OECD/IEA 1997) (Table A-207). Six of these climate zones occur in the conterminous United States and Hawaii (Eve et al. 2001).

Table A-207: Characteristics of the IPCC Climate Zones that Occur in the United States

Climate Zone	Annual Average Temperature (˚C)	Average Annual Precipitation (mm)	Length of Dry Season (months)
Cold Temperate, Dry	< 10	< Potential Evapotranspiration	NA
Cold Temperate, Moist	< 10	≥ Potential Evapotranspiration	NA
Warm Temperate, Dry	10 – 20	< 600	NA
Warm Temperate, Moist	10 – 20	≥ Potential Evapotranspiration	NA
Sub-Tropical, Dry*	> 20	< 1,000	Usually long
Sub-Tropical, Moist (w/short dry season)*	> 20	1,000 – 2,000	< 5

* The climate characteristics listed in the table for these zones are those that correspond to the tropical dry and tropical moist zones of the IPCC. They have been renamed "sub-tropical" here.

Mean climate (1961-1990) variables from the PRISM data set (Daly et al. 1994) were used to classify climate zones. Average annual precipitation and average annual temperature (4×4 km grid) were averaged (weighted by area) for each of the 180 MLRAs in the United States to assign an IPCC climate zone to each MLRA (Figure A-10). MLRAs represent geographic units with relatively similar soils, climate, water resources, and land uses (NRCS 1981).

Figure A-10. Major Land Resource Areas by IPCC Climate Zone

Soils were classified into one of seven classes based upon texture, morphology, and ability to store organic matter (IPCC/UNEP/OECD/IEA 1997). Six of the categories are mineral types and one is organic (i.e., Histosol). Reference C stocks, representing estimates from conventionally managed cropland, were computed for each of the mineral soil types across the various climate zones, based on pedon data from the National Soil Survey Characterization Database (NRCS 1997) (Table A-208). These stocks are used in conjunction with management factors to compute the change in SOC stocks that result from management and land-use activity. PDFs, which represent the variability in the stock estimates, were constructed as normal densities based on the mean and variance from the pedon data. Pedon locations were clumped in various parts of the country, which reduces the statistical independence of individual pedon estimates. To account for this lack of independence, samples from each climate by soil zone were tested for spatial autocorrelation using the Moran's I test, and variance terms were inflated by 10 percent for all zones with significant p-values.

Table A-208: U.S. Soil Groupings Based on the IPCC Categories and Dominant Taxonomic Soil, and Reference Carbon Stocks (Metric Tons C/ha)

		Reference Carbon Stock in Climate Regions					
IPCC Inventory Soil Categories	USDA Taxonomic Soil Orders	Cold Temperate, Dry	Cold Temperate, Moist	Warm Temperate, Dry	Warm Temperate, Moist	Sub-Tropical, Dry	Sub-Tropical, Moist
High Clay Activity Mineral Soils	Vertisols, Mollisols, Inceptisols, Aridisols, and high base status Alfisols	42 (n = 133)	65 (n = 526)	37 (n = 203)	51 (n = 424)	42 (n = 26)	57 (n = 12)
Low Clay Activity Mineral Soils	Ultisols, Oxisols, acidic Alfisols, and many Entisols	45 (n = 37)	52 (n = 113)	25 (n = 86)	40 (n = 300)	39 (n = 13)	47 (n = 7)
Sandy Soils	Any soils with greater than 70 percent sand and less than 8 percent clay (often Entisols)	24 (n = 5)	40 (n = 43)	16 (n = 19)	30 (n = 102)	33 (n = 186)	50 (n = 18)
Volcanic Soils	Andisols	124 (n = 12)	114 (n = 2)	124 (n = 12)	124 (n = 12)	124 (n = 12)	128 (n = 9)
Spodic Soils	Spodosols	86 (n=20)	74 (n = 13)	86 (n=20)	107 (n = 7)	86 (n=20)	86 (n=20)
Aquic Soils	Soils with Aquic suborder	86 (n = 4)	89 (n = 161)	48 (n = 26)	51 (n = 300)	63 (n = 503)	48 (n = 12)
Organic Soils*	Histosols	NA	NA	NA	NA	NA	NA

* C stocks are not needed for organic soils.

Notes: C stocks are for the top 30 cm of the soil profile, and were estimated from pedon data available in the National Soil Survey Characterization database (NRCS 1997); sample size provided in parentheses (i.e., 'n' values refer to sample size).

Step 2: Estimate Organic C Stock Changes for Agricultural Lands on Mineral Soils Simulated with the Tier 3 Century Model

This part of the inventory methods description is divided into two sub-steps. First, the model was used to establish the initial conditions and C stocks for 1979 which was the last year before the first inventory time period (i.e., 1980-84). In the second sub-step, Century was used to estimate changes in soil organic C stocks based on the land-use and management histories recorded in the NRI (USDA-NRCS 2000), including the reporting period starting in 1990.

Step 2a: Simulate Initial Conditions (Pre-NRI Conditions)

Century model initialization involves two steps, with the goal of estimating the most accurate stock for the pre-NRI history, and the distribution of organic C among the pools represented in the model (e.g., Structural, Metabolic, Active, Slow, Passive). Each pool has a different turnover rate (representing the heterogeneous nature of soil organic matter), and the amount of C in each pool at any point in time influences the forward trajectory of the total soil organic C storage. There is currently no national set of soil C measurements that can be used for establishing initial conditions in the model. Sensitivity analysis of the Century model showed that the rate of change of soil organic matter is relatively insensitive to the *amount* of total soil organic C but is highly sensitive to the relative *distribution* of C among different pools (Parton et al. 1987). By simulating the historical land use prior to the inventory period, initial pool distributions are estimated in an unbiased way.

The first step involves running the model to a steady-state condition (e.g., equilibrium) under native vegetation, with long-term mean climate represented by 30-yr averages of the PRISM data (1960-1990), and the soil physical attributes for the NRI points. Native vegetation is represented at the MLRA level for pre-settlement time periods in the United States. The model was run for 7000 years to achieve a steady-state condition.

The second step is to run the model for the period of time from settlement to the beginning of the NRI survey, representing the influence of historic land-use change and management, particularly the conversion of native vegetation to agricultural uses. This encompasses a varying time period from land conversion (depending on historical settlement patterns) to 1979. The information on historical cropping practices used for Century simulations was gathered from a variety of sources, ranging from the historical accounts of farming practices reported in the literature (e.g., Miner 1998) to national level databases (e.g., NASS 2004). A detailed description of the data sources and assumptions used in constructing the base history scenarios of agricultural practices can be found in Williams and Paustian (2005).

Step 2b—Estimate Soil Organic C Stock Changes and Uncertainties

After estimating model initialization, the model is used to simulate the NRI land use and management histories through to 2000.[60] The simulation system incorporates a dedicated MySQL database server and a 24-node parallel processing computer cluster. Input/output operations are managed by a set of run executive programs written in PERL. The assessment framework for this analysis is illustrated in Figure A-11.

Figure A-11. Uncertainty in Data Inputs

Inventory points within a county that shared the same land-use/management histories and soil type were aggregated for simulation purposes (i.e., data inputs to the model were identical and, therefore, the model results were also identical). For the 370,738 NRI points representing non-federal cropland and grassland, there was a total

[60] The estimated soil C stock change in 2000 is currently assumed to represent the changes between 2001 and 2005. Estimates will be re-calculated upon release and incorporation of new NRI data into the inventory assessment, which is a planned improvement.

of 170,279 clustered points that were simulated using Century (i.e., which represent the unique combinations of climate, soils, land use, and management in the modeled data set).

Evaluating uncertainty was an integral part of the analysis, and was comprised of three components: 1) uncertainty in the main activity data inputs affecting soil C balance (input uncertainty); 2) uncertainty in the model formulation and parameterization (structural uncertainty); and 3) uncertainty in the land-use and management system areas (scaling uncertainty). For component 1, input uncertainty was evaluated for fertilization management, manure applications, and tillage, which are the primary management activity data that were supplemental to the NRI observations and have significant influence on soil C dynamics. As described in Step 1b, probability density functions (PDFs) were derived from surveys at the county scale in most cases. To represent uncertainty in these inputs, a Monte-Carlo Analysis was used with 100 iterations for each NRI cluster-point in which random draws were made from PDFs for fertilizer, manure application, and tillage. As described above, an adjustment factor was also selected from PDFs with normal densities to represent the dependence between manure amendments and N fertilizer application rates. The total number of Century simulations was over 14 million for the Monte Carlo Analysis with 100 iterations.

The second component dealt with uncertainty inherent in model formulation and parameterization. An empirically-based procedure was employed to develop a structural uncertainty estimator from the relationship between modeled results and field measurements from agricultural experiments (Ogle et al. 2006b). The Century model was initialized for 45 long-term field experiments with over 800 treatments in which soil C was measured under a variety of management conditions (e.g., variation in crop rotation, tillage, fertilization rates, manure amendments). These studies were obtained from an extensive search of published studies. All studies located in North America that met minimum criteria of having sufficient site level information and experimental designs were used, including C stock estimates, texture data, experimental designs with control plots, and land-use and management records for the experimental time period and pre-experiment condition. The inputs to the model were essentially known in the simulations for the long-term experiments, and, therefore, the analysis was designed to evaluate uncertainties associated with the model structure (i.e., model algorithms and parameterization).

The relationship between modeled soil C stocks and field measurements was statistically analyzed using linear-mixed effect modeling techniques. Additional fixed effects were included in the mixed effect model, which explained significant variation in the relationship between modeled and measured stocks (if they met an alpha level of 0.05 for significance). Several variables were tested including: land-use class; type of tillage; cropping system; geographic location; climate; soil texture; time since the management change; original land cover (i.e., forest or grassland); grain harvest as predicted by the model compared to the experimental values; and variation in fertilizer and residue management. The final model included variables for organic matter amendments, fertilizer rates, inclusion of hay/pasture in cropping rotations, use of no-till, and inclusion of bare fallow in the rotation, which were significant at an alpha level of 0.05. These fixed effects were used to make an adjustment to modeled values due to biases that were creating significant mismatches between the modeled and measured stock values. Random effects captured the statistical dependence (i.e., the data are not fully independent) in time series and data collected from the same long-term experimental site. Accounting for this statistical dependency is needed to estimate appropriate standard deviations for parameter coefficients.

A Monte Carlo approach was used to apply the uncertainty estimator. Parameter values for the statistical equation (i.e., fixed effects) were selected from their joint probability distribution, as well as random error associated with fine-scale estimates at NRI points. The stock estimate and associated management information was then used as input into the equation, and adjusted stock values were computed for each C stock estimate produced in the evaluation of input uncertainty for *Cropland Remaining Cropland* (Component 1 of the uncertainty analysis). Note that the uncertainty estimator needs further development for application to *Grassland Remaining Grassland* and the land use change categories. This development is a planned improvement for the soil C inventory. The variance of the adjusted C stock estimates were computed from the 100 simulated values from the Monte Carlo analysis.

The third element was the uncertainty associated with scaling the Century results for each NRI cluster to the entire land base, using the expansion factors provided with the NRI database. The expansion factors represent the number of hectares associated with the land-use and management history of a particular cluster. This uncertainty was determined by computing the variances of the expanded estimates, accounting for the two-stage sampling design of the NRI.

