3. Energy

Energy-related activities were the primary sources of U.S. anthropogenic greenhouse gas emissions, accounting for 82.1 percent of total greenhouse gas emissions on a carbon dioxide (CO_2) equivalent basis in 2021.[1] This included 96.5, 41.6, and 10.3 percent of the nation's CO_2, methane (CH_4), and nitrous oxide (N_2O) emissions, respectively. Energy-related CO_2 emissions alone constituted 76.7 percent of U.S. greenhouse gas emissions from all sources on a CO_2-equivalent basis, while the non-CO_2 emissions from energy-related activities represented a much smaller portion of total national emissions (5.4 percent collectively).

Emissions from fossil fuel combustion comprise the vast majority of energy-related emissions, with CO_2 being the primary gas emitted (see Figure 3-1 and Figure 3-2). Globally, approximately 33,000 million metric tons (MMT) of CO_2 were added to the atmosphere through the combustion of fossil fuels in 2021, of which the United States accounted for approximately 14 percent.[2] Due to their relative importance over time (see Figure 3-2), fossil fuel combustion-related CO_2 emissions are considered in more detail than other energy-related emissions in this report (see Figure 3-3).

Fossil fuel combustion also emits CH_4 and N_2O. Stationary combustion of fossil fuels was the second largest source of N_2O emissions in the United States and mobile fossil fuel combustion was the fifth largest source. Energy-related activities other than fuel combustion, such as the production, transmission, storage, and distribution of fossil fuels, also emit greenhouse gases. These emissions consist primarily of fugitive CH_4 emissions from natural gas systems, coal mining, and petroleum systems.

[1] Estimates are presented in units of million metric tons of carbon dioxide equivalent (MMT CO_2 Eq.), which weight each gas by its global warming potential, or GWP, value. See section on global warming potentials in the Executive Summary.

[2] Global CO_2 emissions from fossil fuel combustion were taken from International Energy Agency *Global energy-related CO_2 emissions, 1990-2021 – Charts* Available at: https://www.iea.org/data-and-statistics/charts/global-energy-related-co2-emissions-1990-2021 (IEA 2022).

Figure 3-1: 2021 Energy Sector Greenhouse Gas Sources

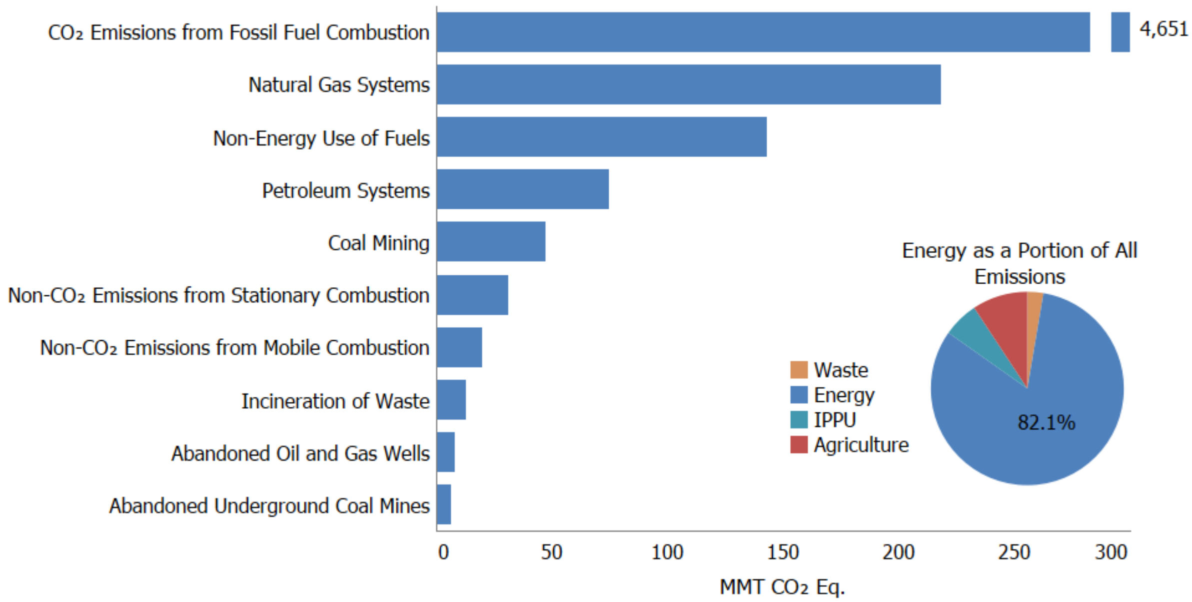


Figure 3-2: Trends in Energy Sector Greenhouse Gas Sources

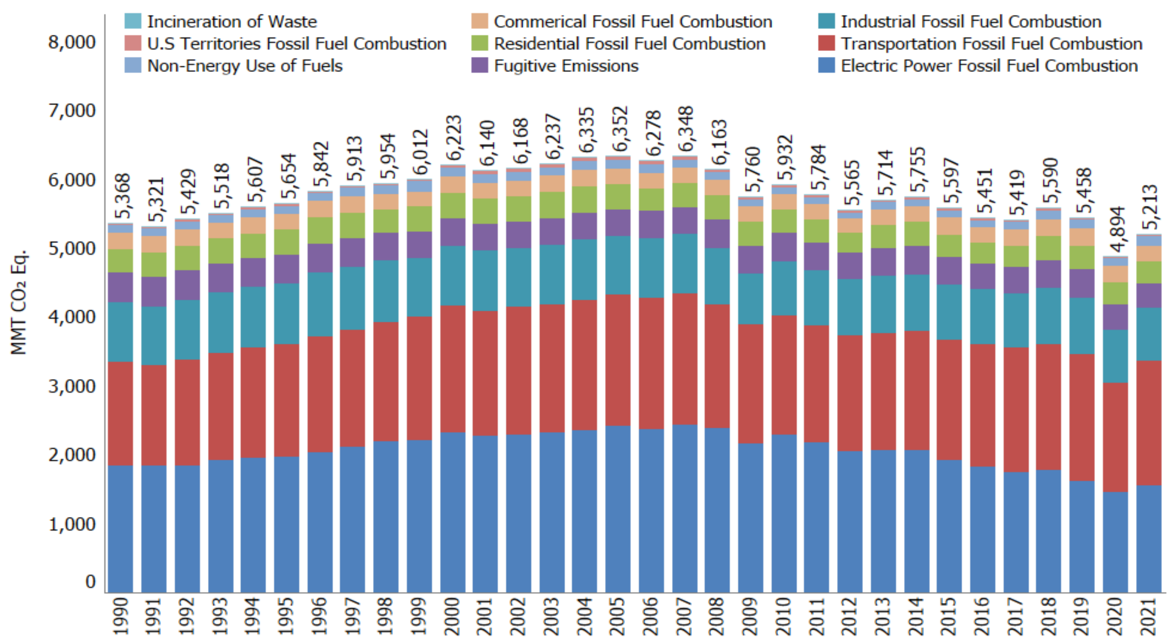


Figure 3-3: 2021 U.S. Fossil Carbon Flows

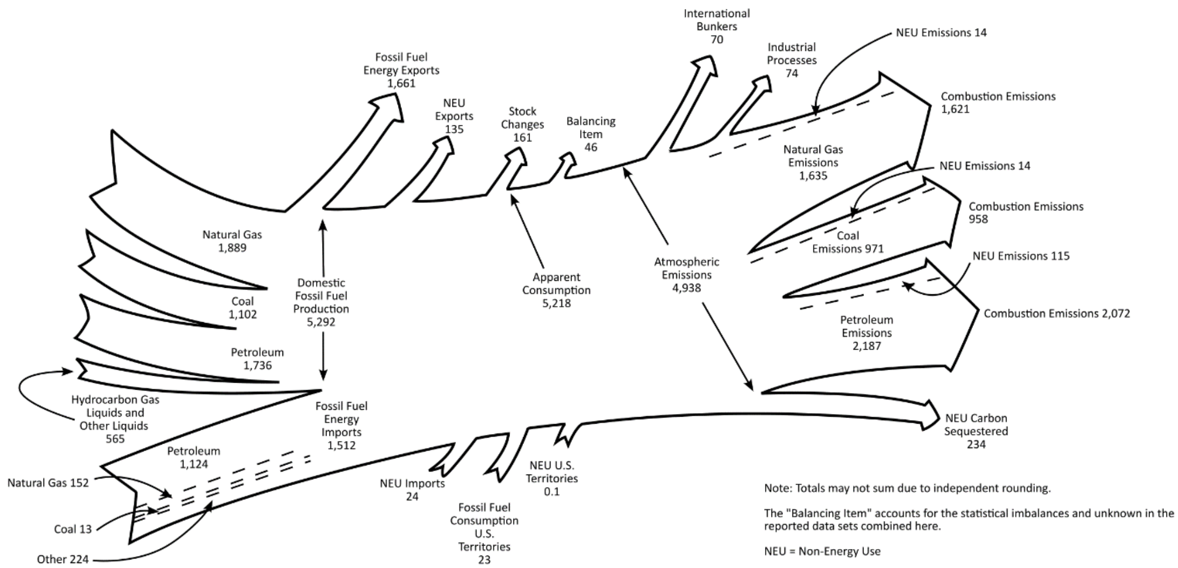


Table 3-1 summarizes emissions from the Energy sector in units of MMT CO_2 Eq., while unweighted gas emissions in kilotons (kt) are provided in Table 3-2. Overall, emissions due to energy-related activities were 5,212.5 MMT CO_2 Eq. in 2021,[3] a decrease of 2.9 percent since 1990 and an increase of 6.5 percent since 2020. The increase in 2021 emissions was due to rebounding activity levels after the coronavirus (COVID-19) pandemic reduced overall demand for fossil fuels across all sectors in 2020. Longer term trends are driven by a number of factors including a shift from coal to natural gas and renewables in the electric power sector.

Table 3-1: CO_2, CH_4, and N_2O Emissions from Energy (MMT CO_2 Eq.)

Gas/Source	1990	2005	2017	2018	2019	2020	2021
CO_2	**4,900.0**	**5,929.1**	**5,037.9**	**5,204.8**	**5,082.5**	**4,544.5**	**4,870.6**
Fossil Fuel Combustion	4,728.2	5,747.3	4,852.5	4,989.8	4,853.4	4,344.8	4,651.0
Transportation	*1,468.9*	*1,858.6*	*1,780.1*	*1,812.9*	*1,813.9*	*1,572.5*	*1,789.4*
Electricity Generation	*1,820.0*	*2,400.1*	*1,732.0*	*1,753.4*	*1,606.7*	*1,439.6*	*1,542.2*
Industrial	*852.4*	*850.8*	*789.0*	*813.5*	*815.9*	*767.9*	*762.4*
Residential	*338.6*	*358.9*	*293.4*	*338.2*	*341.4*	*313.2*	*310.1*
Commercial	*228.3*	*227.1*	*232.0*	*245.8*	*250.7*	*228.5*	*223.9*
U.S. Territories	*20.0*	*51.9*	*25.9*	*25.9*	*24.8*	*23.2*	*23.0*
Non-Energy Use of Fuels	112.4	128.9	112.8	129.4	127.6	119.2	143.2
Natural Gas Systems	32.4	25.2	31.8	33.0	38.7	36.3	36.8
Petroleum Systems	9.5	10.2	24.5	36.1	46.9	29.1	24.7
Incineration of Waste	12.9	13.3	13.2	13.3	12.9	12.9	12.5
Coal Mining	4.6	4.2	3.2	3.1	3.0	2.2	2.5
Abandoned Oil and Gas Wells	+	+	+	+	+	+	+
Biomass-Wood[a]	*215.2*	*206.9*	*212.0*	*220.0*	*217.7*	*200.4*	*202.8*
Biofuels-Ethanol[a]	*4.2*	*22.9*	*82.1*	*81.9*	*82.6*	*71.8*	*79.1*
International Bunker Fuels[b]	*103.6*	*113.3*	*120.2*	*122.2*	*116.1*	*69.6*	*69.3*
Biofuels-Biodiesel[a]	*0.0*	*0.9*	*18.7*	*17.9*	*17.1*	*17.7*	*16.1*
Biomass-MSW[a]	*18.5*	*14.7*	*16.1*	*16.1*	*15.7*	*15.6*	*15.3*
CH_4	**407.0**	**354.8**	**336.7**	**341.8**	**334.2**	**312.0**	**302.3**
Natural Gas Systems	215.1	203.4	186.4	194.4	193.6	185.4	181.4

[3] Following the current reporting requirements under the UNFCCC, this Inventory report presents CO_2 equivalent values based on the IPCC *Fifth Assessment Report* (AR5) GWP values. See Chapter 1, Introduction for more information.

Petroleum Systems	51.3	50.9	61.9	60.6	59.9	54.5	50.2
Coal Mining	108.1	71.8	61.4	59.1	53.0	46.2	44.7
Stationary Combustion	9.6	8.8	8.6	9.6	9.8	8.8	8.9
Abandoned Oil and Gas Wells	7.7	8.1	8.3	8.3	8.3	8.2	8.2
Abandoned Underground Coal Mines	8.1	7.4	7.2	6.9	6.6	6.5	6.4
Mobile Combustion	7.2	4.4	2.9	2.9	2.9	2.6	2.6
Incineration of Waste	+	+	+	+	+	+	+
International Bunker Fuels[b]	*0.2*	*0.1*	*0.1*	*0.1*	*0.1*	*0.1*	*0.1*
N_2O	**61.1**	**67.9**	**44.2**	**43.1**	**41.5**	**37.2**	**39.6**
Stationary Combustion	22.3	30.5	25.3	25.1	22.2	20.7	22.1
Mobile Combustion	38.4	37.0	18.5	17.5	19.0	16.1	17.1
Incineration of Waste	0.4	0.3	0.4	0.4	0.4	0.3	0.4
Petroleum Systems	+	+	+	+	+	+	+
Natural Gas Systems	+	+	+	+	+	+	+
International Bunker Fuels[b]	*0.8*	*0.9*	*0.9*	*1.0*	*0.9*	*0.5*	*0.5*
Total	**5,368.2**	**6,351.8**	**5,418.8**	**5,589.7**	**5,458.3**	**4,893.8**	**5,212.5**

+ Does not exceed 0.05 MMT CO_2 Eq.

[a] Emissions from Biomass and Biofuel Consumption are not included specifically in summing energy sector totals. Net carbon fluxes from changes in biogenic carbon reservoirs are accounted for in the estimates for LULUCF.

[b] Emissions from International Bunker Fuels are not included in totals. These values are presented for informational purposes only, in line with the *2006 IPCC Guidelines* and UNFCCC reporting obligations.

Note: Totals may not sum due to independent rounding.

Table 3-2: CO_2, CH_4, and N_2O Emissions from Energy (kt)

Gas/Source	1990	2005	2017	2018	2019	2020	2021
CO_2	**4,899,997**	**5,929,084**	**5,037,909**	**5,204,849**	**5,082,550**	**4,544,547**	**4,870,614**
Fossil Fuel Combustion	4,728,194	5,747,307	4,852,515	4,989,843	4,853,402	4,344,837	4,650,953
Non-Energy Use of Fuels	112,407	128,920	112,841	129,441	127,621	119,208	143,209
Natural Gas Systems	32,363	25,206	31,770	32,974	38,705	36,296	36,846
Petroleum Systems	9,519	10,221	24,462	36,102	46,874	29,081	24,667
Incineration of Waste	12,900	13,254	13,161	13,339	12,948	12,921	12,476
Coal Mining	4,606.5	4,169.7	3,153.1	3,141.4	2,992.3	2,197.6	2,456.0
Abandoned Oil and Gas Wells	7	7	7	7	8	7	7
Biomass-Wood[a]	*215,186*	*206,901*	*211,965*	*220,005*	*217,692*	*200,421*	*202,841*
Biofuels-Ethanol[a]	*4,227*	*22,943*	*82,088*	*81,917*	*82,578*	*71,848*	*79,064*
International Bunker Fuels[b]	*103,634*	*113,328*	*120,192*	*122,179*	*116,132*	*69,638*	*69,280*
Biofuels-Biodiesel[a]	*0*	*856*	*18,705*	*17,936*	*17,080*	*17,678*	*16,112*
Biomass-MSW[a]	*18,534*	*14,722*	*16,130*	*16,115*	*15,709*	*15,614*	*15,329*
CH_4	**14,537**	**12,671**	**12,024**	**12,208**	**11,934**	**11,145**	**10,798**
Natural Gas Systems	7,682	7,263	6,657	6,943	6,915	6,620	6,479
Petroleum Systems	1,833	1,819	2,209	2,165	2,138	1,945	1,791
Coal Mining	3,860	2,566	2,192	2,110	1,893	1,648	1,595
Stationary Combustion	344	313	307	344	351	313	316

Abandoned Oil and Gas Wells	274	289	295	296	297	295	295
Abandoned Underground Coal Mines	288	264	257	247	237	232	228
Mobile Combustion	258	158	105	102	103	92	94
Incineration of Waste	+	+	+	+	+	+	+
International Bunker Fuels[b]	*7*	*5*	*4*	*4*	*4*	*3*	*3*
N_2O	**231**	**256**	**167**	**163**	**157**	**140**	**149**
Stationary Combustion	84	115	95	95	84	78	83
Mobile Combustion	145	140	70	66	72	61	65
Incineration of Waste	2	1	1	1	1	1	1
Petroleum Systems	+	+	+	+	+	+	+
Natural Gas Systems	+	+	+	+	+	+	+
International Bunker Fuels[b]	*3*	*3*	*4*	*4*	*3*	*2*	*2*

+ Does not exceed 0.5 kt.

[a] Emissions from Biomass and Biofuel Consumption are not included specifically in summing Energy sector totals. Net carbon fluxes from changes in biogenic carbon reservoirs are accounted for in the estimates for LULUCF.

[b] Emissions from International Bunker Fuels are not included in totals. These values are presented for informational purposes only, in line with the *2006 IPCC Guidelines* and UNFCCC reporting obligations.

Note: Totals by gas may not sum due to independent rounding.

Emissions estimates reported in the Energy chapter from fossil fuel combustion and fugitive sources include those from all 50 states, including Hawaii and Alaska, and the District of Columbia. Emissions are also included from U.S. Territories to the extent they are known to occur (e.g., coal mining does not occur in U.S. Territories). For some sources there is a lack of detailed information on U.S. Territories including some non-CO_2 emissions from biomass combustion. As part of continuous improvement efforts, EPA reviews this on an ongoing basis to ensure emission sources are included across all geographic areas including U.S. Territories if they are occurring. See Annex 5 for more information on EPA's assessment of the sources not included in this Inventory.

Each year, some emission and sink estimates in the Inventory are recalculated and revised with improved methods and/or data. In general, recalculations are made to the U.S. greenhouse gas emission estimates either to incorporate new methodologies or, most commonly, to update recent historical data. These improvements are implemented consistently across the previous Inventory's time series (i.e., 1990 to 2020) to ensure that the trend is accurate. Key updates in this year's Inventory include, updates to the transportation methodology which use distributions ofehicle miles traveled (VMT) and fuel use from EPA's MOVES3 model to estimate vehicle emissions by vehicle class, updates to the CH_4 and N_2O emission factors for alternative fuel vehicles based on the GREET2022 model, , revisions to methods for estimating CH_4 from both Natural Gas Systems and Petroleum Systems now incorporate additional basin-level data from GHGRP Subpart W for several emission sources in the onshore production segment, changes to the Non-Energy Use of Fossil Fuel methodology (e.g., updated energy consumption statistics, updated polyester fiber and acetic acid production data, updated import and export data, and updated shipment data from the U.S census Bureau), and accounting for biogenic emissions from combusted MSW within Biomass estimates. In addition, the GWPs for calculating CO_2-equivalent totals emissions of CH_4 and N_2O have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth Assessment Report* (AR4) (IPCC 2007) (used in the previous Inventories). The combined impact of these recalculations averaged 9.6 MMT CO_2 Eq. (+0.2 percent) per year across the time series. For more information on specific methodological updates, please see the Recalculations Discussion section for each category in this chapter.

Box 3-1: Methodological Approach for Estimating and Reporting U.S. Emissions and Removals, including Relationship to EPA's Greenhouse Gas Reporting Program

In following the United Nations Framework Convention on Climate Change (UNFCCC) requirement under Article 4.1 to develop and submit national greenhouse gas emission inventories, the emissions and removals presented in this report and this chapter are organized by source and sink categories and calculated using internationally-accepted methods provided by the Intergovernmental Panel on Climate Change (IPCC) in the *2006 IPCC Guidelines for National Greenhouse Gas Inventories (2006 IPCC Guidelines)*. Additionally, the calculated emissions and removals in a given year for the United States are presented in a common format in line with the UNFCCC reporting guidelines for the reporting of inventories under this international agreement. The use of consistent methods to calculate emissions and removals by all nations providing their inventories to the UNFCCC ensures that these reports are comparable. The presentation of emissions and removals provided in the Energy chapter do not preclude alternative examinations, but rather, this chapter presents emissions and removals in a common format consistent with how countries are to report Inventories under the UNFCCC. The report itself, and this chapter, follows this standardized format, and provides an explanation of the application of methods used to calculate emissions and removals from energy-related activities.

Energy Data from EPA's Greenhouse Gas Reporting Program

EPA's Greenhouse Gas Reporting Program (GHGRP)[4] dataset and the data presented in this Inventory are complementary. The Inventory was used to guide the development of the GHGRP, particularly in terms of scope and coverage of both sources and gases. The GHGRP dataset continues to be an important resource for the Inventory, providing not only annual emissions information, but also other annual information, such as activity data and emission factors that can improve and refine national emission estimates and trends over time. GHGRP data also allow EPA to disaggregate national inventory estimates in new ways that can highlight differences across regions and sub-categories of emissions, along with enhancing application of QA/QC procedures and assessment of uncertainties.

EPA uses annual GHGRP data in a number of Energy sector categories to improve the national estimates presented in this Inventory consistent with IPCC guidelines (see Box 3-3 of this chapter, and Sections 3.3 Incineration of Waste, 3.4 Coal Mining, 3.6 Petroleum Systems, and 3.7 Natural Gas Systems).[5] Methodologies used in EPA's GHGRP are consistent with IPCC guidelines, including higher tier methods. Under EPA's GHGRP, facilities collect detailed information specific to their operations according to detailed measurement standards. It should be noted that the definitions and provisions for reporting fuel types in EPA's GHGRP may differ from those used in the Inventory in meeting the UNFCCC reporting guidelines. In line with the UNFCCC reporting guidelines, the Inventory report is a comprehensive accounting of all emissions from fuel types identified in the IPCC guidelines and provides a separate reporting of emissions from biomass.

In addition to using GHGRP data to estimate emissions (Sections 3.3 Incineration of Waste, 3.4 Coal Mining, 3.6 Petroleum Systems, and 3.7 Natural Gas Systems), EPA also uses the GHGRP fuel consumption activity data in the Energy sector to disaggregate industrial end-use sector emissions in the category of CO_2 Emissions from Fossil Fuel Combustion, for use in reporting emissions in Common Reporting Format (CRF) tables (See Box 3-3). The industrial end-use sector activity data collected for the Inventory (EIA 2022) represent aggregated data for the industrial end-use sector. EPA's GHGRP collects industrial fuel consumption activity data by individual categories within the industrial end-use sector. Therefore, GHGRP data are used to provide a more detailed breakout of total emissions in the industrial end-use sector within that source category.

[4] On October 30, 2009, the U.S. Environmental Protection Agency (EPA) published a rule requiring annual reporting of greenhouse gas data from large greenhouse gas emission sources in the United States. Implementation of the rule, codified at 40 CFR Part 98, is referred to as EPA's Greenhouse Gas Reporting Program (GHGRP).

[5] See http://www.ipcc-nggip.iges.or.jp/public/tb/TFI_Technical_Bulletin_1.pdf.

As indicated in the respective Planned Improvements sections for source categories in this chapter, EPA continues to examine the uses of facility-level GHGRP data to improve the national estimates presented in this Inventory. See Annex 9 for more information on use of EPA's GHGRP in the Inventory.

3.1 Fossil Fuel Combustion (CRF Source Category 1A)

Emissions from the combustion of fossil fuels for energy include the greenhouse gases CO_2, CH_4, and N_2O. Given that CO_2 is the primary gas emitted from fossil fuel combustion and represents the largest share of U.S. total emissions, CO_2 emissions from fossil fuel combustion are discussed at the beginning of this section. An overview of CH_4 and N_2O emissions from the combustion of fuels in stationary sources is then presented, followed by fossil fuel combustion emissions for all three gases by sector: electric power, industrial, residential, commercial, U.S. Territories, and transportation.

Methodologies for estimating CO_2 emissions from fossil fuel combustion differ from the estimation of CH_4 and N_2O emissions from stationary combustion and mobile combustion. Thus, three separate descriptions of methodologies, uncertainties, recalculations, and planned improvements are provided at the end of this section. Total CO_2, CH_4, and N_2O emissions from fossil fuel combustion are presented in Table 3-3 and Table 3-4.

Table 3-3: CO_2, CH_4, and N_2O Emissions from Fossil Fuel Combustion (MMT CO_2 Eq.)

Gas	1990	2005	2017	2018	2019	2020	2021
CO_2	4,728.2	5,747.3	4,852.5	4,989.8	4,853.4	4,344.8	4,651.0
CH_4	16.8	13.2	11.5	12.5	12.7	11.3	11.5
N_2O	60.7	67.6	43.8	42.6	41.1	36.8	39.2
Total	**4,805.7**	**5,828.0**	**4,907.9**	**5,045.0**	**4,907.3**	**4,392.9**	**4,701.7**

Note: Totals may not sum due to independent rounding.

Table 3-4: CO_2, CH_4, and N_2O Emissions from Fossil Fuel Combustion (kt)

Gas	1990	2005	2017	2018	2019	2020	2021
CO_2	4,728,194	5,747,307	4,852,515	4,989,843	4,853,402	4,344,837	4,650,953
CH_4	601	471	412	446	454	405	410
N_2O	229	255	165	161	155	139	148

CO_2 from Fossil Fuel Combustion

Carbon dioxide is the primary gas emitted from fossil fuel combustion and represents the largest share of U.S. total greenhouse gas emissions. Carbon dioxide emissions from fossil fuel combustion are presented in Table 3-5. In 2021, CO_2 emissions from fossil fuel combustion increased by 7.0 percent relative to the previous year (as shown in Table 3-6). The increase in CO_2 emissions from fossil fuel consumption was a result of a 5.9 percent increase in fossil fuel energy use. This increase in fossil fuel consumption was due primarily rebounding economic activity after the COVID-19 pandemic. Carbon dioxide emissions from natural gas increased by 8.3 MMT CO_2 Eq., a 0.5 percent increase from 2020. In a shift from recent trends, CO_2 emissions from coal consumption increased by 122.1 MMT CO_2 Eq., a 14.6 percent increase from 2020. The increase in natural gas consumption and emissions in 2021 is observed across all sectors except the Electric Power sector and U.S. Territories, while the coal increase is primarily in the Electric Power sector. Emissions from petroleum use also increased 175.8 MMT CO_2 Eq. (9.3 percent) from

2020 to 2021. In 2021, CO_2 emissions from fossil fuel combustion were 4,651.0 MMT CO_2 Eq., or 1.6 percent below emissions in 1990 (see Table 3-5).[6]

Table 3-5: CO_2 Emissions from Fossil Fuel Combustion by Fuel Type and Sector (MMT CO_2 Eq.)

Fuel/Sector	1990	2005	2017	2018	2019	2020	2021
Coal	**1,719.8**	**2,113.7**	**1,270.0**	**1,211.6**	**1,028.2**	**835.6**	**957.7**
Residential	3.0	0.8	NO	NO	NO	NO	NO
Commercial	12.0	9.3	2.0	1.8	1.6	1.4	1.4
Industrial	157.8	117.8	58.7	54.4	49.5	43.0	43.7
Transportation	NO	NO	NO	NO	NO	NO	NO
Electric Power	1,546.5	1,982.8	1,207.1	1,152.9	973.5	788.2	909.7
U.S. Territories	0.5	3.0	2.3	2.6	3.6	3.1	2.9
Natural Gas	**998.6**	**1,166.2**	**1,433.2**	**1,592.0**	**1,649.3**	**1,612.4**	**1,620.7**
Residential	237.8	262.2	241.5	273.8	275.5	256.4	258.6
Commercial	142.0	162.9	173.2	192.5	192.9	173.8	180.9
Industrial	407.4	387.8	468.1	493.5	501.5	486.1	498.4
Transportation	36.0	33.1	42.3	50.9	58.9	58.7	65.1
Electric Power	175.4	318.9	505.6	577.9	616.6	634.8	615.1
U.S. Territories	NO	1.3	2.5	3.3	3.8	2.6	2.6
Petroleum	**2,009.2**	**2,467.0**	**2,148.8**	**2,185.8**	**2,175.6**	**1,896.4**	**2,072.2**
Residential	97.8	95.9	51.9	64.4	65.9	56.8	51.5
Commercial	74.3	54.9	56.8	51.5	56.2	52.8	41.4
Industrial	287.1	345.2	262.2	265.6	264.9	238.9	220.3
Transportation	1,432.9	1,825.5	1,737.8	1,762.0	1,754.9	1,513.9	1,724.3
Electric Power	97.5	98.0	18.9	22.2	16.2	16.2	17.1
U.S. Territories	19.5	47.6	21.1	20.1	17.5	17.5	17.5
Geothermal[a]	**0.5**	**0.5**	**0.4**	**0.4**	**0.4**	**0.4**	**0.4**
Electric Power	0.5	0.5	0.4	0.4	0.4	0.4	0.4
Total	**4,728.2**	**5,747.3**	**4,852.5**	**4,989.8**	**4,853.4**	**4,344.8**	**4,651.0**

NO (Not Occurring)

[a] Although not technically a fossil fuel, geothermal energy-related CO_2 emissions are included for reporting purposes. The source of CO_2 is non-condensable gases in subterranean heated water.

Note: Totals may not sum due to independent rounding.

Trends in CO_2 emissions from fossil fuel combustion are influenced by many long-term and short-term factors. On a year-to-year basis, the overall demand for fossil fuels in the United States and other countries generally fluctuates in response to changes in general economic conditions, energy prices, weather, and the availability of non-fossil alternatives. For example, in a year with increased consumption of goods and services, low fuel prices, severe summer and winter weather conditions, nuclear plant closures, and lower precipitation feeding hydroelectric dams, there would likely be proportionally greater fossil fuel consumption than a year with poor economic performance, high fuel prices, mild temperatures, and increased output from nuclear and hydroelectric plants. The 2020 to 2021 trends were particularly impacted by the COVID-19 pandemic which generally led to a reduction in demand for fossil fuels in 2020, but an increase in demand as activities rebounded in 2021.

Longer-term changes in energy usage patterns, however, tend to be more a function of aggregate societal trends that affect the scale of energy use (e.g., population, number of cars, size of houses, and number of houses), the efficiency with which energy is used in equipment (e.g., cars, HVAC systems, power plants, steel mills, and light bulbs), and social planning and consumer behavior (e.g., walking, bicycling, or telecommuting to work instead of driving).

[6] An additional discussion of fossil fuel emission trends is presented in the Trends in U.S. Greenhouse Gas Emissions chapter.

Carbon dioxide emissions also depend on the source of energy and its carbon (C) intensity. The amount of C in fuels varies significantly by fuel type. For example, coal contains the highest amount of C per unit of useful energy. Petroleum has roughly 75 percent of the C per unit of energy as coal, and natural gas has only about 55 percent.[7] Table 3-6 shows annual changes in emissions during the last five years for coal, petroleum, and natural gas in selected sectors.

Table 3-6: Annual Change in CO_2 Emissions and Total 2021 CO_2 Emissions from Fossil Fuel Combustion for Selected Fuels and Sectors (MMT CO_2 Eq. and Percent)

Sector	Fuel Type	2017 to 2018		2018 to 2019		2019 to 2020		2020 to 2021		Total 2021
Transportation	Petroleum	24.1	1.4%	-7.0	-0.4%	-241.1	-13.7%	210.4	13.9%	1,724.3
Electric Power	Coal	-54.2	-4.5%	-179.3	-15.6%	-185.4	-19.0%	121.6	15.4%	909.7
Electric Power	Natural Gas	72.3	14.3%	38.7	6.7%	18.2	3.0%	-19.8	-3.1%	634.8
Industrial	Natural Gas	25.3	5.4%	8.0	1.6%	-15.5	-3.1%	12.3	2.5%	498.4
Residential	Natural Gas	32.3	13.4%	1.7	0.6%	-19.1	-6.9%	2.3	0.9%	258.6
Commercial	Natural Gas	19.3	11.2%	0.4	0.2%	-19.1	-9.9%	7.0	4.0%	180.9
Transportation	**All Fuels[a]**	**32.8**	**1.8%**	**1.0**	**0.1%**	**-241.3**	**-13.3%**	**216.9**	**13.8%**	**1,789.4**
Electric Power	**All Fuels[a]**	**21.4**	**1.2%**	**-146.7**	**-8.4%**	**-167.2**	**-10.4%**	**102.6**	**7.1%**	**1,542.2**
Industrial	**All Fuels[a]**	**24.5**	**3.1%**	**2.4**	**0.3%**	**-48.0**	**-5.9%**	**-5.5**	**-0.7%**	**762.4**
Residential	**All Fuels[a]**	**44.8**	**15.3%**	**3.2**	**0.9%**	**-28.2**	**-8.3%**	**-3.1**	**-1.0%**	**310.1**
Commercial	**All Fuels[a]**	**13.8**	**6.0%**	**4.9**	**2.0%**	**-22.2**	**-8.9%**	**-4.6**	**-2.0%**	**223.9**
All Sectors[a,b]	**All Fuels[a]**	**137.3**	**2.8%**	**-136.4**	**-2.7%**	**-508.6**	**-10.5%**	**306.1**	**7.0%**	**4,651.0**

[a] Includes sector and fuel combinations not shown in this table.

[b] Includes U.S. Territories.

As shown in Table 3-6, recent trends in CO_2 emissions from fossil fuel combustion show a 2.8 percent increase from 2017 to 2018, a 2.7 percent decrease from 2018 to 2019, a 10.5 percent decrease from 2019 to 2020, and a 7.0 percent increase from 2020 to 2021. These changes contributed to an overall 4.2 percent decrease in CO_2 emissions from fossil fuel combustion from 2017 to 2021.

Recent trends in CO_2 emissions from fossil fuel combustion are largely driven by the electric power sector, which until 2017 has accounted for the largest portion of these emissions. The types of fuels consumed to produce electricity have changed in recent years. Electric power sector consumption of natural gas primarily increased due to increased production capacity as natural gas-fired plants replaced coal-fired plants and increased electricity demand related to heating and cooling needs (EIA 2018; EIA 2022e). Total net electric power generation from all fossil and non-fossil sources increased by 3.6 percent from 2017 to 2018, decreased by 1.3 percent from 2018 to 2019, decreased by 2.9 percent from 2019 to 2020, and increased by 2.8 percent from 2020 to 2021 (EIA 2022a). Carbon dioxide emissions from the electric power sector increased from 2020 to 2021 by 7.1 percent due to increased production and the increased use of coal for electric power generation. Carbon dioxide emissions from coal consumption for electric power generation decreased by 24.6 percent overall since 2017, but increased by 15.4 percent from 2020 to 2021.

The recent trends in CO_2 emissions from fossil fuel combustion also follow changes in heating degree days (see Box 3-2). Emissions from natural gas consumption in the residential and commercial sectors increased by 7.1 percent and 4.4 percent from 2017 to 2021, respectively. This trend can be partially attributed to a 2.6 percent increase in heating degree days from 2017 to 2021, which led to an increased demand for heating fuel and electricity for heat in these sectors. Industrial consumption of natural gas is dependent on market effects of supply and demand in addition to weather-related heating needs.