For the land base that was simulated with the Century model (Table A-204), soil organic C stocks ranged from losses of 13.9 Tg CO_2 Eq. to gains of 66.4 Tg CO_2 Eq. annually, depending on the land-use/land-use change category and inventory time period. Estimates and uncertainties are provided in Table A-209.

Table A-209: Annual Change in Soil Organic Carbon Stocks (95% Confidence Interval) for the Land Base Simulated with the Tier 3 Century Model-Based Approach (Tg CO_2 Eq.)

Soil Type	1990-1994	1995-2005
Mineral Soils		
Cropland Remaining Cropland	(58.59)	(66.44)
95% C.I.	(72.5) to (44.7)	(76.96) to (55.92)
Land Converted to Cropland	2.1	0.4
95% C.I.	1.5 to 2.6	0.28 to 0.56
Grassland Remaining Grassland	(2.8)	13.9
95% C.I.	(4.7) to (0.9)	13.5 to 14.2
Land Converted to Grassland	(10.5)	(12.2)
95% C.I.	(10.8) to (10.2)	(12.3) to (12.1)

Note: Does not include the change in storage resulting from the annual application of sewage sludge, or the additional Conservation Reserve Program enrollment.

Step 3: Estimate C Stock Changes in Agricultural Lands on Mineral Soils Approximated with the Tier 2 Approach, in addition to CO_2 Emissions from Agricultural Lands on Drained Organic Soils

Mineral and organic soil calculations were made for each climate by soil zone across the United States. Mineral stock values were derived for non-major crop rotations and land converted from non-agricultural uses to cropland in 1982, 1992, and 1997 based on the land-use and management activity data in conjunction with appropriate reference C stocks, land-use change, tillage, input, and wetland restoration factors. C losses from organic soils were computed based on 1992 and 1997 land use and management in conjunction with the appropriate C loss rate. Each input to the inventory calculations for the Tier 2 approach had some level of uncertainty that was quantified in probability density functions, including the land-use and management activity data, reference C stocks, and management factors. A Monte Carlo Analysis was used to quantify uncertainty in SOC change for the inventory period based on uncertainty in the inputs. Input values were randomly selected from probability density functions in an iterative process to estimate SOC change for 50,000 times and produce a 95 percent confidence interval for the inventory results.

Step 3a: Derive Mineral Soil Stock Change and Organic Soil Emission Factors

Stock change factors representative of U.S. conditions were estimated from published studies (Ogle et al. 2003, Ogle et al. 2006). The numerical factors quantify the impact of changing land use and management on SOC storage in mineral soils, including tillage practices, cropping rotation or intensification, and land conversions between cultivated and native conditions (including set-asides in the Conservation Reserve Program), as well as the net loss of SOC from organic soils attributed to agricultural production on drained soils. Studies from the United States and Canada were used in this analysis under the assumption that they would best represent management impacts for this inventory.

For mineral soils, studies had to report SOC stocks (or information to compute stocks), depth of sampling, and the number of years since a management change to be included in the analysis. The data were analyzed using linear mixed-effect modeling, accounting for both fixed and random effects. Fixed effects included depth, number of years since a management change, climate, and the type of management change (e.g., reduced tillage vs. no-till). For depth increments, the data were not aggregated for the C stock measurements; each depth increment (e.g., 0-5 cm, 5-10 cm, and 10-30 cm) was included as a separate point in the dataset. Similarly, time series data were not aggregated in these datasets. Consequently, random effects were needed to account for the dependence in time series data and the dependence among data points representing different depth increments from the same study. Factors were estimated for the effect of management practices at 20 years for the top 30 cm of the soil (Table A-210). Variance was calculated for each of the U.S. factor values, and used to construct PDFs with a normal density. In the IPCC method, specific factor values are given for improved grassland, high input cropland with organic amendments, and for wetland rice, each of which influences the C balance of soils. Specifically, higher stocks are

associated with increased productivity and C inputs (relative to native grassland) on improved grassland with both medium and high input.[61] Organic amendments in annual cropping systems also increase SOC stocks due to greater C inputs, while high SOC stocks in rice cultivation are associated with reduced decomposition due to periodic flooding. There were insufficient field studies to re-estimate factor values for these systems, and, thus, estimates from IPCC (2003) were used under the assumption that they would best approximate the impacts, given the lack of sufficient data to derive U.S.-specific factors. A measure of uncertainty was provided for these factors in IPCC (2003), which was used to construct PDFs.

Table A- 210: Stock Change Factors for the United States and the IPCC Default Values Associated with Management Impacts on Mineral Soils

		U.S. Factor			
	IPCC default	Warm Moist Climate	Warm Dry Climate	Cool Moist Climate	Cool Dry Climate
Land-Use Change Factors					
Cultivated[a]	1	1	1	1	1
General Uncult.[a,b] (n=251)	1.4	1.42±0.06	1.37±0.05	1.24±0.06	1.20±0.06
Set-Aside[a] (n=142)	1.25	1.31±0.06	1.26±0.04	1.14±0.06	1.10±0.05
Improved Grassland Factors[c]					
Medium Input	1.1	1.14±0.06	1.14±0.06	1.14±0.06	1.14±0.06
High Input	Na	1.11±0.04	1.11±0.04	1.11±0.04	1.11±0.04
Wetland Rice Production Factor[c]	1.1	1.1	1.1	1.1	1.1
Tillage Factors					
Conv. Till	1	1	1	1	1
Red. Till (n=93)	1.05	1.08±0.03	1.01±0.03	1.08±0.03	1.01±0.03
No-till (n=212)	1.1	1.13±0.02	1.05±0.03	1.13±0.02	1.05±0.03
Cropland Input Factors					
Low (n=85)	0.9	0.94±0.01	0.94±0.01	0.94±0.01	0.94±0.01
Medium	1	1	1	1	1
High (n=22)	1.1	1.07±0.02	1.07±0.02	1.07±0.02	1.07±0.02
High with amendment[c]	1.2	1.38±0.06	1.34±0.08	1.38±0.06	1.34±0.08

Note: The "n" values refer to sample size.
[a] Factors in the IPCC documentation (IPCC/UNEP/OECD/IEA 1997) were converted to represent changes in SOC storage from a cultivated condition rather than a native condition.
[b] Default factor was higher for aquic soils at 1.7. The U.S. analysis showed no significant differences between aquic and non-aquic soils, so a single U.S. factor was estimated for all soil types.
[c] U.S.-specific factors were not estimated for land improvements, rice production, or high input with amendment because of few studies addressing the impact of legume mixtures, irrigation, or manure applications for crop and grassland in the United States, or the impact of wetland rice production in the US. Factors provided in IPCC (2003) were used as the best estimates of these impacts.

Wetland restoration management also influences SOC storage in mineral soils, because restoration leads to higher water tables and inundation of the soil for at least part of the year. A stock change factor was estimated assessing the difference in SOC storage between restored and unrestored wetlands enrolled in the Conservation Reserve Program (Euliss and Gleason 2002), which represents an initial increase of C in the restored soils over the first 10 years (Table A-211). A PDF with a normal density was constructed from these data based on results from a linear regression model. Following the initial increase of C, natural erosion and deposition leads to additional accretion of C in these wetlands. The mass accumulation rate of organic C was estimated using annual sedimentation rates (cm/yr) in combination with percent organic C, and soil bulk density (g/cm^3) (Euliss and Gleason 2002). Procedures for calculation of mass accumulation rate are described in Dean and Gorham (1998); the resulting rate and variance were used to construct a PDF with a normal density (Table A-211).

[61] Improved grasslands are identified in the *1997 National Resources Inventory* as grasslands that were irrigated or seeded with legumes, in addition to those reclassified as improved with manure amendments.

Table A-211: Factor Estimate for the Initial and Subsequent Increase in Organic Soil C Following Wetland Restoration of Conservation Reserve Program

Variable	Value
Factor (Initial Increase—First 10 Years)	1.22±0.18
Mass Accumulation (After Initial 10 Years)	0.79±0.05 Mg C/ha-yr

Note: Mass accumulation rate represents additional gains in C for mineral soils after the first 10 years (Euliss and Gleason 2002).

In addition, C loss rates were estimated for cultivated organic soils based on subsidence studies in the United States and Canada (Table A-212). Probability density functions were constructed as normal densities based on the mean C loss rates and associated variances.

Table A-212: Carbon Loss Rates for Organic Soils Under Agricultural Management in the United States, and IPCC Default Rates (Metric Ton C/ha-yr)

	Cropland		Grassland	
Region	IPCC	U.S. Revised	IPCC	U.S. Revised
Cold Temperate, Dry & Cold Temperate, Moist	1	11.2±2.5	0.25	2.8±0.5[a]
Warm Temperate, Dry & Warm Temperate, Moist	10	14.0±2.5	2.5	3.5±0.8[a]
Sub-Tropical, Dry & Sub-Tropical, Moist	20	14.0±3.3	5	3.5±0.8[a]

[a] There were not enough data available to estimate a U.S. value for C losses from grassland. Consequently, estimates are 25 percent of the values for cropland, which was an assumption used for the IPCC default organic soil C losses on grassland.

Step 3b: Estimate Annual Changes in Mineral Soil Organic C Stocks and CO_2 Emissions from Organic Soils

In accordance with IPCC methodology, annual changes in mineral soil C were calculated by subtracting the beginning stock from the ending stock and then dividing by 20.[62] For this analysis, the base inventory estimate for 1990 through 1992 is the annual average of 1992 stock minus the 1982 stock. The annual average change between 1993 and 2005 is the difference between the 1997 and 1992 C stocks. Using the Monte Carlo approach, SOC stock change for mineral soils was estimated 50,000 times between 1982 and 1992, and between 1992 and 1997. From the final distribution of 50,000 values, a 95 percent confidence interval was generated based on the simulated values at the 2.5 and 97.5 percentiles in the distribution (Ogle et al. 2003).

For organic soils, annual losses of CO_2 were estimated for 1992 and 1997 by applying the Monte Carlo approach to 1992 and 1997 land-use data in the United States. The results for 1992 were applied to the years 1990 through 1992, and the results for 1997 were applied to the years 1993 through 2005.

Mineral soils for the land base estimated with the Tier 2 approach sequestered about 1.7 to 3.0 Tg CO_2 Eq. annually in *Cropland Remaining Cropland*, while mineral soils in *Land Converted to Cropland* lost C at a rate of about 4.1 to 4.2 Tg CO_2 Eq. annually. Minerals soils in *Grassland Remaining Grassland* had small gains of about 0.2 to 0.3 Tg CO_2 Eq. annually and sequestered from 4.5 to 5.0 Tg CO_2 Eq. annually in *Land Converted to Grassland*. Organic soils lost about 27.4 to 27.7 Tg CO_2 Eq. annually in *Cropland Remaining Cropland* and 2.4 to 2.6 Tg CO_2 Eq. annually in *Land Converted to Cropland*, as well as an additional 3.7 to 3.9 Tg CO_2 Eq. annually in *Grassland Remaining Grassland* (Table A-213) and 0.5 to 0.9 Tg CO_2 Eq. annually in *Land Converted to Grassland*. Estimates and uncertainties are provided in Table A-213.

Table A-213: Annual Change in Soil Organic Carbon Stocks (95% Confidence Interval) for the Land Base Estimated with the Tier 2 Analysis using U.S. Factor Values, Reference Carbon Stocks, and Carbon Loss Rates (Tg CO_2 Eq. /yr)

Soil Type	1990-1992	1993-2005
Mineral Soils		
Cropland Remaining Cropland	(1.7)	(3.0)
95% C.I.	(2.6) to 5.8	(-6.9) to 0.8
Land Converted to Cropland	4.2	4.1
95% C.I.	2.5 to 5.98	2.3 to 5.8
Grassland Remaining Grassland	(0.3)	(0.2)
95% C.I.	(0.6) to (0.1)	(0.3) to 0.04
Land Converted to Grassland	(4.5)	(5.0)
95% C.I.	(6.5) to (2.74)	(6.96) to (2.84)

[62] The difference in C stocks is divided by 20 because the stock change factors represent change over a 20-year time period.

Organic Soils		
Cropland Remaining Cropland	27.4	27.7
95% C.I.	18.3 to 39.4	15.8 to 36.9
Land Converted to Cropland	2.4	2.6
95% C.I.	1.4 to 3.8	1.2 to 3.7
Grassland Remaining Grassland	3.9	3.7
95% C.I.	1.97 to 6.4	1.2 to 5.5
Land Converted to Grassland	0.5	0.9
95% C.I.	0.22 to 0.8	0.2 to 1.8

Step 4: Compute Additional Changes in Soil Organic C Stocks Due to Organic Amendments and CRP Enrollment after 1997

There are two additional land-use and management activities in U.S. agriculture lands that were not estimated in Steps 2 and 3. The first activity involved the application of sewage sludge to agricultural lands. Minimal data exist on where and how much sewage sludge is applied to U.S. agricultural soils, but national estimates of mineral soil land area receiving sewage sludge can be approximated based on sewage sludge N production data, and the assumption that amendments are applied at a rate equivalent to the assimilative capacity from Kellogg et al. (2000). It was assumed that sewage sludge for agricultural land application was applied to grassland because of the high heavy metal content and other pollutants found in human waste, which limits it application to crops. The impact of organic amendments on SOC was calculated as 0.38 metric tonnes C/ha-yr. This rate is based on the IPCC default method and country-specific factors (see Table A-214), by calculating the effect of converting nominal, medium-input grassland to high input improved grassland (assuming a reference C stock of 50 metric tonnes C/ha, which represents a mid-range value for the dominant cropland soils in the United States, the land use factor for grassland (1.4) and the country-specific factor of 1.11 for high input improved grassland, with the change in stocks occurring over a 20 year (default value) time period; i.e., [50×1.4×1.11 – 50×1.4]/20 = 0.38)). From 1990 through 2005, sewage sludge applications in agricultural lands increased SOC storage from 0.6 to 1.3 Tg CO_2 Eq./year (Table A-214). A nominal ±50 percent uncertainty was attached to these estimates due to limited information on application and the rate of change in soil C stock change with sewage sludge amendments.

The second activity was the change in enrollment for the Conservation Reserve Program after 1997 for mineral soils. Relative to the enrollment in 1997, the total area in the Conservation Reserve Program declined in 1998 through 2000, and then increased 2001 to 2005, leading to an additional enrollment of 893,377 ha over the five-year period (Barbarika 2006). An average annual change in SOC of 0.5 metric tonnes C/ha-yr was used to estimate the effect of the enrollment changes. This rate is based on the IPCC default method and country-specific factors (see Table A- 210) by calculating the impact of setting aside a medium input cropping system in the Conservation Reserve Program (assuming a reference C stock of 50 metric tonnes C/ha, which represents a mid-range value for the dominant cropland soils in the United States and the average country-specific factor of 1.2 for setting-aside cropland from production, with the change in stocks occurring over a 20 yr (default value) time period; i.e., [50×1.2 – 50]/20 = 0.5)). The change in enrollment generated emissions in 1998 through 2000, but with increased enrollment from 2001 to 2005, agricultural land sequestered an additional 0.7 to 1.6 Tg CO_2 Eq. annually between 2001 and 2005, respectively (Table A-215). A nominal ±50 percent uncertainty was also attached to these estimates due to limited information about the enrollment trends at subregional scales, which creates uncertainty in the rate of the soil C stock change (stock change factors for set-aside lands vary by climate region).

Step 5: Compute Net CO_2 Emissions and Removals from Agricultural Lands

The sum of total CO_2 emissions and removals from the Tier 3 Century Model Approach (Step 2), Tier 2 IPCC Methods (Step 3) and additional land-use and management considerations (Step 4) are presented in Table A-215. Agricultural soils, both organic and mineral, were estimated to sequester 39 Tg CO_2 Eq. in 1990, but this rate had declined slightly by the end of the reporting period in 2005 to a net sequestration rate of 36 Tg CO_2 Eq.

Table A-214: Assumptions and Calculations to Estimate the Contribution to Soil Organic Carbon Stocks from Application of Sewage Sludge to Mineral Soils

	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Sewage Sludge N Applied to Agricultural Land (Mg N)[a]	51,222	57,794	64,688	71,906	78,442	85,129	87,245	88,547	89,849	93,430	97,076	100,012	103,028	106,125	109,306	112,573
Assimilative Capacity (Mg N/ha)[b]	0.120	0.120	0.120	0.122	0.122	0.122	0.122	0.122	0.122	0.122	0.122	0.122	0.122	0.122	0.122	0.122
Area covered by Available Sewage Sludge N (ha)[c]	426,848	481,618	539,064	589,391	642,964	697,777	715,122	725,795	736,469	765,821	795,706	819,772	844,490	869,878	895,952	922,729
Average Annual Rate of C storage (Mg C/ha-yr)[d]	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38	0.38
Contribution to Soil C ($TgCO_2$/yr)[e,f]	**(0.59)**	**(0.67)**	**(0.75)**	**(0.82)**	**(0.90)**	**(0.97)**	**(1.00)**	**(1.01)**	**(1.03)**	**(1.07)**	**(1.11)**	**(1.14)**	**(1.18)**	**(1.21)**	**(1.25)**	**(1.29)**
95% C.I. Range	**(0.89) to (0.30)**	**(1.01) to (0.34)**	**(1.13) to (0.38)**	**(1.23) to (0.41)**	**(1.35) to (0.45)**	**(1.46) to (0.49)**	**(1.50) to (0.50)**	**(1.52) to (0.51)**	**(1.55) to (0.52)**	**(1.61) to (0.54)**	**(1.67) to (0.56)**	**(1.71) to (0.57)**	**(1.77) to (0.59)**	**(1.82) to (0.61)**	**(1.88) to (0.63)**	**(1.94) to (0.65)**

Values in parentheses indicate net C storage.

[a] N applied to soils described in Step 1d.

[b] Assimilative Capacity is the national average amount of manure-derived N that can be applied on cropland without buildup of nutrients in the soil (Kellogg et al., 2000).

[c] Area covered by sewage sludge N available for application to soils is the available N applied at the assimilative capacity rate. The 1992 assimilative capacity rate was applied to 1990 – 1992 and the 1997 rate was applied to 1993-2000.

[d] Annual rate of C storage based on national average increase in C storage for grazing lands that is attributed to organic matter amendments (0.38 Mg/ha-yr)

[e] Contribution to Soil C is estimated as the product of the area covered by the available sewage sludge N and the average annual C storage attributed to an organic matter amendment.

[f] Note: Some small, undetermined fraction of this applied N is probably not applied to agricultural soils, but instead is applied to forests, home gardens, and other lands

Table A-215: Annual Soil C Stock Change in *Cropland Remaining Cropland* (CRC), *Land Converted to Cropland* (LCC), *Grassland Remaining Grassland* (GRG), and *Land Converted to Grassland* (LCG), in U.S. Agricultural Soils (Tg CO_2 Eq.)

	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Net emissions based on Tier 3 Century-based analysis (Step 2)																
CRC	(58.59)	(58.59)	(58.59)	(58.59)	(58.59)	(66.44)	(66.44)	(66.44)	(66.44)	(66.44)	(66.44)	(66.44)	(66.44)	(66.44)	(66.44)	(66.44)
LCC	2.08	2.08	2.08	2.08	2.08	0.42	0.42	0.42	0.42	0.42	0.42	0.42	0.42	0.42	0.42	0.42
GRG	(2.81)	(2.81)	(2.81)	(2.81)	(2.81)	13.85	13.85	13.85	13.85	13.85	13.85	13.85	13.85	13.85	13.85	13.85
LCG	(10.49)	(10.49)	(10.49)	(10.49)	(10.49)	12.22)	12.22)	12.22)	12.22)	12.22)	12.22)	12.22)	12.22)	12.22)	12.22)	12.22)
Net emissions based on the IPCC Tier 2 analysis (Step 3)																
Mineral Soils																
CRC	(1.65)	(1.65)	(1.65)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)	(3.02)
LCC	4.16	4.16	4.16	4.14	4.14	4.14	4.14	4.14	4.14	4.14	4.14	4.14	4.14	4.14	4.14	4.14
GRG	(0.31)	(0.31)	(0.31)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)	(0.15)
LCG	(4.54)	(4.54)	(4.54)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)	(4.99)
Organic Soils																
CRC	27.43	27.43	27.43	27.67	27.67	27.67	27.67	27.67	27.67	27.67	27.67	27.67	27.67	27.67	27.67	27.67
LCC	2.42	2.42	2.42	2.64	2.64	2.64	2.64	2.64	2.64	2.64	2.64	2.64	2.64	2.64	2.64	2.64
GRG	3.85	3.85	3.85	3.69	3.69	3.69	3.69	3.69	3.69	3.69	3.69	3.69	3.69	3.69	3.69	3.69
LCG	0.47	0.47	0.47	0.88	0.88	0.88	0.88	0.88	0.88	0.88	0.88	0.88	0.88	0.88	0.88	0.88
Additional changes in net emissions from mineral soils based on application of sewage sludge to agricultural land (Step 4)																
GRG	(0.59)	(0.67)	(0.75)	(0.82)	(0.90)	(0.97)	(1.00)	(1.01)	(1.03)	(1.07)	(1.11)	(1.14)	(1.18)	(1.21)	(1.25)	(1.29)
Additional changes in net emissions from mineral soils based on additional enrollment of CRP land (Step 4)																
CRC	-	-	-	-	-	-	-	-	1.91	2.13	0.94	(0.66)	(0.94)	(1.08)	(1.49)	(1.64)
Total Stock Changes by Land Use/Land-Use Change Category (Step 5)																
CRC	(32.81)	(32.81)	(32.81)	(33.95)	(33.95)	(41.80)	(41.80)	(41.80)	(39.89)	(39.66)	(40.85)	(42.46)	(42.74)	(42.88)	(43.28)	(43.43)
LCC	8.67	8.67	8.67	8.87	8.87	7.21	7.21	7.21	7.21	7.21	7.21	7.21	7.21	7.21	7.21	7.21
GRG	0.13	0.05	(0.03)	(0.09)	(0.17)	16.42	16.40	16.38	16.37	16.32	16.28	16.25	16.22	16.18	16.14	16.11
LCG	(14.55)	(14.55)	(14.55)	(14.59)	(14.59)	(16.32)	(16.32)	(16.32)	(16.32)	(16.32)	(16.32)	(16.32)	(16.32)	(16.32)	(16.32)	(16.32)
Total	(38.6)	(38.6)	(38.7)	(39.8)	(39.8)	(34.5)	(34.5)	(34.5)	(32.6)	(32.5)	(33.7)	(35.3)	(35.6)	(35.8)	(36.3)	(36.4)

Regionally, the total stock change (see Figures 7-4, 7-5, 7-6, and 7-7 in the *Land Use, Land-Use Change, and Forestry* chapter) as well as per hectare rate of change varies among MLRAs (Figure A-12 and Figure A-13). On a per hectare basis, the highest sequestration rates occurred in the southeastern and north-central United States. Other MLRAs with relatively high sequestration rates occur in the southern Great Plains, southern Corn Belt Region, and Pacific Northwest. For organic soils, emission rates were highest in the regions that contain the majority of the drained organic soils, including the southeastern Coastal Region, central and northern portions of the Pacific Coast, and Great Lakes Region. On a per hectare basis, the emission rate patterns were very similar to the total emissions from MLRAs, with the highest rates in those regions with warmer climates and a larger proportion of the drained organic soil managed for crop production.