Petroleum use in the transportation sector is another major driver of emissions, representing the largest source of CO_2 emissions from fossil fuel combustion in 2021. Emissions from petroleum consumption for transportation have decreased by 0.8 percent since 2017 and are primarily attributed to a 0.5 percent decrease in VMT over the same

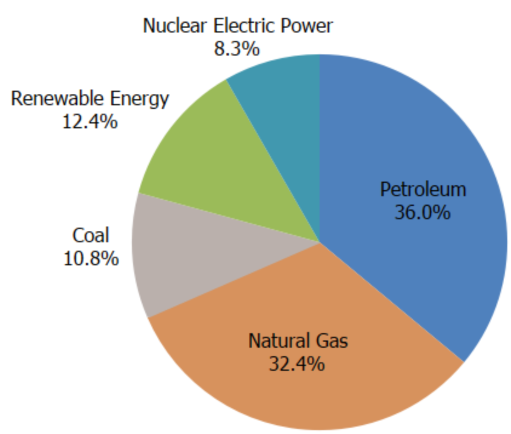
[7] Based on national aggregate carbon content of all coal, natural gas, and petroleum fuels combusted in the United States. See Annex 2.2 for more details on fuel carbon contents.

time period. Beginning with 2017, the transportation sector is the largest source of national CO_2 emissions–whereas in prior years, electric power was the largest source sector.

The overall 2020 to 2021 trends were largely driven by the gradual recovery from the COVID-19 pandemic, which saw reduced economic activity in 2020 and caused changes in energy demand and supply patterns across different sectors. The recovery from the COVID-19 pandemic generally led to increased energy use and emissions across all economic sectors from 2020 to 2021. The increase in emissions from 2020 to 2021 was also due to a reversal in recent trends in coal use. In recent years the trend has been one of decreased coal use however, from 2020 to 2021 overall use of coal increased by 14.6 percent (EIA 2022a).

In the United States, 79.3 percent of the energy used in 2021 was produced through the combustion of fossil fuels such as petroleum, natural gas, and coal (see Figure 3-4 and Figure 3-5). Specifically, petroleum supplied the largest share of domestic energy demands, accounting for 36 percent of total U.S. energy used in 2021. Natural gas and coal followed in order of fossil fuel energy demand importance, accounting for approximately 32 percent and 11 percent of total U.S. energy used, respectively. Petroleum was consumed primarily in the transportation end-use sector and the majority of coal was used in the electric power sector. Natural gas was broadly consumed in all end-use sectors except transportation (see Figure 3-6) (EIA 2021c). The remaining portion of energy used in 2021 was supplied by nuclear electric power (8 percent) and by a variety of renewable energy sources (12 percent), primarily wind energy, hydroelectric power, solar, geothermal and biomass (EIA 2021c).[8]

Figure 3-4: 2021 U.S. Energy Use by Energy Source



[8] Renewable energy, as defined in EIA's energy statistics, includes the following energy sources: hydroelectric power, geothermal energy, biomass, solar energy, and wind energy.

Figure 3-5: Annual U.S. Energy Use

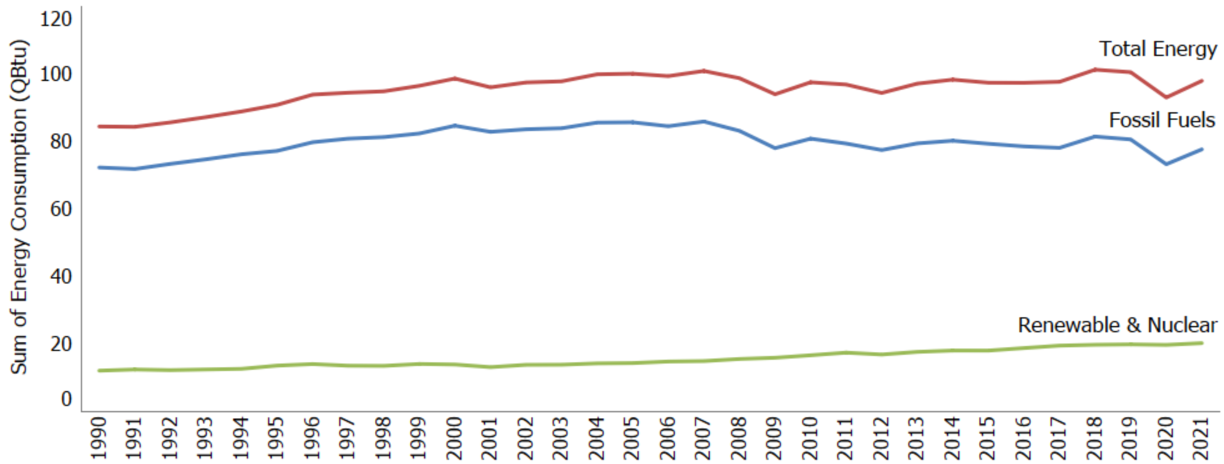
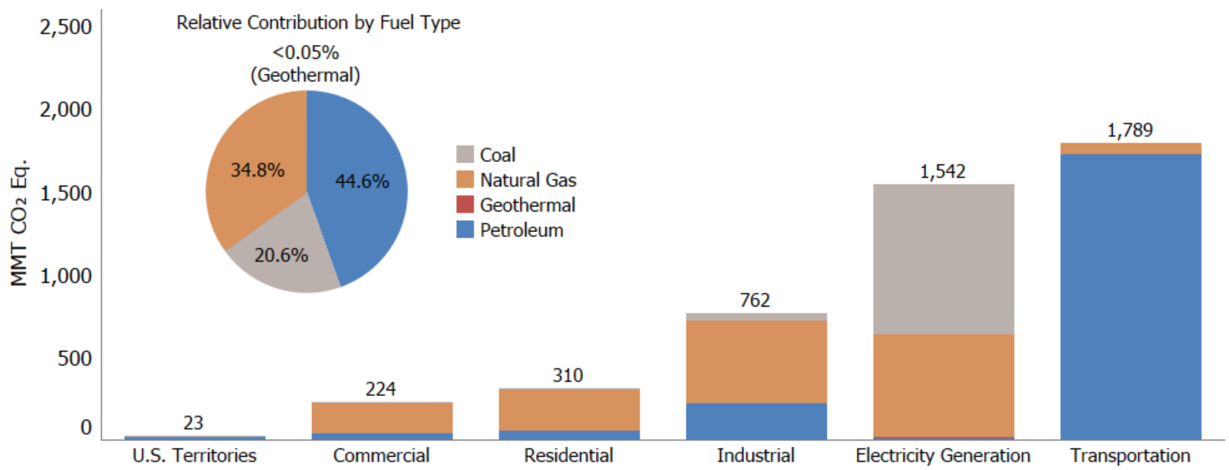


Figure 3-6: 2021 CO_2 Emissions from Fossil Fuel Combustion by Sector and Fuel Type



Fossil fuels are generally combusted for the purpose of producing energy for useful heat and work. During the combustion process, the C stored in the fuels is oxidized and emitted as CO_2 and smaller amounts of other gases, including CH_4, carbon monoxide (CO), and non-methane volatile organic compounds (NMVOCs).[9] These other C-containing non-CO_2 gases are emitted as a byproduct of incomplete fuel combustion, but are, for the most part, eventually oxidized to CO_2 in the atmosphere. Therefore, as per IPCC guidelines it is assumed all of the C in fossil fuels used to produce energy is eventually converted to atmospheric CO_2.

Box 3-2: Weather and Non-Fossil Energy Effects on CO_2 Emissions from Fossil Fuel Combustion Trends

The United States in 2021 experienced a colder winter overall compared to 2020, as heating degree days increased 0.5 percent. Colder winter conditions compared to 2020 impacted the amount of energy required for heating. In 2021 heating degree days in the United States were 9.3 percent below normal (see Figure 3-7). Cooling degree days decreased by 1.9 percent compared to 2020, which decreased demand for air conditioning in the residential and commercial sector. Cooler summer conditions compared to 2020 impacted the amount of

[9] See the sections entitled Stationary Combustion and Mobile Combustion in this chapter for information on non-CO_2 gas emissions from fossil fuel combustion.

energy required for cooling. 2020 cooling degree days in the United States were 11.8 percent above normal (see Figure 3-8) (EIA 2022a).[10] The combination of colder winter and summer conditions led to overall residential and commercial energy consumption decrease of 1.0 and 2.0 percent, respectively relative to 2020.

Figure 3-7: Annual Deviations from Normal Heating Degree Days for the United States (1950–2021, Index Normal = 100)

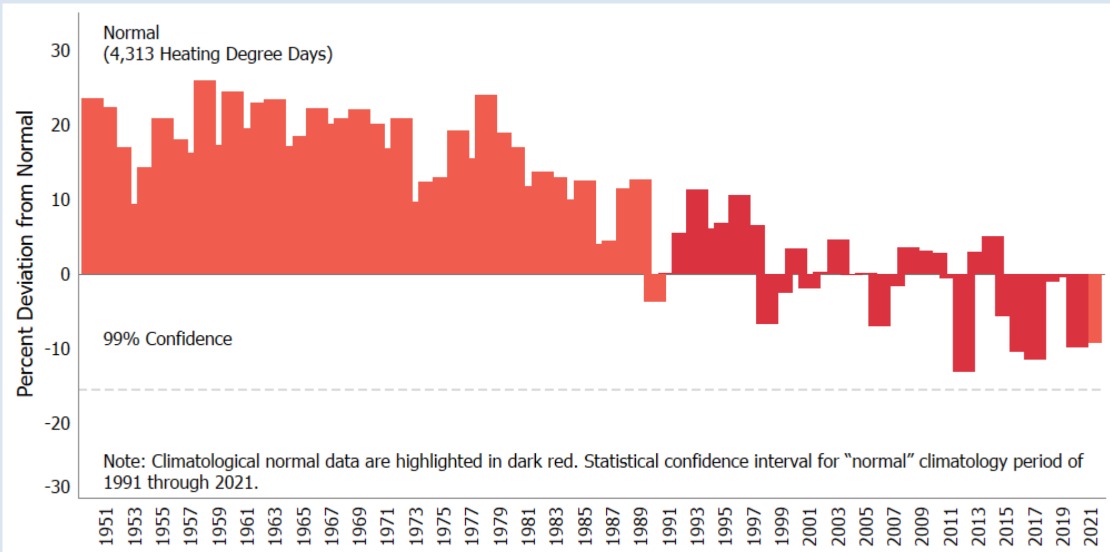
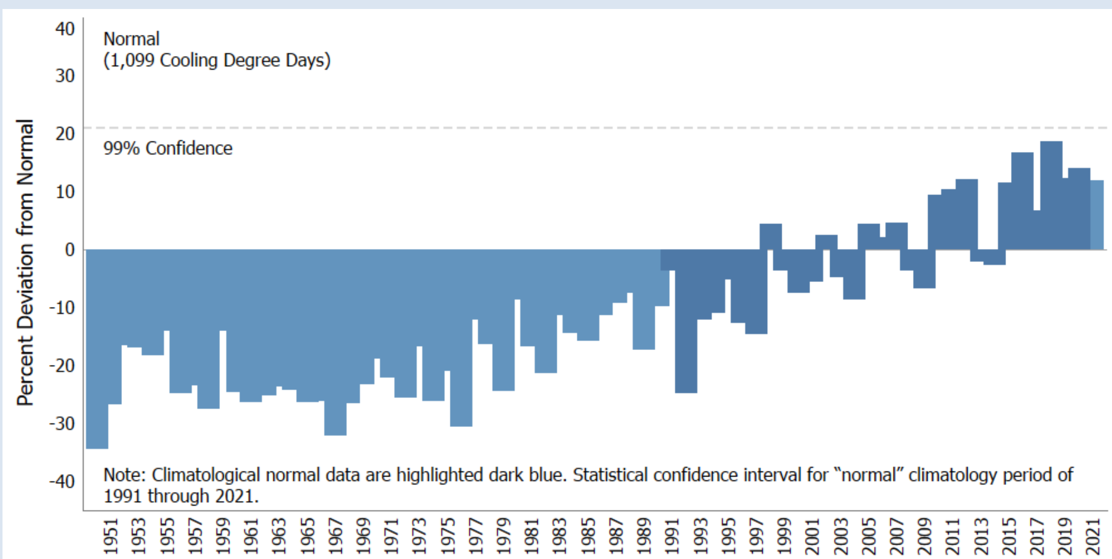


Figure 3-8: Annual Deviations from Normal Cooling Degree Days for the United States (1950–2021, Index Normal = 100)



10 Degree days are relative measurements of outdoor air temperature. Heating degree days are deviations of the mean daily temperature below 65 degrees Fahrenheit, while cooling degree days are deviations of the mean daily temperature above 65 degrees Fahrenheit. Heating degree days have a considerably greater effect on energy demand and related emissions than do cooling degree days. Excludes Alaska and Hawaii. Normals are based on data from 1991 through 2020. The variation in these normals during this time period was ±16 percent and ±27 percent for heating and cooling degree days, respectively (99 percent confidence interval).

The carbon intensity of the electric power sector is impacted by the amount of non-fossil energy sources of electricity. The utilization (i.e., capacity factors)[11] of nuclear power plants in 2021 remained high at 93 percent. In 2021, nuclear power represented 20 percent of total electricity generation. Since 1990, the wind and solar power sectors have shown strong growth and have become relatively important electricity sources. Between 1990 and 2021, renewable energy generation (in kWh) from solar and wind energy have increased from 0.1 percent in 1990 to 12 percent in 2021 of total electricity generation, which helped drive the decrease in the carbon intensity of the electricity supply in the United States.

Stationary Combustion

The direct combustion of fuels by stationary sources in the electric power, industrial, commercial, and residential sectors represent the greatest share of U.S. greenhouse gas emissions. Table 3-7 presents CO_2 emissions from fossil fuel combustion by stationary sources. The CO_2 emitted is closely linked to the type of fuel being combusted in each sector (see Methodology section of CO_2 from Fossil Fuel Combustion). In addition to the CO_2 emitted from fossil fuel combustion, CH_4 and N_2O are emitted as well. Table 3-8 and Table 3-9 present CH_4 and N_2O emissions from the combustion of fuels in stationary sources. The CH_4 and N_2O emissions are linked to the type of fuel being combusted as well as the combustion technology (see Methodology section for CH_4 and N_2O from Stationary Combustion).

Table 3-7: CO_2 Emissions from Stationary Fossil Fuel Combustion (MMT CO_2 Eq.)

Sector/Fuel Type	1990	2005	2017	2018	2019	2020	2021
Electric Power	**1,820.0**	**2,400.1**	**1,732.0**	**1,753.4**	**1,606.7**	**1,439.6**	**1,542.2**
Coal	1,546.5	1,982.8	1,207.1	1,152.9	973.5	788.2	909.7
Natural Gas	175.4	318.9	505.6	577.9	616.6	634.8	615.1
Fuel Oil	97.5	98.0	18.9	22.2	16.2	16.2	17.1
Geothermal	0.5	0.5	0.4	0.4	0.4	0.4	0.4
Industrial	**852.4**	**850.8**	**789.0**	**813.5**	**815.9**	**767.9**	**762.4**
Coal	157.8	117.8	58.7	54.4	49.5	43.0	43.7
Natural Gas	407.4	387.8	468.1	493.5	501.5	486.1	498.4
Fuel Oil	287.1	345.2	262.2	265.6	264.9	238.6	220.2
Commercial	**228.3**	**227.1**	**232.0**	**245.8**	**250.7**	**228.5**	**223.9**
Coal	12.0	9.3	2.0	1.8	1.6	1.4	1.4
Natural Gas	142.0	162.9	173.2	192.5	192.9	173.8	180.9
Fuel Oil	74.3	54.9	56.8	51.5	56.2	52.8	41.4
Residential	**338.6**	**358.9**	**293.4**	**338.2**	**341.4**	**313.2**	**310.1**
Coal	3.0	0.8	NO	NO	NO	NO	NO
Natural Gas	237.8	262.2	241.5	273.8	275.5	256.4	258.6
Fuel Oil	97.8	95.9	51.9	64.4	65.9	56.8	51.5
U.S. Territories	**20.0**	**51.9**	**25.9**	**25.9**	**24.8**	**23.2**	**23.0**
Coal	0.5	3.0	2.3	2.6	3.6	3.1	2.9
Natural Gas	NO	1.3	2.5	3.3	3.8	2.6	2.6
Fuel Oil	19.5	47.6	21.1	20.1	17.5	17.5	17.5
Total	**3,259.3**	**3,888.8**	**3,072.4**	**3,176.9**	**3,039.5**	**2,772.3**	**2,861.6**

NO (Not Occurring)

Note: Totals may not sum due to independent rounding.

[11] The capacity factor equals generation divided by net summer capacity. Summer capacity is defined as "The maximum output that generating equipment can supply to system load, as demonstrated by a multi-hour test, at the time of summer peak demand (period of June 1 through September 30)" (EIA 2020a). Data for both the generation and net summer capacity are from EIA (2022a).

1 **Table 3-8: CH_4 Emissions from Stationary Combustion (MMT CO_2 Eq.)**

Sector/Fuel Type	1990	2005	2017	2018	2019	2020	2021
Electric Power	**0.5**	**1.0**	**1.2**	**1.4**	**1.4**	**1.4**	**1.4**
Coal	0.4	0.4	0.3	0.3	0.2	0.2	0.2
Fuel Oil	+	+	+	+	+	+	+
Natural gas	0.1	0.5	1.0	1.1	1.2	1.2	1.2
Wood	+	+	+	+	+	+	+
Industrial	**2.0**	**1.9**	**1.7**	**1.7**	**1.7**	**1.6**	**1.6**
Coal	0.5	0.3	0.2	0.2	0.1	0.1	0.1
Fuel Oil	0.2	0.2	0.2	0.2	0.2	0.2	0.1
Natural gas	0.2	0.2	0.2	0.2	0.3	0.2	0.3
Wood	1.2	1.2	1.2	1.1	1.1	1.1	1.1
Commercial	**1.2**	**1.2**	**1.3**	**1.4**	**1.4**	**1.3**	**1.3**
Coal	+	+	+	+	+	+	+
Fuel Oil	0.3	0.2	0.2	0.2	0.2	0.2	0.2
Natural gas	0.4	0.4	0.4	0.5	0.5	0.4	0.5
Wood	0.5	0.6	0.7	0.7	0.7	0.7	0.7
Residential	**5.9**	**4.5**	**4.2**	**5.1**	**5.3**	**4.4**	**4.6**
Coal	0.3	0.1	NO	NO	NO	NO	NO
Fuel Oil	0.4	0.4	0.2	0.3	0.3	0.2	0.2
Natural Gas	0.6	0.7	0.6	0.7	0.7	0.6	0.6
Wood	4.6	3.4	3.4	4.2	4.4	3.5	3.7
U.S. Territories	**+**	**0.1**	**+**	**+**	**+**	**+**	**+**
Coal	+	+	+	+	+	+	+
Fuel Oil	+	0.1	+	+	+	+	+
Natural Gas	0.0	+	+	+	+	+	+
Wood	NE	NE	NE	NE	NE	NE	NE
Total	**9.6**	**8.8**	**8.6**	**9.6**	**9.8**	**8.8**	**8.9**

+ Does not exceed 0.05 MMT CO_2 Eq.
NO (Not Occurring)
NE (Not Estimated)
Note: Totals may not sum due to independent rounding.

2 **Table 3-9: N_2O Emissions from Stationary Combustion (MMT CO_2 Eq.)**

Sector/Fuel Type	1990	2005	2017	2018	2019	2020	2021
Electric Power	**18.2**	**26.7**	**22.0**	**21.7**	**18.8**	**17.5**	**19.0**
Coal	17.9	24.9	18.8	18.1	14.8	13.5	15.1
Fuel Oil	0.1	0.1	+	+	+	+	+
Natural Gas	0.3	1.7	3.2	3.6	3.9	4.0	3.9
Wood	+	+	+	+	+	+	+
Industrial	**2.7**	**2.6**	**2.3**	**2.2**	**2.2**	**2.1**	**2.0**
Coal	0.7	0.5	0.2	0.2	0.2	0.2	0.2
Fuel Oil	0.4	0.5	0.3	0.3	0.3	0.3	0.2
Natural Gas	0.2	0.2	0.2	0.2	0.2	0.2	0.2
Wood	1.5	1.5	1.5	1.4	1.4	1.4	1.4
Commercial	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**	**0.3**
Coal	+	+	+	+	+	+	+
Fuel Oil	0.2	0.1	0.1	0.1	0.1	0.1	0.1
Natural Gas	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Wood	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Residential	**0.9**	**0.8**	**0.7**	**0.8**	**0.8**	**0.7**	**0.7**
Coal	+	+	NO	NO	NO	NO	NO
Fuel Oil	0.2	0.2	0.1	0.2	0.2	0.1	0.1

Natural Gas	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Wood	0.6	0.4	0.4	0.5	0.5	0.4	0.5
U.S. Territories	**+**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**
Coal	+	+	+	+	+	+	+
Fuel Oil	+	0.1	+	+	+	+	+
Natural Gas	0.0	+	+	+	+	+	+
Wood	NE	NE	NE	NE	NE	NE	NE
Total	**22.3**	**30.5**	**25.3**	**25.1**	**22.2**	**20.7**	**22.1**

+ Does not exceed 0.05 MMT CO_2 Eq.
NO (Not Occurring)
NE (Not Estimated)
Note: Totals may not sum due to independent rounding.

Fossil Fuel Combustion Emissions by Sector

Table 3-10 provides an overview of the CO_2, CH_4, and N_2O emissions from fossil fuel combustion by sector, including transportation, electric power, industrial, residential, commercial, and U.S. Territories.

Table 3-10: CO_2, CH_4, and N_2O Emissions from Fossil Fuel Combustion by Sector (MMT CO_2 Eq.)

End-Use Sector	1990	2005	2017	2018	2019	2020	2021
Transportation	**1,514.6**	**1,900.0**	**1,801.6**	**1,833.3**	**1,835.7**	**1,591.2**	**1,809.2**
CO_2	1,468.9	1,858.6	1,780.1	1,812.9	1,813.9	1,572.5	1,789.4
CH_4	7.2	4.4	2.9	2.9	2.9	2.6	2.6
N_2O	38.4	37.0	18.5	17.5	19.0	16.1	17.1
Electric Power	**1,838.7**	**2,427.8**	**1,755.3**	**1,776.5**	**1,626.9**	**1,458.5**	**1,562.6**
CO_2	1,820.0	2,400.1	1,732.0	1,753.4	1,606.7	1,439.6	1,542.2
CH_4	0.5	1.0	1.2	1.4	1.4	1.4	1.4
N_2O	18.2	26.7	22.0	21.7	18.8	17.5	19.0
Industrial	**857.2**	**855.4**	**793.0**	**817.5**	**819.8**	**771.6**	**765.9**
CO_2	852.4	850.8	789.0	813.5	815.9	767.9	762.4
CH_4	2.0	1.9	1.7	1.7	1.7	1.6	1.6
N_2O	2.7	2.6	2.3	2.2	2.2	2.1	2.0
Residential	**345.4**	**364.2**	**298.3**	**344.2**	**347.6**	**318.3**	**315.4**
CO_2	338.6	358.9	293.4	338.2	341.4	313.2	310.1
CH_4	5.9	4.5	4.2	5.1	5.3	4.4	4.6
N_2O	0.9	0.8	0.7	0.8	0.8	0.7	0.7
Commercial	**229.8**	**228.6**	**233.6**	**247.5**	**252.4**	**230.1**	**225.4**
CO_2	228.3	227.1	232.0	245.8	250.7	228.5	223.9
CH_4	1.2	1.2	1.3	1.4	1.4	1.3	1.3
N_2O	0.3	0.3	0.3	0.3	0.3	0.3	0.3
U.S. Territories[a]	**20.1**	**52.1**	**26.0**	**26.0**	**24.9**	**23.3**	**23.1**
Total	**4,805.7**	**5,828.0**	**4,907.9**	**5,045.0**	**4,907.3**	**4,392.9**	**4,701.7**

[a] U.S. Territories are not apportioned by sector, and emissions shown in the table are total greenhouse gas emissions from all fuel combustion sources.
Note: Totals may not sum due to independent rounding.

Other than greenhouse gases CO_2, CH_4, and N_2O, gases emitted from stationary combustion include the greenhouse gas precursors nitrogen oxides (NO_x), CO, NMVOCs, and SO_2. Methane and N_2O emissions from stationary combustion sources depend upon fuel characteristics, size and vintage of combustion device, along with combustion technology, pollution control equipment, ambient environmental conditions, and operation and maintenance practices. Nitrous oxide emissions from stationary combustion are closely related to air-fuel mixes and combustion temperatures, as well as the characteristics of any pollution control equipment that is employed.

Methane emissions from stationary combustion are primarily a function of the CH_4 content of the fuel and combustion efficiency.

Mobile combustion also produces emissions of CH_4, N_2O, and greenhouse gas precursors including NO_x, CO, and NMVOCs. As with stationary combustion, N_2O and NO_x emissions from mobile combustion are closely related to fuel characteristics, air-fuel mixes, combustion temperatures, and the use of pollution control equipment. Nitrous oxide from mobile sources, in particular, can be formed by the catalytic processes used to control NO_x, CO, and hydrocarbon emissions. Carbon monoxide emissions from mobile combustion are significantly affected by combustion efficiency and the presence of post-combustion emission controls. Carbon monoxide emissions are highest when air-fuel mixtures have less oxygen than required for complete combustion. These emissions occur especially in vehicle idle, low speed, and cold start conditions. Methane and NMVOC emissions from motor vehicles are a function of the CH_4 content of the motor fuel, the amount of hydrocarbons passing uncombusted through the engine, and any post-combustion control of hydrocarbon emissions (such as catalytic converters).

An alternative method of presenting combustion emissions is to allocate emissions associated with electric power to the sectors in which it is used. Four end-use sectors are defined: transportation, industrial, residential, and commercial. In Table 3-11 below, electric power emissions have been distributed to each end-use sector based upon the sector's share of national electricity use, with the exception of CH_4 and N_2O from transportation electricity use.[12] Emissions from U.S. Territories are also calculated separately due to a lack of end-use-specific consumption data.[13] This method assumes that emissions from combustion sources are distributed across the four end-use sectors based on the ratio of electricity use in that sector. The results of this alternative method are presented in Table 3-11.

Table 3-11: CO_2, CH_4, and N_2O Emissions from Fossil Fuel Combustion by End-Use Sector with Electricity Emissions Distributed (MMT CO_2 Eq.)

End-Use Sector	1990		2005		2017	2018	2019	2020	2021
Transportation	**1,517.6**		**1,904.7**		**1,805.9**	**1,838.1**	**1,840.6**	**1,595.3**	**1,814.2**
CO_2	1,472.0		1,863.3		1,784.4	1,817.7	1,818.7	1,576.6	1,794.5
CH_4	7.2		4.4		2.9	2.9	2.9	2.6	2.6
N_2O	38.4		37.0		18.5	17.5	19.0	16.1	17.1
Industrial	**1,550.7**		**1,600.2**		**1,304.2**	**1,325.5**	**1,291.1**	**1,186.8**	**1,212.3**
CO_2	1,538.8		1,587.1		1,293.4	1,314.9	1,281.4	1,177.7	1,202.8
CH_4	2.2		2.2		2.1	2.1	2.1	2.0	2.0
N_2O	9.6		10.8		8.7	8.5	7.6	7.1	7.5
Residential	**944.2**		**1,230.1**		**923.8**	**994.9**	**938.6**	**870.8**	**900.3**
CO_2	931.3		1,214.9		910.5	980.5	925.1	858.5	887.3
CH_4	6.0		4.9		4.7	5.6	5.8	4.9	5.1
N_2O	6.9		10.3		8.5	8.8	7.7	7.4	7.9
Commercial	**773.1**		**1,040.9**		**848.0**	**860.5**	**812.0**	**716.8**	**751.8**
CO_2	766.0		1,030.1		838.2	850.9	803.4	708.8	743.3
CH_4	1.3		1.5		1.8	1.8	1.9	1.8	1.8
N_2O	5.7		9.3		8.0	7.8	6.8	6.2	6.7
U.S. Territories[a]	**20.1**		**52.1**		**26.0**	**26.0**	**24.9**	**23.3**	**23.1**
Total	**4,805.7**		**5,828.0**		**4,907.9**	**5,045.0**	**4,907.3**	**4,392.9**	**4,701.7**

[a] U.S. Territories are not apportioned by sector, and emissions are total greenhouse gas emissions from all fuel combustion sources.

Notes: Totals may not sum due to independent rounding. Emissions from fossil fuel combustion by electric power are allocated based on aggregate national electricity use by each end-use sector.

[12] Separate calculations are performed for transportation-related CH_4 and N_2O. The methodology used to calculate these emissions is discussed in the Mobile Combustion section.

[13] U.S. Territories (including American Samoa, Guam, Puerto Rico, U.S. Virgin Islands, Wake Island, and other outlying U.S. Pacific Islands) consumption data obtained from EIA are only available at the aggregate level and cannot be broken out by end-use sector. The distribution of emissions to each end-use sector for the 50 states does not apply to territories data.

Electric Power Sector

The process of generating electricity is the largest stationary source of CO_2 emissions in the United States, representing 28.5 percent of total CO_2 emissions from all CO_2 emissions sources across the United States. Methane and N_2O accounted for a small portion of total greenhouse gas emissions from electric power, representing 0.1 percent and 1.2 percent, respectively. Electric power also accounted for 33.2 percent of CO_2 emissions from fossil fuel combustion in 2021. Methane and N_2O from electric power represented 12.2 and 48.6 percent of total CH_4 and N_2O emissions from fossil fuel combustion in 2021, respectively.

For the underlying energy data used in this chapter, the Energy Information Administration (EIA) places electric power generation into three functional categories: the electric power sector, the commercial sector, and the industrial sector. The energy use and emissions associated with the electric power sector are included here. The electric power sector consists of electric utilities and independent power producers whose primary business is the production of electricity. This includes both regulated utilities and non-utilities (e.g., independent power producers, qualifying co-generators, and other small power producers). Energy use and emissions associated with electric generation in the commercial and industrial sectors is reported in those other sectors where the producer of the power indicates that its primary business is something other than the production of electricity.[14]

Total greenhouse gas emissions from the electric power sector have decreased by 15.0 percent since 1990. From 1990 to 2007, electric power sector emissions increased by 33 percent, driven by a significant increase in electricity demand (39 percent) while the carbon intensity of electricity generated showed a modest decline (3.2 percent). From 2008 to 2021, as electricity demand increased by 1.6 percent, electric power sector emissions decreased by 35 percent, driven by a significant drop (22 percent) in the carbon intensity of electricity generated. Overall, the carbon intensity of the electric power sector, in terms of CO_2 Eq. per QBtu, decreased by 25 percent from 1990 to 2020 with additional trends detailed in Box 3-4. This decoupling of electric power generation and the resulting CO_2 emissions is shown in Figure 3-9. This recent decarbonization of the electric power sector is a result of several key drivers.

Coal-fired electric generation (in kilowatt-hours [kWh]) decreased from 54 percent of generation in 1990 to 23 percent in 2021.[15] This corresponded with an increase in natural gas generation and renewable energy generation, largely from wind and solar energy. Natural gas generation (in kWh) represented 11 percent of electric power generation in 1990 and increased over the 32-year period to represent 37 percent of electric power sector generation in 2021 (see Table 3-12). Natural gas has a much lower carbon content than coal and is generated in power plants that are generally more efficient in terms of kWh produced per Btu of fuel combusted, which has led to lower emissions as natural gas replaces coal-powered electricity generation. Natural gas and coal used in the United States in 2021 had an average carbon content of 14.43 MMT C/QBtu and 26.13 MMT C/QBtu respectively.

Table 3-12: Electric Power Generation by Fuel Type (Percent)

Fuel Type	1990		2005		2017	2018	2019	2020	2021
Coal	54.1%		51.1%		30.9%	28.4%	24.2%	19.9%	22.5%
Natural Gas	10.7%		17.5%		30.9%	34.0%	37.3%	39.5%	37.2%
Nuclear	19.9%		20.0%		20.8%	20.1%	20.4%	20.5%	19.6%
Renewables	11.3%		8.3%		16.8%	16.8%	17.6%	19.5%	20.1%
Petroleum	4.1%		3.0%		0.5%	0.6%	0.4%	0.4%	0.4%
Other Gases[a]	0.0%		0.1%		0.1%	0.1%	0.1%	0.1%	0.1%
Net Electricity Generation (Billion kWh)[b]	2,905		3,902		3,878	4,020	3,966	3,851	3,961

+ Does not exceed 0.05 percent.

[14] Utilities primarily generate power for the U.S. electric grid for sale to retail customers. Non-utilities typically generate electricity for sale on the wholesale electricity market (e.g., to utilities for distribution and resale to retail customers). Where electricity generation occurs outside the EIA-defined electric power sector, it is typically for the entity's own use.

[15] Values represent electricity *net* generation from the electric power sector (EIA 2022a).

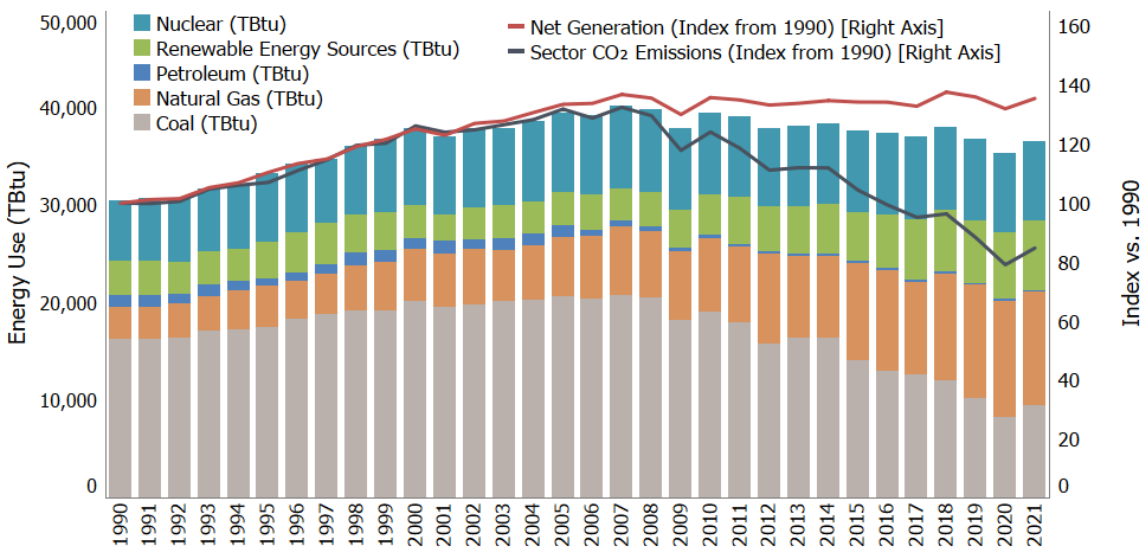
[a] Other gases include blast furnace gas, propane gas, and other manufactured and waste gases derived from fossil fuels.

[b] Represents net electricity generation from the electric power sector. Excludes net electricity generation from commercial and industrial combined-heat-and-power and electricity-only plants. Does not include electricity generation from purchased steam as the fuel used to generate the steam cannot be determined.

In 2021, CO_2 emissions from the electric power sector increased by 7.1 percent relative to 2020. This increase in CO_2 emissions was primarily driven by an increase in coal consumed to produce electricity in the electric power sector. Consumption of coal for electric power increased by 15.4 percent while consumption of natural gas decreased 3.1 percent from 2020 to 2021, leading to an overall increase in emissions. There has also been a rapid increase in renewable energy electricity generation in the electric power sector in recent years. Electricity generation from renewable sources increased by 6 percent from 2020 to 2021 (see Table 3-12). A decrease in coal-powered electricity generation and increase in natural gas and renewable energy electricity generation contributed to a decoupling of emissions trends from electric power generation trends over the recent time series (see Figure 3-9).

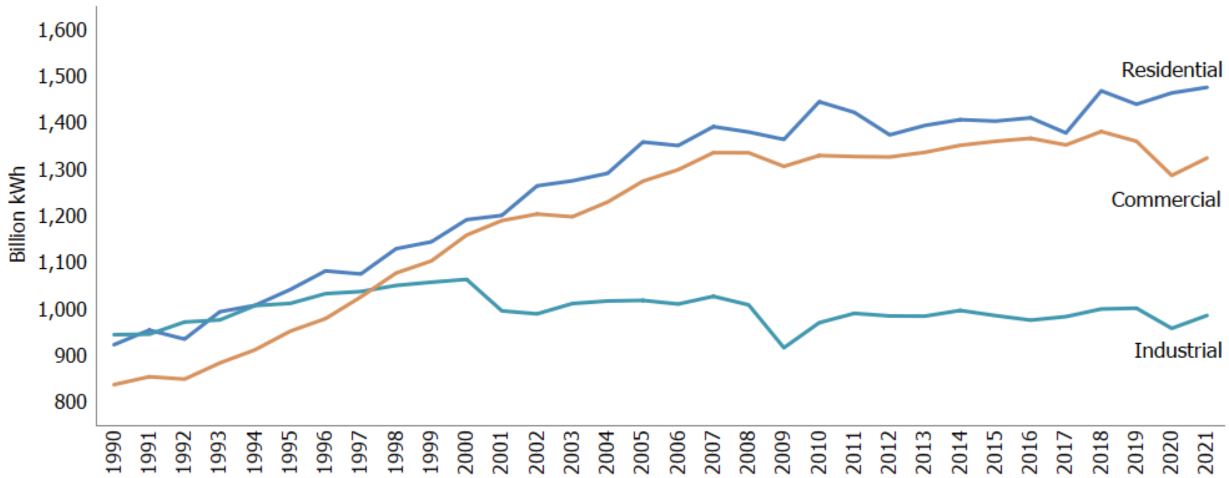
Decreases in natural gas prices and the associated increase in natural gas generation, particularly between 2005 and 2021, was one of the main drivers of the recent fuel switching and decrease in electric power sector carbon intensity. During this time period, the cost of natural gas (in $/MMBtu) decreased by 25 percent while the cost of coal (in $/MMBtu) increased by 71 percent (EIA 2021c). Also, between 1990 and 2021, renewable energy generation (in kWh) from wind and solar energy increased from 0.1 percent of total generation in 1990 to 12 percent in 2021, which also helped drive the decrease in electric power sector carbon intensity. This decrease in carbon intensity occurred even as total electricity retail sales increased 40 percent, from 2,713 billion kWh in 1990 to 3,795 billion kWh in 2021.

Figure 3-9: Fuels Used in Electric Power Generation and Total Electric Power Sector CO_2 Emissions



Electricity was used primarily in the residential, commercial, and industrial end-use sectors for lighting, heating, electric motors, appliances, electronics, and air conditioning (see Figure 3-10). Note that transportation is an end-use sector as well but is not shown in Figure 3-10 due to the sector's relatively low percentage of electricity use. Table 3-13 provides a break-out of CO_2 emissions from electricity use in the transportation end-use sector.

Figure 3-10: Electric Power Retail Sales by End-Use Sector



In 2021, electricity sales to the residential and commercial end-use sectors, as presented in Figure 3-10, increased by 0.8 percent and 2.9 percent relative to 2020, respectively. Electricity sales to the industrial sector in 2021 increased by approximately 2.9 percent relative to 2020. The sections below describe end-use sector energy use in more detail. Overall, in 2021, the amount of electricity retail sales (in kWh) increased by 2.1 percent relative to 2020.

Industrial Sector

Industrial sector CO_2, CH_4, and N_2O emissions accounted for 16, 14, and 5 percent of CO_2, CH_4, and N_2O emissions from fossil fuel combustion, respectively in 2021. Carbon dioxide, CH_4, and N_2O emissions resulted from the direct consumption of fossil fuels for steam and process heat production.

The industrial end-use sector, per the underlying energy use data from EIA, includes activities such as manufacturing, construction, mining, and agriculture. The largest of these activities in terms of energy use is manufacturing, of which six industries—Petroleum Refineries, Chemicals, Paper, Primary Metals, Food, and Nonmetallic Mineral Products—represent the majority of the energy use (EIA 2021c; EIA 2009b).

There are many dynamics that impact emissions from the industrial sector including economic activity, changes in the make-up of the industrial sector, changes in the emissions intensity of industrial processes, and weather-related impacts on heating and cooling of industrial buildings.[16] Structural changes within the U.S. economy that lead to shifts in industrial output away from energy-intensive manufacturing products to less energy-intensive products (e.g., from steel to computer equipment) have had a significant effect on industrial emissions.

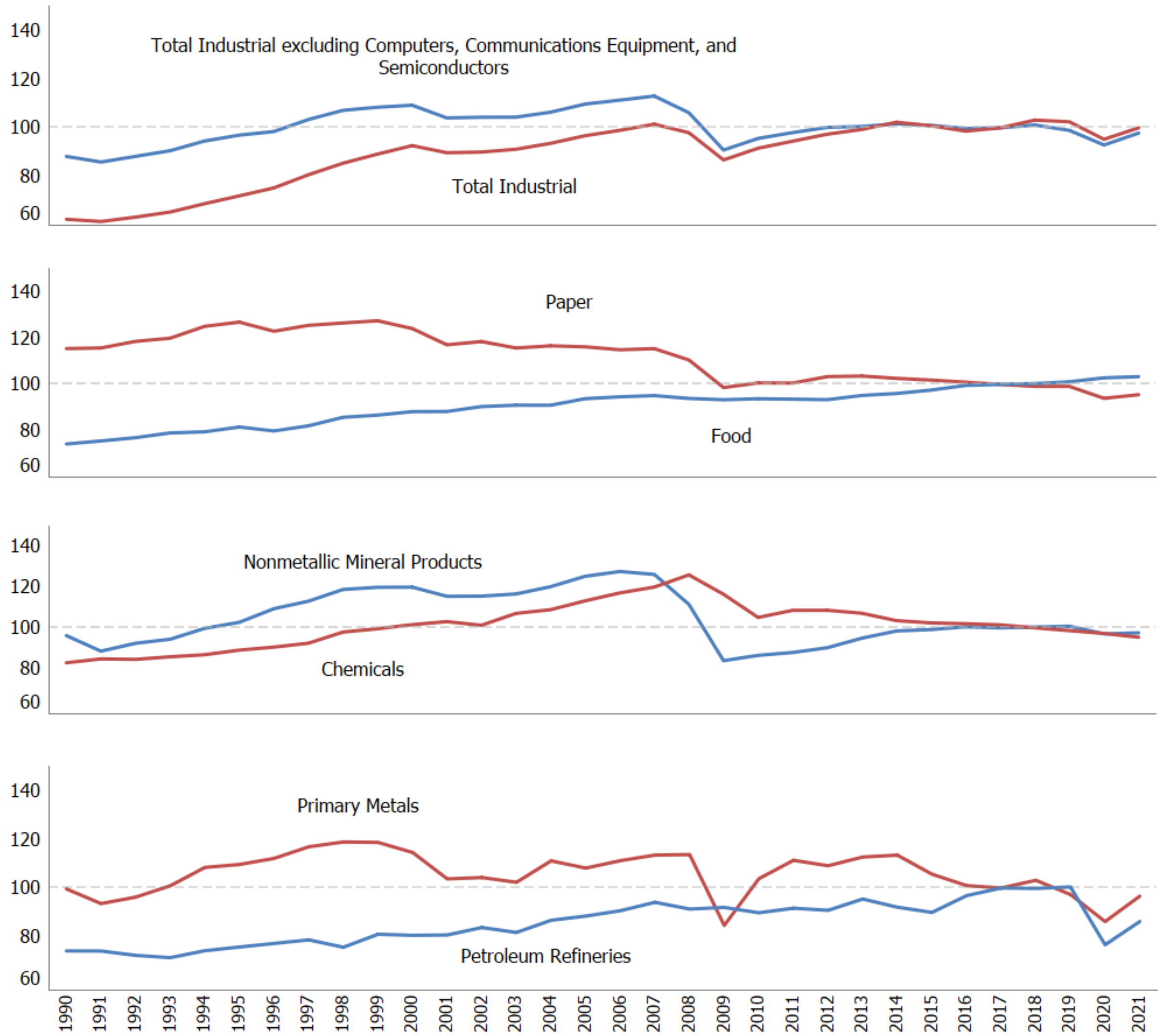
From 2020 to 2021, total industrial production and manufacturing output increased by 4.9 percent (FRB 2022). Over this period, output increased slightly across production indices for Food, Nonmetallic Mineral Products, Paper, Petroleum Refineries, and Primary Metals. Production of chemicals declined slightly between 2020 and 2021 (see Figure 3-11). From 2020 to 2021, energy use from fossil fuels in the industrial sector decreased by less than half a percent. Total energy use in the industrial sector increased by 0.7 percent, driven mainly by a 2.9 percent increase in the consumption of renewables. Due to the relative increases and decreases of individual indices there was an increase in natural gas and an increase in electricity used by the sector (see Figure 3-12). In

[16] Some commercial customers are large enough to obtain an industrial price for natural gas and/or electricity and are consequently grouped with the industrial end-use sector in U.S. energy statistics. These misclassifications of large commercial customers likely cause the industrial end-use sector to appear to be more sensitive to weather conditions.

2021, CO_2, CH_4, and N_2O emissions from fossil fuel combustion and electricity use within the industrial end-use sector totaled 1,212.3 MMT CO_2 Eq., a 2.1 percent increase from 2020 emissions.

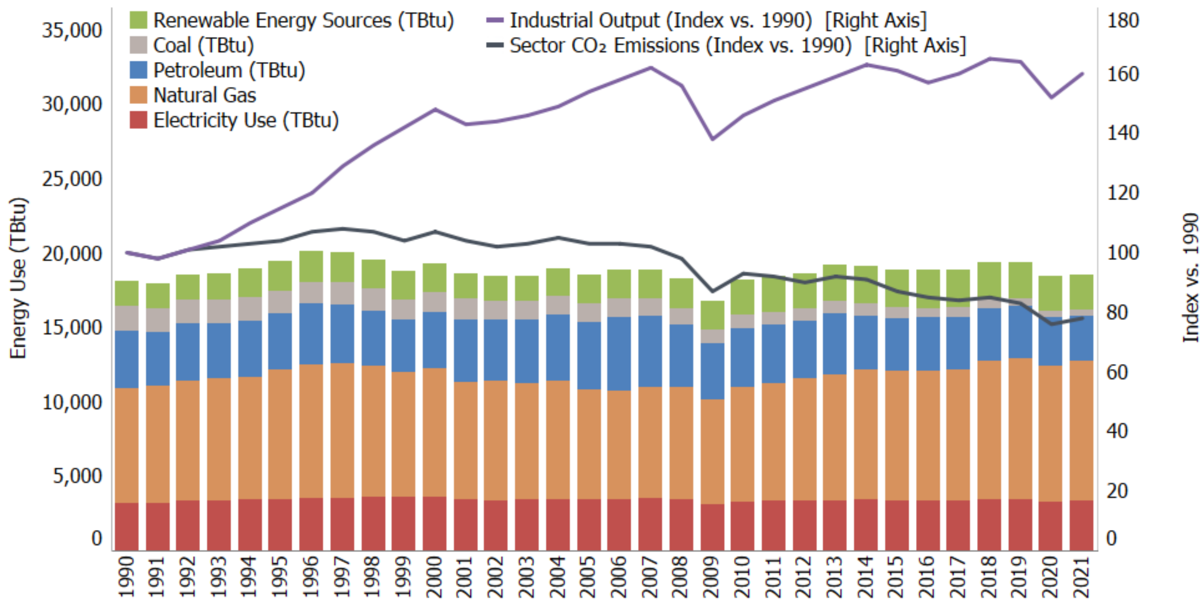
Through EPA's Greenhouse Gas Reporting Program (GHGRP), specific industrial sector trends can be discerned from the overall total EIA industrial fuel consumption data used for these calculations. For example, from 2020 to 2021, the underlying EIA data showed increased consumption of coal and natural gas in the industrial sector. The GHGRP data highlights that several industries contributed to these trends, including chemical manufacturing; pulp, paper and print; food processing, beverages and tobacco; minerals manufacturing; and agriculture-forest-fisheries.[17]

Figure 3-11: Industrial Production Indices (Index 2017=100)



[17] Further details on industrial sector combustion emissions are provided by EPA's GHGRP. See http://ghgdata.epa.gov/ghgp/main.do.

Figure 3-12: Fuels and Electricity Used in Industrial Sector, Industrial Output, and Total Sector CO_2 Emissions (Including Electricity)



Despite the growth in industrial output (60 percent) and the overall U.S. economy (109 percent) from 1990 to 2021, direct CO_2 emissions from fossil fuel combustion in the industrial sector decreased by 10.6 percent over the same time series. A number of factors are assumed to result in decoupling of growth in industrial output from industrial greenhouse gas emissions, for example: (1) more rapid growth in output from less energy-intensive industries relative to traditional manufacturing industries, and (2) energy-intensive industries such as steel are employing new methods, such as electric arc furnaces, that are less carbon intensive than the older methods.

Box 3-3: Uses of Greenhouse Gas Reporting Program Data and Improvements in Reporting Emissions from Industrial Sector Fossil Fuel Combustion

As described in the calculation methodology, total fossil fuel consumption for each year is based on aggregated end-use sector consumption published by the EIA. The availability of facility-level combustion emissions through EPA's GHGRP has provided an opportunity to better characterize the industrial sector's energy consumption and emissions in the United States, through a disaggregation of EIA's industrial sector fuel consumption data from select industries.

For GHGRP 2010 through 2021 reporting years, facility-level fossil fuel combustion emissions reported through EPA's GHGRP were categorized and distributed to specific industry types by utilizing facility-reported NAICS codes (as published by the U.S. Census Bureau). As noted previously in this report, the definitions and provisions for reporting fuel types in EPA's GHGRP include some differences from the Inventory's use of EIA national fuel statistics to meet the UNFCCC reporting guidelines. The IPCC has provided guidance on aligning facility-level reported fuels and fuel types published in national energy statistics, which guided this exercise.[18]

As with previous Inventory reports, the current effort represents an attempt to align, reconcile, and coordinate the facility-level reporting of fossil fuel combustion emissions under EPA's GHGRP with the national-level approach presented in this report. Consistent with recommendations for reporting the Inventory to the UNFCCC, progress was made on certain fuel types for specific industries and has been included in the CRF tables

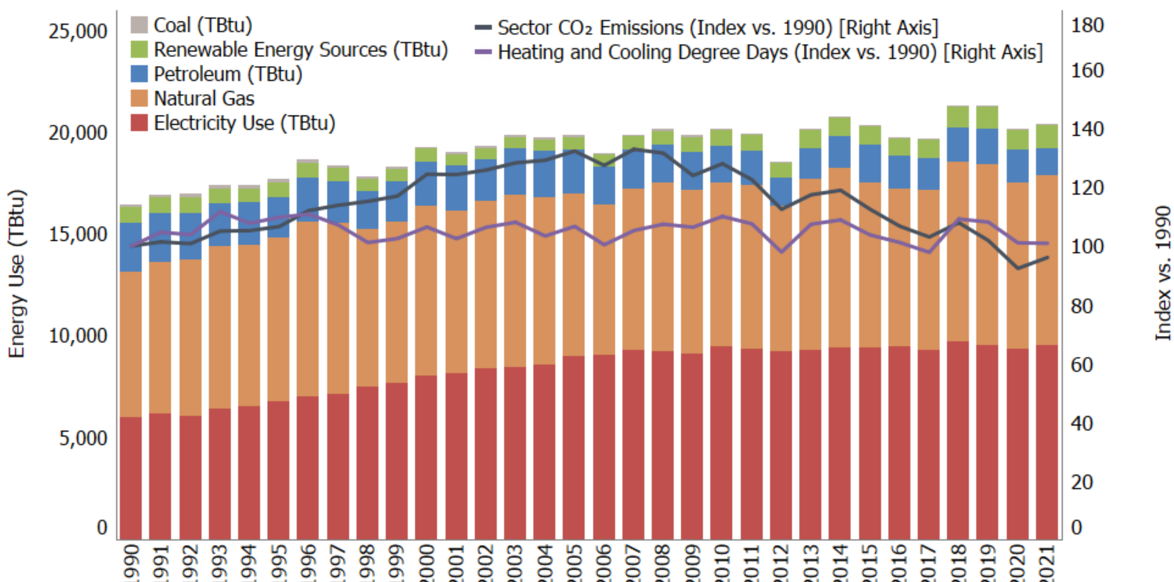
[18] See Section 4 "Use of Facility-Level Data in Good Practice National Greenhouse Gas Inventories" of the IPCC meeting report, and specifically the section on using facility-level data in conjunction with energy data, at http://www.ipcc-nggip.iges.or.jp/public/tb/TFI_Technical_Bulletin_1.pdf.

that are submitted to the UNFCCC along with this report.[19] The efforts in reconciling fuels focus on standard, common fuel types (e.g., natural gas, distillate fuel oil) where the fuels in EIA's national statistics aligned well with facility-level GHGRP data. For these reasons, the current information presented in the Common Reporting Format (CRF) tables should be viewed as an initial attempt at this exercise. Additional efforts will be made for future Inventory reports to improve the mapping of fuel types and examine ways to reconcile and coordinate any differences between facility-level data and national statistics. The current analysis includes the full time series presented in the CRF tables. Analyses were conducted linking GHGRP facility-level reporting with the information published by EIA in its MECS data in order to disaggregate the full 1990 through 2021 time period in the CRF tables. It is believed that the current analysis has led to improvements in the presentation of data in the Inventory, but further work will be conducted, and future improvements will be realized in subsequent Inventory reports. This includes incorporating the latest MECS data as it becomes available.

Residential and Commercial Sectors

Emissions from the residential and commercial sectors have generally decreased since 2005. Short-term trends are often correlated with seasonal fluctuations in energy use caused by weather conditions, rather than prevailing economic conditions. Population growth and a trend towards larger houses has led to increasing energy use over the time series, while population migration to warmer areas and improved energy efficiency and building insulation have slowed the increase in energy use in recent years. Starting in around 2014, energy use and emissions begin to decouple due to decarbonization of the electric power sector (see Figure 3-13).

Figure 3-13: Fuels and Electricity Used in Residential and Commercial Sectors, Heating and Cooling Degree Days, and Total Sector CO_2 Emissions (Including Electricity)



In 2021 the residential and commercial sectors accounted for 7 and 5 percent of CO_2 emissions from fossil fuel combustion, respectively; 40 and 11 percent of CH_4 emissions from fossil fuel combustion, respectively; and 2 and 1 percent of N_2O emissions from fossil fuel combustion, respectively. Emissions from these sectors were largely due to the direct consumption of natural gas and petroleum products, primarily for heating and cooking needs. Coal consumption was a minor component of energy use in the commercial sector and did not contribute to any energy use in the residential sector. In 2021, total emissions (CO_2, CH_4, and N_2O) from fossil fuel combustion and

[19] See https://www.epa.gov/ghgemissions/inventory-us-greenhouse-gas-emissions-and-sinks.

electricity use within the residential and commercial end-use sectors were 900.3 MMT CO_2 Eq. and 751.8 MMT CO_2 Eq., respectively. Total CO_2, CH_4, and N_2O emissions from combined fossil fuel combustion and electricity use within the residential and commercial end-use sectors increased by 3.4 and 4.9 percent from 2020 to 2021, respectively. An increase in heating degree days (0.5 percent) increased energy demand for heating in the residential and commercial sectors. This was partially offset by a 1.9 percent decrease in cooling degree days compared to 2020, which impacted demand for air conditioning in the residential and commercial sectors. This resulted in a 0.8 percent increase in residential sector electricity use. From 2020 to 2021 there was a 0.7 percent lower direct energy use in the commercial sector. In addition, a shift toward energy efficient products and more stringent energy efficiency standards for household equipment has contributed to a decrease in energy demand in households (EIA 2022g), resulting in a decrease in energy-related emissions. In the long term, the residential sector is also affected by population growth, migration trends toward warmer areas, and changes in total housing units and building attributes (e.g., larger sizes and improved insulation).

In 2021, combustion emissions from natural gas consumption represented 83 and 81 percent of the direct fossil fuel CO_2 emissions from the residential and commercial sectors, respectively. Carbon dioxide emissions from natural gas combustion in the residential and commercial sectors in 2021 increased by 0.9 percent and increased by 4.0 percent from 2020 to 2021, respectively.

U.S. Territories

Emissions from U.S. Territories are based on the fuel consumption in American Samoa, Guam, Puerto Rico, U.S. Virgin Islands, Wake Island, and other outlying U.S. Pacific Islands. As described in the Methodology section of CO_2 from Fossil Fuel Combustion, this data is collected separately from the sectoral-level data available for the general calculations. As sectoral information is not available for U.S. Territories, CO_2, CH_4, and N_2O emissions are not presented for U.S. Territories in the tables above by sector, though the emissions will occur across all sectors and sources including stationary, transportation and mobile combustion sources. Due to data availability limitations, 2021 and 2020 energy consumption for U.S. Territories for petroleum is proxied to 2019 consumption data.

Transportation Sector and Mobile Combustion

This discussion of transportation emissions follows the alternative method of presenting combustion emissions by allocating emissions associated with electricity generation to the transportation end-use sector, as presented in Table 3-11. Table 3-10 presents direct CO_2, CH_4, and N_2O emissions from all transportation sources (i.e., excluding emissions allocated to electricity consumption in the transportation end-use sector).

The transportation end-use sector and other mobile combustion accounted for 1,814 MMT CO_2 Eq. in 2021, which represented 37 percent of CO_2 emissions, 26 percent of CH_4 emissions, and 39 percent of N_2O emissions from fossil fuel combustion, respectively.[20] Fuel purchased in the United States for international aircraft and marine travel accounted for an additional 69.9 MMT CO_2 Eq. in 2021;[21] these emissions are recorded as international bunkers and are not included in U.S. totals according to UNFCCC reporting protocols.

Transportation End-Use Sector

From 1990 to 2019, transportation emissions from fossil fuel combustion rose by 21 percent, followed by a 13 percent reduction from 2019 to 2020. Overall, from 1990 to 2021, transportation emissions from fossil fuel combustion increased by 20 percent. The increase in transportation emissions from fossil fuel combustion from 1990 to 2021 was due, in large part, to increased demand for travel (see Figure 3-14). The number of vehicle miles traveled (VMT) by light-duty motor vehicles (passenger cars and light-duty trucks) increased 48 percent from 1990

[20] Note that these totals include CO_2, CH_4 and N_2O emissions from some sources in the U.S. Territories (ships and boats, recreational boats, non-transportation mobile sources) and CH_4 and N_2O emissions from transportation rail electricity.

[21] Some bunker fuels data are not yet available and has been proxied for 2021. This value will be updated for the Final Report published in April 2023.

to 2021,[22] as a result of a confluence of factors including population growth, economic growth, urban sprawl, and periods of low fuel prices. Between 2019 and 2020, emissions from light-duty vehicles fell by 11 percent, primarily the result of the COVID-19 pandemic and associated restrictions, such as people working from home and traveling less. Light-duty vehicle VMT rebounded in 2021 but is still estimated to be 1 percent below 2019 levels.[23]

Emissions from commercial aircraft for 2021 will be estimated in the Final Report published in April 2023, which will incorporate the latest data from FAA and other data sources. Here, commercial aircraft emissions are proxied to remain the same between 2020 and 2021. Commercial aircraft emissions have decreased 35 percent since 2007 (FAA 2022 and DOT 1991 through 2021).[24] Decreases in jet fuel emissions (excluding bunkers) started in 2007 due in part to improved operational efficiency that results in more direct flight routing, improvements in aircraft and engine technologies to reduce fuel burn and emissions, and the accelerated retirement of older, less fuel-efficient aircraft; however, the sharp decline in commercial aircraft emissions from 2019 to 2020 is primarily due to COVID-19 impacts on scheduled passenger air travel.

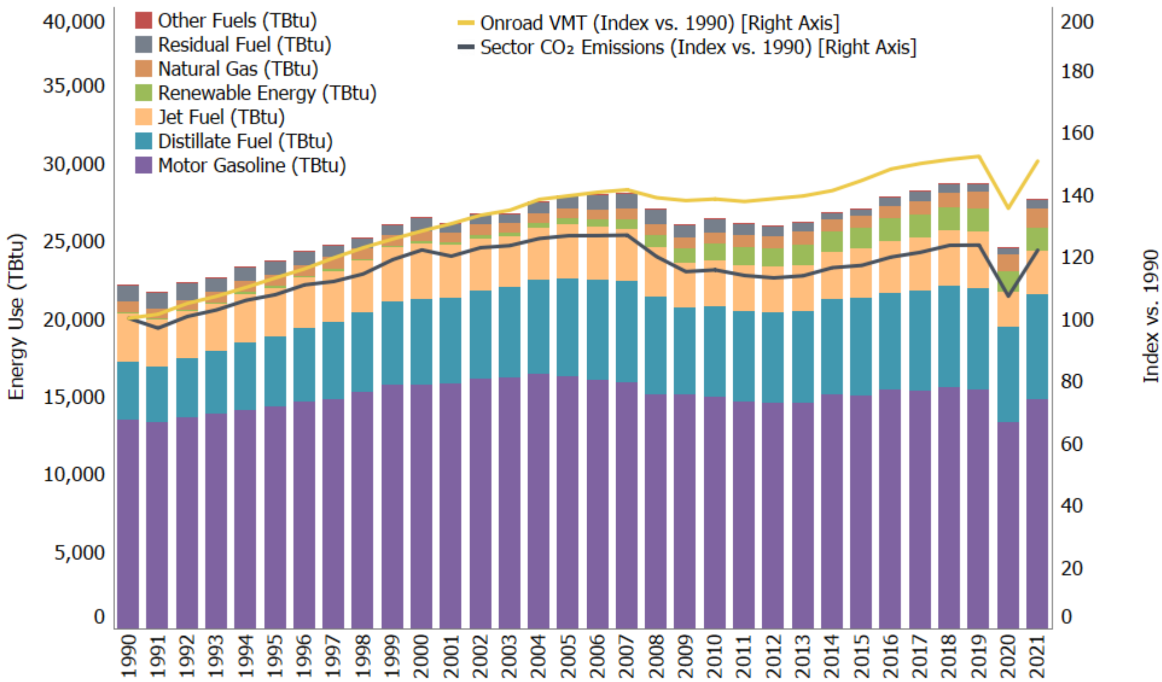
Almost all of the energy consumed for transportation was supplied by petroleum-based products, with more than half being related to gasoline consumption in automobiles and other highway vehicles. Other fuel uses, especially diesel fuel for freight trucks and jet fuel for aircraft, accounted for the remainder. The primary driver of transportation-related emissions was CO_2 from fossil fuel combustion, which increased by 22 percent from 1990 to 2021. Annex 3.2 presents the total emissions from all transportation and mobile sources, including CO_2, N_2O, CH_4, and HFCs.

[22] VMT estimates are based on data from FHWA Highway Statistics Table VM-1 (FHWA 1996 through 2020). VMT estimates from FHWA are allocated to vehicle type using ratios of VMT per vehicle type to total VMT, derived from EPA's MOVES3 model (see Annex 3.2 for information about the MOVES model). Data for 2021 has been proxied using FHWA Traffic Volume Trends.

[23] 2021 VMT is estimated based on FHWA Traffic Volume Trends data and will be updated when the 2021 data are released by FHWA.

[24] Commercial aircraft consists of passenger aircraft, cargo, and other chartered flights.

Figure 3-14: Fuels Used in Transportation Sector, On-road VMT, and Total Sector CO_2 Emissions



Notes: Distillate fuel, residual fuel, and jet fuel include adjustments for international bunker fuels. Distillate fuel and motor gasoline include adjustments for the sectoral allocation of these fuels. Other Fuels includes aviation gasoline and propane.
Source: Information on fuel consumption was obtained from EIA (2022).

Transportation Fossil Fuel Combustion CO_2 Emissions

Domestic transportation CO_2 emissions increased by 22 percent (323 MMT CO_2) between 1990 and 2021, an annualized increase of 0.7 percent. This includes a 24 percent increase in CO_2 emissions between 1990 and 2019, followed by a 13 percent decrease in 2020. Carbon dioxide emissions then increased by 14 percent between 2020 and 2021. Among domestic transportation sources in 2021, light-duty vehicles (including passenger cars and light-duty trucks) represented 57 percent of CO_2 emissions from fossil fuel combustion, medium- and heavy-duty trucks and buses 25 percent, commercial aircraft 5 percent, and other sources 13 percent. See Table 3-13 for a detailed breakdown of transportation CO_2 emissions by mode and fuel type.

Almost all of the energy consumed by the transportation sector is petroleum-based, including motor gasoline, diesel fuel, jet fuel, and residual oil. Carbon dioxide emissions from the combustion of ethanol and biodiesel for transportation purposes, along with the emissions associated with the agricultural and industrial processes involved in the production of biofuel, are captured in other Inventory sectors.[25] Ethanol consumption by the transportation sector has increased from 0.7 billion gallons in 1990 to 13.2 billion gallons in 2021, while biodiesel consumption has increased from 0.01 billion gallons in 2001 to 1.7 billion gallons in 2021. For additional information, see Section 3.10 on biofuel consumption at the end of this chapter and Table A-76 in Annex 3.2.

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[25] Biofuel estimates are presented in the Energy chapter for informational purposes only, in line with IPCC methodological guidance and UNFCCC reporting obligations. Net carbon fluxes from changes in biogenic carbon reservoirs in croplands are accounted for in the estimates for Land Use, Land-Use Change, and Forestry (see Chapter 6). More information and additional analyses on biofuels are available at EPA's Renewable Fuels Standards website. See https://www.epa.gov/renewable-fuel-standard-program.

Carbon dioxide emissions from passenger cars and light-duty trucks totaled 1,031.4 MMT CO_2 in 2021, an increase of 13 percent (117 MMT CO_2) from 1990 to 2021. The increase in CO_2 emissions from passenger cars and light-duty trucks from 1990 to 2021 was due, in large part, to increased demand for travel as fleet-wide light-duty vehicle fuel economy was relatively stable (average new vehicle fuel economy declined slowly from 1990 through 2004 and then increased more rapidly from 2005 through 2021). Carbon dioxide emissions from passenger cars and light-duty trucks peaked at 1,145.7 MMT CO_2 in 2004, and since then have declined about 10 percent. The decline in new light-duty vehicle fuel economy between 1990 and 2004 (Figure 3-15) reflects the increasing market share of light-duty trucks, which grew from about 30 percent of new vehicle sales in 1990 to 48 percent in 2004. Starting in 2005, average new vehicle fuel economy began to increase while light-duty vehicle VMT grew only modestly for much of the period. Light-duty vehicle VMT grew by less than one percent or declined each year between 2005 and 2013, and again between 2017 and 2019.[26] VMT grew at faster rates of 2.6 percent from 2014 to 2015 and 2.5 percent from 2015 to2016. From 2019 to 2020, light-duty vehicle VMT declined by 11 percent due to the COVID-19 pandemic; from 2020 to 2021 light-duty vehicle VMT rebounded, increasing by 11.2 percent.

Average new vehicle fuel economy has increased almost every year since 2005, while the light-duty truck share of new vehicle sales decreased to about 33 percent in 2009 and has since varied from year to year between 36 and 61 percent. Since 2014, the light-duty truck share has steadily increased, reaching 61 percent of new vehicles sales in model year 2021 (EPA 2022b). See Annex 3.2 for data by vehicle mode and information on VMT and the share of new vehicles (in VMT).

Medium- and heavy-duty truck CO_2 emissions increased by 81 percent from 1990 to 2021. This increase was largely due to a substantial growth in medium- and heavy-duty truck VMT, which increased by 71 percent between 1990 and 2021.

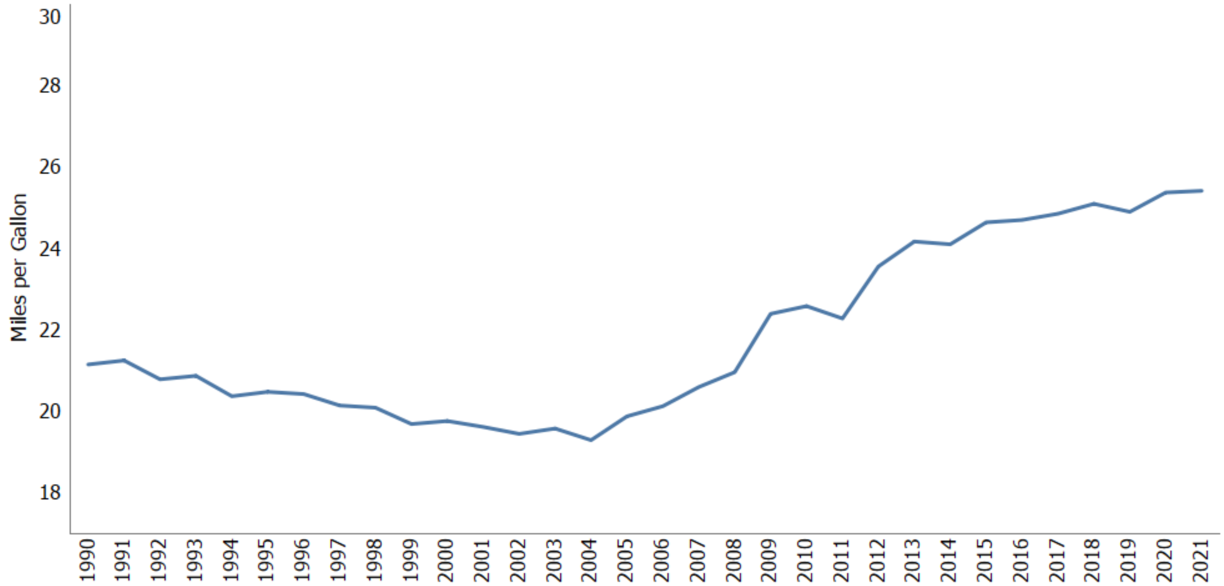
Carbon dioxide emissions from the domestic operation of commercial aircraft decreased by 17 percent (18.6 MMT CO_2) from 1990 to 2021. Across all categories of aviation, excluding international bunkers, CO_2 emissions decreased by 12 percent (21.8 MMT CO_2) between 1990 and 2021.[27] Emissions from military aircraft decreased 68 percent between 1990 and 2021. Commercial aircraft emissions increased 27 percent between 1990 and 2007, dropped 4 percent from 2007 to 2019, and then dropped 32 percent from 2019 to 2020, a change of approximately 17 percent between 1990 and 2021. Commercial aircraft emissions are proxied to remain the same between 2020 and 2021 and will be updated in the Final Report published in April 2023.

Transportation sources also produce CH_4 and N_2O; these emissions are included in Figure 3-14 and Table 3-15 and in the CH_4 and N_2O from Mobile Combustion section. Annex 3.2 presents total emissions from all transportation and mobile sources, including CO_2, CH_4, N_2O, and HFCs.

[26] VMT estimates are based on data from FHWA Highway Statistics Table VM-1 (FHWA 1996 through 2020). VMT estimates from FHWA are allocated to vehicle type using ratios of VMT per vehicle type to total VMT, derived from EPA's MOVES3 model (see Annex 3.2 for information about the MOVES model). Data for 2021 has been proxied using FHWA Traffic Volume Trends.