Figure A-12: Net C Stock Change, per Hectare, for Mineral Soils Under Agricultural Management, 2005

Figure A-13: Net C Stock Change, per Hectare, for Organic Soils Under Agricultural Management, 2005

3.14. Methodology for Estimating CH_4 Emissions from Landfills

Landfill gas is a mixture of substances generated when bacteria decompose the organic materials contained in municipal solid waste (MSW). By volume, MSW landfill gas is about half CH_4 and half CO_2.[63] The amount and rate of CH_4 generation depends upon the quantity and composition of the landfilled material, as well as the surrounding landfill environment.

Not all CH_4 generated within a landfill is emitted to the atmosphere. The CH_4 can be extracted and either flared or utilized for energy, thus oxidizing to CO_2 during combustion. Of the remaining CH_4, a portion oxidizes to CO_2 as it travels through the top layer of the landfill cover. In general, landfill-related CO_2 emissions are of biogenic origin and primarily result from the decomposition, either aerobic or anaerobic, of organic matter such as food or yard wastes.[64]

To estimate the amount of CH_4 produced in a landfill in a given year, information is needed on the type and quantity of waste in the landfill, as well as the landfill characteristics (e.g., size, aridity, waste density). However, this information is not available for all landfills in the United States. Consequently, to estimate CH_4 generation, a methodology was developed based on the quantity of waste placed in landfills nationwide each year, the first order decay model, and model parameters from the analysis of measured CH_4 generation rates for U.S. landfills with gas recovery systems.

From various studies and surveys of the generation and disposal of municipal solid waste, estimates of the amount of waste placed in landfills were developed. A database of measured CH_4 generation rates at landfills with gas recovery systems was compiled and analyzed. The results of this analysis and other studies were used to develop an estimate of the CH_4 generation potential for use in the first order decay model. In addition, the analysis and other studies provided estimates of the CH_4 generation rate constant as a function of precipitation. The first order decay model was applied to annual waste disposal estimates for each year and for three ranges of precipitation to estimate CH_4 generation rates nationwide for the years of interest. Based on the organic content of industrial wastes and the estimates of the fraction of these wastes sent to industrial landfills, CH_4 emissions from industrial landfills were also estimated using the first order decay model. Total CH_4 emissions were estimated by adding the CH_4 from MSW and industrial landfills and subtracting the amounts recovered for energy or flared and the amount oxidized in the soil. The steps taken to estimate CH_4 emissions from U.S. landfills for the years 1990 through 2005 are discussed in greater detail below.

Figure A-14 presents the CH_4 emissions process—from waste generation to emissions—in graphical format.

Step 1: Estimate Annual Quantities of Solid Waste Placed in Landfills

For 1989 to 2005, estimates of the annual quantity of waste placed in MSW landfills were developed from a survey of State agencies as reported in BioCycle's *State of Garbage in America* (BioCycle 2006), adjusted to include U.S. territories.[65] Table A-216 shows estimates of MSW contributing to CH_4 emissions. The table shows BioCycle estimates of total waste landfilled each year from 1990 through 2000, 2002, and 2004, adjusted for U.S. territories. A linear interpolation was used for 2001 and 2003 because there were no BioCycle surveys for those years. An estimate was made for 2005 based on the increase in population since BioCycle data were not yet available at the time this report was published. The estimate for 2005 will be updated when the next BioCycle survey results become available.

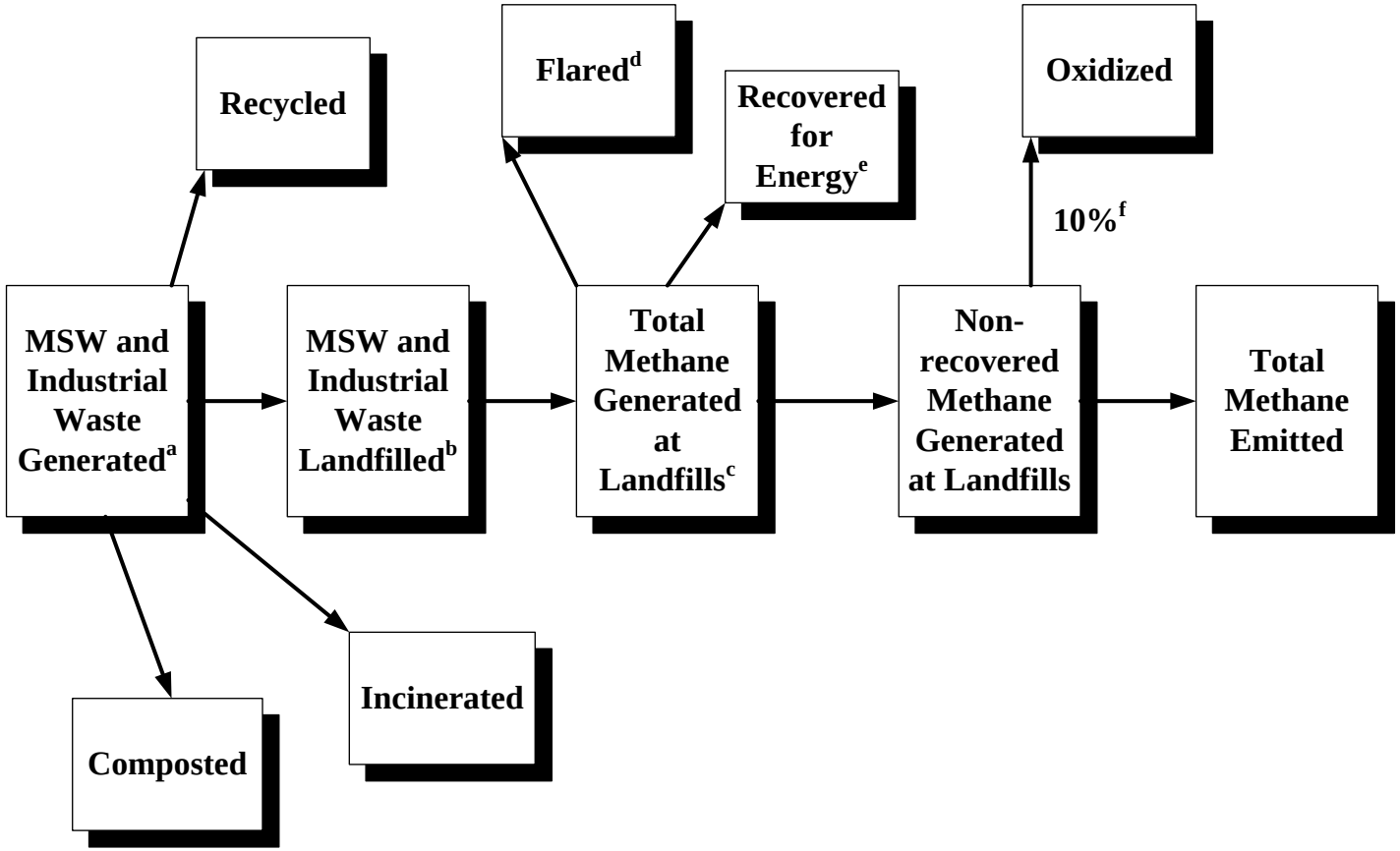
---

[63] Typically, landfill gas also contains small amounts of nitrogen, oxygen, and hydrogen, less than 1 percent nonmethane volatile organic compounds (NMVOCs), and trace amounts of inorganic compounds.

[64] See Box 3-3 in the Energy chapter for additional background on how biogenic emissions of landfill CO_2 are addressed in the U.S. Inventory.

[65] Since the BioCycle survey does not include U.S. territories, waste landfilled in U.S. territories was estimated using population data for the U.S territories (U.S. Census Bureau 2006) and the per capita rate for waste landfilled from BioCycle (2006).

Figure A-14: Methane Emissions Resulting from Landfilling Municipal and Industrial Waste



[a] BioCycle 2006 for MSW and activity factors for industrial waste.
[b] 1960 through 1988 based on EPA 1988 and EPA 1993; 1989 through 2005 based on BioCycle 2006.
[c] 2006 IPCC Revised Guidance – First Order Decay Model.
[d] EIA 2006 and flare vendor database.
[e] EIA 2006 and EPA (LMOP) 2006.
[f] 2006 IPCC Revised Guidance; Mancinelli and McKay 1985; Czepiel et al 1996.

Table A-216: Municipal Solid Waste (MSW) Contributing to CH_4 Emissions (Tg unless otherwise noted)

Description	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
Total MSW Generated[a]	271	259	269	283	298	302	302	314	346	353	377	416	455	462	470	474
Percent of MSW Landfilled[a]	77%	76%	72%	71%	67%	63%	62%	61%	61%	60%	61%	63%	66%	65%	64%	64%
Total MSW Landfilled	209	197	194	201	200	190	187	192	211	212	230	263	298	300	301	304
Waste in Place (30 years)[b]	4,674	4,768	4,848	4,922	5,001	5,075	5,137	5,194	5,252	5,327	5,400	5,488	5,608	5,759	5,909	6,058
MSW Contributing to Emissions[c]	6,815	7,012	7,206	7,407	7,606	7,796	7,984	8,175	8,386	8,598	8,828	9,092	9,390	9,690	9,991	10,295

[a] Source: *BioCycle* (2006), adjusted for missing U.S. territories using U.S. Census Bureau (2006) population data and per capita generation rate from *BioCycle*. The data, originally reported in short tons, are converted to metric tons. Estimates shown for 2001 and 2003 are based on an interpolation because there were no surveys in 2001 and 2003; estimate shown for 2005 based on the increase in population.
[b] This estimate represents the waste that has been in place for 30 years or less, which contributes about 90 percent of the CH_4 generation.
[c] This estimate represents the cumulative amount of waste that has been placed in landfills from 1940 to the year indicated and is the sum of the annual disposal rates used in the first order decay model.