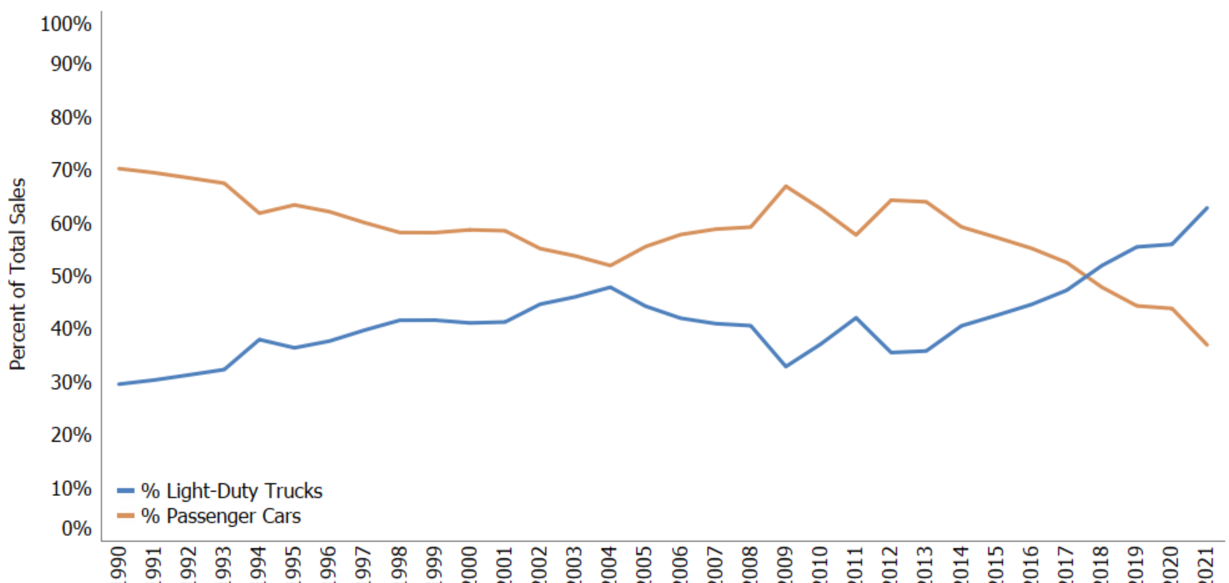
[27] Includes consumption of jet fuel and aviation gasoline. Does not include aircraft bunkers, which are not included in national emission totals, in line with IPCC methodological guidance and UNFCCC reporting obligations.

Figure 3-15: Sales-Weighted Fuel Economy of New Passenger Cars and Light-Duty Trucks, 1990–2021



Source: EPA (2022a).

Figure 3-16: Sales of New Passenger Cars and Light-Duty Trucks, 1990–2021



Source: EPA (2022b).

Table 3-13: CO_2 Emissions from Fossil Fuel Combustion in Transportation End-Use Sector (MMT CO_2 Eq.)

Fuel/Vehicle Type	1990		2005		2017	2018	2019	2020	2021
Gasoline[a]	**958.9**		**1,150.1**		**1,081.8**	**1,097.0**	**1,086.5**	**937.0**	**1,040.3**
Passenger Cars	612.8		518.9		375.2	382.5	380.0	328.0	364.6

Light-Duty Trucks	283.6	583.4	661.5	667.6	658.6	565.8	627.0
Medium- and Heavy-Duty Trucks[b]	42.8	28.1	24.9	26.2	27.0	24.1	27.7
Buses	2.1	1.1	2.5	2.7	2.8	2.5	2.9
Motorcycles	3.4	4.9	7.0	7.3	7.4	6.6	7.5
Recreational Boats[c]	14.3	13.7	10.6	10.7	10.7	10.1	10.6
Distillate Fuel Oil (Diesel)[a]	**274.6**	**472.1**	**474.9**	**486.6**	**484.1**	**455.0**	**502.2**
Passenger Cars	9.4	2.2	3.0	2.8	2.7	2.5	2.8
Light-Duty Trucks	8.4	30.4	31.1	31.2	31.2	30.2	34.4
Medium- and Heavy-Duty Trucks[b]	189.0	357.2	362.0	371.5	373.0	353.4	392.6
Buses	11.1	15.5	19.7	20.4	20.7	19.8	22.1
Rail	35.5	46.1	37.4	38.5	36.0	31.0	32.1
Recreational Boats[c]	2.7	2.9	2.8	2.8	2.9	2.7	2.8
Ships and Non-Recreational Boats[c]	6.8	8.4	10.0	9.3	7.5	7.4	7.7
International Bunker Fuels[e]	11.7	9.5	9.0	10.0	10.1	7.8	7.4
Jet Fuel	**222.3**	**249.5**	**249.4**	**253.1**	**258.5**	**160.4**	**203.8**
Commercial Aircraft[f]	109.9	132.7	128.0	129.6	134.2	91.3	91.3
Military Aircraft	35.7	19.8	12.5	12.1	12.1	10.7	11.3
General Aviation Aircraft	38.5	36.8	31.2	30.6	31.4	18.6	61.3
International Bunker Fuels[e]	*38.2*	*60.2*	77.8	80.9	80.8	39.8	39.9
International Bunker Fuels from Commercial Aviation	*30.0*	*55.6*	*74.5*	*77.7*	*77.6*	*36.7*	*36.7*
Aviation Gasoline	**3.1**	**2.4**	**1.4**	**1.5**	**1.6**	**1.4**	**1.5**
General Aviation Aircraft	3.1	2.4	1.4	1.5	1.6	1.4	1.5
Residual Fuel Oil	**76.3**	**62.9**	**49.9**	**45.4**	**39.7**	**29.4**	**45.5**
Ships and Non-Recreational Boats[e]	22.6	19.3	16.5	14.0	14.5	7.3	23.6
International Bunker Fuels[e]	53.7	43.6	33.4	31.4	25.2	22.1	21.9
Natural Gas[i]	**36.0**	**33.1**	**42.3**	**50.9**	**58.9**	**58.7**	**65.1**
Passenger Cars	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	+	+	+	+	+
Medium- and Heavy-Duty Trucks	+	0.1	0.1	0.1	0.1	0.1	0.1
Buses	+	0.3	0.5	0.6	0.6	0.6	0.7
Pipeline[g]	36.0	32.6	41.6	50.2	58.2	57.9	64.2
LPG[i]	**1.4**	**1.8**	**0.6**	**0.6**	**0.5**	**0.3**	**0.3**
Passenger Cars	+	+	+	+	+	+	+
Light-Duty Trucks	0.2	0.3	+	0.1	0.1	+	0.1
Medium- and Heavy-Duty Trucks[b]	1.1	1.1	0.4	0.4	0.4	0.2	0.2
Buses	0.2	0.3	0.1	0.1	0.1	+	+
Electricity[l]	**3.0**	**4.7**	**4.3**	**4.8**	**4.8**	**4.1**	**5.1**
Passenger Cars	+	+	0.8	1.2	1.4	1.3	1.8
Light-Duty Trucks	+	+	0.1	0.2	0.2	0.3	0.7
Buses	+	+	+	+	+	0.1	0.1
Rail	3.0	4.7	3.4	3.4	3.1	2.4	2.5
Total (Excluding Bunkers)[e]	**1,472.0**	**1,863.3**	**1,784.4**	**1,817.7**	**1,818.7**	**1,576.6**	**1,794.5**
Total (Including Bunkers)[j]	**1,575.6**	**1976.6**	**1,904.6**	**1,939.8**	**1,934.8**	**1,646.3**	**1,863.7**
Biofuels-Ethanol[h]	*4.1*	*21.6*	*77.7*	*78.6*	*78.7*	*68.1*	*76.3*
Biofuels-Biodiesel[h]	*+*	*0.9*	*18.7*	*17.9*	*17.1*	*17.7*	*16.1*

+ Does not exceed 0.05 MMT CO_2 Eq.

[a] On-road fuel consumption data from FHWA Table MF-21 and MF-27 were used to determine total on-road use of motor gasoline and diesel fuel (FHWA 1996 through 2020). Data for 2021 is proxied using FHWA Traffic Volume Travel Trends. Ratios developed from MOVES3 output are used to apportion FHWA fuel consumption data to vehicle type and fuel type (see Annex 3.2 for information about the MOVES model).

[b] Includes medium- and heavy-duty trucks over 8,500 lbs.

[c] In 2014, EPA incorporated the NONROAD2008 model into the MOVES model framework. The current Inventory uses the Nonroad component of MOVES3 for years 1999 through 2021. See Annex 3.2 for information about the MOVES model.

[d] Note that large year over year fluctuations in emission estimates partially reflect nature of data collection for these sources.

[e] Official estimates exclude emissions from the combustion of both aviation and marine international bunker fuels; however, estimates including international bunker fuel-related emissions are presented for informational purposes.

[f] Commercial aircraft, as modeled in FAA's Aviation Environmental Design Tool (AEDT), consists of passenger aircraft, cargo, and other chartered flights.

[g] Pipelines reflect CO_2 emissions from natural gas-powered pipelines transporting natural gas.

[h] Ethanol and biodiesel estimates are presented for informational purposes only. See Section 3.10 of this chapter and the estimates in Land Use, Land-Use Change, and Forestry (see Chapter 6), in line with IPCC methodological guidance and UNFCCC reporting obligations, for more information on ethanol and biodiesel.

[i] Transportation sector natural gas and LPG consumption are based on data from EIA (2021b). Prior to the 1990 to 2015 Inventory, data from DOE TEDB were used to estimate each vehicle class's share of the total natural gas and LPG consumption. Since TEDB does not include estimates for natural gas use by medium and heavy-duty trucks or LPG use by passenger cars, EIA Alternative Fuel Vehicle Data (Browning 2017) is now used to determine each vehicle class's share of the total natural gas and LPG consumption. These changes were first incorporated in the 1990 to 2016 Inventory and apply to the 1990 to 2021 time period.

[j] Includes emissions from rail electricity.

[k] Electricity consumption by passenger cars, light-duty trucks (SUVs), and buses is based on plug-in electric vehicle sales and engine efficiency data, as outlined in Browning (2018a). In prior Inventory years, CO_2 emissions from electric vehicle charging were allocated to the residential and commercial sectors. They are now allocated to the transportation sector. These changes apply to the 2010 through 2021 time period.

Notes: This table does not include emissions from non-transportation mobile sources, such as agricultural equipment and construction/mining equipment; it also does not include emissions associated with electricity consumption by pipelines or lubricants used in transportation. In addition, this table does not include CO_2 emissions from U.S. Territories, since these are covered in a separate chapter of the Inventory. Totals may not sum due to independent rounding.

Mobile Fossil Fuel Combustion CH_4 and N_2O Emissions

Mobile combustion includes emissions of CH_4 and N_2O from all transportation sources identified in the U.S. Inventory with the exception of pipelines and electric locomotives;[28] mobile sources also include non-transportation sources such as construction/mining equipment, agricultural equipment, vehicles used off-road, and other sources (e.g., snowmobiles, lawnmowers, etc.).[29] Annex 3.2 includes a summary of all emissions from both transportation and mobile sources. Table 3-14 and Table 3-15 provide mobile fossil fuel CH_4 and N_2O emission estimates in MMT CO_2 Eq.[30]

[28] Emissions of CH_4 from natural gas systems are reported separately. More information on the methodology used to calculate these emissions are included in this chapter and Annex 3.4.

[29] See the methodology sub-sections of the CO_2 from Fossil Fuel Combustion and CH_4 and N_2O from Mobile Combustion sections of this chapter. Note that N_2O and CH_4 emissions are reported using different categories than CO_2. CO_2 emissions are reported by end-use sector (Transportation, Industrial, Commercial, Residential, U.S. Territories), and generally adhere to a top-down approach to estimating emissions. CO_2 emissions from non-transportation sources (e.g., lawn and garden equipment, farm equipment, construction equipment) are allocated to their respective end-use sector (i.e., construction equipment CO_2 emissions are included in the Industrial end-use sector instead of the Transportation end-use sector). CH_4 and N_2O emissions are reported using the "Mobile Combustion" category, which includes non-transportation mobile sources. CH_4 and N_2O emission estimates are bottom-up estimates, based on total activity (fuel use, VMT) and emissions factors by source and technology type. These reporting schemes are in accordance with IPCC guidance. For informational purposes only, CO_2 emissions from non-transportation mobile sources are presented separately from their overall end-use sector in Annex 3.2.

[30] See Annex 3.2 for a complete time series of emission estimates for 1990 through 2021.

Mobile combustion was responsible for a small portion of national CH_4 emissions (0.4 percent) and was the fifth largest source of national N_2O emissions (4.5 percent). From 1990 to 2021, mobile source CH_4 emissions declined by 64 percent, to 2.6 MMT CO_2 Eq. (94 kt CH_4), due largely to emissions control technologies employed in on-road vehicles since the mid-1990s to reduce CO, NO_x, NMVOC, and CH_4 emissions. Mobile source emissions of N_2O decreased by 55 percent from 1990 to 2021, to 17.1 MMT CO_2 Eq. (65 kt N_2O). Earlier generation emissions control technologies initially resulted in higher N_2O emissions, causing a 31 percent increase in N_2O emissions from mobile sources between 1990 and 1997. Improvements in later-generation emissions control technologies have reduced N_2O emissions, resulting in a 66 percent decrease in mobile source N_2O emissions from 1997 to 2021 (Figure 3-17). Overall, CH_4 and N_2O emissions were predominantly from gasoline-fueled passenger cars and light-duty trucks and non-highway sources. See Annex 3.2 for data by vehicle mode and information on VMT and the share of new vehicles.

Figure 3-17: Mobile Source CH_4 and N_2O Emissions

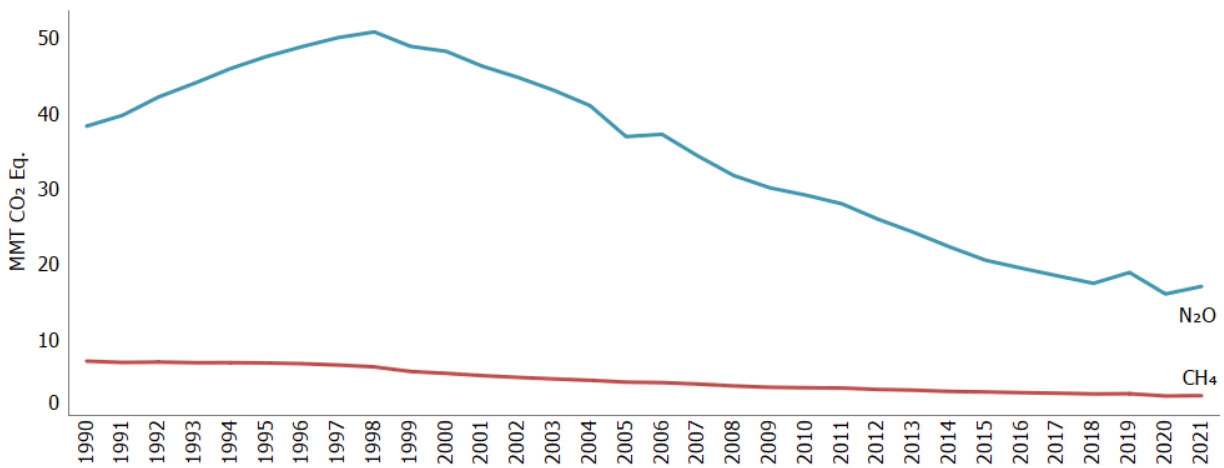


Table 3-14: CH_4 Emissions from Mobile Combustion (MMT CO_2 Eq.)

Fuel Type/Vehicle Type[a]	1990	2005	2017	2018	2019	2020	2021
Gasoline On-Road[b]	**5.8**	**2.4**	**1.0**	**0.9**	**1.0**	**0.8**	**0.8**
Passenger Cars	3.8	1.2	0.3	0.3	0.3	0.2	0.3
Light-Duty Trucks	1.5	1.0	0.6	0.5	0.6	0.5	0.5
Medium- and Heavy-Duty Trucks and Buses	0.5	0.1	+	+	+	+	+
Motorcycles	+	+	+	+	+	+	+
Diesel On-Road[b]	**+**	**+**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**
Passenger Cars	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	+	+	+	+	+
Medium- and Heavy-Duty Trucks	+	+	0.1	0.1	0.1	0.1	0.1
Medium- and Heavy-Duty Buses	+	+	+	+	+	+	+
Alternative Fuel On-Road	**+**	**0.2**	**0.1**	**0.1**	**0.1**	**0.1**	**0.1**
Non-Road[c]	**1.4**	**1.8**	**1.7**	**1.7**	**1.7**	**1.6**	**1.6**
Ships and Boats	0.4	0.5	0.5	0.5	0.4	0.4	0.5
Rail[d]	0.1	0.1	0.1	0.1	0.1	0.1	0.1
Aircraft	0.1	0.1	+	+	+	+	+
Agricultural Equipment[e]	0.2	0.2	0.1	0.1	0.1	0.1	0.1
Construction/Mining Equipment[f]	0.2	0.3	0.2	0.2	0.2	0.2	0.2
Other[g]	0.5	0.7	0.8	0.8	0.8	0.8	0.7
Total	**7.2**	**4.4**	**2.9**	**2.9**	**2.9**	**2.6**	**2.6**

+ Does not exceed 0.05 MMT CO_2 Eq.

[a] See Annex 3.2 for definitions of on-road vehicle types.

[b] Gasoline and diesel highway vehicle mileage estimates are based on data from FHWA Highway Statistics Table VM-1. VMT estimates from FHWA are allocated to vehicle type using ratios of VMT per vehicle type to total VMT, derived from EPA's MOVES3 model (see Annex 3.2 for information about the MOVES model). Data for 2021 is proxied using FHWA Traffic Volume Trends Data.

[c] Nonroad fuel consumption estimates for 2020 are adjusted to account for the COVID-19 pandemic and associated restrictions. For agricultural equipment and airport equipment, sector specific adjustment factors were applied to the 2019 data. For all other sectors, a 7.7 percent reduction factor is used, based on transportation diesel use (EIA 2022

[d]Rail emissions do not include emissions from electric powered locomotives. Class II and Class III diesel consumption data for 2014 to 2021 is estimated by applying the historical average fuel usage per carload factor to the annual number of carloads.

[e] Includes equipment, such as tractors and combines, as well as fuel consumption from trucks that are used off-road in agriculture.

[f] Includes equipment, such as cranes, dumpers, and excavators, as well as fuel consumption from trucks that are used off-road in construction.

[g] "Other" includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment, as well as fuel consumption from trucks that are used off-road for commercial/industrial purposes.

Table 3-15: N_2O Emissions from Mobile Combustion (MMT CO_2 Eq.)

Fuel Type/Vehicle Type[a]	1990	2005	2017	2018	2019	2020	2021
Gasoline On-Road[b]	**32.0**	**28.5**	**8.4**	**7.0**	**8.1**	**6.4**	**6.2**
Passenger Cars	22.4	13.3	2.9	2.5	2.5	2.0	2.0
Light-Duty Trucks	8.7	14.0	5.2	4.3	5.4	4.2	4.0
Medium- and Heavy-Duty Trucks and Buses	0.8	1.2	0.2	0.2	0.2	0.1	0.1
Motorcycles	+	+	0.1	0.1	0.1	0.1	0.1
Diesel On-Road[b]	**0.2**	**0.4**	**2.8**	**3.0**	**3.2**	**3.0**	**3.4**
Passenger Cars	+	+	+	+	+	+	+
Light-Duty Trucks	+	+	0.2	0.3	0.2	0.2	0.3
Medium- and Heavy-Duty Trucks	0.2	0.3	2.5	2.8	3.0	2.8	3.2
Medium- and Heavy-Duty Buses	+	+	0.2	0.2	0.3	0.2	0.3
Alternative Fuel On-Road	**+**	**+**	**+**	**+**	**+**	**+**	**+**
Non-Road[c]	**6.2**	**8.1**	**7.4**	**7.5**	**7.6**	**6.8**	**7.5**
Ships and Boats	0.2	0.2	0.2	0.2	0.2	0.1	0.3
Rail[d]	0.2	0.3	0.3	0.3	0.2	0.2	0.2
Aircraft	1.5	1.6	1.4	1.4	1.5	1.0	1.3
Agricultural Equipment[e]	1.2	1.4	1.1	1.1	1.1	1.1	1.
Construction/Mining Equipment[f]	1.2	1.9	1.6	1.6	1.7	1.6	1.7
Other[g]	1.8	2.8	2.8	2.9	2.9	2.8	2.9
Total	**38.4**	**37.0**	**18.5**	**17.5**	**19.0**	**16.1**	**17.1**

+ Does not exceed 0.05 MMT CO_2 Eq.

[a] See Annex 3.2 for definitions of on-road vehicle types.

[b] Gasoline and diesel highway vehicle mileage estimates are based on data from FHWA Highway Statistics Table VM-1. VMT estimates from FHWA are allocated to vehicle type using ratios of VMT per vehicle type to total VMT, derived from EPA's MOVES3 model (see Annex 3.2 for information about the MOVES model). Data for 2021 is proxied using FHWA Traffic Volume Trends Data.

[c] Nonroad fuel consumption estimates for 2020 are adjusted to account for the COVID-19 pandemic and associated restrictions. For agricultural equipment and airport equipment, sector specific adjustment factors were applied to the 2019 data. For all other sectors, a 7.7 percent reduction factor is used, based on transportation diesel use (EIA 2022).

[d] Rail emissions do not include emissions from electric powered locomotives. Class II and Class III diesel consumption data for 2014 through 2021 is estimated by applying the historical average fuel usage per carload factor to the annual number of carloads.

[e] Includes equipment, such as tractors and combines, as well as fuel consumption from trucks that are used off-road in agriculture.

[f] Includes equipment, such as cranes, dumpers, and excavators, as well as fuel consumption from trucks that are used off-road in construction.

[g] “Other” includes snowmobiles and other recreational equipment, logging equipment, lawn and garden equipment, railroad equipment, airport equipment, commercial equipment, and industrial equipment, as well as fuel consumption from trucks that are used off-road for commercial/industrial purposes.
Note: Totals may not sum due to independent rounding.

CO_2 from Fossil Fuel Combustion

Methodology and Time-Series Consistency

CO_2 emissions from fossil fuel combustion are estimated in line with a Tier 2 method described by the IPCC in the *2006 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC 2006) with some exceptions as discussed below.[31] A detailed description of the U.S. methodology is presented in Annex 2.1, and is characterized by the following steps:

1. *Determine total fuel consumption by fuel type and sector*. Total fossil fuel consumption for each year is estimated by aggregating consumption data by end-use sector (e.g., commercial, industrial), primary fuel type (e.g., coal, petroleum, gas), and secondary fuel category (e.g., motor gasoline, distillate fuel oil). Fuel consumption data for the United States were obtained directly from the EIA of the U.S. Department of Energy (DOE), primarily from the *Monthly Energy Review* (EIA 2022a). EIA data include fuel consumption statistics from the 50 U.S. states and the District of Columbia, including tribal lands. The EIA does not include territories in its national energy statistics, so fuel consumption data for territories were collected separately from EIA’s International Energy Statistics (EIA 2022b).[32]

   For consistency of reporting, the IPCC has recommended that countries report energy data using the International Energy Agency (IEA) reporting convention and/or IEA data. Data in the IEA format are presented “top down”—that is, energy consumption for fuel types and categories are estimated from energy production data (accounting for imports, exports, stock changes, and losses). The resulting quantities are referred to as “apparent consumption.” The data collected in the United States by EIA on an annual basis and used in this Inventory are predominantly from mid-stream or conversion energy consumers such as refiners and electric power generators. These annual surveys are supplemented with end-use energy consumption surveys, such as the Manufacturing Energy Consumption Survey, that are conducted on a periodic basis (every four years). These consumption datasets help inform the annual surveys to arrive at the national total and sectoral breakdowns for that total.[33]

   Also, note that U.S. fossil fuel energy statistics are generally presented using gross calorific values (GCV) (i.e., higher heating values). Fuel consumption activity data presented here have not been adjusted to correspond to international standards, which are to report energy statistics in terms of net calorific values (NCV) (i.e., lower heating values).[34]

   Subtract uses accounted for in the Industrial Processes and Product Use chapter. Portions of the fuel consumption data for seven fuel categories—coking coal, distillate fuel, industrial other coal, petroleum coke, natural gas, residual fuel oil, and other oil—were reallocated to the Industrial Processes and Product Use chapter, as they were consumed during non-energy-related industrial activity. To make these

[31] The IPCC Tier 3B methodology is used for estimating emissions from commercial aircraft.

[32] Fuel consumption by U.S. Territories (i.e., American Samoa, Guam, Puerto Rico, U.S. Virgin Islands, Wake Island, and other U.S. Pacific Islands) is included in this report and contributed total emissions of 23.0 MMT CO_2 Eq. in 2020. Data is only available for EIA’s International Energy Statistics through 2020 for coal and natural gas consumption and through 2019 for petroleum consumption. For this reason, data for the 2020 U.S. Territories emission estimates is proxied to the most recent data available.

[33] See IPCC Reference Approach for Estimating CO_2 Emissions from Fossil Fuel Combustion in Annex 4 for a comparison of U.S. estimates using top-down and bottom-up approaches.

[34] A crude convention to convert between gross and net calorific values is to multiply the heat content of solid and liquid fossil fuels by 0.95 and gaseous fuels by 0.9 to account for the water content of the fuels. Biomass-based fuels in U.S. energy statistics, however, are generally presented using net calorific values.

adjustments, additional data were collected from AISI (2004 through 2021), Coffeyville (2012), U.S. Census Bureau (2001 through 2011), EIA (2022a, 2022d, 2022f), USAA (2008 through 2021), USGS (1991 through 2020), (USGS 2019), USGS (2014 through 2021a), USGS (2014 through 2021b), USGS (1995 through 2013), USGS (1995, 1998, 2000, 2001, 2002, 2007), USGS (2021a), USGS (1991 through 2015a), USGS (1991 through 2020), USGS (2014 through 2021a), USGS (1991 through 2015b), USGS (2021b), USGS (1991 through 2020).[35]

2. *Adjust for biofuels and petroleum denaturant*. Fossil fuel consumption estimates are adjusted downward to exclude fuels with biogenic origins and avoid double counting in petroleum data statistics. Carbon dioxide emissions from ethanol added to motor gasoline and biodiesel added to diesel fuel are not included specifically in summing energy sector totals. Net carbon fluxes from changes in biogenic carbon reservoirs are accounted for in the estimates for LULUCF, therefore, fuel consumption estimates are adjusted to remove ethanol and biodiesel.[36] For the years 1993 through 2008, petroleum denaturant is currently included in EIA statistics for both natural gasoline and finished motor gasoline. To avoid double counting, petroleum denaturant is subtracted from finished motor gasoline for these years.[37]

3. *Adjust for exports of CO_2.* Since October 2000, the Dakota Gasification Plant has been exporting CO_2 produced in the coal gasification process to Canada by pipeline. Because this CO_2 is not emitted to the atmosphere in the United States, the associated fossil fuel (lignite coal) that is gasified to create the exported CO_2 is subtracted from EIA (2022f) coal consumption statistics that are used to calculate greenhouse gas emissions from the Energy Sector. The associated fossil fuel is the total fossil fuel burned at the plant with the CO_2 capture system multiplied by the fraction of the plant's total site-generated CO_2 that is recovered by the capture system. To make these adjustments, data for CO_2 exports were collected from Environment and Climate Change Canada (2022). A discussion of the methodology used to estimate the amount of CO_2 captured and exported by pipeline is presented in Annex 2.1.

4. *Adjust sectoral allocation of distillate fuel oil and motor gasoline.* EPA conducted a separate bottom-up analysis of transportation fuel consumption based on data from the Federal Highway Administration that indicated that the amount of distillate and motor gasoline consumption allocated to the transportation sector in the EIA statistics should be adjusted. Therefore, for these estimates, the transportation sector's distillate fuel and motor gasoline consumption were adjusted to match the value obtained from the bottom-up analysis. As the total distillate and motor gasoline consumption estimate from EIA are considered to be accurate at the national level, the distillate and motor gasoline consumption totals for the residential, commercial, and industrial sectors were adjusted proportionately. The data sources used in the bottom-up analysis of transportation fuel consumption include AAR (2008 through 2022), Benson (2002 through 2004), DOE (1993 through 2020), EIA (2007), EIA (1991 through 2022), EPA (2022c), and FHWA (1996 through 2021).[38]

5. *Adjust for fuels consumed for non-energy uses*. U.S. aggregate energy statistics include consumption of fossil fuels for non-energy purposes. These are fossil fuels that are manufactured into plastics, asphalt, lubricants, or other products. Depending on the end-use, this can result in storage of some or all of the C contained in the fuel for a period of time. As the emission pathways of C used for non-energy purposes are vastly different than fuel combustion (since the C in these fuels ends up in products instead of being combusted), these emissions are estimated separately in Section 3.2 – Carbon Emitted and Stored in Products from Non-Energy Uses of Fossil Fuels. Therefore, the amount of fuels used for non-energy

[35] See sections on Iron and Steel Production and Metallurgical Coke Production, Ammonia Production and Urea Consumption, Petrochemical Production, Titanium Dioxide Production, Ferroalloy Production, Aluminum Production, and Silicon Carbide Production and Consumption in the Industrial Processes and Product Use chapter.

[36] Natural gas energy statistics from EIA (2022e) are already adjusted downward to account for biogas in natural gas.

[37] These adjustments are explained in greater detail in Annex 2.1.

[38] Bottom-up gasoline and diesel highway vehicle fuel consumption estimates are based on data from FHWA Highway Statistics Table MF-21, MF-27, and VM-1 (FHWA 1996 through 2021).

purposes was subtracted from total fuel consumption. Data on non-fuel consumption were provided by EIA (2022d).

6. *Subtract consumption of international bunker fuels.* According to the UNFCCC reporting guidelines emissions from international transport activities, or bunker fuels, should not be included in national totals. U.S. energy consumption statistics include these bunker fuels (e.g., distillate fuel oil, residual fuel oil, and jet fuel) as part of consumption by the transportation end-use sector, however, so emissions from international transport activities were calculated separately following the same procedures used to calculate emissions from consumption of all fossil fuels (i.e., estimation of consumption, and determination of carbon content).[39] The Office of the Under Secretary of Defense (Installations and Environment) and the Defense Logistics Agency Energy (DLA Energy) of the U.S. Department of Defense (DoD) (DLA Energy 2022) supplied data on military jet fuel and marine fuel use. Commercial jet fuel use was estimated based on data from FAA (2022) and DOT (1991 through 2022); residual and distillate fuel use for civilian marine bunkers was obtained from DOC (1991 through 2022) for 1990 through 2001 and 2007 through 2020, and DHS (2008) for 2003 through 2006.[40] Consumption of these fuels was subtracted from the corresponding fuels totals in the transportation end-use sector. Estimates of international bunker fuel emissions for the United States are discussed in detail in Section 3.9 – International Bunker Fuels.

7. *Determine the total carbon content of fuels consumed.* Total C was estimated by multiplying the amount of fuel consumed by the amount of C in each fuel. This total C estimate defines the maximum amount of C that could potentially be released to the atmosphere if all of the C in each fuel was converted to CO_2. A discussion of the methodology and sources used to develop the C content coefficients are presented in Annexes 2.1 and 2.2.

8. *Estimate CO_2 Emissions.* Total CO_2 emissions are the product of the adjusted energy consumption (from the previous methodology steps 1 through 6), the carbon content of the fuels consumed, and the fraction of C that is oxidized. The fraction oxidized was assumed to be 100 percent for petroleum, coal, and natural gas based on guidance in IPCC (2006) (see Annex 2.1). Carbon emissions were multiplied by the molecular-to-atomic weight ratio of CO_2 to C (44/12) to obtain total CO_2 emitted from fossil fuel combustion in million metric tons (MMT).

9. *Allocate transportation emissions by vehicle type.* This report provides a more detailed accounting of emissions from transportation because it is such a large consumer of fossil fuels in the United States. For fuel types other than jet fuel, fuel consumption data by vehicle type and transportation mode were used to allocate emissions by fuel type calculated for the transportation end-use sector. Heat contents and densities were obtained from EIA (2022d) and USAF (1998).[41]

    - For on-road vehicles, annual estimates of combined motor gasoline and diesel fuel consumption by vehicle category were obtained from FHWA (1996 through 2021); for each vehicle category, the

[39] See International Bunker Fuels section in this chapter for a more detailed discussion.

[40] Data for 2002 were interpolated due to inconsistencies in reported fuel consumption data.

[41] For a more detailed description of the data sources used for the analysis of the transportation end use sector see the Mobile Combustion (excluding CO_2) and International Bunker Fuels sections of the Energy chapter, Annex 3.2, and Annex 3.8, respectively.

percent gasoline, diesel, and other (e.g., CNG, LPG) fuel consumption are estimated using data from EPA's MOVES model and DOE (1993 through 2022). [42,43]

- For non-road vehicles, activity data were obtained from AAR (2008 through 2022), APTA (2007 through 2021), APTA (2006), BEA (1991 through 2015), Benson (2002 through 2004), DLA Energy (2022), DOC (1991 through 2022), DOE (1993 through 2022), DOT (1991 through 2022), EIA (2009a), EIA (2022e), EIA (2002), EIA (1991 through 2022), EPA (2022c),[44] and Gaffney (2007).

- For jet fuel used by aircraft, CO_2 emissions from commercial aircraft were developed by the U.S. Federal Aviation Administration (FAA) using a Tier 3B methodology, consistent IPCC (2006) (see Annex 3.3). Carbon dioxide emissions from other aircraft were calculated directly based on reported consumption of fuel as reported by EIA. Allocation to domestic military uses was made using DoD data (see Annex 3.8). General aviation jet fuel consumption is calculated as the remainder of total jet fuel use (as determined by EIA) nets all other jet fuel use as determined by FAA and DoD. For more information, see Annex 3.2.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2021. Due to data availability and sources, some adjustments outlined in the methodology above are not applied consistently across the full 1990 to 2021 time series. As described in greater detail in Annex 2.1, to align with EIA's methodology for calculating motor gasoline consumption, petroleum denaturant adjustments are applied to motor gasoline consumption only for the period 1993 through 2008. In addition to ensuring time-series consistency, to ensure consistency in reporting between the Inventory and the Canadian National Greenhouse Gas Inventory, the amount of associated fossil fuel (lignite coal) that is gasified to create the exported CO_2 from the Dakota Gasification Plant is adjusted to align with the Canadian National Greenhouse Gas Inventory (Environment and Climate Change Canada 2022). This adjustment is explained in greater detail in Annex 2.1. As discussed in Annex 5, data are unavailable to include estimates of CO_2 emissions from any liquid fuel used in pipeline transport or non-hazardous industrial waste incineration, but those emissions are assumed to be insignificant.

Box 3-4: Carbon Intensity of U.S. Energy Consumption

The amount of C emitted from the combustion of fossil fuels is dependent upon the carbon content of the fuel and the fraction of that C that is oxidized. Fossil fuels vary in their average carbon content, ranging from about 53 MMT CO_2 Eq./QBtu for natural gas to upwards of 95 MMT CO_2 Eq./QBtu for coal and petroleum coke (see Tables A-42 and A-43 in Annex 2.1 for carbon contents of all fuels). In general, the carbon content per unit of energy of fossil fuels is the highest for coal products, followed by petroleum, and then natural gas. The overall carbon intensity of the U.S. economy is thus dependent upon the quantity and combination of fuels and other energy sources employed to meet demand.

Table 3-16 provides a time series of the carbon intensity of direct emissions for each sector of the U.S. economy. The time series incorporates only the energy from the direct combustion of fossil fuels in each sector. For

[42] On-road fuel consumption data from FHWA Table MF-21 and MF-27 were used to determine total on-road use of motor gasoline and diesel fuel (FHWA 1996 through 2020). Data for 2021 is proxied using FHWA Traffic Volume Travel Trends. Ratios developed from MOVES3 output are used to apportion FHWA fuel consumption data to vehicle type and fuel type (see Annex 3.2 for information about the MOVES model).