Estimates of the annual quantity of waste placed in landfills from 1960 through 1988 were developed from EPA's 1993 Report to Congress (EPA 1993) and a 1986 survey of MSW landfills (EPA 1988). Based on the national survey and estimates of the growth of commercial, residential and other wastes, the annual quantity of waste placed in landfills averaged 127 million metric tons in the 1960s, 154 million metric tons in the 1970s, and 190 million metric tons in the 1990s. Estimates of waste placed in landfills in the 1940s and 1950s were developed based on U.S. population for each year and the per capital disposal rates from the 1960s.

Step 2: Estimate CH_4 Generation at Municipal Solid Waste Landfills

The CH_4 generation was estimated from the integrated form of the first order decay (FOD) model using the procedures and spreadsheets from IPCC (2006) for estimating CH_4 emissions from solid waste disposal. The form of the FOD model that was applied incorporates a time delay of 6 months after waste disposal before the generation of methane begins.

The input parameters needed for the FOD model equations are the mass of waste disposed each year, which was discussed in the previous section, degradable organic carbon (DOC), and the decay rate constant (k). The DOC is determined from the CH_4 generation potential (L_0 in m^3 CH_4/Mg waste), which is discussed in more detail in subsequent paragraphs, and the following equation:

$$DOC = [L_0 \times 6.74 \times 10^{-4}] \div [F \times 16/12 \times DOC_f \times MCF]$$

Where,
DOC = degradable organic carbon (fraction, Gg C/Gg waste),
L_0 = CH_4 generation potential (m^3 CH_4/Mg waste),
6.74×10^{-4} = CH_4 density (m^3/Mg),
F = fraction of CH_4 by volume in generated landfill gas (equal to 0.5)
16/12 = molecular weight ratio CH_4/C,
DOC_f = fraction of DOC that can decompose in the anaerobic conditions in the landfill (fraction equal to 0.5 for MSW), and
MCF = methane correction factor for year of disposal (fraction equal to 1 for anaerobic managed sites).

The DOC value used in the CH_4 generation estimates from MSW landfills is 0.203 based on the methane generation potential of 100 m^3 CH_4/Mg waste as described below.

Values for the CH_4 generation potential (L_o) were evaluated from landfill gas recovery data for 52 landfills, which resulted in a best fit value for L_o of 99 m^3/Mg of waste (RTI 2004) and from other studies. This value compares favorably with a range of 50 to 162 (midrange of 106) m^3/Mg presented by Peer, Thorneloe, and Epperson (1993); a range of 87 to 91 m^3/Mg from a detailed analysis of 18 landfills sponsored by the Solid Waste Association of North America (SWANA 1998); a value of 100 m^3/Mg recommended in EPA's compilation of emission factors (EPA 1998) based on data from 21 landfills; and a range of 50 to 150 (midrange 100) m^3/Mg based on landfill studies conducted by SCS Engineers. Based on the results from these studies, a value of 100 m^3/Mg appears to be a reasonable best estimate to use in the FOD model for the national inventory.

The FOD model was applied to the gas recovery data for the 52 landfills to calculate the rate constant (k) directly for L_0 = 100 m^3/Mg. The rate constant was found to increase with annual average precipitation; consequently, average values of k were developed for three ranges of precipitation, shown in Table A- 217.

Table A- 217. Average Values for Rate Constant (k) by Precipitation Range (yr^{-1})

Precipitation range (inches/year)	k (yr^{-1})
<20	0.020
20-40	0.038
>40	0.057

These values for k show reasonable agreement with the results of other studies. For example, EPA's compilation of emission factors (EPA 1998) recommends a value of 0.02 yr^{-1} for arid areas (less than 20 inches/year of precipitation) and 0.04 yr^{-1} for non-arid areas. The SWANA study of 18 landfills reported a range in values of k from 0.03 to 0.06 yr^{-1} based on CH_4 recovery data collected generally in the time frame of 1986 to 1995.

Using data collected primarily for the year 2000, the distribution of waste in place vs. precipitation was developed from over 400 landfills (RTI 2004). A distribution was also developed for population vs. precipitation for comparison. The two distributions were very similar and indicated that population in areas or regions with a given precipitation range was a reasonable proxy for waste landfilled in regions with the same range of precipitation. Using U.S. census data and rainfall data, the distributions of population vs. rainfall were developed for each census decade from 1950 through 2000. The distributions showed that the U.S. population has shifted to more arid areas over the past several decades. Consequently, the population distribution was used to apportion the waste landfilled in each decade according to the precipitation ranges developed for k, as shown in Table A-218.

Table A-218. Percent of U.S. Population within Precipitation Ranges (%)

Precipitation Range (inches/year)	1950	1960	1970	1980	1990	2000
<20	11	13	14	16	19	20
20-40	40	39	38	36	34	33
>40	49	48	48	48	47	47

Source: RTI (2004) using population data from the U.S. Bureau of Census and precipitation data from the National Climatic Data Center's National Oceanic and Atmospheric Administration.

Step 3: Estimate CH_4 Generation at Industrial Landfills

Industrial landfills receive waste from factories, processing plants, and other manufacturing activities. In previous national inventories, CH_4 generation at industrial landfills was estimated as seven percent of the total CH_4 generation from MSW landfills, based on a study conducted by EPA (1993). For the current inventory, the methodology was updated and improved by using activity factors (industrial production levels) to estimate the amount of industrial waste landfilled each year and by applying the FOD model to estimate CH_4 generation. A nationwide survey of industrial waste landfills found that over 99 percent of the organic waste placed in industrial landfills originated from two industries: food processing (meat, vegetables, fruits) and pulp and paper (EPA 1993). Data for annual nationwide production for the food processing and pulp and paper industries were taken from industry and government sources for recent years; estimates were developed for production for the earlier years for which data were not available based on the change in U.S. population. For the pulp and paper industry, production data published by the Lockwood-Post's Directory (ERG 2006) and U.S. Department of Agriculture (2003) were the primary sources for years 1965-2005. An extrapolation based on U.S. population was used for years 1940 through 1964. For the food processing industry, production levels were obtained or developed from the U.S. Department of Agriculture for the years 1990 through 2005 (ERG 2006). An extrapolation based on U.S. population was used for the years 1940 through 1989.

In addition to production data for the pulp and paper and food processing industries, the following inputs were needed to use the FOD model for estimating CH_4 generation from industrial landfills: 1) quantity of waste that is disposed in industrial landfills (as a function of production), 2) CH_4 generation potential (L_0) or DOC, and 3) FOD decay constant (k). Research into waste generation and disposal in landfills for the pulp and paper industry indicated that the quantity of waste landfilled was about 0.050 Mg/Mg of product compared to 0.046 Mg/Mg product for the food processing industry (Weitz and Bahner 2006). These factors were applied to estimates of annual production to estimate annual waste disposal in landfills. Estimates for DOC and k were taken from the default values in the 2006 IPCC revised guidelines. The DOC value is 0.20 (L_o of 99 m^3/Mg) for industrial pulp and

paper waste and 0.29 (L_o of 143 m^3/Mg) for industrial food waste. The value of k given for food waste with disposal in a wet temperate climate is 0.185 yr^{-1}, and the value given for paper waste is 0.06 yr^{-1}.

The parameters discussed above were used in the integrated form of the FOD model to estimate CH_4 generation from industrial landfills. Although this is a completely new and independent approach from the simpler one used in previous inventories, the effect of this change was small (a decrease of 2 percent in the estimate of CH_4 generation from industrial landfills over the time series).

Step 4: Estimate CH_4 Emissions Avoided

The estimate of CH_4 emissions avoided (e.g., combusted) was based on landfill-specific data on landfill gas-to-energy (LFGTE) projects and flares. A destruction efficiency of 99 percent was applied to CH_4 recovered to estimate CH_4 emissions avoided. The value for efficiency was selected based on the range of efficiencies (98 to 100 percent) recommended for flares in EPA's "AP-42 Compilation of Air Pollutant Emission Factors, Chapter 2.4," efficiencies used to establish new source performance standards (NSPS) for landfills, and in recommendations for closed flares used in the Landfill Methane Outreach Program (LMOP).

Step 4a: Estimate CH_4 Emissions Avoided Through Landfill Gas-to-Energy (LFGTE) Projects

The quantity of CH_4 avoided due to LFGTE systems was estimated based on information from two sources: (1) a database maintained by the Energy Information Administration (EIA) for the voluntary reporting of greenhouse gases (EIA 2006) and (2) a database compiled by LMOP (EPA 2006). The EIA database included location information for landfills with LFGTE projects, estimates of CH_4 reductions, descriptions of the projects, and information on the methodology used to determine the CH_4 reductions. Generally the CH_4 reductions for each reporting year were based on the measured amount of landfill gas collected and the percent CH_4 in the gas. For the LMOP database, data on landfill gas flow and energy generation (i.e., MW capacity) were used to estimate the total direct CH_4 emissions avoided due to the LFGTE project. Detailed information on the landfill name, owner or operator, city, and state were available for both the EIA and LMOP databases; consequently, it was straightforward to identify landfills that were in both databases. The EIA database was given priority because reductions were reported for each year and were based on direct measurements. Landfills in the LMOP database that were also in the EIA database were dropped to avoid double counting.

Step 4b: Estimate CH_4 Emissions Avoided Through Flaring

The quantity of CH_4 flared was based on data from the EIA database and on information provided by flaring equipment vendors. To avoid double-counting, flares associated with landfills in the EIA and LMOP databases were excluded from the flare vendor database. As with the LFGTE projects, reductions from flaring landfill gas in the EIA database were based on measuring the volume of gas collected and the percent of CH_4 in the gas. The information provided by the flare vendors included information on the number of flares, flare design flow rates or flare dimensions, year of installation, and generally the city and state location of the landfill. When a range of design flare flow rates was provided by the flare vendor, the median landfill gas flow rate was used to estimate CH_4 recovered from each remaining flare (i.e., for each flare not associated with a landfill in the EIA or LMOP databases). Several vendors provided information on the size of the flare rather than the flare design gas flow rate. To estimate a median flare gas flow rate for flares associated with these vendors, the size of the flare was matched with the size and corresponding flow rates provided by other vendors. Some flare vendors reported the maximum capacity of the flare. An analysis of flare capacity versus measured CH_4 flow rates from the EIA database showed that the flares operated at 51 percent of capacity when averaged over the time series and at 72 percent of capacity for the highest flow rate for a given year. For those cases when the flare vendor supplied maximum capacity, the actual flow was estimated as 50 percent of capacity. Total CH_4 avoided through flaring from the flare vendor database was estimated by summing the estimates of CH_4 recovered by each flare for each year.

Step 4c: Reduce CH_4 Emissions Avoided Through Flaring

As mentioned in Step 4b, flares in the flare vendor database associated with landfills in the EIA and LMOP databases were excluded from the flare reduction estimates in the flare vendor database. If comprehensive data on flares was available, each LFGTE project in the EIA and LMOP databases would have an identified flare because

most LFGTE projects have flares. However, given that the flare vendor data only covers approximately 50 to 75 percent of the flare population, an associated flare was not identified for all LFGTE projects. These LFGTE projects likely have flares; however, flares were unable to be identified due to one of two reasons: 1) inadequate identifier information in the flare vendor data; or 2) the lack of the flare in the flare vendor database. For those projects for which a flare was not identified due to inadequate information, CH_4 avoided would be overestimated, as both the CH_4 avoided from flaring and the LFGTE project would be counted. To avoid overestimating emissions avoided from flaring, the CH_4 avoided from LFGTE projects with no identified flares was determined and the flaring estimate from the flare vendor database was reduced by this quantity on a state-by-state basis. This step likely underestimates CH_4 avoided due to flaring. This approach was applied to be conservative in the estimates of CH_4 emissions avoided.