[43] Transportation sector natural gas and LPG consumption are based on data from EIA (2022a). In previous Inventory years, data from DOE (1993 through 2022) TEDB was used to estimate each vehicle class's share of the total natural gas and LPG consumption. Since TEDB does not include estimates for natural gas use by medium- and heavy-duty trucks or LPG use by passenger cars, EIA Alternative Fuel Vehicle Data (Browning 2017) is now used to determine each vehicle class's share of the total natural gas and LPG consumption. These changes were first incorporated in the 1990 through 2015 Inventory and apply to the time period from 1990 to 2015.

[44] In 2014, EPA incorporated the NONROAD2008 model into the MOVES model framework (EPA 2022c). The current Inventory uses the Nonroad component of MOVES3 for years 1999 through 2021.

example, the carbon intensity for the residential sector does not include the energy from or emissions related to the use of electricity for lighting, as it is instead allocated to the electric power sector. For the purposes of maintaining the focus of this section, renewable energy and nuclear energy are not included in the energy totals used in Table 3-16 in order to focus attention on fossil fuel combustion as detailed in this chapter. Looking only at this direct consumption of fossil fuels, the residential sector exhibited the lowest carbon intensity, which is related to the large percentage of its energy derived from natural gas for heating. The carbon intensity of the commercial sector has predominantly declined since 1990 as commercial businesses shift away from petroleum to natural gas. The industrial sector was more dependent on petroleum and coal than either the residential or commercial sectors, and thus had higher C intensities over this period. The carbon intensity of the transportation sector was closely related to the carbon content of petroleum products (e.g., motor gasoline and jet fuel, both around 70 MMT CO_2 Eq./QBtu), which were the primary sources of energy. Lastly, the electric power sector had the highest carbon intensity due to its heavy reliance on coal for generating electricity.

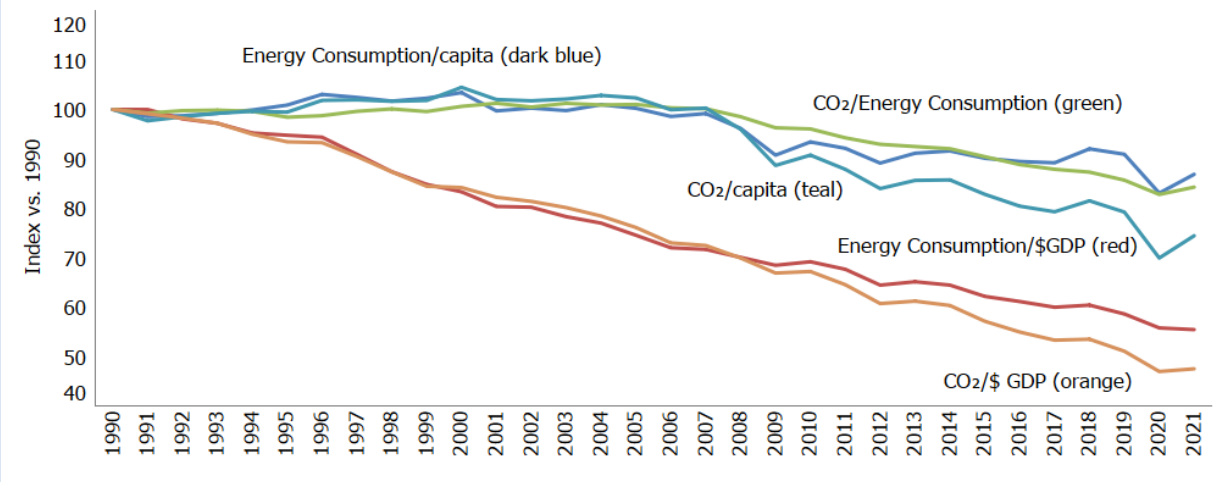
Table 3-16: Carbon Intensity from Direct Fossil Fuel Combustion by Sector (MMT CO_2 Eq./QBtu)

Sector	1990	2005	2017	2018	2019	2020	2021
Residential[a]	57.4	56.8	55.1	55.3	55.2	55.1	54.8
Commercial[a]	59.7	57.8	56.6	56.0	56.1	56.2	55.5
Industrial[a]	64.6	64.7	60.8	60.5	60.3	59.8	59.4
Transportation[a]	71.1	71.5	71.2	71.0	70.9	70.8	70.9
Electric Power[b]	87.3	85.8	77.3	75.5	72.9	70.5	72.3
U.S. Territories[c]	73.1	73.4	71.0	70.4	70.8	71.6	71.5
All Sectors[c]	**73.1**	**73.6**	**69.1**	**68.3**	**67.3**	**66.3**	**67.0**

[a] Does not include electricity or renewable energy consumption.
[b] Does not include electricity produced using nuclear or renewable energy.
[c] Does not include nuclear or renewable energy consumption.
Note: Excludes non-energy fuel use emissions and consumption.

For the time period of 1990 through about 2008, the carbon intensity of U.S. energy consumption was fairly constant, as the proportion of fossil fuels used by the individual sectors did not change significantly over that time. Starting in 2008 the carbon intensity has decreased, reflecting the shift from coal to natural gas in the electric power sector during that time period. Per capita energy consumption fluctuated little from 1990 to 2007, but then started decreasing after 2007 and, in 2021, was approximately 13.1 percent below levels in 1990 (see Figure 3-18). To differentiate these estimates from those of Table 3-16, the carbon intensity trend shown in Figure 3-18 and described below includes nuclear and renewable energy EIA data to provide a comprehensive economy-wide picture of energy consumption. Due to a general shift from a manufacturing-based economy to a service-based economy, as well as overall increases in efficiency, energy consumption and energy-related CO_2 emissions per dollar of gross domestic product (GDP) have both declined since 1990 (BEA 2022).

Figure 3-18: U.S. Energy Consumption and Energy-Related CO_2 Emissions Per Capita and Per Dollar GDP



Carbon intensity estimates were developed using nuclear and renewable energy data from EIA (2022d), EPA (2010), and fossil fuel consumption data as discussed above and presented in Annex 2.1.

Uncertainty

For estimates of CO_2 from fossil fuel combustion, the amount of CO_2 emitted is directly related to the amount of fuel consumed, the fraction of the fuel that is oxidized, and the carbon content of the fuel. Therefore, a careful accounting of fossil fuel consumption by fuel type, average carbon contents of fossil fuels consumed, and production of fossil fuel-based products with long-term carbon storage should yield an accurate estimate of CO_2 emissions.

Nevertheless, there are uncertainties in the consumption data, carbon content of fuels and products, and carbon oxidation efficiencies. For example, given the same primary fuel type (e.g., coal, petroleum, or natural gas), the amount of carbon contained in the fuel per unit of useful energy can vary. For the United States, however, the impact of these uncertainties on overall CO_2 emission estimates is believed to be relatively small. See, for example, Marland and Pippin (1990). See also Annex 2.2 for a discussion of uncertainties associated with fuel carbon contents. Recent updates to carbon factors for natural gas and coal utilized the same approach as previous Inventories with updated recent data, therefore, the uncertainty estimates around carbon contents of the different fuels as outlined in Annex 2.2 were not impacted and the historic uncertainty ranges still apply.

Although national statistics of total fossil fuel and other energy consumption are relatively accurate, the allocation of this consumption to individual end-use sectors (i.e., residential, commercial, industrial, and transportation) is less certain. For example, for some fuels the sectoral allocations are based on price rates (i.e., tariffs), but a commercial establishment may be able to negotiate an industrial rate or a small industrial establishment may end up paying an industrial rate, leading to a misallocation of emissions. Also, the deregulation of the natural gas industry and the more recent deregulation of the electric power industry have likely led to some minor challenges in collecting accurate energy statistics as firms in these industries have undergone significant restructuring.

To calculate the total CO_2 emission estimate from energy-related fossil fuel combustion, the amount of fuel used in non-energy production processes were subtracted from the total fossil fuel consumption. The amount of CO_2 emissions resulting from non-energy related fossil fuel use has been calculated separately and reported in the Carbon Emitted from Non-Energy Uses of Fossil Fuels section of this report (Section 3.2). These factors all contribute to the uncertainty in the CO_2 estimates. Detailed discussions on the uncertainties associated with C emitted from Non-Energy Uses of Fossil Fuels can be found within that section of this chapter.

Various sources of uncertainty surround the estimation of emissions from international bunker fuels, which are subtracted from the U.S. totals (see the detailed discussions on these uncertainties provided in Section 3.9 – International Bunker Fuels). Another source of uncertainty is fuel consumption by U.S. Territories. The United States does not collect energy statistics for its territories at the same level of detail as for the fifty states and the District of Columbia. Therefore, estimating both emissions and bunker fuel consumption by these territories is difficult.

Uncertainties in the emission estimates presented above also result from the data used to allocate CO_2 emissions from the transportation end-use sector to individual vehicle types and transport modes. In many cases, bottom-up estimates of fuel consumption by vehicle type do not match aggregate fuel-type estimates from EIA. Further research is planned to improve the allocation into detailed transportation end-use sector emissions.

The uncertainty analysis was performed by primary fuel type for each end-use sector, using the IPCC-recommended Approach 2 uncertainty estimation methodology, Monte Carlo Stochastic Simulation technique, with @RISK software. For this uncertainty estimation, the inventory estimation model for CO_2 from fossil fuel combustion was integrated with the relevant variables from the inventory estimation model for International Bunker Fuels, to realistically characterize the interaction (or endogenous correlation) between the variables of these two models. About 170 input variables were modeled for CO_2 from energy-related Fossil Fuel Combustion (including about 20 for non-energy fuel consumption and about 20 for International Bunker Fuels).

In developing the uncertainty estimation model, uniform distributions were assumed for all activity-related input variables and emission factors, based on the SAIC/EIA (2001) report.[45] Triangular distributions were assigned for the oxidization factors (or combustion efficiencies). The uncertainty ranges were assigned to the input variables based on the data reported in SAIC/EIA (2001) and on conversations with various agency personnel.[46]

The uncertainty ranges for the activity-related input variables were typically asymmetric around their inventory estimates; the uncertainty ranges for the emissions factors were symmetric. Bias (or systematic uncertainties) associated with these variables accounted for much of the uncertainties associated with these variables (SAIC/EIA 2001).[47] For purposes of this uncertainty analysis, each input variable was simulated 10,000 times through Monte Carlo sampling.

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-17. Fossil fuel combustion CO_2 emissions in 2021 were estimated to be between 4,553.9 and 4,856.1 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 2 percent below to 4 percent above the 2021 emission estimate of 4,651.0 MMT CO_2 Eq.

[45] SAIC/EIA (2001) characterizes the underlying probability density function for the input variables as a combination of uniform and normal distributions (the former to represent the bias component and the latter to represent the random component). However, for purposes of the current uncertainty analysis, it was determined that uniform distribution was more appropriate to characterize the probability density function underlying each of these variables.

[46] In the SAIC/EIA (2001) report, the quantitative uncertainty estimates were developed for each of the three major fossil fuels used within each end-use sector; the variations within the sub-fuel types within each end-use sector were not modeled. However, for purposes of assigning uncertainty estimates to the sub-fuel type categories within each end-use sector in the current uncertainty analysis, SAIC/EIA (2001)-reported uncertainty estimates were extrapolated.

[47] Although, in general, random uncertainties are the main focus of statistical uncertainty analysis, when the uncertainty estimates are elicited from experts, their estimates include both random and systematic uncertainties. Hence, both these types of uncertainties are represented in this uncertainty analysis.

Table 3-17: Approach 2 Quantitative Uncertainty Estimates for CO_2 Emissions from Energy-Related Fossil Fuel Combustion by Fuel Type and Sector (MMT CO_2 Eq. and Percent)

Fuel/Sector	2021 Emission Estimate (MMT CO_2 Eq.)	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(%)	
		Lower Bound	Upper Bound	Lower Bound	Upper Bound
Coal[b]	**957.7**	**925.0**	**1,048.0**	**-3%**	**9%**
Residential	NO	NO	NO	NO	NO
Commercial	1.4	1.4	1.7	-5%	15%
Industrial	43.7	41.6	50.6	-5%	16%
Transportation	NO	NO	NO	NO	NO
Electric Power	909.7	874.9	997.4	-4%	10%
U.S. Territories	2.9	2.5	3.4	-12%	19%
Natural Gas[b]	**1,620.7**	**1,600.4**	**1,695.0**	**-1%**	**5%**
Residential	258.6	251.3	276.8	-3%	7%
Commercial	180.9	175.8	193.6	-3%	7%
Industrial	498.4	482.0	535.2	-3%	7%
Transportation	65.1	63.3	69.7	-3%	7%
Electric Power	615.1	597.2	646.4	-3%	5%
U.S. Territories	2.6	2.3	3.1	-12%	17%
Petroleum[b]	**2,072.2**	**1,947.0**	**2,195.5**	**-6%**	**6%**
Residential	51.5	48.4	54.5	-6%	6%
Commercial	41.6	39.4	43.5	-5%	5%
Industrial	220.3	166.7	272.0	-24%	23%
Transportation	1,724.3	1,616.1	1,831.2	-6%	6%
Electric Power	17.1	16.2	18.5	-5%	8%
U.S. Territories	17.5	16.3	19.4	-7%	11%
Total (excluding Geothermal)[b]	**4,650.6**	**4,553.4**	**4,855.6**	**-2%**	**4%**
Geothermal	**0.4**	**NE**	**NE**	**NE**	**NE**
Electric Power	0.4	NE	NE	NE	NE
Total (including Geothermal)[b,c]	**4,651.0**	**4,553.9**	**4,856.1**	**-2%**	**4%**

NO (Not Occurring)
NE (Not Estimated)
[a] Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval.
[b] The low and high estimates for total emissions were calculated separately through simulations and, hence, the low and high emission estimates for the sub-source categories do not sum to total emissions.
[c] Geothermal emissions added for reporting purposes, but an uncertainty analysis was not performed for CO_2 emissions from geothermal production.
Note: Totals may not sum due to independent rounding.

QA/QC and Verification

In order to ensure the quality of the CO_2 emission estimates from fossil fuel combustion, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and methodology used for estimating CO_2 emissions from fossil fuel combustion in the United States. Emission totals for the different sectors and fuels were compared and trends were investigated to determine whether any corrective actions were needed. Minor corrective actions were taken.

One area of QA/QC and verification is to compare the estimates and emission factors used in the Inventory with other sources of CO_2 emissions reporting. Two main areas and sources of data were considered. The first is a comparison with the EPA GHGRP combustion data (subpart C) for stationary combustion sources excluding the electric power sector. This mainly focused on considering carbon factors for natural gas. The second comparison is with the EPA Air Markets Program data for electric power production. This considered carbon factors for coal and natural gas used in electric power production.

The EPA GHGRP collects greenhouse gas emissions data from large emitters including information on fuel combustion. This excludes emissions from mobile sources and smaller residential and commercial sources, those emissions are covered under supplier reporting (subparts MM and NN) and are areas for further research. Fuel combustion CO_2 data reported in 2021 was 2,084.0 MMT CO_2. Of that, 1,581.4 MMT CO_2 was from electricity production. Therefore, the non electric power production fuel combustion reporting was a fraction of the total covered by the Inventory under fossil fuel combustion. Furthermore, reporters under the GHGRP can use multiple methods of calculating emissions; one method is to use the default emission factors provided in the rule, while another is based on a tier 3 approach using their own defined emission factors. Based on data from reporters on approach used, it was determined that only about 10 percent of natural gas combustion emissions were based on a tier 3 approach. Given the small sample size compared to the overall Inventory calculations for natural gas combustion EPA determined it was not reasonable to consider the GHGRP tier 3 natural gas factors at this time.

EPA collects detailed sulfur dioxide (SO_2), nitrogen oxides (NO_X), and carbon dioxide (CO_2) emissions data and other information from power plants across the country as part of the Acid Rain Program (ARP), the Cross-State Air Pollution Rule (CSAPR), the CSAPR Update, and the Revised CSAPR Update (RCU). The CO_2 data from these Air Market Programs (AMP) can be compared to the electric power sector emissions calculated from the Inventory as shown in Table 3-18 for the three most recent years of data.

Table 3-18: Comparison of Electric Power Sector Emissions (MMT CO_2 Eq. and Percent)

	CO_2 Emissions (MMT CO_2 Eq.)			% Change	
Fuel/Sector	**2019**	**2020**	**2021**	**19-20**	**20-21**
Inventory Electric Power Sector	**1,606.7**	**1,439.0**	**1,542.22**	**-10.4%**	**7.1%**
Coal	973.5	788.2	909.7	-19.0%	15.4%
Natural Gas	616.6	634.3	615.1	3.0%	-3.1%
Petroleum	16.2	16.2	17.1	0.0%	5.6%
AMP Electric Power Sector	**1,605.4**	**1,437.7**	**1,538.6**	**-10.4%**	**7.0%**
Coal	980.9	796.3	917.2	-18.8%	15.2%
Natural Gas	616.4	632.6	612.7	2.6%	-3.2%
Petroleum	8.1	8.8	8.7	7.8%	-0.6%

Note: Totals may not sum due to independent rounding.

In general the emissions and trends from the two sources line up well. There are differences expected based on coverage and scope of each source. The Inventory covers all emissions from the electric power sector as defined above. The EPA AMP data covers emissions from electricity generating units of a certain size so in some respects it could cover more sources (like electric power units at industrial facilities that would be covered under the industrial sector in the Inventory) and not as many sources (since smaller units are excluded). The EPA AMP data also includes heat input for different fuel types. That data can be combined with emissions to calculate implied emission factors.[48] The following Table 3-19 shows the implied emissions factors for coal and natural gas from the EPA AMP data compared to the factors used in the Inventory for the three most recent years of data.

Table 3-19: Comparison of Emissions Factors (MMT Carbon/QBtu)

Fuel Type	**2019**	**2020**	**2021**
EPA AMP			
Coal	25.52	25.52	25.55
Natural Gas	14.43	14.47	14.50

[48] These emission factors can be converted from MMT Carbon/QBtu to MMT CO_2 Eq./QBtu by multiplying the emission factor by 44/12, the molecular-to-atomic weight ratio of CO_2 to C. This would assume the fraction oxidized to be 100 percent, which is the guidance in IPCC (2006) (see Annex 2.1).

EPA Inventory			
Electric Power Coal	26.08	26.12	26.13
Natural Gas	14.43	14.43	14.43

The factors for natural gas line up reasonably well. For coal the EPA emissions factors are roughly 2 percent higher than those calculated from the EPA AMP data. One possible reason for the difference is that the EPA Inventory factors are based on all coal used in electric power production while the factors from the EPA AMP data are based on only units where coal is the only source of fuel used. There are units that use coal and other fuel sources but emissions for each foul type could not be calculated. This is an area of further research but given current data available the approach to develop carbon factors as outlined in Annex 2 is still felt to be the most appropriate to represent total fuel combustion in the United States.

The UNFCCC reporting guidelines also require countries to complete a "top-down" reference approach for estimating CO_2 emissions from fossil fuel combustion in addition to their “bottom-up” sectoral methodology. The reference approach (detailed in Annex 4) uses alternative methodologies and different data sources than those contained in this section of the report. The reference approach estimates fossil fuel consumption by adjusting national aggregate fuel production data for imports, exports, and stock changes rather than relying on end-user consumption surveys. The reference approach assumes that once carbon-based fuels are brought into a national economy, they are either saved in some way (e.g., stored in products, kept in fuel stocks, or left unoxidized in ash) or combusted, and therefore the carbon in them is oxidized and released into the atmosphere. In the reference approach, accounting for actual consumption of fuels at the sectoral or sub-national level is not required. One difference between the two approaches is that emissions from carbon that was not stored during non-energy use of fuels are subtracted from the sectoral approach and reported separately (see Section 3.2). These emissions, however, are not subtracted in the reference approach. As a result, the reference approach emission estimates are comparable to those of the sectoral approach, with the exception that the Non-Energy Use (NEU) source category emissions are included in the reference approach (see Annex 4 for more details).

Recalculations Discussion

Several updates to activity data and emission factors lead to recalculations of previous year results. The major updates are as follows:

- EIA (2022a) updated energy consumption statistics across the time series relative to the previous Inventory. This includes an update to transportation sector propane consumption data post 2010.
- EIA (2022a) updated industrial energy sector activity data post 2010 relative to the previous Inventory. This caused the annually variable carbon contents for HGL (energy use) and HGL (non-energy use) to be updated across the time series, because post 2010 data is used to back-cast data for prior years. EIA (2022a) updated petroleum statistics in coordination with its Petroleum Supply Annual 2021. This impacted the HGL category across the time series.
- EPA revised territories data to correct for an error in how LPG data was pulled. The values for LPG were previously referencing the values for Other Petroleum from the EIA’s International Energy Statistics (EIA 2022b) and have been corrected to reflect the values for Liquified Petroleum Gas from the same source.
- Natural gas consumption data from EIA’s *Monthly Energy Review* (EIA 2022a) Table 10b was updated, which impacted years 2018-2020.
- The carbon content for propylene was updated from 65.95 kg CO_2/MMBtu to 67.77 kg CO_2/MMBtu to reflect values used in the EPA Greenhouse Gas Emission Factors Hub.
- Fuel consumption changes for the U.S. Territories provided by EIA’s International Energy Statistics (EIA 2022b) was updated across the time series.
- Updated values of natural gas used for ammonia production across the time series relative to the previous Inventory.

All of the revisions discussed above resulted in the following impacts on emissions over time:

- From 1990 to 2020, petroleum emissions from the residential sector decreased by an average annual amount of 0.09 MMT CO_2 Eq. (less than half a percent). Petroleum emissions from the commercial, industrial, and transportation sectors increased by an average annual amount of 0.05 MMT CO_2 Eq. (less than half a percent), 0.15 MMT CO_2 Eq. (less than half a percent), and 0.01 MMT CO_2 Eq. (less than half a percent), respectively. These changes are due to changes in EIA consumption statistics for petroleum, changes in EIA industrial energy sector activity data, and the change in carbon content for propylene.

- Petroleum emissions from U.S. Territories decreased by an average annual amount of 1.82 MMT CO_2 Eq. (5.51 percent) due to the correction in data pulled for LPG from 1990 to 2020, change in carbon content for propylene, and change in fuel consumption data for U.S. Territories.
- Natural gas emissions across the residential, commercial, transportation, and electric power sectors for years 2018-2020 increased by an average annual amount of 0.19 MMT CO_2 Eq. (less than half a percent) due to an update in natural gas consumption for these sectors in EIA's *Monthly Energy Review* (EIA 2022a) Table 10b.
- Natural gas emissions for the industrial sector from 1990-2017 decreased by an average annual amount of 1.00 MMT CO_2 Eq. (less than half a percent) due to an update in the correction for natural gas used for ammonia production. Natural gas emissions for the industrial sector from 2018-2020 decreased by 0.03 MMT CO_2 Eq. (less than half a percent) due to updates to both ammonia production and MER table 10b.
- Coal emissions from U.S. Territories decreased by an average annual amount of less than 0.01 MMT CO_2 Eq. (less than half a percent) due to the change in fuel consumption data for U.S. Territories.

Overall, these changes resulted in an average annual decrease of 2.5 MMT CO_2 Eq. (less than 0.05 percent) in CO_2 emissions from fossil fuel combustion for the period 1990 through 2020, relative to the previous Inventory. However, there were bigger absolute changes across the time series as discussed above.

Planned Improvements

To reduce uncertainty of CO_2 from fossil fuel combustion estimates for U.S. Territories, further expert elicitation may be conducted to better quantify the total uncertainty associated with emissions from U.S. Territories. Additionally, although not technically a fossil fuel, since geothermal energy-related CO_2 emissions are included for reporting purposes, further expert elicitation may be conducted to better quantify the total uncertainty associated with CO_2 emissions from geothermal energy use.

The availability of facility-level combustion emissions through EPA's GHGRP will continue to be examined to help better characterize the industrial sector's energy consumption in the United States and further classify total industrial sector fossil fuel combustion emissions by business establishments according to industrial economic activity type. Most methodologies used in EPA's GHGRP are consistent with IPCC methodologies, though for EPA's GHGRP, facilities collect detailed information specific to their operations according to detailed measurement standards, which may differ with the more aggregated data collected for the Inventory to estimate total, national U.S. emissions. In addition, and unlike the reporting requirements for this chapter under the UNFCCC reporting guidelines, some facility-level fuel combustion emissions reported under the GHGRP may also include industrial process emissions.[49] In line with UNFCCC reporting guidelines, fuel combustion emissions are included in this chapter, while process emissions are included in the Industrial Processes and Product Use chapter of this report. In examining data from EPA's GHGRP that would be useful to improve the emission estimates for the CO_2 from fossil fuel combustion category, particular attention will also be made to ensure time-series consistency, as the facility-level reporting data from EPA's GHGRP are not available for all inventory years as reported in this Inventory.

Additional analyses will be conducted to align reported facility-level fuel types and IPCC fuel types per the national energy statistics. For example, additional work will look at CO_2 emissions from biomass to ensure they are separated in the facility-level reported data and maintaining consistency with national energy statistics provided by EIA. In implementing improvements and integration of data from EPA's GHGRP, the latest guidance from the IPCC on the use of facility-level data in national inventories will continue to be relied upon.[50]

[49] See https://unfccc.int/resource/docs/2013/cop19/eng/10a03.pdf#page=2.

[50] See http://www.ipcc-nggip.iges.or.jp/public/tb/TFI_Technical_Bulletin_1.pdf.

An ongoing planned improvement is to develop improved estimates of domestic waterborne fuel consumption. The Inventory estimates for residual and distillate fuel used by ships and boats is based in part on data on bunker fuel use from the U.S. Department of Commerce. Domestic fuel consumption is estimated by subtracting fuel sold for international use from the total sold in the United States. It may be possible to more accurately estimate domestic fuel use and emissions by using detailed data on marine ship activity. The feasibility of using domestic marine activity data to improve the estimates will continue to be investigated.

EPA is also evaluating the methods used to adjust for conversion of fuels and exports of CO_2. EPA is exploring the approach used to account for CO_2 transport, injection, and geologic storage, as part of this there may be changes made to accounting for CO_2 exports.

Finally, another ongoing planned improvement is to evaluate data availability to update the carbon and heat content of more fuel types accounted for in this Inventory. This update will impact consumption and emissions across all sectors and will improve consistency with EIA data as carbon and heat contents of fuels will be accounted for as annually variable and therefore improve accuracy across the time series. Some of the fuels considered in this effort include petroleum coke, residual fuel, and woody biomass.

CH_4 and N_2O from Stationary Combustion

Methodology and Time-Series Consistency

Methane and N_2O emissions from stationary combustion were estimated by multiplying fossil fuel and wood consumption data by emission factors (by sector and fuel type for industrial, residential, commercial, and U.S. Territories; and by fuel and technology type for the electric power sector). The electric power sector utilizes a Tier 2 methodology, whereas all other sectors utilize a Tier 1 methodology. The activity data and emission factors used are described in the following subsections.

More detailed information on the methodology for calculating emissions from stationary combustion, including emission factors and activity data, is provided in Annex 3.1.

Industrial, Residential, Commercial, and U.S. Territories

National coal, natural gas, fuel oil, and wood consumption data were grouped by sector: industrial, commercial, residential, and U.S. Territories. For the CH_4 and N_2O emission estimates, consumption data for each fuel were obtained from EIA's *Monthly Energy Review* (EIA 2022a). Because the United States does not include territories in its national energy statistics, fuel consumption data for territories were provided separately by EIA's International Energy Statistics (EIA 2022b).[51] Fuel consumption for the industrial sector was adjusted to subtract out mobile source construction and agricultural use, which is reported under mobile sources. Construction and agricultural mobile source fuel use was obtained from EPA (2022b) and FHWA (1996 through 2022). Estimates for wood biomass consumption for fuel combustion do not include municipal solid waste, tires, etc., that are reported as biomass by EIA. Non-CO_2 emissions from combustion of the biogenic portion of municipal solid waste and tires is included under waste incineration (Section 3.2). Estimates for natural gas combustion do not include biogas, and therefore non-CO_2 emissions from biogas are not included (see the Planned Improvements section, below). Tier 1 default emission factors for the industrial, commercial, and residential end-use sectors were provided by the *2006 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC 2006). U.S. Territories' emission factors were estimated using the U.S. emission factors for the primary sector in which each fuel was combusted.

[51] U.S. Territories data also include combustion from mobile activities because data to allocate territories' energy use were unavailable. For this reason, CH_4 and N_2O emissions from combustion by U.S. Territories are only included in the stationary combustion totals.

Electric Power Sector

The electric power sector uses a Tier 2 emission estimation methodology as fuel consumption for the electric power sector by control-technology type was based on EPA's Acid Rain Program Dataset (EPA 2022a). Total fuel consumption in the electric power sector from EIA (2022a) was apportioned to each combustion technology type and fuel combination using a ratio of fuel consumption by technology type derived from EPA (2022a) data. The combustion technology and fuel use data by facility obtained from EPA (2022a) were only available from 1996 to 2020, so the consumption estimates from 1990 to 1995 were estimated by applying the 1996 consumption ratio by combustion technology type from EPA (2022a) to the total EIA (2022a) consumption for each year from 1990 to 1995.

Emissions were estimated by multiplying fossil fuel and wood consumption by technology-, fuel-, and country-specific Tier 2 emission factors. The Tier 2 emission factors used are based in part on emission factors published by EPA, and EPA's Compilation of Air Pollutant Emission Factors, AP-42 (EPA 1997) for coal wall-fired boilers, residual fuel oil, diesel oil and wood boilers, natural gas-fired turbines, and combined cycle natural gas units.[52]

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2021 as discussed below. As discussed in Annex 5, data are unavailable to include estimates of CH_4 and N_2O emissions from biomass use in Territories, but those emissions are assumed to be insignificant.

Uncertainty

Methane emission estimates from stationary sources exhibit high uncertainty, primarily due to difficulties in calculating emissions from wood combustion (i.e., fireplaces and wood stoves). The estimates of CH_4 and N_2O emissions presented are based on broad indicators of emissions (i.e., fuel use multiplied by an aggregate emission factor for different sectors), rather than specific emission processes (i.e., by combustion technology and type of emission control).

An uncertainty analysis was performed by primary fuel type for each end-use sector, using the IPCC-recommended Approach 2 uncertainty estimation methodology, Monte Carlo Stochastic Simulation technique, with @RISK software.

The uncertainty estimation model for this source category was developed by integrating the CH_4 and N_2O stationary source inventory estimation models with the model for CO_2 from fossil fuel combustion to realistically characterize the interaction (or endogenous correlation) between the variables of these three models. About 55 input variables were simulated for the uncertainty analysis of this source category (about 20 from the CO_2 emissions from fossil fuel combustion inventory estimation model and about 35 from the stationary source inventory models).

In developing the uncertainty estimation model, uniform distribution was assumed for all activity-related input variables and N_2O emission factors, based on the SAIC/EIA (2001) report.[53] For these variables, the uncertainty ranges were assigned to the input variables based on the data reported in SAIC/EIA (2001).[54] However, the CH_4

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[52] Several of the U.S. Tier 2 emission factors were used in IPCC (2006) as Tier 1 emission factors. See Table A-69 in Annex 3.1 for emission factors by technology type and fuel type for the electric power sector.

[53] SAIC/EIA (2001) characterizes the underlying probability density function for the input variables as a combination of uniform and normal distributions (the former distribution to represent the bias component and the latter to represent the random component). However, for purposes of the current uncertainty analysis, it was determined that uniform distribution was more appropriate to characterize the probability density function underlying each of these variables.

[54] In the SAIC/EIA (2001) report, the quantitative uncertainty estimates were developed for each of the three major fossil fuels used within each end-use sector; the variations within the sub-fuel types within each end-use sector were not modeled. However, for purposes of assigning uncertainty estimates to the sub-fuel type categories within each end-use sector in the current uncertainty analysis, SAIC/EIA (2001)-reported uncertainty estimates were extrapolated.

emission factors differ from those used by EIA. These factors and uncertainty ranges are based on IPCC default uncertainty estimates (IPCC 2006).

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-20. Stationary combustion CH_4 emissions in 2021 (including biomass) were estimated to be between 5.8 and 20.3 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 35 percent below to 129 percent above the 2021 emission estimate of 8.9 MMT CO_2 Eq.[55] Stationary combustion N_2O emissions in 2021 (including biomass) were estimated to be between 16.3 and 33.2 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 26 percent below to 50 percent above the 2021 emission estimate of 22.1 MMT CO_2 Eq.

Table 3-20: Approach 2 Quantitative Uncertainty Estimates for CH_4 and N_2O Emissions from Energy-Related Stationary Combustion, Including Biomass (MMT CO_2 Eq. and Percent)

Source	Gas	2021 Emission Estimate (MMT CO_2 Eq.)	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Stationary Combustion	CH_4	8.9	5.8	20.3	-35%	129%
Stationary Combustion	N_2O	22.1	16.3	33.2	-26%	50%

[a] Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval.

The uncertainties associated with the emission estimates of CH_4 and N_2O are greater than those associated with estimates of CO_2 from fossil fuel combustion, which mainly rely on the carbon content of the fuel combusted. Uncertainties in both CH_4 and N_2O estimates are due to the fact that emissions are estimated based on emission factors representing only a limited subset of combustion conditions. For the indirect greenhouse gases, uncertainties are partly due to assumptions concerning combustion technology types, age of equipment, emission factors used, and activity data projections.

QA/QC and Verification

In order to ensure the quality of the non-CO_2 emission estimates from stationary combustion, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and emission factor sources and methodology used for estimating CH_4, N_2O, and the greenhouse gas precursors from stationary combustion in the United States. Emission totals for the different sectors and fuels were compared and trends were investigated.

Recalculations Discussion

EIA (2022a) updated petroleum statistics in coordination with its Petroleum Supply Annual 2021. This impacted the HGL category across the time series.

Fuel consumption data for U.S. Territories provided by EIA's International Energy Statistics (EIA 2022b) was updated across the timeseries. Non-CO_2 emissions from U.S Territories decreased by an average annual amount of less than 0.01 MMT CO_2 Eq. (less than half a percent) for coal and less than 0.01 MMT CO_2 Eq. (5.76 percent) for fuel oil due to the update in fuel consumption data for U.S. Territories.

[55] The low emission estimates reported in this section have been rounded down to the nearest integer values and the high emission estimates have been rounded up to the nearest integer values.

Wood and natural gas consumption data from EIA's *Monthly Energy Review* (EIA 2022a) Table 10b was updated, which impacted years 2018-2020. Non-CO_2 emissions across the residential, commercial, industrial, and electric power sectors decreased by an average annual amount of less than 0.04 MMT CO_2 Eq. (less than half a percent) for wood and increased by an average annual amount of less than 0.03 MMT CO_2 Eq. (less than half a percent) for natural gas due to the update in MER table 10b.

In addition, for the current Inventory, CO_2-equivalent emissions of CH_4 and N_2O from stationary combustion have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth Assessment Report* (AR4), used in previous Inventories (IPCC 2007). The AR5 GWPs have been applied across the entire time series for consistency. Prior inventories used GWPs of 25 and 298 for CH_4 and N_2O, respectively. These values have been updated to 28 and 265, respectively. Compared to the previous Inventory which applied 100-year GWP values from AR4, the average annual change in CO_2-equivalent CH_4 emissions was a 12 percent increase and the average annual change in CO_2-equivalent N_2O emissions was an 11 percent decrease for the time series. As a result of the change in methodology, total emissions across the timeseries changed by an average annual decrease of 2.3 MMT CO_2 Eq. (6.1 percent) relative to emissions results calculated using the prior GWPs. Further discussion on this update and the overall impacts of updating the Inventory GWP values to reflect the IPCC AR5 can be found in Chapter 9, Recalculations and Improvements.