Step 5: Estimate CH_4 Oxidation

A portion of the CH_4 escaping from a landfill oxidizes to CO_2 in the top layer of the soil. The amount of oxidation depends upon the characteristics of the soil and the environment. For purposes of this analysis, it was assumed that of the CH_4 generated, minus the amount of gas recovered for flaring or LFGTE projects, ten percent was oxidized in the soil (Jensen and Pipatti 2002; Mancinelli and McKay 1985; Czepiel et al 1996). The factor of 10 percent is consistent with the value recommended in the 2006 IPCC revised guidelines for managed and covered landfills. This oxidation factor was applied to the estimates of CH_4 generation minus recovery for both MSW and industrial landfills.

Step 6: Estimate Total CH_4 Emissions

Total CH_4 emissions were calculated by adding emissions from MSW and industrial waste, and subtracting CH_4 recovered and oxidized, as shown in Table A-219.

Table A-219: CH_4 Emissions from Landfills (Gg)

Activity	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005
MSW Generation	8,985	9,175	9,324	9,458	9,607	9,745	9,850	9,944	10,045	10,198	10,348	10,541	10,820	11,188	11,543	11,885
Industrial Generation	614	620	628	638	649	664	678	692	706	718	731	744	749	757	761	767
Potential Emissions	**9,599**	**9,796**	**9,953**	**10,096**	**10,256**	**10,410**	**10,529**	**10,635**	**10,752**	**10,917**	**11,079**	**11,285**	**11,570**	**11,944**	**12,303**	**12,653**
Emissions Avoided	(1,079)	(1,243)	(1,165)	(1,348)	(1,612)	(2,099)	(2,415)	(2,886)	(3,401)	(3,751)	(4,101)	(4,531)	(4,670)	(4,805)	(5,312)	(5,668)
Landfill Gas-to-Energy	(840)	(861)	(717)	(800)	(851)	(1,061)	(1,256)	(1,566)	(1,898)	(2,155)	(2,335)	(2,588)	(2,590)	(2,614)	(2,720)	(2,790)
Flare	(239)	(382)	(448)	(548)	(762)	(1,039)	(1,159)	(1,320)	(1,503)	(1,595)	(1,766)	(1,943)	(2,080)	(2,192)	(2,593)	(2,877)
Oxidation at MSW Landfills	(791)	(793)	(816)	(811)	(799)	(765)	(744)	(706)	(664)	(645)	(625)	(601)	(615)	(638)	(623)	(622)
Oxidation at Industrial Landfills	(61)	(62)	(63)	(64)	(65)	(66)	(68)	(69)	(71)	(72)	(73)	(74)	(75)	(76)	(76)	(77)
Net Emissions	**7,668**	**7,697**	**7,909**	**7,873**	**7,779**	**7,479**	**7,303**	**6,974**	**6,615**	**6,449**	**6,280**	**6,078**	**6,210**	**6,425**	**6,292**	**6,286**

Note: Totals may not sum due to independent rounding.
Note: MSW generation in Table A-219 represents emissions before oxidation. In other tables throughout the text, MSW generation estimates account for oxidation.
Note: Parentheses denote negative values.

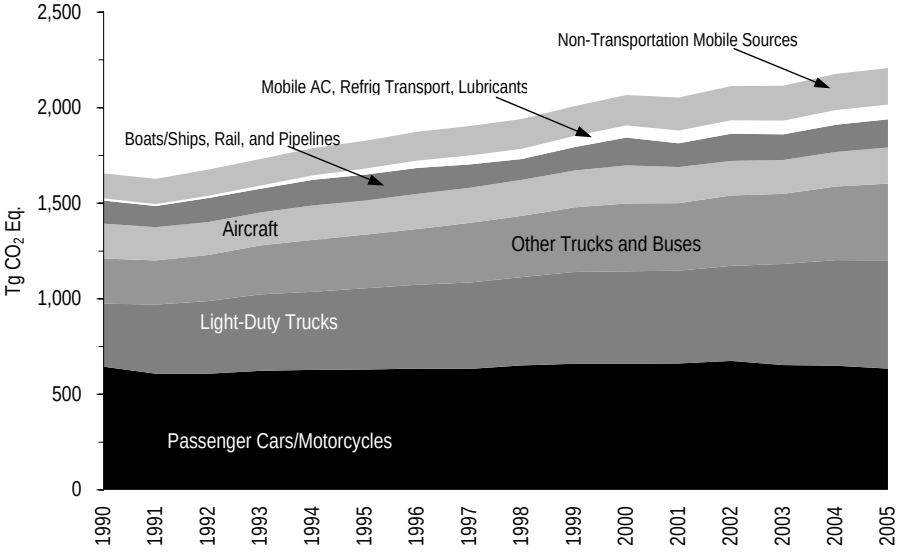


Figure A- 4: Domestic Greenhouse Gas Emissions by Mode and Vehicle Type, 1990 to 2005 (Tg CO2 Eq.)

Figure A-5

Major Crops, Direct N_2O Emissions Estimated Using the DAYCENT Model, 1990-2005 (kg N/ha/yr)

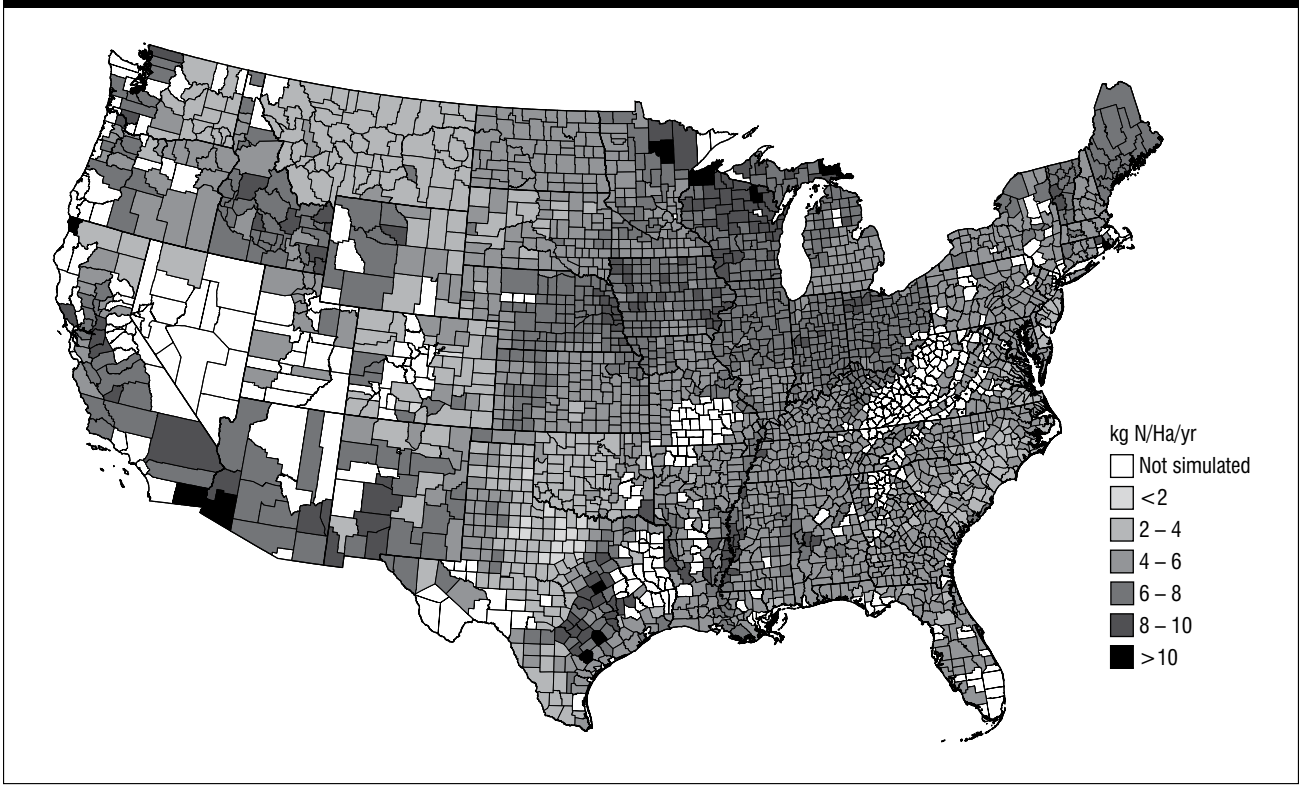


Figure A-6

Major Crops, N Losses Leading to Indirect N_2O Emissions Estimated Using the DAYCENT Model, 1990-2005 (kg N/ha/yr)

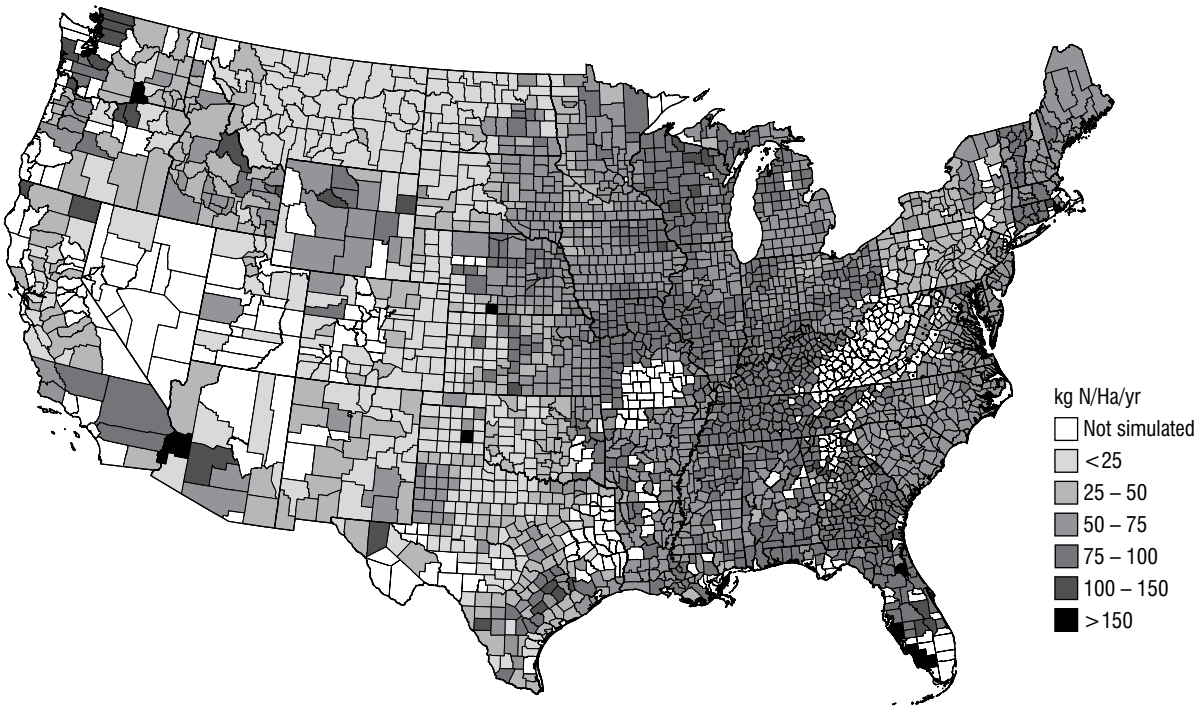


Figure A-7

Grasslands, Direct N_2O Emissions Estimated Using the DAYCENT Model, 1990-2005 (kg N/ha/yr)