Planned Improvements

Several items are being evaluated to improve the CH_4 and N_2O emission estimates from stationary combustion and to reduce uncertainty for U.S. Territories. Efforts will be taken to work with EIA and other agencies to improve the quality of the U.S. Territories data. Because these data are not broken out by stationary and mobile uses, further research will be aimed at trying to allocate consumption appropriately. In addition, the uncertainty of biomass emissions will be further investigated because it was expected that the exclusion of biomass from the estimates would reduce the uncertainty; and in actuality the exclusion of biomass increases the uncertainty. These improvements are not all-inclusive but are part of an ongoing analysis and efforts to continually improve these stationary combustion estimates from U.S. Territories.

Other forms of biomass-based gas consumption include biogas. As an additional planned improvement, EPA will examine EIA and GHGRP data on biogas collected and burned for energy use and determine if CH_4 and N_2O emissions from biogas can be included in future Inventories. EIA (2022a) natural gas data already deducts biogas used in the natural gas supply, so no adjustments are needed to the natural gas fuel consumption data to account for biogas.

CH_4 and N_2O from Mobile Combustion

Methodology and Time-Series Consistency

Estimates of CH_4 and N_2O emissions from mobile combustion were calculated by multiplying emission factors by measures of activity for each fuel and vehicle type (e.g., light-duty gasoline trucks). Activity data included vehicle miles traveled (VMT) for on-road vehicles and fuel consumption for non-road mobile sources. The activity data and emission factors used in the calculations are described in the subsections that follow. A complete discussion of the methodology used to estimate CH_4 and N_2O emissions from mobile combustion and the emission factors used in the calculations is provided in Annex 3.2.

On-Road Vehicles

Estimates of CH_4 and N_2O emissions from gasoline and diesel on-road vehicles are based on VMT and emission factors (in grams of CH_4 and N_2O per mile) by vehicle type, fuel type, model year, and emission control technology.

Emission estimates for alternative fuel vehicles (AFVs) are based on VMT and emission factors (in grams of CH_4 and N_2O per mile) by vehicle and fuel type.[56]

CH_4 and N_2O emissions factors by vehicle type and emission tier for newer (starting with model year 2004) on-road gasoline vehicles were calculated by Browning (2019) from annual vehicle certification data compiled by EPA. CH_4 and N_2O emissions factors for older (model year 2003 and earlier) on-road gasoline vehicles were developed by ICF (2004). These earlier emission factors were derived from EPA, California Air Resources Board (CARB) and Environment and Climate Change Canada (ECCC) laboratory test results of different vehicle and control technology types. The EPA, CARB and ECCC tests were designed following the Federal Test Procedure (FTP). The procedure covers three separate driving segments, since vehicles emit varying amounts of greenhouse gases depending on the driving segment. These driving segments are: (1) a transient driving cycle that includes cold start and running emissions, (2) a cycle that represents running emissions only, and (3) a transient driving cycle that includes hot start and running emissions. For each test run, a bag was affixed to the tailpipe of the vehicle and the exhaust was collected; the content of this bag was then analyzed to determine quantities of gases present. The emissions characteristics of driving segment 2 tests were used to define running emissions. Running emissions were subtracted from the total FTP emissions to determine start emissions. These were then recombined to approximate average driving characteristics, based upon the ratio of start to running emissions for each vehicle class from MOBILE6.2, an EPA emission factor model that predicts grams per mile emissions of CO_2, CO, HC, NO_x, and PM from vehicles under various conditions.[57]

Diesel on-road vehicle emission factors were developed by ICF (2006a). CH_4 and N_2O emissions factors for newer (starting with model year 2007) on-road diesel vehicles (those using engine aftertreatment systems) were calculated from annual vehicle certification data compiled by EPA.

CH_4 and N_2O emission factors for AFVs were developed based on the 2021 Greenhouse gases, Regulated Emissions, and Energy use in Transportation (GREET) model (ANL 2022). For light-duty trucks, EPA used travel fractions for LDT1 and LDT2 (MOVES Source Type 31 for LDT1 and MOVES Source Type 32 for LDT2; see Annex 3.2 for information about the MOVES model) to determine light-duty truck emission factors. For medium-duty vehicles, EPA used emission factors for light heavy-duty vocational trucks. For heavy-duty vehicles, EPA used emission factors for long-haul combination trucks. For buses, EPA used emission factors for transit buses. These values represent vehicle operations only (tank-to-wheels); upstream well-to-tank emissions are calculated elsewhere in the Inventory. Biodiesel CH_4 emission factors were corrected from GREET values to be the same as CH_4 emission factors for diesel vehicles. GREET overestimated biodiesel CH_4 emission factors based upon an incorrect CH_4-to-THC ratio for diesel vehicles with aftertreatment technology.

Annual VMT data for 1990 through 2020 were obtained from the Federal Highway Administration's (FHWA) Highway Performance Monitoring System database as reported in Highway Statistics (FHWA 1996 through 2020).[58] VMT estimates were then allocated to vehicle type using ratios of VMT per vehicle type to total VMT, derived from EPA's MOVES3 model (see Annex 3.2 for information about the MOVES model). This corrects time series inconsistencies in FHWA definitions of vehicle types (Browning 2022a). VMT for alternative fuel vehicles (AFVs) were estimated based on Browning (2022b). The age distributions of the U.S. vehicle fleet were obtained from EPA (2004, 2021b), and the average annual age-specific vehicle mileage accumulation of U.S. vehicles were obtained from EPA (2021b).

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[56] Alternative fuel and advanced technology vehicles are those that can operate using a motor fuel other than gasoline or diesel. This includes electric or other bi-fuel or dual-fuel vehicles that may be partially powered by gasoline or diesel.

[57] Additional information regarding the MOBILE model can be found at https://www.epa.gov/moves/description-and-history-mobile-highway-vehicle-emission-factor-model.

[58] Note that VMT for 2021 is estimated with FHWA Traffic Volume Trends data for this public review, but actual data for 2021 will be included in the Final Report when it is released.

Control technology and standards data for on-road vehicles were obtained from EPA's Office of Transportation and Air Quality (EPA 2021c, 2021d, and 1998) and Browning (2005). These technologies and standards are defined in Annex 3.2, and were compiled from EPA (1994a, 1994b, 1998, 1999a) and IPCC (2006) sources.

Non-Road Mobile Sources

The nonroad mobile category for CH_4 and N_2O includes ships and boats, aircraft, locomotives, and other mobile non-road sources (e.g., construction or agricultural equipment). For locomotives, aircraft, ships and non-recreational boats , fuel-based emission factors are applied to data on fuel consumption, following the IPCC Tier 1 approach, The Tier 2 approach for these sources would require separate fuel-based emissions factors by technology, for which data are not currently available. For other non-road sources, EPA uses the Nonroad component of the MOVES model to estimate fuel use. Emission factors by horsepower bin are estimated from EPA engine certification data. Because separate emission factors are applied to specific engine technologies; these non-road sources utilize a Tier 2 approach.

To estimate CH_4 and N_2O emissions from non-road mobile sources, fuel consumption data were employed as a measure of activity and multiplied by fuel-specific emission factors (in grams of N_2O and CH_4 per kilogram of fuel consumed). [59] Activity data were obtained from AAR (2008 through 2022), APTA (2007 through 2022), Rail Inc (2014 through 2022), APTA (2006), BEA (1991 through 2015), Benson (2002 through 2004), DLA Energy (2022), DOC (1991 through 2022), DOE (1993 through 2022), DOT (1991 through 2022), EIA (2002, 2007, 2022), EIA (2022f), EIA (1991 through 2022), EPA (2022b), Esser (2003 through 2004), FAA (2022), FHWA (1996 through 2022),[60] Gaffney (2007), and Whorton (2006 through 2014). Emission factors for non-road modes were taken from IPCC (2006) and Browning (2020a and 2018b).

Uncertainty

A quantitative uncertainty analysis was conducted for the mobile source sector using the IPCC-recommended Approach 2 uncertainty estimation methodology, Monte Carlo Stochastic Simulation technique, using @RISK software. The uncertainty analysis was performed on 2021 estimates of CH_4 and N_2O emissions, incorporating probability distribution functions associated with the major input variables. For the purposes of this analysis, the uncertainty was modeled for the following four major sets of input variables: (1) VMT data, by on-road vehicle and fuel type, (2) emission factor data, by on-road vehicle, fuel, and control technology type, (3) fuel consumption, data, by non-road vehicle and equipment type, and (4) emission factor data, by non-road vehicle and equipment type.

Uncertainty analyses were not conducted for NO_x, CO, or NMVOC emissions. Emission factors for these gases have been extensively researched because emissions of these gases from motor vehicles are regulated in the United States, and the uncertainty in these emission estimates is believed to be relatively low. For more information, see Section 3.11. However, a much higher level of uncertainty is associated with CH_4 and N_2O emission factors due to limited emission test data, and because, unlike CO_2 emissions, the emission pathways of CH_4 and N_2O are highly complex.

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[59] The consumption of international bunker fuels is not included in these activity data, but emissions related to the consumption of international bunker fuels are estimated separately under the International Bunker Fuels source category.

[60] This Inventory uses FHWA's Agriculture, Construction, and Commercial/Industrial MF-24 fuel volumes along with the MOVES model gasoline volumes to estimate non-road mobile source CH_4 and N_2O emissions for these categories. For agriculture, the MF-24 gasoline volume is used directly because it includes both non-road trucks and equipment. For construction and commercial/industrial category gasoline estimates, the 2014 and older MF-24 volumes represented non-road trucks only; therefore, the MOVES gasoline volumes for construction and commercial/industrial categories are added to the respective categories in the Inventory. Beginning in 2015, this addition is no longer necessary since the FHWA updated its methods for estimating on-road and non-road gasoline consumption. Among the method updates, FHWA now incorporates MOVES equipment gasoline volumes in the construction and commercial/industrial categories.

Based on the uncertainty analysis, mobile combustion CH_4 emissions from all mobile sources in 2021 were estimated to be 2.5 and 3.4 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 4 percent below to 30 percent above the corresponding 2021 emission estimate of 2.6 MMT CO_2 Eq. Mobile combustion N_2O emissions from mobile sources in 2021 were estimated to be between 16.0 and 20.8 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 7 percent below to 21 percent above the corresponding 2021 emission estimate of 17.2 MMT CO_2 Eq.

Table 3-21: Approach 2 Quantitative Uncertainty Estimates for CH_4 and N_2O Emissions from Mobile Sources (MMT CO_2 Eq. and Percent)

Source	Gas	2021 Emission Estimate (MMT CO_2 Eq.)	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(Percent)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Mobile Sources	CH_4	2.6	2.5	3.4	-4%	+30%
Mobile Sources	N_2O	17.2	16.0	20.8	-7%	+21%

[a] Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval.

This uncertainty analysis is a continuation of a multi-year process for developing quantitative uncertainty estimates for this source category using the IPCC Approach 2 uncertainty estimation methodology. As a result, as new information becomes available, uncertainty characterization of input variables may be improved and revised. For additional information regarding uncertainty in emission estimates for CH_4 and N_2O please refer to the Uncertainty Annex. As discussed in Annex 5, data are unavailable to include estimates of CH_4 and N_2O emissions from any liquid fuel used in pipeline transport or some biomass used in transportation sources, but those emissions are assumed to be insignificant.

QA/QC and Verification

In order to ensure the quality of the emission estimates from mobile combustion, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The specific plan used for mobile combustion was updated prior to collection and analysis of this current year of data. The Tier 2 procedures focused on the emission factor and activity data sources, as well as the methodology used for estimating emissions. These procedures included a qualitative assessment of the emission estimates to determine whether they appear consistent with the most recent activity data and emission factors available. A comparison of historical emissions between the current Inventory and the previous Inventory was also conducted to ensure that the changes in estimates were consistent with the changes in activity data and emission factors.

Recalculations Discussion

In previous inventories, on-highway greenhouse gas emissions were calculated using FHWA fuel consumption and vehicle miles traveled (VMT) data delineating by FHWA vehicle classes. These fuel consumption estimates were then combined with estimates of fuel shares by vehicle type from Oak Ridge National Laboratory's Transportation Energy Data Book (TEDB), to develop an estimate of fuel consumption for each vehicle type in the Inventory (i.e., passenger cars, light-duty trucks, buses, medium- and heavy-duty trucks, motorcycles). However, in 2011, FHWA changed its methods for estimating VMT and related data. These methodological changes included how vehicles are classified, moving from a system based on body-type to one that is based on wheelbase. These changes were first incorporated in the 1990 through 2008 Inventory and applied to the time series beginning in 2007. The FHWA methodology update resulted in large changes in VMT and fuel consumption by vehicle class, leading to a shift in emissions among vehicle classes. For example, FHWA replaced the vehicle category "Passenger Cars" with "Light-duty Vehicles-Short Wheelbase" and the "Other 2 axle-4 Tire Vehicles" category was replaced by "Light-duty Vehicles, Long Wheelbase." FHWA changed the definition of light-duty vehicles to less than 10,000 lbs. GVWR instead of 8,500 lbs. GVWR pushed some single-unit heavy-duty trucks to the light-duty class. This change in vehicle classification also moved some smaller trucks and sport utility vehicles from the light truck category to the passenger cars category in this Inventory. These updates resulted in a disconnect in FHWA VMT and fuel consumption data in the 2006 to 2007 timeframe, generating a large drop in the light-duty truck VMT and fuel consumption trend lines between 2006 and 2007, and a corresponding increase in the passenger cars trend lines.

To address this inconsistency in the time series, EPA updated the methodology to divide FHWA VMT data into vehicle classes and fuel type using distributions from EPA's Motor Vehicle Emission Simulator, MOVES. The MOVES model is a nationally recognized model based on vehicle registration, travel activity, and emission rates that are updated with each model release. MOVES3 is the latest version of MOVES and uses forecast growth factors which provide EPA's best estimate of likely future activity based on historical data (see Annex 3.2 for more information about the MOVES model). Thus, dividing FHWA total VMT data into vehicle class and fuel type using MOVES3 ratios provides a more consistent estimate of vehicle activity over the Inventory time series. MOVES3 ratios are also used to reallocate FHWA gasoline and diesel fuel use data (Browning 2022a). For this update, the MOVES3 model was run for calendar years 1990 and 1999 through 2021 for all vehicle types. Calendar years 1991 through 1998 were linearly interpolated from 1990 and 1999 calendar year MOVES3 outputs. Model outputs of VMT and fuel consumption were binned by calendar year, MOVES vehicle type, and fuel type; MOVES vehicle types were then mapped to the vehicle types used in the Inventory. Only outputs of gasoline and diesel fuel consumption from MOVES3 were used; alternative fuel VMT and fuel consumption outputs are ignored because they are calculated for the Inventory under a separate methodology. Total gasoline and diesel fuel consumption values from FHWA were then allocated to Inventory vehicle types using gasoline and diesel fuel consumption ratios by vehicle type from MOVES3. Similarly, VMT by vehicle type and fuel type was calculated by multiplying the total VMT from FHWA by VMT ratios by vehicle and fuel type generated by MOVES3. Overall, because total fuel consumption and VMT values are conserved, the changes in total emissions are small, within 0.1 percent. Observed differences in total emissions are due to changes in CH_4 and N_2O emissions, as the methodology for calculating these non-CO_2 emissions utilizes more detailed activity data and is therefore sensitive to the re-allocation of activity data. While total emissions estimates are not significantly impacted by this methodology update, there are significant changes in the allocation of emissions by vehicle type. The share of emissions allocated to passenger cars now generally decline through the time series while the share of emissions allocated to light-duty trucks increase over time.

In addition, the methodology for estimating emissions from alternative fuel vehicles was revised. In previous Inventories, EPA used Energy Information Administration (EIA) surveys of fleet vehicles used by electricity providers, federal agencies, natural gas providers, propane providers, state agencies and transit agencies to determine fuel use and vehicle counts for most alternative fuel vehicles. However, EIA stopped conducting these surveys in 2017. To address this data void, EPA used various methods to determine vehicle counts. Beginning with the 1990 through 2018 Inventory, electric, plug-in electric, and fuel cell vehicle counts were determined from vehicles sales data published by Wards Intelligence. Beginning with this Inventory, electric and fuel cell heavy-duty bus counts are determined from Zukowski, D. (2022) for calendar years 2018 through 2021. Vehicle counts for other fuels (methanol, ethanol, natural gas, and LPG) for 2018 onward were estimated via regression analysis (Browning 2022b).

In addition, the latest version of Argonne National Laboratory's *Greenhouse Gas, Regulated Emissions, and Energy Use in Transportation Model (GREET2022)* provided updated emission factors for all alternative fuel vehicle classes (ANL 2022). Updated emission factors from GREET2022 were implemented in this Inventory, across the entire time series.

The updated vehicle counts and emission factors resulted in a 16 percent reduction in CO_2, a 51 percent reduction in CH_4, and a 92 percent reduction in N_2O in calendar year 2020 for alternative fuel vehicles compared with the previous methodology. This resulted in a 21 percent overall reduction in CO_2 Eq. for alternative fuel vehicles compared with the previous methodology.

In addition, for the current Inventory, CO_2-equivalent estimates of CH_4 and N_2O emissions from transportation and mobile combustion have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC

Fourth Assessment Report (AR4) (IPCC 2007) (used in the previous inventories). The AR5 GWPs have been applied across the entire time series for consistency.

The GWP of CH_4 increased, leading to an overall increase in CH_4 emissions reported in CO_2 equivalent. The GWP of N_2O decreased, leading to a decrease in emissions from N_2O reported in CO_2 equivalent. Compared to the previous Inventory which applied 100-year GWP values from AR4, the average annual change in CO_2-equivalent CH_4 emissions was a 12 percent increase and the average annual change in CO_2-equivalent N_2O emissions was 11 percent decrease for the time series. The net impact from these updates was an average annual 0.1 percent decrease in total CO_2 Eq. emissions for the time series in recent years. Further discussion on this update and the overall impacts of updating the Inventory GWP values to reflect the IPCC AR5 can be found in Chapter 9, Recalculations and Improvements.

Planned Improvements

While the data used for this report represent the most accurate information available, several areas for improvement have been identified.

- Update emission factors for ships and non-recreational boats using residual fuel and distillate fuel, emission factors for locomotives using ultra low sulfur diesel, and emission factors for aircraft using jet fuel. The Inventory currently uses IPCC default values for these emission factors.
- Continue to explore potential improvements to estimates of domestic waterborne fuel consumption for future Inventories. The Inventory estimates for residual and distillate fuel used by ships and boats is based in part on data on bunker fuel use from the U.S. Department of Commerce. Domestic fuel consumption is estimated by subtracting fuel sold for international use from the total sold in the United States. Since 2015, all ships travelling within 200 nautical miles of the U.S. coastlines must use distillate fuels thereby overestimating the residual fuel used by U.S. vessels and underestimating distillate fuel use in these ships.

3.2 Carbon Emitted from Non-Energy Uses of Fossil Fuels (CRF Source Category 1A)

In addition to being combusted for energy, fossil fuels are also consumed for non-energy uses (NEU) in the United States. The fuels used for these purposes are diverse, including natural gas, hydrocarbon gas liquids (HGL),[61] asphalt (a viscous liquid mixture of heavy crude oil distillates), petroleum coke (manufactured from heavy oil), and coal (metallurgical) coke (manufactured from coking coal). The non-energy applications of these fuels are equally diverse, including feedstocks for the manufacture of plastics, rubber, synthetic fibers and other materials; reducing agents for the production of various metals and inorganic products; and products such as lubricants, waxes, and asphalt (IPCC 2006). Emissions from non-energy use of lubricants, paraffin waxes, bitumen / asphalt, and solvents are reported in the Energy sector, as opposed to the Industrial Processes and Product Use (IPPU) sector, to reflect national circumstances in its choice of methodology and to increase transparency of this source category's unique country-specific data sources and methodology (see Box 3-5). In addition, estimates of non-energy use emissions included here do not include emissions already reflected in the IPPU sector, e.g., fuels used as reducing agents. To avoid double counting, the "raw" non-energy fuel consumption data reported by EIA are reduced to account for these emissions already included under IPPU.

[61] HGL (formerly referred to as liquefied petroleum gas, or LPG) are hydrocarbons that occur as gases at atmospheric pressure and as liquids under higher pressures. HGLs include paraffins, such as ethane, propane, butanes, isobutane, and natural gasoline (formerly referred to as pentanes plus), and HGLs include olefins, such as ethylene, propylene, butylene and isobutylene.

Carbon dioxide emissions arise from non-energy uses via several pathways. Emissions may occur during the manufacture of a product, as is the case in producing plastics or rubber from fuel-derived feedstocks. Additionally, emissions may occur during the product's lifetime, such as during solvent use. Overall, throughout the time series and across all uses, about 62 percent of the total C consumed for non-energy purposes was stored in products (e.g., plastics), and not released to the atmosphere; the remaining 38 percent was emitted.

There are several areas in which non-energy uses of fossil fuels are closely related to other parts of this Inventory. For example, some of the non-energy use products release CO_2 at the end of their commercial life when they are combusted after disposal; these emissions are reported separately within the Energy chapter in the Incineration of Waste source category. There are also net exports of petrochemical intermediate products that are not completely accounted for in the EIA data, and the Inventory calculations adjust for the effect of net exports on the mass of C in non-energy applications.

As shown in Table 3-22, fossil fuel emissions in 2021 from the non-energy uses of fossil fuels were 143.2 MMT CO_2 Eq., which constituted approximately 2.8 percent of overall fossil fuel emissions. In 2021, the consumption of fuels for non-energy uses (after the adjustments described above) was 5,938.1 TBtu (see Table 3-23). A portion of the C in the 5,938.1 TBtu of fuels was stored (234.4 MMT CO_2 Eq.), while the remaining portion was emitted (143.2 MMT CO_2 Eq.). Non-energy use emissions increased by 20.1 percent from 2020 to 2021, mainly due to an increase in HGL and industrial coking coal fuel consumption, which contributed to an 18.3 MMT CO_2 Eq. increase in emissions from 2020 to 2021. Although a rise in consumption of some fuels was potentially due to a bounce back in production following the early effects of the COVID-19 pandemic (e.g., naphtha and special naphtha production returned closer to pre-2020 levels), the overall increase in 2021 emissions for select industries exceeds pre-pandemic levels. See Annex 2.3 for more details.

Table 3-22: CO_2 Emissions from Non-Energy Use Fossil Fuel Consumption (MMT CO_2 Eq. and Percent C)

Year	1990	2005	2017	2018	2019	2020	2021
Potential Emissions	305.8	366.9	332.4	352.6	355.9	350.2	377.6
C Stored	193.4	238.0	219.6	223.2	228.2	231.0	234.4
Emissions as a % of Potential	37%	35%	34%	37%	36%	34%	38%
C Emitted	**112.4**	**128.9**	**112.8**	**129.4**	**127.6**	**119.2**	**143.2**

Note: NEU emissions presented in this table differ from the NEU emissions presented in CRF table 1.A(a)s4 as the CRF NEU emissions do not include NEU of lubricants and other petroleum in U.S. Territories. NEU emissions from U.S. Territories are reported under U.S. Territories in the CRF table 1.A(a)s4.

Methodology and Time-Series Consistency

The first step in estimating C stored in products was to determine the aggregate quantity of fossil fuels consumed for non-energy uses. The C content of these feedstock fuels is equivalent to potential emissions, or the product of consumption and the fuel-specific C content values. Both the non-energy fuel consumption and C content data were supplied by the EIA (2022) (see Annex 2.1). Consumption values for industrial coking coal, petroleum coke, other oils, and natural gas in Table 3-23 and Table 3-24 have been adjusted to subtract non-energy uses that are included in the source categories of the Industrial Processes and Product Use chapter.[62] Consumption of natural gas, HGL, naphthas, other oils, and special naphtha were adjusted to subtract out net exports of these products that are not reflected in the raw data from EIA. Consumption values were also adjusted to subtract net exports of HGL components (e.g., propylene, ethane).

[62] These source categories include Iron and Steel Production, Lead Production, Zinc Production, Ammonia Manufacture, Carbon Black Manufacture (included in Petrochemical Production), Titanium Dioxide Production, Ferroalloy Production, Silicon Carbide Production, and Aluminum Production.

For the remaining non-energy uses, the quantity of C stored was estimated by multiplying the potential emissions by a storage factor.

- For several fuel types—petrochemical feedstocks (including natural gas for non-fertilizer uses, HGL, naphthas, other oils, still gas, special naphtha, and industrial other coal), asphalt and road oil, lubricants, and waxes—U.S. data on C stocks and flows were used to develop C storage factors, calculated as the ratio of (a) the C stored by the fuel's non-energy products to (b) the total C content of the fuel consumed. A lifecycle approach was used in the development of these factors in order to account for losses in the production process and during use. Because losses associated with municipal solid waste management are handled separately in the Energy sector under the Incineration of Waste source category, the storage factors do not account for losses at the disposal end of the life cycle.
- For industrial coking coal and distillate fuel oil, storage factors were taken from Marland and Rotty (1984).
- For the remaining fuel types (petroleum coke, miscellaneous products and other petroleum), IPCC (2006) does not provide guidance on storage factors, and assumptions were made based on the potential fate of C in the respective non-energy use products. Carbon dioxide emissions from carbide production are implicitly accounted for in the storage factor calculation for the non-energy use of petroleum coke.

Table 3-23: Adjusted Consumption of Fossil Fuels for Non-Energy Uses (TBtu)

Year	1990	2005	2017	2018	2019	2020	2021
Industry	**4,317.8**	**5,115.1**	**5,089.8**	**5,448.0**	**5,484.1**	**5,444.8**	**5,815.9**
Industrial Coking Coal	NO	80.4	113.0	124.7	112.8	70.0	160.3
Industrial Other Coal	7.6	11.0	9.5	9.5	9.5	9.5	9.5
Natural Gas to Chemical Plants	282.4	260.9	588.0	676.4	667.6	663.3	667.3
Asphalt & Road Oil	1,170.2	1,323.2	849.2	792.8	843.9	832.3	898.1
HGL[a]	1,218.0	1,610.1	2,193.7	2,506.9	2,550.7	2,658.0	2,819.6
Lubricants	186.3	160.2	124.9	122.0	118.3	111.1	113.9
Natural Gasoline[b]	117.5	95.4	81.7	105.3	155.0	163.7	202.4
Naphtha (<401 °F)	327.0	679.5	413.0	421.2	369.5	329.4	331.1
Other Oil (>401 °F)	663.6	499.5	242.9	219.1	212.1	195.6	196.3
Still Gas	36.7	67.7	163.8	166.9	158.7	145.4	152.8
Petroleum Coke	29.1	104.2	NO	NO	NO	NO	NO
Special Naphtha	101.1	60.9	95.3	87.0	89.5	80.8	76.1
Distillate Fuel Oil	7.0	16.0	5.8	5.8	5.8	5.8	5.8
Waxes	33.3	31.4	10.2	12.4	10.4	9.2	11.8
Miscellaneous Products	137.8	112.8	198.8	198.0	180.2	170.7	170.8
Transportation	**176.0**	**151.3**	**142.0**	**137.0**	**131.3**	**115.6**	**118.6**
Lubricants	176.0	151.3	142.0	137.0	131.3	115.6	118.6
U.S. Territories	**50.8**	**114.9**	**3.5**	**3.6**	**3.6**	**3.6**	**3.6**
Lubricants	0.7	4.6	1.0	1.0	1.0	1.0	1.0
Other Petroleum (Misc. Prod.)	50.1	110.3	2.4	2.5	2.6	2.6	2.6
Total	**4,544.6**	**5,379.4**	**5,235.3**	**5,588.5**	**5,619.1**	**5,564.0**	**5,938.1**

NO (Not Occurring)

[a] Excludes natural gasoline.

[b] Formerly referred to as "Pentanes Plus." This source has been adjusted and is reported separately from HGL to align with historic data and revised EIA terminology.

Table 3-24: 2021 Adjusted Non-Energy Use Fossil Fuel Consumption, Storage, and Emissions

Sector/Fuel Type	Adjusted Non-Energy Use[a] (TBtu)	Carbon Content Coefficient (MMT C/QBtu)	Potential Carbon (MMT C)	Storage Factor	Carbon Stored (MMT C)	Carbon Emissions (MMT C)	Carbon Emissions (MMT CO_2 Eq.)
Industry	**5,815.9**	**NA**	**100.5**	**NA**	**63.7**	**36.8**	**135.0**
Industrial Coking Coal	160.3	25.60	4.1	0.10	0.4	3.7	13.5
Industrial Other Coal	9.5	26.10	0.2	0.59	0.1	0.1	0.4
Natural Gas to Chemical Plants	667.3	14.47	9.6	0.59	5.7	3.9	14.4
Asphalt & Road Oil	898.1	20.55	18.5	1.00	18.4	0.1	0.3
HGL[b]	2,819.6	16.83	47.4	0.59	28.0	19.4	71.1
Lubricants	113.9	20.20	2.3	0.09	0.2	2.1	7.7
Natural Gasoline[c]	202.4	18.24	3.7	0.59	2.2	1.5	5.5
Naphtha (<401° F)	331.1	18.55	6.1	0.59	3.6	2.5	9.2
Other Oil (>401° F)	196.3	20.17	4.0	0.59	2.3	1.6	5.9
Still Gas	152.8	17.51	2.7	0.59	1.6	1.1	4.0
Petroleum Coke	NO	27.85	NO	0.30	NO	NO	NO
Special Naphtha	76.1	19.74	1.5	0.59	0.9	0.6	2.3
Distillate Fuel Oil	5.8	20.22	0.1	0.50	0.1	0.1	0.2
Waxes	11.8	19.80	0.2	0.58	0.1	0.1	0.4
Miscellaneous Products	170.8	NO	NO	NO	NO	NO	NO
Transportation	**118.6**	**NA**	**2.4**	**NA**	**0.2**	**2.2**	**8.0**
Lubricants	118.6	20.20	2.4	0.09	0.2	2.2	8.0
U.S. Territories	**3.6**	**NA**	**0.1**	**NA**	**+**	**0.1**	**0.2**
Lubricants	1.0	20.20	+	0.09	+	+	0.1
Other Petroleum (Misc. Prod.)	2.6	20.00	0.1	0.10	+	+	0.2
Total	**5,938.1**		**103.0**		**63.9**	**39.1**	**143.2**

+ Does not exceed 0.05 TBtu, MMT C, or MMT CO_2 Eq.
NA (Not Applicable)
NO (Not Occurring)
[a] To avoid double counting, net exports have been deducted.
[b] Excludes natural gasoline.
[c] Formerly referred to as "Pentanes Plus." This source has been adjusted and is reported separately from HGL to align with historic data and revised EIA terminology.
Note: Totals may not sum due to independent rounding.

Lastly, emissions were estimated by subtracting the C stored from the potential emissions (see Table 3-22). More detail on the methodology for calculating storage and emissions from each of these sources is provided in Annex 2.3.

Where storage factors were calculated specifically for the United States, data were obtained on (1) products such as asphalt, plastics, synthetic rubber, synthetic fibers, cleansers (soaps and detergents), pesticides, food additives, antifreeze and deicers (glycols), and silicones; and (2) industrial releases including energy recovery (waste gas from chemicals), Toxics Release Inventory (TRI) releases, hazardous waste incineration, and volatile organic compound, solvent, and non-combustion CO emissions. Data were taken from a variety of industry sources, government reports, and expert communications. Sources include EPA reports and databases such as compilations of air emission factors (EPA 2001), *National Emissions Inventory (NEI) Air Pollutant Emissions Trends Data* (EPA 2022), *Toxics Release Inventory, 1998* (EPA 2000b), *Biennial Reporting System* (EPA 2000a, 2009), *Resource Conservation and Recovery Act Information System* (EPA 2013b, 2015, 2016b, 2018b, 2021), pesticide sales and use estimates (EPA 1998, 1999, 2002, 2004, 2011, 2017), and the Chemical Data Access Tool (EPA 2014b); the EIA Manufacturer's Energy Consumption Survey (MECS) (EIA 1994, 1997, 2001, 2005, 2010, 2013, 2017, 2021); the National

Petrochemical & Refiners Association (NPRA 2002); the U.S. Census Bureau (1999, 2004, 2009, 2014, 2021); Bank of Canada (2012, 2013, 2014, 2016, 2017, 2018, 2019, 2020, 2021, 2022); Financial Planning Association (2006); INEGI (2006); the United States International Trade Commission (2022); Gosselin, Smith, and Hodge (1984); EPA's *Municipal Solid Waste (MSW) Facts and Figures* (EPA 2013, 2014a, 2016a, 2018a, 2019); the U.S. Tire Manufacturers Association (USTMA2012, 2013, 2014, 2016, 2018, 2020, 2022); the International Institute of Synthetic Rubber Products (IISRP 2000, 2003); the Fiber Economics Bureau (FEB 2001, 2003, 2005, 2007, 2009, 2010, 2011, 2012, 2013); the Independent Chemical Information Service (ICIS 2008, 2016); the EPA Chemical Data Access Tool (CDAT) (EPA 2014b); the American Chemistry Council (ACC 2003 through 2011, 2013, 2014, 2015, 2016, 2017, 2018, 2019, 2020, 2021, 2022a); the *Guide to the Business of Chemistry* (ACC 2022b); and the Chemistry Industry Association of Canada (CIAC 2022). Specific data sources are listed in full detail in Annex 2.3.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2021 as discussed below.

Box 3-5: Reporting of Lubricants, Waxes, and Asphalt and Road Oil Product Use in Energy Sector

IPCC (2006) provides methodological guidance to estimate emissions from the first use of fossil fuels as a product for primary purposes other than combustion for energy purposes (including lubricants, paraffin waxes, bitumen / asphalt, and solvents) under the IPPU sector.[63] In this Inventory, C storage and C emissions from product use of lubricants, waxes, and asphalt and road oil are reported under the Energy sector in the Carbon Emitted from Non-Energy Uses of Fossil Fuels source category (CRF Source Category 1A5).[64]

The emissions are reported in the Energy sector, as opposed to the IPPU sector, to reflect national circumstances in its choice of methodology and to increase transparency of this source category's unique country-specific data sources and methodology. Although emissions from these non-energy uses are reported in the Energy chapter the methodologies used to determine emissions are compatible with the *2006 IPCC Guidelines*. The country-specific methodology used for the Carbon Emitted from Non-Energy Uses of Fossil Fuels source category is based on a carbon balance (i.e., C inputs-outputs) calculation of the aggregate amount of fossil fuels used for non-energy uses, including inputs of lubricants, waxes, asphalt and road oil (see Table 3-24).

For those inputs, U.S. country-specific data on C stocks and flows are used to develop carbon storage factors, which are calculated as the ratio of the C stored by the fossil fuel non-energy products to the total C content of the fuel consumed, taking into account losses in the production process and during product use.[65] The country-specific methodology to reflect national circumstances starts with the aggregate amount of fossil fuels used for non-energy uses and applies a C balance calculation, breaking out the C emissions from non-energy use of lubricants, waxes, and asphalt and road oil. The emissions are reported under the Energy chapter to improve transparency, report a more complete carbon balance and to avoid double counting. Due to U.S. national circumstances, reporting these C emissions separately under IPPU would involve making artificial adjustments to allocate both the C inputs and C outputs of the non-energy use C balance. For example, only the emissions from the first use of lubricants and waxes are to be reported under the IPPU sector, emissions from use of lubricants in 2-stroke engines and emissions from secondary use of lubricants and waxes in waste incineration with energy recovery are to be reported under the Energy sector. Reporting these non-energy use emissions from only first use of lubricants and waxes under IPPU would involve making artificial adjustments to the non-energy use C carbon balance and could potentially result in double counting of emissions. These artificial adjustments would also be required for asphalt and road oil and solvents (which are captured as part of petrochemical feedstock emissions) and could also potentially result in double counting of emissions. To avoid

[63] See for example Volume 3: Industrial Processes and Product Use, and Chapter 5: Non-Energy Products from Fuels and Solvent Use of the *2006 IPCC Guidelines for National Greenhouse Gas Inventories* (IPCC 2006).

[64] Non-methane volatile organic compound (NMVOC) emissions from solvent use are reported separately in the IPPU sector, following Chapter 5 of the *2006 IPCC Guidelines*.

[65] Data and calculations for lubricants and waxes and asphalt and road oil are in Annex 2.3 – Methodology for Estimating Carbon Emitted from Non-Energy Uses of Fossil Fuels.

presenting an incomplete C balance and a less transparent approach for the Carbon Emitted from Non-Energy Uses of Fossil Fuels source category calculation, the entire calculation of C storage and C emissions is therefore conducted in the Non-Energy Uses of Fossil Fuels category calculation methodology, and both the C storage and C emissions for lubricants, waxes, and asphalt and road oil are reported under the Energy sector.

However, emissions from non-energy uses of fossil fuels as feedstocks or reducing agents (e.g., petrochemical production, aluminum production, titanium dioxide, and zinc production) are reported in the IPPU chapter, unless otherwise noted due to specific national circumstances.

Uncertainty

An uncertainty analysis was conducted to quantify the uncertainty surrounding the estimates of emissions and storage factors from non-energy uses. This analysis, performed using @RISK software and the IPCC-recommended Approach 2 methodology (Monte Carlo Stochastic Simulation technique), provides for the specification of probability density functions for key variables within a computational structure that mirrors the calculation of the inventory estimate. The results presented below provide the 95 percent confidence interval, the range of values within which emissions are likely to fall, for this source category.

As noted above, the non-energy use analysis is based on U.S.-specific storage factors for (1) feedstock materials (natural gas, HGL, natural gasoline, naphthas, other oils, still gas, special naphthas, and other industrial coal), (2) asphalt, (3) lubricants, and (4) waxes. For the remaining fuel types (the "other" category in Table 3-23 and Table 3-24) the storage factors were taken directly from IPCC (2006), where available, and otherwise assumptions were made based on the potential fate of carbon in the respective NEU products. To characterize uncertainty, five separate analyses were conducted, corresponding to each of the five categories. In all cases, statistical analyses or expert judgments of uncertainty were not available directly from the information sources for all the activity variables; thus, uncertainty estimates were determined using assumptions based on source category knowledge.

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-25 (emissions) and Table 3-26 (storage factors). Carbon emitted from non-energy uses of fossil fuels in 2021 was estimated to be between 84.0 and 205.5 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 41 percent below to 43 percent above the 2021 emission estimate of 143.2 MMT CO_2 Eq. The uncertainty in the emission estimates is a function of uncertainty in both the quantity of fuel used for non-energy purposes and the storage factor.

Table 3-25: Approach 2 Quantitative Uncertainty Estimates for CO_2 Emissions from Non-Energy Uses of Fossil Fuels (MMT CO_2 Eq. and Percent)

Source	Gas	2021 Emission Estimate (MMT CO_2 Eq.)	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Feedstocks	CO_2	112.9	59.3	178.6	-48%	58%
Asphalt	CO_2	0.3	0.1	0.7	-58%	+125%
Lubricants	CO_2	15.7	13.0	18.2	-17%	+16%
Waxes	CO_2	0.4	0.3	0.7	-24%	+83%
Other	CO_2	13.9	2.5	16.2	-82%	+16%
Total	**CO_2**	**143.2**	**84.0**	**205.5**	**-41%**	**+43%**

[a] Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval.

Note: Totals may not sum due to independent rounding.

Table 3-26: Approach 2 Quantitative Uncertainty Estimates for Storage Factors of Non-Energy Uses of Fossil Fuels (Percent)

Source	Gas	2021 Storage Factor (%)	Uncertainty Range Relative to Emission Estimate[a] (%)		(%, Relative)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Feedstocks	CO_2	59.1%	47.5%	72.2%	-20%	+22%
Asphalt	CO_2	99.6%	99.0%	99.8%	-0.5%	+0.3%
Lubricants	CO_2	9.2%	4.0%	17.4%	-57%	+90%
Waxes	CO_2	57.8%	47.3%	67.5%	-18%	+17%
Other	CO_2	11.1%	6.4%	83.3%	-42%	+650%

[a] Range of emission estimates predicted by Monte Carlo Stochastic Simulation for a 95 percent confidence interval, as a percentage of the inventory value (also expressed in percent terms).

As shown in Table 3-26, waxes and asphalt contribute least to overall storage factor uncertainty on a percentage basis. Although the feedstocks category—the largest use category in terms of total carbon flows—also appears to have relatively tight confidence limits, this is to some extent an artifact of the way the uncertainty analysis was structured. As discussed in Annex 2.3, the storage factor for feedstocks is based on an analysis of six fates that result in long-term storage (e.g., plastics production), and eleven that result in emissions (e.g., volatile organic compound emissions). Rather than modeling the total uncertainty around all of these fate processes, the current analysis addresses only the storage fates, and assumes that all C that is not stored is emitted. As the production statistics that drive the storage values are relatively well-characterized, this approach yields a result that is probably biased toward understating uncertainty.

As is the case with the other uncertainty analyses discussed throughout this document, the uncertainty results above address only those factors that can be readily quantified. More details on the uncertainty analysis are provided in Annex 2.3.

QA/QC and Verification

In order to ensure the quality of the emission estimates from non-energy uses of fossil fuels, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. This effort included a general analysis, as well as portions of a category specific analysis for non-energy uses involving petrochemical feedstocks and for imports and exports. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and methodology for estimating the fate of C (in terms of storage and emissions) across the various end-uses of fossil C. Emission and storage totals for the different subcategories were compared, and trends across the time series were analyzed to determine whether any corrective actions were needed. Corrective actions were taken to rectify minor errors and to improve the transparency of the calculations, facilitating future QA/QC.

For petrochemical import and export data, special attention was paid to NAICS numbers and titles to verify that none had changed or been removed. Import and export totals were compared with 2020 totals as well as their trends across the time series.

It is important to ensure no double counting of emissions between fuel combustion, non-energy use of fuels and industrial process emissions. For petrochemical feedstock production, our review of the categories suggests this is not a significant issue since the non-energy use industrial release data includes different categories of sources and sectors than those included in the Industrial Processes and Product Use (IPPU) emissions category for petrochemicals. Further data integration is not available at his time because feedstock data from the EIA used to estimate non-energy uses of fuels are aggregated by fuel type, rather than disaggregated by both fuel type and particular industries. Also, GHGRP-reported data on quantities of fuel consumed as feedstocks by petrochemical producers are unable to be used due to the data failing GHGRP CBI aggregation criteria.

Recalculations Discussion

Several updates to activity data factors lead to recalculations of previous year results. The major updates are as follows:

- ACC (2022b) updated polyester, polyolefin and nylon fiber, ethylene glycol, maleic anhydride, adipic acid, and acetic acid production in 2020, which resulted in a slight decrease in emissions relative to the previous Inventory.
- U.S. International Trade Commission (2022) updated historical import and export data from 1996 to 2020, resulting in fewer net exports relative to the previous Inventory.
- Updates to the petrochemical feedstock production and stocks led to an increase to the annually variable storage factor from 1996 to 2020 for feedstocks, leading to less carbon emitted and a decrease in emissions, most notably from HGL.
- CIAC (2022) revised shipments for years 2017 to 2020, which resulted in a slight increase in emissions from plastics from 2017 to 2020.

Overall, these changes resulted in an average annual decrease of 0.2 MMT CO_2 Eq. (0.2 percent) in carbon emissions from non-energy uses of fossil fuels for the period 1990 through 2020, relative to the previous Inventory.

Planned Improvements

There are several future improvements planned:

- More accurate accounting of C in petrochemical feedstocks. EPA has worked with EIA to determine the cause of input/output discrepancies in the C mass balance contained within the NEU model. In the future, two strategies to reduce or eliminate this discrepancy will continue to be pursued as part of quality control procedures. First, accounting of C in imports and exports will be improved. The import/export adjustment methodology will be examined to ensure that net exports of intermediaries such as ethylene and propylene are fully accounted for. Second, the use of top-down C input calculation in estimating emissions will be reconsidered. Alternative approaches that rely more substantially on the bottom-up C output calculation will be considered instead.
- Improving the uncertainty analysis. Most of the input parameter distributions are based on professional judgment rather than rigorous statistical characterizations of uncertainty.
- Better characterizing flows of fossil C. Additional fates may be researched, including the fossil C load in organic chemical wastewaters, plasticizers, adhesives, films, paints, and coatings. There is also a need to further clarify the treatment of fuel additives and backflows (especially methyl tert-butyl ether, MTBE).
- Reviewing the trends in fossil fuel consumption for non-energy uses. Annual consumption for several fuel types is highly variable across the time series, including industrial coking coal and other petroleum. A better understanding of these trends will be pursued to identify any mischaracterized or misreported fuel consumption for non-energy uses.
- Updating the average C content of solvents was researched, since the entire time series depends on one year's worth of solvent composition data. The data on C emissions from solvents that were readily available do not provide composition data for all categories of solvent emissions and also have conflicting definitions for volatile organic compounds, the source of emissive C in solvents. Additional sources of solvents data will be investigated in order to update the C content assumptions.
- Updating the average C content of cleansers (soaps and detergents) was researched; although production and consumption data for cleansers are published every 5 years by the Census Bureau, the composition (C content) of cleansers has not been recently updated. Recently available composition data sources may

facilitate updating the average C content for this category.

- Revising the methodology for consumption, production, and C content of plastics was researched; because of recent changes to the type of data publicly available for plastics, the NEU model for plastics applies data obtained from personal communications. Potential revisions to the plastics methodology to account for the recent changes in published data will be investigated.
- Although U.S.-specific storage factors have been developed for feedstocks, asphalt, lubricants, and waxes, default values from IPCC are still used for two of the non-energy fuel types (industrial coking coal, distillate oil), and broad assumptions are being used for miscellaneous products and other petroleum. Over the long term, there are plans to improve these storage factors by analyzing C fate similar to those described in Annex 2.3 or deferring to more updated default storage factors from IPCC where available.
- Reviewing the storage of carbon black across various sectors in the Inventory; in particular, the carbon black abraded and stored in tires.
- Assess the current method and/or identify new data sources (e.g., EIA) for estimating emissions from ammonia/fertilizer use of natural gas.
- Investigate EIA NEU and MECS data to update, as needed, adjustments made for ammonia production and "natural gas to chemical plants, other uses" and "natural gas to other" non-energy uses, including iron and steel production, in energy uses and IPPU.

3.3 Incineration of Waste (CRF Source Category 1A5)

Combustion is used to manage about 7 to 19 percent of the solid wastes generated in the United States, depending on the source of the estimate and the scope of materials included in the definition of solid waste (EPA 2000; EPA 2020; Goldstein and Madtes 2001; Kaufman et al. 2004; Simmons et al. 2006; van Haaren et al. 2010). In the context of this section, waste includes all municipal solid waste (MSW) as well as scrap tires. In the United States, combustion of MSW tends to occur at waste-to-energy facilities or industrial facilities where useful energy is recovered, and thus emissions from waste combustion are accounted for in the Energy chapter. Similarly, scrap tires are combusted for energy recovery in industrial and utility boilers, pulp and paper mills, and cement kilns. Combustion of waste results in conversion of the organic inputs to CO_2. According to the *2006 IPCC Guidelines*, when the CO_2 emitted is of fossil origin, it is counted as a net anthropogenic emission of CO_2 to the atmosphere. Thus, the emissions from waste combustion are calculated by estimating the quantity of waste combusted and the fraction of the waste that is C derived from fossil sources.

Most of the organic materials in MSW are of biogenic origin (e.g., paper, yard trimmings), and have their net C flows accounted for under the Land Use, Land-Use Change, and Forestry chapter. However, some components of MSW and scrap tires—plastics, synthetic rubber, synthetic fibers, and carbon black—are of fossil origin. Plastics in the U.S. waste stream are primarily in the form of containers, packaging, and durable goods. Rubber is found in durable goods, such as carpets, and in non-durable goods, such as clothing and footwear. Fibers in MSW are predominantly from clothing and home furnishings. As noted above, scrap tires (which contain synthetic rubber and carbon black) are also considered a "non-hazardous" waste and are included in the waste combustion estimate, though waste disposal practices for tires differ from MSW. Estimates on emissions from hazardous waste combustion can be found in Annex 2.3 and are accounted for as part of the C mass balance for non-energy uses of fossil fuels.

Approximately 27.8 million metric tons of MSW were combusted in 2021 (EPA 2021). Carbon dioxide emissions from combustion of waste decreased 3.3 percent since 1990, to an estimated 12.5 MMT CO_2 (12,476 kt) in 2021. Emissions across the time series are shown in Table 3-27**Error! Reference source not found.** and Table 3-28**Error! Reference source not found.**.

Waste combustion is also a source of CH_4 and N_2O emissions (De Soete 1993; IPCC 2006). Methane emissions from the combustion of waste were estimated to be less than 0.05 MMT CO_2 Eq. (less than 0.05 kt CH_4) in 2021 and have remained steady since 1990. Nitrous oxide emissions from the combustion of waste were estimated to be 0.4 MMT CO_2 Eq. (1.3 kt N_2O) in 2021 and have decreased by 13 percent since 1990. This decrease is driven by the decrease in total MSW combusted.

Table 3-27: CO_2, CH_4, and N_2O Emissions from the Combustion of Waste (MMT CO_2 Eq.)

Gas	1990	2005	2017	2018	2019	2020	2021
CO_2	12.9	13.3	13.2	13.3	12.9	12.9	12.5
CH_4	+	+	+	+	+	+	+
N_2O	0.4	0.3	0.4	0.4	0.4	0.3	0.4
Total	**13.3**	**13.6**	**13.5**	**13.7**	**13.3**	**13.3**	**12.8**

+ Does not exceed 0.05 MMT CO_2 Eq.
Note: Totals may not sum due to independent rounding.

Table 3-28: CO_2, CH_4, and N_2O Emissions from the Combustion of Waste (kt)

Gas	1990	2005	2017	2018	2019	2020	2021
CO_2	12,900	13,254	13,161	13,339	12,948	12,921	12,476
CH_4	+	+	+	+	+	+	+
N_2O	2	1	1	1	1	1	1

+ Does not exceed 0.05 kt.

Methodology and Time-Series Consistency

Municipal Solid Waste Combustion

To determine both CO_2 and non-CO_2 emissions from the combustion of waste, the tonnage of waste combusted and an estimated emissions factor are needed. Emission estimates from the combustion of tires are discussed separately. Data for total waste combusted was derived from *BioCycle* (van Haaren et al. 2010), EPA Facts and Figures Report, Energy Recovery Council (ERC), EPA's Greenhouse Gas Reporting Program (GHGRP), and the U.S. Energy Information Administration (EIA). Multiple sources were used to ensure a complete, quality dataset, as each source encompasses a different timeframe.

EPA determined the MSW tonnages based on data availability and accuracy throughout the time series.

- 1990-2006: MSW combustion tonnages are from Biocycle combustion data. Tire combustion data from the U.S. Tire Manufacturers Association (USTMA) are removed to arrive at MSW combusted without tires
- 2006-2010: MSW combustion tonnages are an average of Biocycle (with USTMA tire data tonnage removed), U.S. EPA Facts and Figures, EIA, and Energy Recovery Council data (with USTMA tire data tonnage removed).
- 2011-2021: MSW combustion tonnages are from EPA's GHGRP data.

Table 3-29 provides the estimated tons of MSW combusted including and excluding tires.

Table 3-29: Municipal Solid Waste Combusted (Short Tons)

Year	Waste Combusted *(excluding tires)*	Waste Combusted *(including tires)*
1990	33,344,839	33,766,239
2005	26,486,414	28,631,054

2017	28,574,258	30,310,598
2018	29,162,364	30,853,949
2019	28,174,311	29,821,141
2020	27,586,271	29,106,686
2021	27,867,446	29,261,446

Sources: BioCycle, EPA Facts and Figures, ERC, GHGRP, EIA, USTMA.

CO_2 Emissions from MSW Excluding Scrap Tires

Fossil CO_2 emission factors were calculated from EPA's GHGRP data for non-biogenic sources. Using GHGRP-reported emissions for CH_4 and N_2O and assumed emission factors, the tonnage of waste combusted, excluding tires, was derived. Methane and N_2O emissions and assumed emission factors were used to estimate the amount of MSW combusted in terms of energy content. The energy content of MSW combusted was then converted into tonnage based on assumed MSW heating value. Two estimates were generated (one for CH_4 and one for N_2O) and the two were averaged together. Dividing fossil CO_2 emissions from GHGRP FLIGHT data for MSW combustors by this estimated tonnage yielded an annual CO_2 emission factor. As this data was only available following 2011, all years prior use an average of the emission factors from 2011 through 2015. See Annex 3.7 for more detail on how MSW C factors were calculated.

Finally, CO_2 emissions were calculated by multiplying the annual tonnage estimates, excluding tires, by the calculated emissions factor. Calculated fossil CO_2 emission factors are shown in Table 3-30**.**

Table 3-30: Calculated Fossil CO_2 Content per Ton Waste Combusted (kg CO_2/Short Ton Combusted)

	1990	2005	2017	2018	2019	2020	2021
CO_2 Emission Factors	366	366	360	361	363	377	365

CO_2 Emissions from Scrap Tires

Scrap tires contain several types of synthetic rubber, carbon black, and synthetic fibers. Each type of synthetic rubber has a discrete C content, and carbon black is 100 percent C. For synthetic rubber and carbon black in scrap tires, information was obtained biannually from U.S. Scrap Tire Management Summary for 2005 through 2021 data (USTMA 2022). Information about scrap tire composition was taken from the Rubber Manufacturers' Association internet site (USTMA 2012a). Emissions of CO_2 were calculated based on the amount of scrap tires used for fuel and the synthetic rubber and carbon black content of scrap tires. The mass of combusted material is multiplied by its C content to calculate the total amount of carbon stored. More detail on the methodology for calculating emissions from each of these waste combustion sources is provided in Annex 3.7. Table 3-31 provides CO_2 emissions from combustion of waste tires.

Table 3-31: CO_2 Emissions from Combustion of Tires (MMT CO_2 Eq.)

	1990	2005	2017	2018	2019	2020	2021
Synthetic Rubber	0.3	1.6	1.3	1.3	1.2	1.1	1.0
C Black	0.4	2.0	1.6	1.5	1.5	1.4	1.3
Total	**0.7**	**3.6**	**2.9**	**2.8**	**2.7**	**2.5**	**2.3**

Non-CO_2 Emissions

Combustion of waste also results in emissions of CH_4 and N_2O. These emissions were calculated by multiplying the total estimated mass of waste combusted, including tires, by the respective emission factors. The emission factors for CH_4 and N_2O emissions per quantity of MSW combusted are default emission factors for the default continuously-fed stoker unit MSW combustion technology type and were taken from IPCC (2006).

Uncertainty

An Approach 2 Monte Carlo analysis was performed to determine the level of uncertainty surrounding the estimates of CO_2 emissions and N_2O emissions from the incineration of waste (given the very low emissions for CH_4, no uncertainty estimate was derived). IPCC Approach 2 analysis allows the specification of probability density functions for key variables within a computational structure that mirrors the calculation of the Inventory estimate. Statistical analyses or expert judgments of uncertainty were not available directly from the information sources for most variables; thus, uncertainty estimates for these variables were determined using assumptions based on source category knowledge and the known uncertainty estimates for the waste generation variables.

The uncertainties in the waste incineration emission estimates arise from both the assumptions applied to the data and from the quality of the data. Key factors include reported CO_2 emissions; N_2O and CH_4 emissions factors, and tire synthetic rubber and black carbon contents. The highest levels of uncertainty surround the reported emissions from GHGRP; the lowest levels of uncertainty surround variables that were determined by quantitative measurements (e.g., combustion efficiency, C content of C black).

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-32. Waste incineration CO_2 emissions in 2021 were estimated to be between 10.4 and 14.9 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 17 percent below to 19 percent above the 2021 emission estimate of 12.5 MMT CO_2 Eq. Also at a 95 percent confidence level, waste incineration N_2O emissions in 2021 were estimated to be between 0.2 and 0.9 MMT CO_2 Eq. This indicates a range of 54 percent below to 163 percent above the 2021 emission estimate of 0.4 MMT CO_2 Eq.

Table 3-32: Approach 2 Quantitative Uncertainty Estimates for CO_2 and N_2O from the Incineration of Waste (MMT CO_2 Eq. and Percent)

Source	Gas	2021 Emission Estimate (MMT CO_2 Eq.)	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Incineration of Waste	CO_2	12.5	10.4	14.9	-17%	19%
Incineration of Waste	N_2O	0.4	0.2	0.9	-54%	163%

[a] Range of emission estimates predicted by Monte Carlo Simulation for a 95 percent confidence interval.

QA/QC and Verification

In order to ensure the quality of the emission estimates from waste combustion, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and specifically focused on the emission factor and activity data sources and methodology used for estimating emissions from combustion of waste. Trends across the time series were analyzed to determine whether any corrective actions were needed. Corrective actions were taken to rectify minor errors in the use of activity data.

Recalculations Discussion

For the current Inventory, CO_2-equivalent emissions of CH_4 and N_2O from waste incineration have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth Assessment Report* (AR4), used in the previous inventories (IPCC 2007). The AR5 GWPs have been applied across the entire time series for consistency. Prior inventories used GWPs of 25 and 298 for CH_4 and N_2O, respectively. These values have been updated to 28 and 265, respectively. Compared to the previous Inventory which applied 100-year GWP values from AR4, the average annual change in CO_2-equivalent CH_4 emissions was a 12 percent increase and the average annual change in CO_2-equivalent N_2O emissions was an 11 percent decrease for the time series. As a result of the change in methodology, total emissions across the timeseries changed by an average annual decrease of less than 0.05 MMT CO_2 Eq. (0.3 percent) relative to emissions results calculated using the prior GWPs. Further discussion on this update and the overall impacts of updating the Inventory GWP values to reflect the IPCC *Fifth Assessment Report* can be found in Chapter 9, Recalculations and Improvements. All other recalculations described in this section are compared using the prior GWPs.

Recalculations were performed for CO_2 estimates from 1990 through 2010. In previous Inventories, for years prior to 2011, fossil CO_2 content per ton of waste was calculated based on the average of 2011 to the current year of data. For this cycle the calculation was updated to be an average of estimates from 2011 – 2015. Earlier data is assumed to more closely approximate the MSW composition for historic years. As a result of the change in methodology, CO_2 emissions in 1990 decreased by less than 0.05 MMT CO_2 Eq. relative to the previous Inventory and there was an average annual decrease by less than 0.05 MMT CO_2 Eq. from 1990 through 2010.

Recalculations were performed on the estimate of combusted scrap tires in 2020. 2020 estimates for the scrap tire market were previously proxied from the 2019 U.S. Scrap Tire Management Summary (USTMA 2020). The 2021 U.S. Scrap Tire Management Summary was released in October 2022, allowing 2020 estimates to now be calculated by linear interpolation between 2019 and 2021 data. As a result of the change in methodology, CO_2 emissions in 2020 decreased by 0.2 MT CO_2 Eq. relative to the previous Inventory.

Planned Improvements

No planned improvements for waste combustion were identified.

3.4 Coal Mining (CRF Source Category 1B1a)

Three types of coal mining-related activities release CH_4 and CO_2 to the atmosphere: underground mining, surface mining, and post-mining (i.e., coal-handling) activities. While surface coal mines account for the majority of U.S. coal production, underground coal mines contribute the largest share of fugitive CH_4 emissions (see Table 3-34 and Table 3-35) due to the higher CH_4 content of coal in the deeper underground coal seams. In 2021, 174 underground coal mines and 332 surface mines were operating in the United States (EIA 2022). In recent years, the total number of active coal mines in the United States has declined. In 2021, the United States was the fourth largest coal producer in the world (539 MMT), after China (3,685 MMT), India (771 MMT), and Indonesia (545 MMT) (IEA 2022).

Table 3-33: Coal Production (kt)

Year	Underground		Surface		Total	
	Number of Mines	Production	Number of Mines	Production	Number of Mines	Production
1990	1,683	384,244	1,656	546,808	3,339	931,052

2005	586	334,399	789	691,447	1,398	1,025,846
2017	237	247,778	434	454,301	671	702,080
2018	236	249,804	430	435,521	666	685,325
2019	226	242,557	432	397,750	658	640,307
2020	196	177,380	350	307,944	546	485,324
2021	174	200,122	332	323,142	506	523,264

Fugitive CH_4 Emissions

Underground coal mines liberate CH_4 from ventilation systems and from degasification systems. Ventilation systems pump air through the mine workings to dilute noxious gases and ensure worker safety; these systems can exhaust significant amounts of CH_4 to the atmosphere in low concentrations. Degasification systems are wells drilled from the surface or boreholes drilled inside the mine that remove large, often highly concentrated volumes of CH_4 before, during, or after mining. Some mines recover and use CH_4 generated from ventilation and degasification systems, thereby reducing emissions to the atmosphere.

Surface coal mines liberate CH_4 as the overburden is removed and the coal is exposed to the atmosphere. Methane emissions are normally a function of coal rank (a classification related to the percentage of carbon in the coal) and depth. Surface coal mines typically produce lower-rank coals and remove less than 250 feet of overburden, so their level of emissions is much lower than from underground mines.

In addition, CH_4 is released during post-mining activities, as the coal is processed, transported, and stored for use.

Total CH_4 emissions in 2021 were estimated to be 1,595 kt (44.7 MMT CO_2 Eq.), a decline of approximately 59 percent since 1990 (see Table 3-34 and Table 3-35). In 2021, underground mines accounted for approximately 74 percent of total emissions, surface mines accounted for 13 percent, and post-mining activities accounted for 13 percent. In 2021, total CH_4 emissions from coal mining decreased by approximately 3 percent relative to the previous year. Total coal production in 2021 increased by 8 percent compared to 2020. This resulted in an increase of 7 percent in CH_4 emissions from surface mining and post-mining activities in 2021. However, surface mining and post-mining activities have a lower impact on total CH_4 compared to underground mining (74 percent of total emissions in 2021). The number of operating underground mines decreased in 2021 resulting in a slight decrease in overall CH_4 emissions (3 percent), compared to 2020. Additionally, the amount of CH_4 recovered and used in 2021 decreased by less than 0.5 percent compared to 2020 levels.

Table 3-34: CH_4 Emissions from Coal Mining (MMT CO_2 Eq.)

Activity	1990	2005	2017	2018	2019	2020	2021
Underground (UG) Mining	**83.1**	**47.1**	**45.6**	**43.6**	**38.5**	**35.2**	**32.9**
Liberated	90.5	66.9	65.1	64.6	56.3	53.5	51.2
Recovered & Used	(7.4)	(19.8)	(19.5)	(21.0)	(17.8)	(18.3)	(18.3)
Surface Mining	**12.0**	**13.3**	**8.1**	**7.8**	**7.2**	**5.4**	**5.7**
Post-Mining (UG)	**10.3**	**8.6**	**6.0**	**5.9**	**5.8**	**4.3**	**4.8**
Post-Mining (Surface)	**2.6**	**2.9**	**1.8**	**1.7**	**1.5**	**1.2**	**1.2**
Total	**108.1**	**71.8**	**61.4**	**59.1**	**53.0**	**46.2**	**44.7**

Note: Parentheses indicate negative values.

Table 3-35: CH_4 Emissions from Coal Mining (kt)

Activity	1990	2005	2017	2018	2019	2020	2021
Underground (UG) Mining	**2,968**	**1,682**	**1,627**	**1,557**	**1,376**	**1,257**	**1,176**
Liberated	3,231	2,388	2,324	2,308	2,012	1,912	1,828
Recovered & Used	(263)	(706)	(697)	(751)	(636)	(654)	(652)
Surface Mining	**430**	**475**	**290**	**280**	**255**	**194**	**205**

Post-Mining (UG)	**368**	**306**	**213**	**212**	**206**	**155**	**170**
Post-Mining (Surface)	**93**	**103**	**63**	**61**	**55**	**42**	**44**
Total	**3,860**	**2,566**	**2,192**	**2,110**	**1,893**	**1,648**	**1,595**

Note: Parentheses indicate negative values.

Methodology and Time-Series Consistency

EPA uses an IPCC Tier 3 method for estimating CH_4 emissions from underground coal mining and an IPCC Tier 2 method for estimating CH_4 emissions from surface mining and post-mining activities (for coal production from both underground mines and surface mines). The methodology for estimating CH_4 emissions from coal mining consists of two steps:

- Estimate CH_4 emissions from underground mines. These emissions have two sources: ventilation systems and degasification systems. They are estimated using mine-specific data, then summed to determine total CH_4 liberated. The CH_4 recovered and used is then subtracted from this total, resulting in an estimate of net emissions to the atmosphere.
- Estimate CH_4 emissions from surface mines and post-mining activities. Unlike the methodology for underground mines, which uses mine-specific data, the methodology for estimating emissions from surface mines and post-mining activities consists of multiplying basin-specific coal production by basin-specific gas content and an emission factor.

Step 1: Estimate CH_4 Liberated and CH_4 Emitted from Underground Mines

Underground mines generate CH_4 from ventilation systems and degasification systems. Some mines recover and use the liberated CH_4, thereby reducing emissions to the atmosphere. Total CH_4 emitted from underground mines equals the CH_4 liberated from ventilation systems, plus the CH_4 liberated from degasification systems, minus the CH_4 recovered and used.

Step 1.1: Estimate CH_4 Liberated from Ventilation Systems

To estimate CH_4 liberated from ventilation systems, EPA uses data collected through its Greenhouse Gas Reporting Program (GHGRP)[66] (Subpart FF, "Underground Coal Mines"), data provided by the U.S. Mine Safety and Health Administration (MSHA) (MSHA 2022), and occasionally data collected from other sources on a site-specific level (e.g., state gas production databases). Since 2011, the nation's "gassiest" underground coal mines—those that liberate more than 36,500,000 actual cubic feet of CH_4 per year (about 17,525 MT CO_2 Eq.)—have been required to report to EPA's GHGRP (EPA 2022).[67] Mines that report to EPA's GHGRP must report quarterly measurements of CH_4 emissions from ventilation systems; they have the option of recording and reporting their own measurements, or using the measurements taken by MSHA as part of that agency's quarterly safety inspections of all mines in the United States with detectable CH_4 concentrations.[68]

Since 2013, ventilation CH_4 emission estimates have been calculated based on both quarterly GHGRP data submitted by underground mines and on quarterly measurement data obtained directly from MSHA. Because not all mines report under EPA's GHGRP, the emissions of the mines that do not report must be calculated using MSHA data. The MSHA data also serves as a quality assurance tool for validating GHGRP data. For GHGRP data, reported

[66] In implementing improvements and integrating data from EPA's GHGRP, EPA followed the latest guidance from the IPCC on the use of facility-level data in national inventories (IPCC 2011).

[67] Underground coal mines report to EPA under Subpart FF of the GHGRP (40 CFR Part 98). In 2021, 60 underground coal mines reported to the program.

[68] MSHA records coal mine CH_4 readings with concentrations of greater than 50 ppm (parts per million) CH_4. Readings below this threshold are considered non-detectable.

quarterly ventilation methane emissions (metric tons) are summed for each mine to develop mine-specific annual ventilation emissions. For MSHA data, the average daily CH_4 emission rate for each mine is determined using the CH_4 total for all data measurement events conducted during the calendar year and total duration of all data measurement events (in days). The calculated average daily CH_4 emission rate is then multiplied by 365 days to estimate annual ventilation CH_4 emissions for the MSHA dataset.

Step 1.2: Estimate CH_4 Liberated from Degasification Systems

Particularly gassy underground mines also use degasification systems (e.g., wells or boreholes) to remove CH_4 before, during, or after mining. This CH_4 can then be collected for use or vented to the atmosphere. Twenty mines used degasification systems in 2021 and all of these mines reported the CH_4 removed through these systems to EPA's GHGRP under Subpart FF (EPA 2022). Based on the weekly measurements reported to EPA's GHGRP, degasification data summaries for each mine are added to estimate the CH_4 liberated from degasification systems. Twelve of the 20 mines with degasification systems had operational CH_4 recovery and use projects, including two mines with two recovery and use projects each (see step 1.3 below).[69]

Degasification data reported to EPA's GHGRP by underground coal mines is the primary source of data used to develop estimates of CH_4 liberated from degasification systems. Data reported to EPA's GHGRP were used exclusively to estimate CH_4 liberated from degasification systems at 15 of the 20 mines that used degasification systems in 2021. Data from state gas well production databases were used to supplement GHGRP degasification data for the remaining five mines (DMME 2022; GSA 2022; WVGES 2022).

For pre-mining wells, cumulative degasification volumes that occur prior to the well being mined through are attributed to the mine in the inventory year in which the well is mined through.[70] EPA's GHGRP does not require gas production from virgin coal seams (coalbed methane) to be reported by coal mines under Subpart FF.[71] Most pre-mining wells drilled from the surface are considered coalbed methane wells prior to mine-through and associated CH_4 emissions are reported under another subpart of the GHGRP (Subpart W, "Petroleum and Natural Gas Systems"). As a result, GHGRP data must be supplemented to estimate cumulative degasification volumes that occurred prior to well mine-through. There were four mines with degasification systems that include pre-mining wells that were mined through in 2021. For all of these mines, GHGRP data were supplemented with historical data from state gas well production databases (DMME 2022; ERG 2022; GSA 2022; WVGES 2022), as well as with mine-specific information regarding the locations and dates on which the pre-mining wells were mined through (JWR 2010; El Paso 2009; ERG 2022).

Step 1.3: Estimate CH_4 Recovered from Ventilation and Degasification Systems, and Utilized or Destroyed (Emissions Avoided)

Twelve mines had a total of fourteen CH_4 recovery and use projects in place in 2021, including two mines that each have two recovery and use projects. Thirteen of these projects involved degasification systems with one mine having a ventilation air methane abatement project (VAM). Ten of these mines sold the recovered CH_4 to a pipeline, including one that also used CH_4 to fuel a thermal coal dryer. One mine destroyed recovered CH_4 using flares. One mine destroyed the recovered CH_4 (VAM) using regenerative thermal oxidation (RTO) without energy recovery and using enclosed flares.

The CH_4 recovered and used (or destroyed) at the twelve mines described above are estimated using the following methods:

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[69] Several of the mines venting CH_4 from degasification systems use a small portion of the gas to fuel gob well blowers in remote locations where electricity is not available. However, this CH_4 use is not considered to be a formal recovery and use project.

[70] A well is "mined through" when coal mining development or the working face intersects the borehole or well.

[71] This applies for pre-drainage in years prior to the well being mined through. Beginning with the year the well is mined through, the annual volume of CH_4 liberated from a pre-drainage well is reported under Subpart FF of EPA's GHGRP.

- EPA's GHGRP data was exclusively used to estimate the CH_4 recovered and used from six of the 12 mines that deployed degasification systems in 2021. Based on weekly measurements, the GHGRP degasification destruction data summaries for each mine are added together to estimate the CH_4 recovered and used from degasification systems.
- State sales data were used to supplement GHGRP data to estimate CH_4 recovered and used from five mines that deployed degasification systems in 2021 (DMME 2022, ERG 2022, GSA 2022, and WVGES 2022). Four of these mines intersected pre-mining wells in 2021. Supplemental information is used for these mines because estimating CH_4 recovery and use from pre-mining wells requires additional data not reported under Subpart FF of EPA's GHGRP (see discussion in step 1.2 above) to account for the emissions avoided prior to the well being mined through. The supplemental data is obtained from state gas production databases as well as mine-specific information on the location and timing of mined-through pre-mining wells.
- For the single mine that employed VAM for CH_4 recovery and use, the estimates of CH_4 recovered and used were obtained from the mine's offset verification statement (OVS) submitted to the California Air Resources Board (CARB) (McElroy OVS 2022). This mine also reported CH_4 reductions from flaring. GHGRP data were used to estimate CH_4 recovered and flared in 2021.