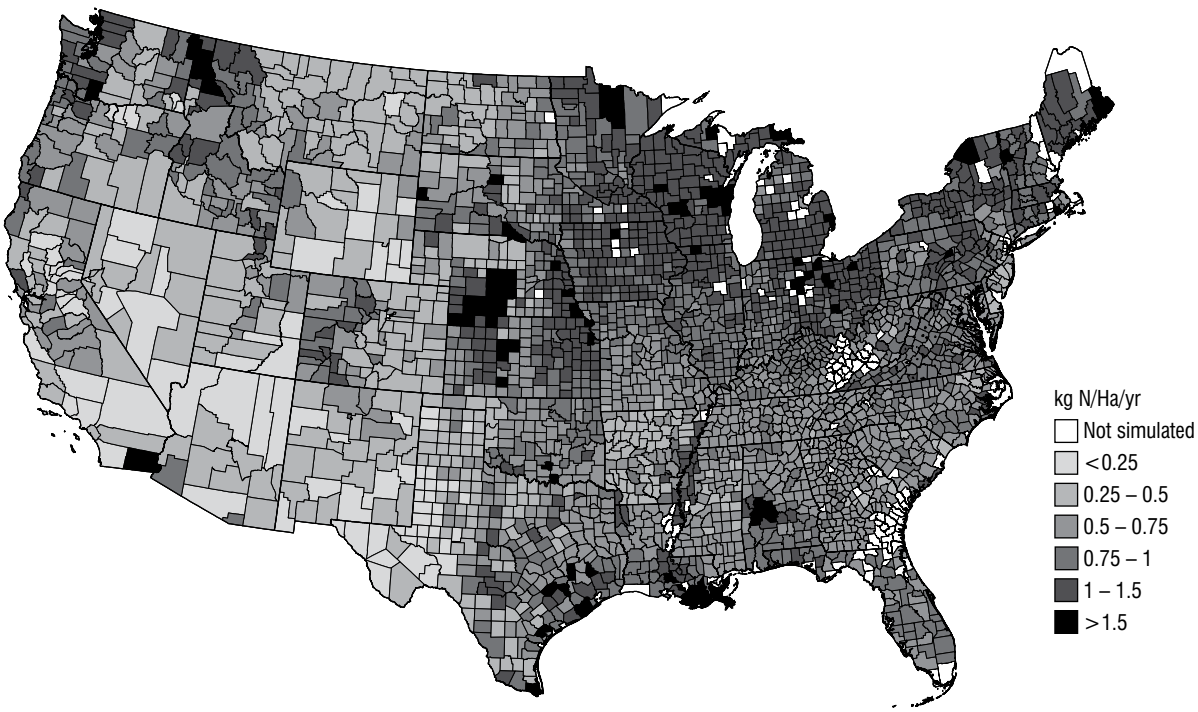


Figure A-8

Grasslands, N Losses Leading to Indirect N_2O Emissions Estimated Using the DAYCENT Model, 1990-2005 (kg N/ha/yr)

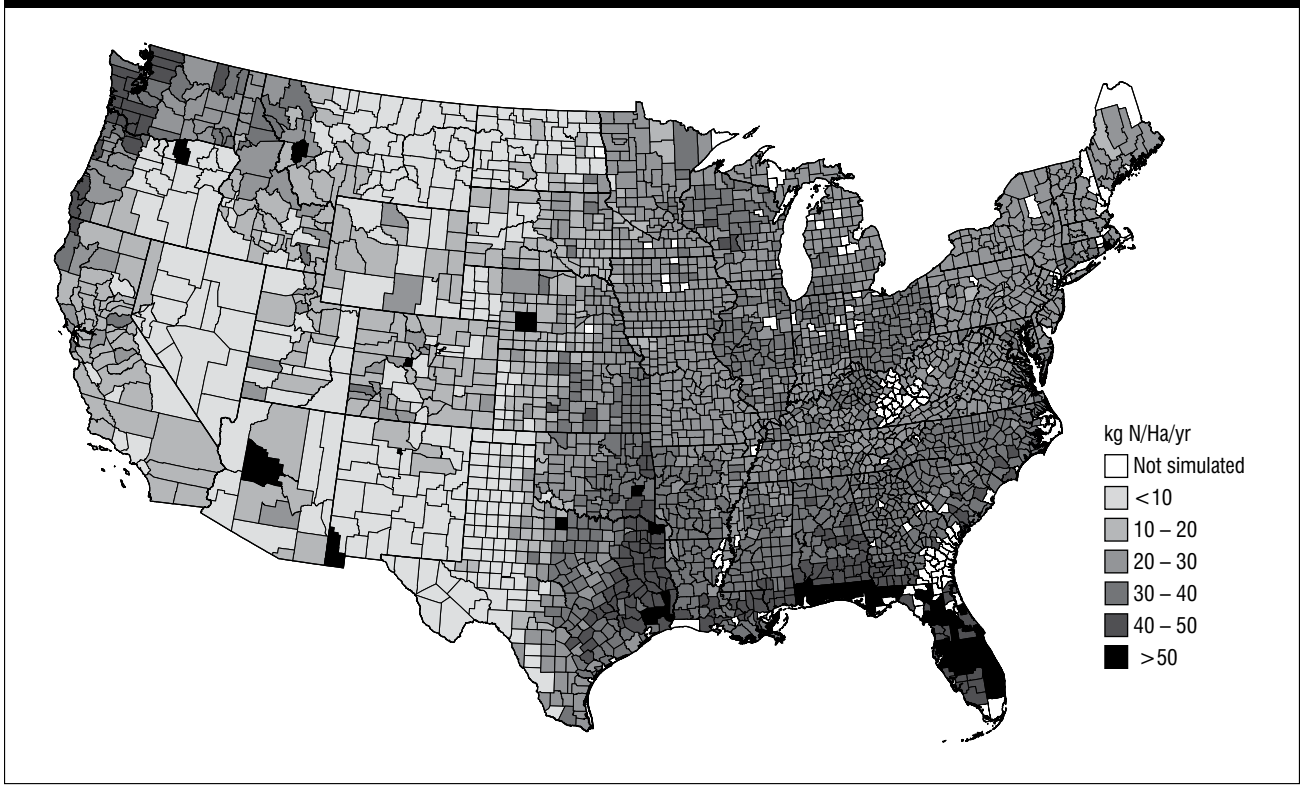
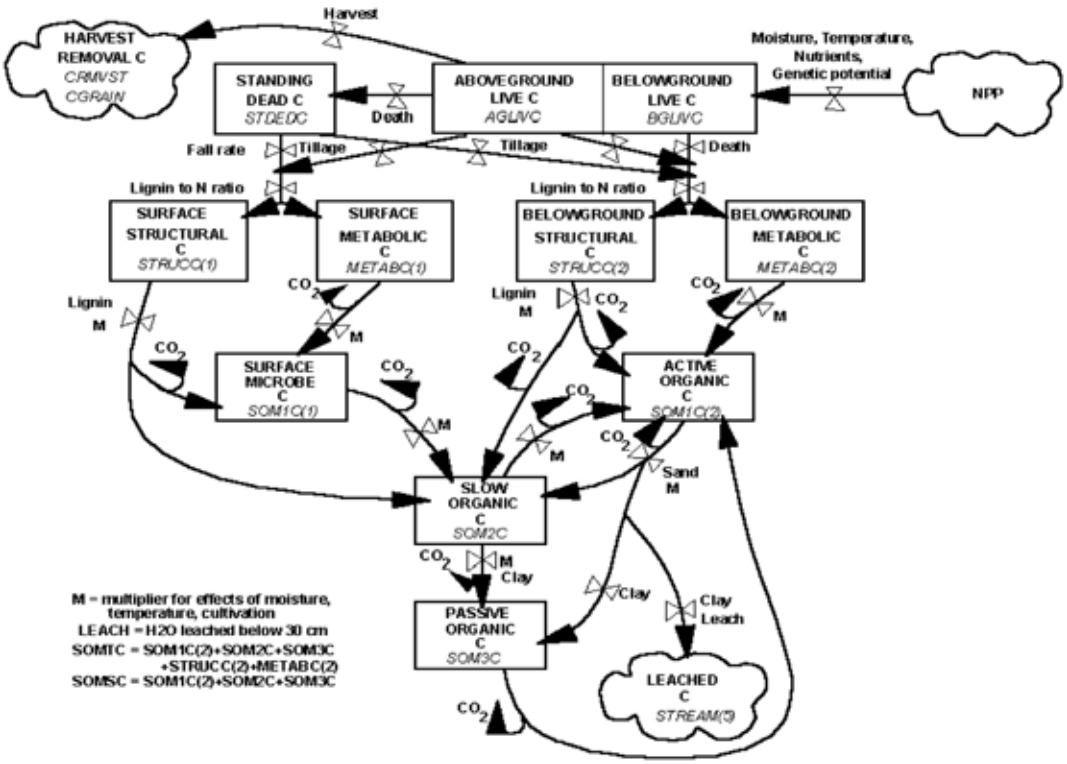


Figure A-9

Flow diagram of Carbon submodel (A) and Nitrogen submodel (B)

A



B

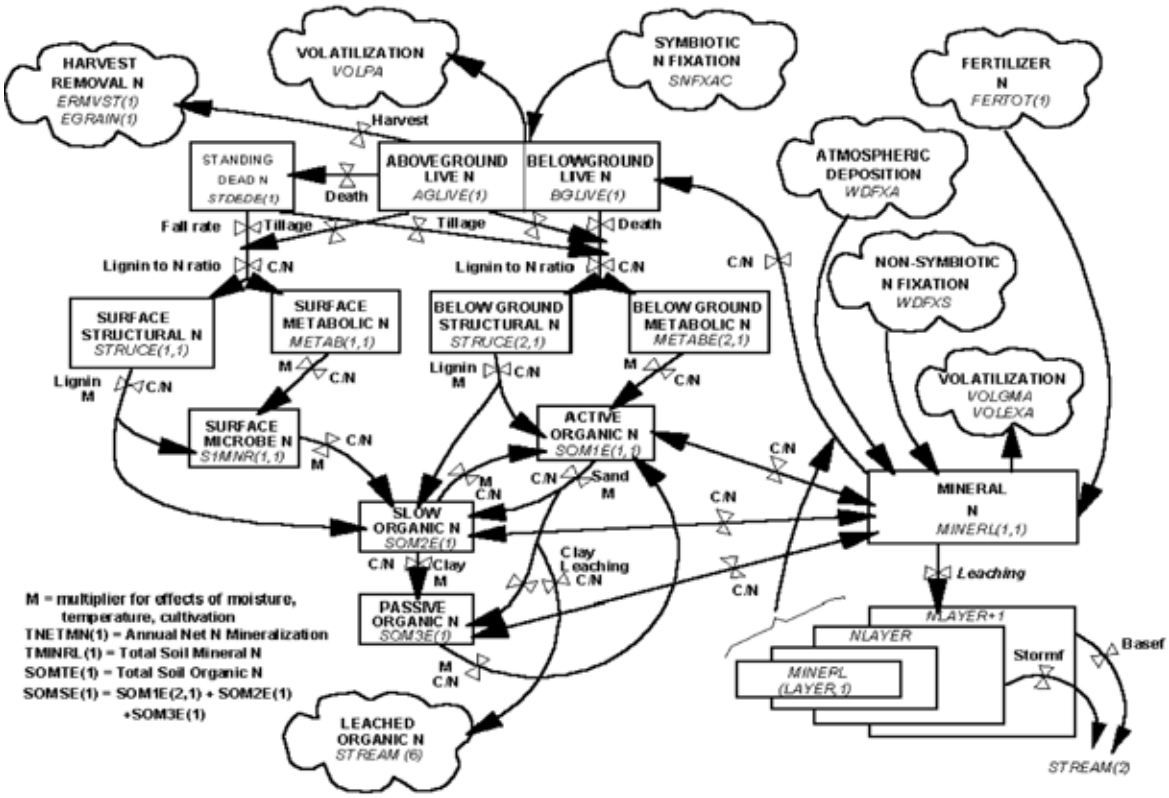
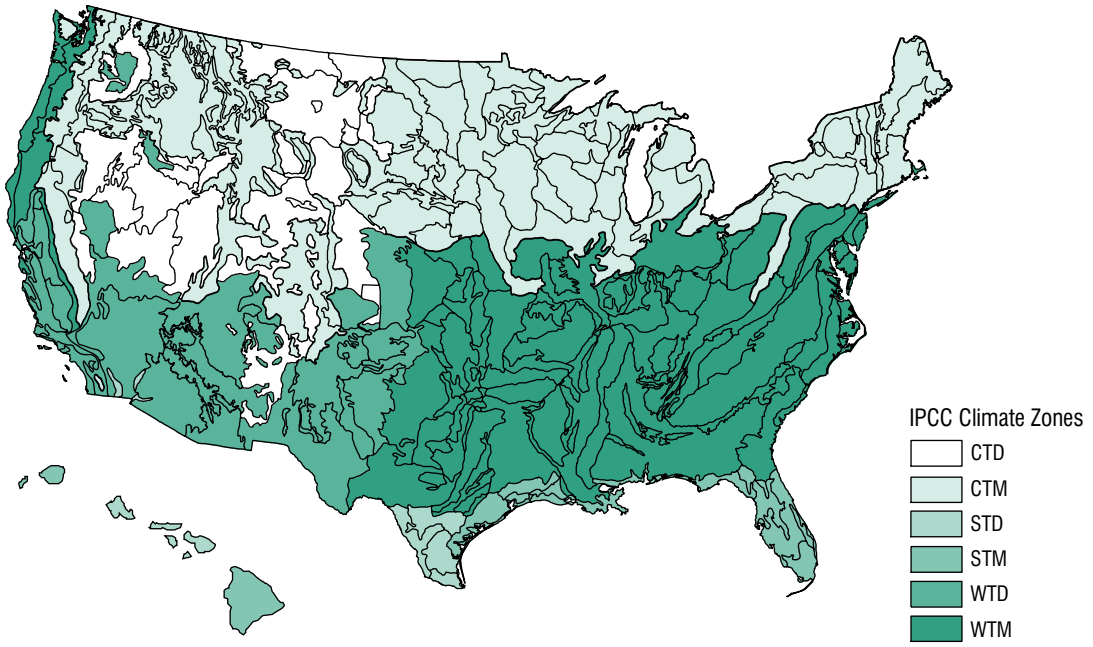


Figure A-10

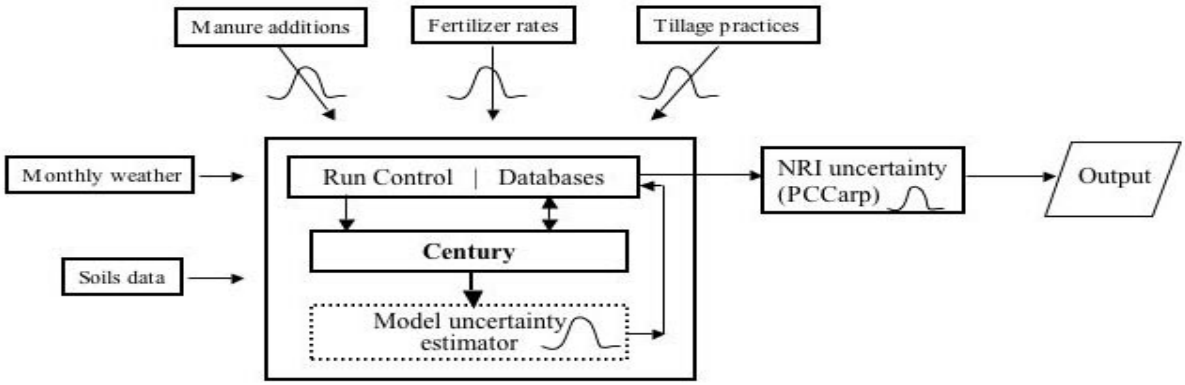
Major Land Resource Areas by IPCC Climate Zone



This figure shows the IPCC climate zone assigned to each of the 180 Major Land Resource Areas (MLRAs) in the United States, based on PRISM climate data averaged for each MRLA.

Figure A-11

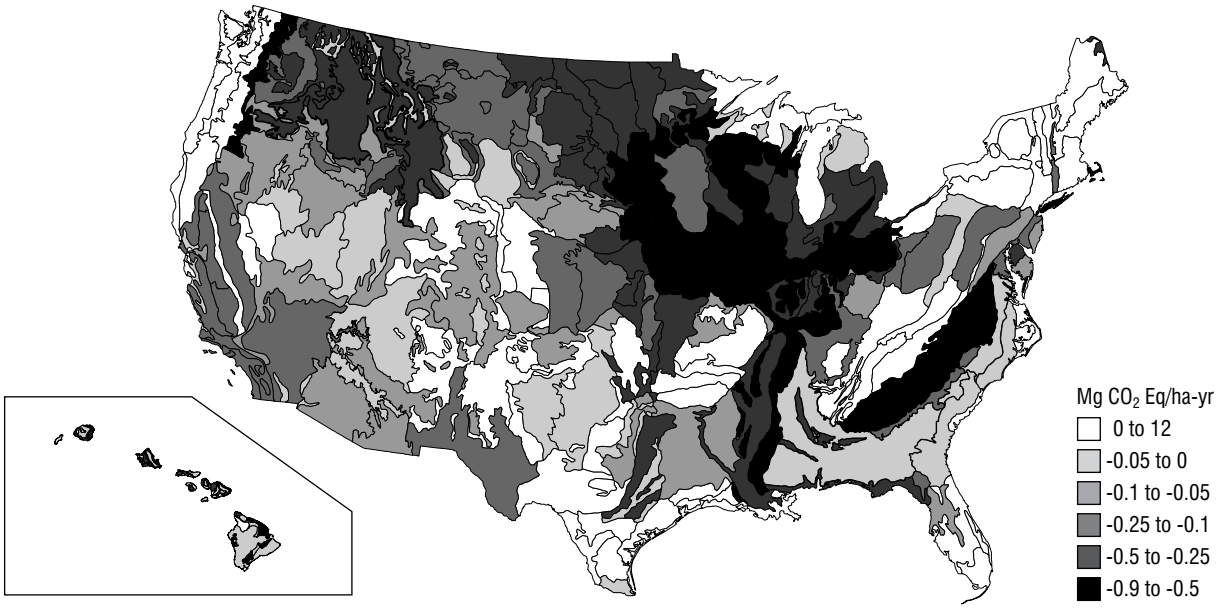
Uncertainty in Data Inputs



Uncertainty in data inputs (i.e. fertilizer, manure, and tillage practices) are estimated using a Monte-Carlo procedure with 100 random draws from input data probability distributions, for each NRI point simulated. Model uncertainty is estimated through an empirically-based approach. Uncertainty in the land representation of NRI is estimated from the statistics compiled from the NRI surveys to determine the land area expansion factors, which are used to upscale data to the national level.

Figure A-12

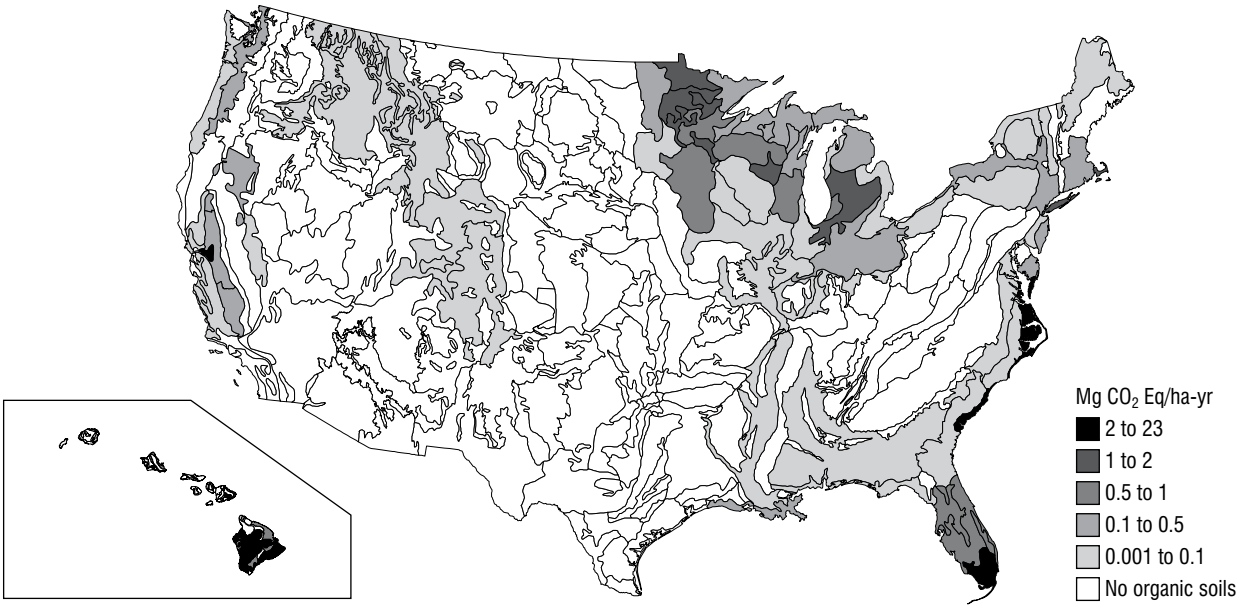
Net Soil C Stock Change, per Hectare, for Mineral Soils Under Agricultural Management, 2005



Note: Values greater than zero represent emissions, and values less than zero represent sequestration. Map accounts for fluxes associated with the Tier 2 and 3 Inventory computations. See Methodology for additional details.

Figure A-13

Net Soil C Stock Change, per Hectare, for Organic Soils Under Agricultural Management, 2005



Note: Values greater than zero represent emissions. Map accounts for fluxes associated with the Tier 2 and 3 Inventory computations. See Methodology for additional details.