Step 2: Estimate CH_4 Emitted from Surface Mines and Post-Mining Activities

Mine-specific data are not available for estimating CH_4 emissions from surface coal mines or for post-mining activities. For surface mines, basin-specific coal production obtained from the Energy Information Administration's *Annual Coal Report* (EIA 2022) is multiplied by basin-specific CH_4 contents (EPA 1996, 2005) and a 150 percent emission factor (to account for CH_4 from over- and under-burden) to estimate CH_4 emissions (King 1994; Saghafi 2013). For post-mining activities, basin-specific coal production is multiplied by basin-specific CH_4 contents and a mid-range 32.5 percent emission factor for CH_4 desorption during coal transportation and storage (Creedy 1993). Basin-specific in situ gas content data were compiled from AAPG (1984) and USBM (1986).

Fugitive CO_2 Emissions

Methane and CO_2 are naturally occurring in coal seams and are collectively referred to as coal seam gas. These gases remain trapped in the coal seam until coal is mined (i.e., coal seam is exposed and fractured during mining operations). Fugitive CO_2 emissions occur during underground coal mining, surface coal mining, and post-mining activities. Methods and data to estimate fugitive CO_2 emissions from underground and surface coal mining are presented in the sections below. Fugitive CO_2 emissions from post-mining activities were not estimated due to the lack of an IPCC method and unavailability of data.

Total fugitive CO_2 emissions in 2021 were estimated to be 2,456 kt (2.5 MMT CO_2 Eq.), a decline of approximately 47 percent since 1990. In 2021, underground mines accounted for approximately 89 percent of total fugitive CO_2 emissions. In 2021, total fugitive CO_2 emissions from coal mining increased by approximately 12 percent relative to the previous year. This increase was due to an increase in annual coal production.

Table 3-36: CO_2 Emissions from Coal Mining (MMT CO_2 Eq.)

Activity	1990	2005	2017	2018	2019	2020	2021
Underground (UG) Mining	**4.2**	**3.6**	**2.8**	**2.8**	**2.7**	**1.9**	**2.2**
Liberated	4.2	3.6	2.7	2.7	2.6	1.9	2.2
Recovered & Used	(+)	(+)	(+)	(+)	(+)	(+)	(+)
Flaring	NO	NO	0.1	0.1	0.1	+	+
Surface Mining	**0.4**	**0.6**	**0.4**	**0.4**	**0.3**	**0.2**	**0.3**
Total	**4.6**	**4.2**	**3.2**	**3.1**	**3.0**	**2.2**	**2.5**

+ Does not exceed 0.05 MMT CO_2 Eq.
NO (Not Occurring)
Note: Parentheses indicate negative values.

Table 3-37: CO_2 Emissions from Coal Mining (kt)

Activity	1990	2005	2017	2018	2019	2020	2021
Underground (UG) Mining	**4,164**	**3,610**	**2,785**	**2,789**	**2,670**	**1,948**	**2,194**
Liberated	4,171	3,630	2,690	2,712	2,633	1,926	2,173
Recovered & Used	(8)	(20)	(19)	(21)	(18)	(18)	(18)
Flaring	NO	NO	114	97	55	41	40
Surface Mining	**443**	**560**	**368**	**353**	**322**	**249**	**262**
Total	**4,606**	**4,170**	**3,153**	**3,141**	**2,992**	**2,198**	**2,456**

NO (Not Occurring)

Notes: Parentheses indicate negative values. Totals may not sum due to independent rounding.

Methodology and Time-Series Consistency

EPA uses an IPCC Tier 1 method for estimating fugitive CO_2 emissions from underground coal mining and surface mining (IPCC 2019). IPCC methods and data to estimate fugitive CO_2 emissions from post-mining activities (for both underground and surface coal mining) are currently not available.

Step 1: Underground Mining

EPA used the following overarching IPCC equation to estimate fugitive CO_2 emissions from underground coal mines (IPCC 2019):

Equation 3-1: Estimating Fugitive CO_2 Emissions from Underground Mines

$$\begin{aligned}&Total\ CO_2\ from\ Underground\ Mines\\ &\quad = CO_2\ from\ underground\ mining\ - Amount\ of\ CO_2\ in\ gas\ recovered\\ &\quad + CO_2\ from\ methane\ flaring\end{aligned}$$

Step 1.1: Estimate Fugitive CO_2 Emissions from Underground Mining

EPA estimated fugitive CO_2 emissions from underground mining using the IPCC Tier 1 emission factor (5.9 m^3/metric ton) and annual coal production from underground mines (EIA 2022). The underground mining default emission factor accounts for all the fugitive CO_2 likely to be emitted from underground coal mining. Therefore, the amount of CO_2 from coal seam gas recovered and utilized for energy is subtracted from underground mining estimates in Step 2, below. Under IPCC methods, the CO_2 emissions from gas recovered and utilized for energy use (e.g., injected into a natural gas pipeline) are reported under other sectors of the Inventory (e.g., stationary combustion of fossil fuel or oil and natural gas systems) and not under the coal mining sector.

Step 1.2: Estimate Amount of CO_2 In Coal Seam Gas Recovered for Energy Purposes

EPA estimated fugitive CO_2 emissions from coal seam gas recovered and utilized for energy purposes by using the IPCC Tier 1 default emission factor (19.57 metric tons CO_2/million cubic meters of coal bed methane (CBM) produced) and quantity of coal seam gas recovered and utilized. Data on annual quantity of coal seam gas recovered and utilized are available from GHGRP and state sales data (GHGRP 2022; DMME 2022; ERG 2022; GSA 2022; WVGES 2022). The quantity of coal seam gas recovered and destroyed without energy recovery (e.g., VAM projects) is deducted from the total coal seam gas recovered quantity (McElroy OVS 2022).

Step 1.3: Estimate Fugitive CO_2 Emissions from Flaring

The IPCC method includes combustion CO_2 emissions from gas recovered for non-energy uses (i.e., flaring, or catalytic oxidation) under fugitive CO_2 emission estimates for underground coal mining. In effect, these emissions, though occurring through stationary combustion, are categorized as fugitive emissions in the Inventory. EPA estimated CO_2 emissions from methane flaring using the following equation:

Equation 3-2: Estimating CO_2 Emissions from Drained Methane Flared or Catalytically Oxidized

$$CO_2\ from\ flaring = 0.98 \times Volume\ of\ methane\ flared \times Conversion\ Factor \times Stoichiometric\ Mass\ Factor$$

Currently there are three mines that report catalytic oxidation of recovered methane through flaring without energy use. Annual data for 2021 were obtained from one mine's offset verification statement (OVS) submitted to the California Air Resources Board (CARB) and the GHGRP for the remaining two mines (McElroy OVS 2022; GHGRP 2022).

Step 2: Surface Mining

EPA estimated fugitive CO_2 emissions from surface mining using the IPCC Tier 1 emission factor (0.44 m^3/metric ton) and annual coal production from surface mines (EIA 2022).

Uncertainty

A quantitative uncertainty analysis was conducted for the coal mining source category using the IPCC-recommended Approach 2 uncertainty estimation methodology. Because emission estimates of CH_4 from underground ventilation systems were based on actual measurement data from EPA's GHGRP or from MSHA, uncertainty is relatively low. A degree of imprecision was introduced because the ventilation air measurements used were not continuous but rather quarterly instantaneous readings that were used to determine the average annual emission rates. Additionally, the measurement equipment used can be expected to have resulted in an average of 10 percent overestimation of annual CH_4 emissions (Mutmansky & Wang 2000). Equipment measurement uncertainty is applied to GHGRP data.

Estimates of CH_4 liberated and recovered by degasification systems are relatively certain for utilized CH_4 because of the availability of EPA's GHGRP data and state gas sales information. Many of the liberation and recovery estimates use data on wells within 100 feet of a mined area. However, uncertainty exists concerning the radius of influence of each well. The number of wells counted, and thus the liberated CH_4 and avoided emissions, may vary if the drainage area is found to be larger or smaller than estimated.

EPA's GHGRP requires weekly CH_4 monitoring of mines that report degasification systems, and continuous CH_4 monitoring is required for CH_4 utilized on- or off-site. Since 2012, GHGRP data have been used to estimate CH_4 emissions from vented degasification wells, reducing the uncertainty associated with prior MSHA estimates used for this sub-source. Beginning in 2013, GHGRP data were also used for determining CH_4 recovery and use at mines without publicly available gas usage or sales records, which has reduced the uncertainty from previous estimation methods that were based on information from coal industry contacts.

Surface mining and post-mining emissions are associated with considerably more uncertainty than underground mines, because of the difficulty in developing accurate emission factors from field measurements. However, since underground coal mining, as a general matter, results in significantly larger CH_4 emissions due to production of higher-rank coal and greater depth, and estimated emissions from underground mining constitute the majority of estimated total coal mining CH_4 emissions, the uncertainty associated with underground emissions is the primary factor that determines overall uncertainty.

The major sources of uncertainty for estimates of fugitive CO_2 emissions are the Tier 1 IPCC default emission factors used for underground mining (-50 percent to +100 percent) and surface mining (-67 percent to +200 percent) (IPCC 2019). Additional sources of uncertainty for fugitive CO_2 emission estimates include EIA's annual coal production data and data used for gas recovery projects, such as GHGRP data, state gas sales data, and VAM estimates for the single mine that operates an active VAM project. Uncertainty ranges for these additional data sources are already available, as these are the same data sources used for CH_4 emission estimates.

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-38. Coal mining CH_4 emissions in 2021 were estimated to be between 40.1 and 54.3 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 10.2 percent below to 21.5 percent above the 2021 emission estimate of 44.7 MMT CO_2 Eq. Coal mining fugitive CO_2 emissions in 2021 were estimated to be between 0.8 and 4.3 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 67.6 percent below to 75.8 percent above the 2021 emission estimate of 2.5 MMT CO_2 Eq.

Table 3-38: Approach 2 Quantitative Uncertainty Estimates for CH_4 and CO_2 Emissions from Coal Mining (MMT CO_2 Eq. and Percent)

Source	Gas	2021 Emission Estimate (MMT CO_2 Eq.)	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Coal Mining	CH_4	44.7	40.1	54.3	-10.2%	+21.5%
Coal Mining	CO_2	2.5	0.8	4.3	-67.6%	+75.8%

[a] Range of emission estimates predicted by Monte Carlo stochastic simulation for a 95 percent confidence interval.

QA/QC and Verification

In order to ensure the quality of the emission estimates for coal mining, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and reported emissions data used for estimating fugitive emissions from coal mining. Trends across the time series were analyzed to determine whether any corrective actions were needed.

Emission estimates for coal mining rely in large part on data reported by coal mines to EPA's GHGRP. EPA verifies annual facility-level reports through a multi-step process to identify potential errors and ensure that data submitted to EPA are accurate, complete, and consistent. All reports submitted to EPA are evaluated by electronic validation and verification checks. If potential errors are identified, EPA will notify the reporter, who can resolve the issue either by providing an acceptable response describing why the flagged issue is not an error or by correcting the flagged issue and resubmitting their annual report. Additional QA/QC and verification procedures occur for each GHGRP subpart. No QA/QC issues or errors were identified in the 2021 Subpart FF data.

Recalculations Discussion

State gas sales production values were updated for five mines, as part of normal updates. This update impacted CH_4 emissions for 1998-2020. As a result of this update, CH_4 emissions from degasification systems and CH_4 emissions avoided increased across the time-series. Degasification CH_4 emissions increased slightly by an average of 0.4 percent and CH_4 emissions avoided increased by an average of 1.6 percent over the 1998 to 2020 time series, compared to the previous Inventory.

Fugitive CO_2 emissions from flaring were recalculated for 2014 through 2020 as a result of adding two flaring projects to the Inventory, as part of normal updates. One of the flaring projects was operational from 2014 onwards and the other one started in 2020. As a result of this update, flaring CO_2 emissions for 2014 to 2020 increased by an average of 230 percent, compared to the previous Inventory, with 2020 emissions increasing by 277 percent. However, as flaring CO_2 emissions only contribute 2 percent of total fugitive CO_2 emissions, this update resulted in a slight increase of overall fugitive CO_2 emissions for 2014 to 2020 by an average of 2 percent, compared to the previous Inventory.

In addition to the above-mentioned updates, for the current Inventory, estimates of CO_2-equivalent CH_4 emissions from coal mining have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth*

Assessment Report (AR4) (IPCC 2007) (used in the previous inventories). The AR5 GWPs have been applied across the entire time series for consistency. The GWP of CH_4 increased from 25 to 28, leading to an overall increase in CO_2-equivalent CH_4 emissions. Compared to the previous Inventory which applied 100-year GWP values from AR4, the average annual change in CO_2-equivalent CH_4 emissions was a 12 percent increase for each year of the time series. Further discussion on this update and the overall impacts of updating the inventory GWPs to reflect the IPCC *Fifth Assessment Report* can be found in Chapter 9, Recalculations and Improvements.

The net impact from the updates listed above was an average annual 12 percent increase in CH_4 emissions and an average annual 0.4 percent increase in CO_2 emissions for the time series.

Planned Improvements

EPA is assessing planned improvements for future reports, but at this time has no specific planned improvements for estimating CH_4 and CO_2 emissions from underground and surface mining and CH_4 emissions from post-mining.

3.5 Abandoned Underground Coal Mines (CRF Source Category 1B1a)

Underground coal mines contribute the largest share of coal mine methane (CMM) emissions, with active underground mines the leading source of underground emissions. However, mines also continue to release CH_4 after closure. As mines mature and coal seams are mined through, mines are closed and abandoned. Many are sealed and some flood through intrusion of groundwater or surface water into the void. Shafts or portals are generally filled with gravel and capped with a concrete seal, while vent pipes and boreholes are plugged in a manner similar to oil and gas wells. Some abandoned mines are vented to the atmosphere to prevent the buildup of CH_4 that may find its way to surface structures through overburden fractures. As work stops within the mines, CH_4 liberation decreases but it does not stop completely. Following an initial decline, abandoned mines can liberate CH_4 at a near-steady rate over an extended period of time, or if flooded, produce gas for only a few years. The gas can migrate to the surface through the conduits described above, particularly if they have not been sealed adequately. In addition, diffuse emissions can occur when CH_4 migrates to the surface through cracks and fissures in the strata overlying the coal mine. The following factors influence abandoned mine emissions:

- Time since abandonment;
- Gas content and adsorption characteristics of coal;
- CH_4 flow capacity of the mine;
- Mine flooding;
- Presence of vent holes; and
- Mine seals.

Annual gross abandoned mine CH_4 emissions ranged from 8.1 to 12.1 MMT CO_2 Eq. from 1990 to 2021, varying, in general, by less than 1 percent to approximately 19 percent from year to year. Fluctuations were due mainly to the number of mines closed during a given year as well as the magnitude of the emissions from those mines when active. Gross abandoned mine emissions peaked in 1996 (12.1 MMT CO_2 Eq.) due to the large number of gassy mine[72] closures from 1994 to 1996 (72 gassy mines closed during the three-year period). In spite of this rapid rise, abandoned mine emissions have been generally on the decline since 1996. Since 2002, there have been fewer than twelve gassy mine closures each year. In 2021 there were two gassy mine closures. Gross abandoned mine emissions decreased slightly from 9.4 MMT CO_2 Eq. (335 kt CH_4) in 2020 to 9.2 (330 kt CH_4) MMT CO_2 Eq. in 2021

[72] A mine is considered a "gassy" mine if it emits more than 100 thousand cubic feet of CH_4 per day (100 Mcfd).

(see Table 3-39 and Table 3-40). Gross emissions are reduced by CH_4 recovered and used at 47 mines, resulting in net emissions in 2021 of 6.4 MMT CO_2 Eq. (228 kt CH_4).

Table 3-39: CH_4 Emissions from Abandoned Coal Mines (MMT CO_2 Eq.)

Activity	1990	2005	2017	2018	2019	2020	2021
Abandoned Underground Mines	8.1	9.3	10.3	9.9	9.6	9.4	9.2
Recovered & Used	NO	(2.0)	(3.1)	(3.0)	(2.9)	(2.9)	(2.9)
Total	**8.1**	**7.4**	**7.2**	**6.9**	**6.6**	**6.5**	**6.4**

NO (Not Occurring)
Notes: Parentheses indicate negative values. Totals may not sum due to independent rounding.

Table 3-40: CH_4 Emissions from Abandoned Coal Mines (kt)

Activity	1990	2005	2017	2018	2019	2020	2021
Abandoned Underground Mines	288	334	367	355	341	335	330
Recovered & Used	NO	(70)	(109)	(107)	(104)	(103)	(103)
Total	**288**	**264**	**257**	**247**	**237**	**232**	**228**

NO (Not Occurring)
Notes: Parentheses indicate negative values. Totals may not sum due to independent rounding.

Methodology and Time-Series Consistency

Estimating CH_4 emissions from an abandoned coal mine requires predicting the emissions of a mine from the time of abandonment through the inventory year of interest. The flow of CH_4 from the coal to the mine void is primarily dependent on the mine's emissions when active and the extent to which the mine is flooded or sealed. The CH_4 emission rate before abandonment reflects the gas content of the coal, the rate of coal mining, and the flow capacity of the mine in much the same way as the initial rate of a water-free conventional gas well reflects the gas content of the producing formation and the flow capacity of the well. A well or a mine that produces gas from a coal seam and the surrounding strata will produce less gas through time as the reservoir of gas is depleted. Depletion of a reservoir will follow a predictable pattern depending on the interplay of a variety of natural physical conditions imposed on the reservoir. The depletion of a reservoir is commonly modeled by mathematical equations and mapped as a type curve. Type curves, which are referred to as decline curves, have been developed for abandoned coal mines. Existing data on abandoned mine emissions through time, although sparse, appear to fit the hyperbolic type of decline curve used in forecasting production from natural gas wells.

To estimate CH_4 emissions over time for a given abandoned mine, it is necessary to apply a decline function, initiated upon abandonment, to that mine. In the analysis, mines were grouped by coal basin with the assumption that they will generally have the same initial pressures, permeability, and isotherm. As CH_4 leaves the system, the reservoir pressure (Pr) declines as described by the isotherm's characteristics. The emission rate declines because the mine pressure (Pw) is essentially constant at atmospheric pressure for a vented mine, and the productivity index (PI), which is expressed as the flow rate per unit of pressure change, is essentially constant at the pressures of interest (atmospheric to 30 psia). The CH_4 flow rate is determined by the laws of gas flow through porous media, such as Darcy's Law. A rate-time equation can be generated that can be used to predict future emissions. This decline through time is hyperbolic in nature and can be empirically expressed as:

Equation 3-3: Decline Function to Estimate Venting Abandoned Mine Methane Emissions

$$q = q_i\,(1 + bD_i t)^{(-1/b)}$$

where,

q	=	Gas flow rate at time t in million cubic feet per day (mmcfd)
q_i	=	Initial gas flow rate at time zero (t_o), mmcfd
b	=	The hyperbolic exponent, dimensionless
D_i	=	Initial decline rate, 1/year

t = Elapsed time from t_o (years)

This equation is applied to mines of various initial emission rates that have similar initial pressures, permeability, and adsorption isotherms (EPA 2004).

The decline curves created to model the gas emission rate of coal mines must account for factors that decrease the rate of emissions after mining activities cease, such as sealing and flooding. Based on field measurement data, it was assumed that most U.S. mines prone to flooding will become completely flooded within eight years and therefore will no longer have any measurable CH_4 emissions. Based on this assumption, an average decline rate for flooded mines was established by fitting a decline curve to emissions from field measurements. An exponential equation was developed from emissions data measured at eight abandoned mines known to be filling with water located in two of the five basins. Using a least squares, curve-fitting algorithm, emissions data were matched to the exponential equation shown below. For this analysis of flooded abandoned mines, there was not enough data to establish basin-specific equations, as was done with the vented, non-flooding mines (EPA 2004). This decline through time can be empirically expressed as:

Equation 3-4: Decline Function to Estimate Flooded Abandoned Mine Methane Emissions

$$q = q_i e^{(-Dt)}$$

where,

q = Gas flow rate at time t in mmcfd
q_i = Initial gas flow rate at time zero (t_o), mmcfd
D = Decline rate, 1/year
t = Elapsed time from t_o (years)

Seals have an inhibiting effect on the rate of flow of CH_4 into the atmosphere compared to the flow rate that would exist if the mine had an open vent. The total volume emitted will be the same, but emissions will occur over a longer period of time. The methodology, therefore, treats the emissions prediction from a sealed mine similarly to the emissions prediction from a vented mine, but uses a lower initial rate depending on the degree of sealing. A computational fluid dynamics simulator was used with the conceptual abandoned mine model to predict the decline curve for inhibited flow. The percent sealed is defined as 100 × (1 – [initial emissions from sealed mine / emission rate at abandonment prior to sealing]). Significant differences are seen between 50 percent, 80 percent, and 95 percent closure. These decline curves were therefore used as the high, middle, and low values for emissions from sealed mines (EPA 2004).

For active coal mines, those mines producing over 100 thousand cubic feet per day (Mcfd) of CH_4 account for about 98 percent of all CH_4 emissions. This same relationship is assumed for abandoned mines. It was determined that the 530 abandoned mines closed since 1972 produced CH_4 emissions greater than 100 Mcfd when active. Further, the status of 306 of the 530 mines (or 58 percent) is known to be either: 1) vented to the atmosphere; 2) sealed to some degree (either earthen or concrete seals); or 3) flooded (enough to inhibit CH_4 flow to the atmosphere). The remaining 42 percent of the mines whose status is unknown were placed in one of these three categories by applying a probability distribution analysis based on the known status of other mines located in the same coal basin (EPA 2004). Table 3-41 presents the count of mines by post-abandonment state, based on EPA's probability distribution analysis.

Table 3-41: Number of Gassy Abandoned Mines Present in U.S. Basins in 2021, Grouped by Class According to Post-Abandonment State

Basin	Sealed	Vented	Flooded	Total Known	Unknown	Total Mines
Central Appl.	43	25	50	118	144	262
Illinois	35	3	14	52	31	83
Northern Appl.	48	23	15	86	39	125
Warrior Basin	0	0	16	16	0	16
Western Basins	28	4	2	34	10	44

Total	**154**	**55**	**97**	**306**	**224**	**530**

Inputs to the decline equation require the average CH_4 emission rate prior to abandonment and the date of abandonment. Generally, these data are available for mines abandoned after 1971; however, such data are largely unknown for mines closed before 1972. Information that is readily available, such as coal production by state and county, is helpful but does not provide enough data to directly employ the methodology used to calculate emissions from mines abandoned before 1972. It is assumed that pre-1972 mines are governed by the same physical, geologic, and hydrologic constraints that apply to post-1971 mines; thus, their emissions may be characterized by the same decline curves.

During the 1970s, 78 percent of CH_4 emissions from coal mining came from seventeen counties in seven states. Mine closure dates were obtained for two states, Colorado and Illinois, for the hundred-year period extending from 1900 through 1999. The data were used to establish a frequency of mine closure histogram (by decade) and applied to the other five states with gassy mine closures. As a result, basin-specific decline curve equations were applied to the 145 gassy coal mines estimated to have closed between 1920 and 1971 in the United States, representing 78 percent of the emissions. State-specific, initial emission rates were used based on average coal mine CH_4 emission rates during the 1970s (EPA 2004).

Abandoned mine emission estimates are based on all closed mines known to have active mine CH_4 ventilation emission rates greater than 100 Mcfd at the time of abandonment. For example, for 1990 the analysis included 145 mines closed before 1972 and 258 mines closed between 1972 and 1990. Initial emission rates based on MSHA reports, time of abandonment, and basin-specific decline curves influenced by a number of factors were used to calculate annual emissions for each mine in the database (MSHA 2022). Coal mine degasification data are not available for years prior to 1990, thus the initial emission rates used reflect only ventilation emissions for pre-1990 closures. Methane degasification amounts were added to the quantity of CH_4 vented to determine the total CH_4 liberation rate for all mines that closed between 1992 and 2021. Since the sample of gassy mines described above is assumed to account for 78 percent of the pre-1972 and 98 percent of the post-1971 abandoned mine emissions, the modeled results were multiplied by 1.22 and 1.02, respectively, to account for all U.S. abandoned mine emissions.

From 1993 through 2021, emission totals were downwardly adjusted to reflect CH_4 emissions avoided from those abandoned mines with CH_4 recovery and use or destruction systems. The Inventory totals were not adjusted for abandoned mine CH_4 emissions avoided from 1990 through 1992, because no data was reported for abandoned coal mine CH_4 recovery and use or destruction projects during that time.

Uncertainty

A quantitative uncertainty analysis was conducted for the abandoned coal mine source category using the IPCC-recommended Approach 2 uncertainty estimation methodology. The uncertainty analysis provides for the specification of probability density functions for key variables within a computational structure that mirrors the calculation of the Inventory estimate. The results provide the range within which, with 95 percent certainty, emissions from this source category are likely to fall.

As discussed above, the parameters for which values must be estimated for each mine to predict its decline curve are: 1) the coal's adsorption isotherm; 2) CH_4 flow capacity as expressed by permeability; and 3) pressure at abandonment. Because these parameters are not available for each mine, a methodological approach to estimating emissions was used that generates a probability distribution of potential outcomes based on the most likely value and the probable range of values for each parameter. The range of values is not meant to capture the extreme values, but rather values that represent the highest and lowest quartile of the cumulative probability density function of each parameter. Once the low, mid, and high values are selected, they are applied to a probability density function.

The results of the Approach 2 quantitative uncertainty analysis are summarized in Table 3-42. Annual abandoned coal mine CH_4 emissions in 2021 were estimated to be between 5.0 and 7.7 MMT CO_2 Eq. at a 95 percent confidence level. This indicates a range of 22 percent below to 21 percent above the 2021 emission estimate of 6.4 MMT CO_2 Eq. One of the reasons for the relatively narrow range is that mine-specific data is available for use in the methodology for mines closed in 1972 and later years. Emissions from mines closed prior to 1972 have the largest degree of uncertainty because no mine-specific CH_4 liberation rates at the time of abandonment exist.

Table 3-42: Approach 2 Quantitative Uncertainty Estimates for CH_4 Emissions from Abandoned Underground Coal Mines (MMT CO_2 Eq. and Percent)

Source	Gas	2021 Emission Estimate (MMT CO_2 Eq.)	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Abandoned Underground Coal Mines	CH_4	6.4	5.0	7.7	-21.7%	+20.6%

[a] Range of emission estimates predicted by Monte Carlo Simulation for a 95 percent confidence interval.

QA/QC and Verification

In order to ensure the quality of the emission estimates for abandoned coal mines, general (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and reported emissions data used for estimating emissions from abandoned coal mines. Trends across the time series were analyzed to determine whether any corrective actions were needed.

Recalculations Discussion

For the current Inventory, estimates of CO_2-equivalent CH_4 emissions from abandoned coal mines have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth Assessment Report* (AR4) (IPCC 2007) (used in the previous inventories). The AR5 GWPs have been applied across the entire time series for consistency. The GWP of CH_4 increased from 25 to 28, leading to an overall increase in CO_2-equivalent CH_4 emissions. Compared to the previous Inventory which applied 100-year GWP values from AR4, the average annual change in CO_2-equivalent CH_4 emissions was 12 percent increase for each year of the time series. Further discussion on this update and the overall impacts of updating the inventory GWPs to reflect the IPCC *Fifth Assessment Report* can be found in Chapter 9, Recalculations and Improvements.

3.6 Petroleum Systems (CRF Source Category 1B2a)

Note that this draft of the Inventory does not yet incorporate updated activity data products for the following data inputs, due to a data base subscription lapse: oil well counts, wells drilled, wells completed, and production. Year 2020 values for activity data are used in place of year 2021. The Final Inventory (to be published April 2023) will incorporate the latest activity data.

This IPCC category (1B2a) is for fugitive emissions from petroleum systems, which per IPCC guidelines include emissions from leaks, venting, and flaring. Methane emissions from petroleum systems are primarily associated with onshore and offshore crude oil exploration, production, transportation, and refining operations. During these activities, CH_4 is released to the atmosphere as emissions from leaks, venting (including emissions from operational upsets), and flaring. Carbon dioxide emissions from petroleum systems are primarily associated with onshore and offshore crude oil production and refining operations. Note, CO_2 emissions in petroleum systems exclude all combustion emissions (e.g., engine combustion) except for flaring CO_2 emissions. All combustion CO_2 emissions (except for flaring) are accounted for in the fossil fuel combustion chapter (see Section 2). Emissions of N_2O from petroleum systems are primarily associated with flaring. Total greenhouse gas emissions (CH_4, CO_2, and N_2O) from petroleum systems in 2021 were 74.8 MMT CO_2 Eq., an increase of 23 percent from 1990, primarily due to increases in CO_2 emissions. Total emissions increased by 10 percent from 2010 levels and have decreased by 10 percent since 2020. Total CO_2 emissions from petroleum systems in 2021 were 24.67 MMT CO_2 (24,667 kt CO_2), 2.6 times higher than in 1990. Total CO_2 emissions in 2021 were 1.8 times higher than in 2010 and 15 percent lower than in 2020. Total CH_4 emissions from petroleum systems in 2021 were 50.2 MMT CO_2 Eq. (1,791 kt CH_4), a decrease of 2 percent from 1990. Since 2010, total CH_4 emissions decreased by 8 percent; and since 2020, CH_4 emissions decreased by 8 percent. Total N_2O emissions from petroleum systems in 2021 were 0.022 MMT CO_2 Eq. (0.082 kt N_2O), 1.7 times higher than in 1990, 1.2 times higher than in 2010, and 34 percent lower than in 2020. Since 1990, U.S. oil production has increased by 46 percent. In 2021, U.S. oil production was 105 percent higher than in 2010 and 1 percent lower than in 2020.

Each year, some estimates in the Inventory are recalculated with improved methods and/or data. These improvements are implemented consistently across the entire Inventory's time series (i.e., 1990 to 2021) to ensure that the trend is representative of changes in emissions levels. Recalculations in petroleum systems in this year's Inventory include:

- Updates to oil and gas production volumes using the most recent data from the United States Energy Information Administration (EIA)
- Recalculations due to Greenhouse Gas Reporting Program (GHGRP) submission revisions
- Recalculations due to methodological updates to four onshore production segment sources - pneumatic controllers, equipment leaks, chemical injection pumps, and storage tanks.
- Recalculations due to updating the global warming potential (GWP) for CH_4 and N_2O to use AR5 values.

Updated well counts and produced water volumes were not available for Public Review estimates, and 2021 data were set equal to 2020. The latest data will be incorporated into the final Inventory.

The Recalculations Discussion section below provides more details on the updated methods.

Exploration. Exploration includes well drilling, testing, and completion. Exploration accounts for less than 0.5 percent of total CH_4 emissions (including leaks, vents, and flaring) from petroleum systems in 2021. The predominant sources of CH_4 emissions from exploration are hydraulically fractured oil well completions. Other sources include well testing, well drilling, and well completions without hydraulic fracturing. Since 1990, exploration CH_4 emissions have decreased 96 percent, and while the number of hydraulically fractured wells completed increased 64 percent, there were decreases in the fraction of such completions without reduced emissions completions (RECs) or flaring. Emissions of CH_4 from exploration were highest in 2012, over 60 times higher than in 2021; and lowest in 2021. Emissions of CH_4 from exploration decreased 52 percent from 2020 to 2021, due to a decrease in emissions from hydraulically fractured oil well completions without RECs, as well as due to hydraulically fractured oil well completions with RECs and venting. Exploration accounts for 2 percent of total CO_2 emissions (including leaks, vents, and flaring) from petroleum systems in 2021. Emissions of CO_2 from exploration in 2021 were 28 percent higher than in 1990, and decreased by 44 percent from 2020, largely due to a decrease in the number of hydraulically fractured oil well completions without RECS or flaring (by 36 percent from 2020). Emissions of CO_2 from exploration were highest in 2014, over 8 times higher than in 2021. Exploration accounts for 1 percent of total N_2O emissions from petroleum systems in 2021. Emissions of N_2O from exploration in 2021 are 35 percent higher than in 1990, and 39 percent lower than in 2020, due to the above-mentioned changes in hydraulically fractured oil well completions with flaring.

Production. Production accounts for 98 percent of total CH_4 emissions (including leaks, vents, and flaring) from petroleum systems in 2021. The predominant sources of emissions from production field operations are pneumatic controllers, offshore oil platforms, equipment leaks, chemical injection pumps, gas engines, produced water, and associated gas flaring. In 2021, these seven sources together accounted for 94 percent of the CH_4 emissions from production. Since 1990, CH_4 emissions from production have increased by 6 percent primarily due to increases in emissions from pneumatic devices. Overall, production segment CH_4 emissions decreased by 8 percent from 2020 levels due primarily to lower pneumatic controller emissions. The number of high- and intermittent-bleed pneumatic controllers decreased from 2020 to 2021 whereas, the number of low-bleed pneumatic controllers increased from 2020 to 2021. Production emissions account for 81 percent of the total CO_2 emissions (including leaks, vents, and flaring) from petroleum systems in 2021. The principal sources of CO_2 emissions are associated gas flaring, miscellaneous production flaring, and oil tanks with flares. In 2021, these three sources together accounted for 97 percent of the CO_2 emissions from production. In 2021, CO_2 emissions from production were 3.4 times higher than in 1990, due to increases in flaring emissions from associated gas flaring, miscellaneous production flaring, and tanks. Overall, in 2021, production segment CO_2 emissions decreased by 17 percent from 2020 levels primarily due to decreases in associated gas flaring and miscellaneous production flaring in the Permian and Williston Basins. Production emissions accounted for 48 percent of the total N_2O emissions from petroleum systems in 2021. The principal sources of N_2O emissions are associated gas flaring, oil tanks with flares, miscellaneous production flaring, and offshore flaring. In 2021, N_2O emissions from production were 115 percent higher than in 1990 and were 51 percent lower than in 2020.

Crude Oil Transportation. Emissions from crude oil transportation account for a very small percentage of the total emissions (including leaks, vents, and flaring) from petroleum systems. Crude oil transportation activities account for 0.4 percent of total CH_4 emissions from petroleum systems. Emissions from tanks, marine loading, and truck loading operations account for 78 percent of CH_4 emissions from crude oil transportation. Since 1990, CH_4 emissions from transportation have increased by 21 percent. In 2021, CH_4 emissions from transportation decreased by 3 percent from 2020 levels. Crude oil transportation activities account for less than 0.01 percent of total CO_2 emissions from petroleum systems. Emissions from tanks, marine loading, and truck loading operations account for 78 percent of CO_2 emissions from crude oil transportation.

Crude Oil Refining. Crude oil refining processes and systems account for 2 percent of total fugitive (including leaks, vents, and flaring) CH_4 emissions from petroleum systems. This low share is because most of the CH_4 in crude oil is removed or escapes before the crude oil is delivered to the refineries. There is a negligible amount of CH_4 in all refined products. Within refineries, flaring accounts for 52 percent of the CH_4 emissions, while delayed cokers, uncontrolled blowdowns, and equipment leaks account for 16, 13 and 9 percent, respectively. Fugitive CH_4 emissions from refining of crude oil have increased by 12 percent since 1990, and decreased 5 percent from 2020; however, like the transportation subcategory, this increase has had little effect on the overall emissions of CH_4 from petroleum systems. Crude oil refining processes and systems account for 17 percent of total fugitive (including leaks, vents, and flaring) CO_2 emissions from petroleum systems. Of the total fugitive CO_2 emissions from refining, almost all (about 99 percent) of it comes from flaring.[73] Since 1990, refinery fugitive CO_2 emissions increased by 28 percent and have decreased by less than 1 percent from the 2020 levels, due to a decrease in flaring. Flaring occurring at crude oil refining processes and systems accounts for 51 percent of total fugitive N_2O emissions from petroleum systems. In 2021, refinery fugitive N_2O emissions increased by 37 percent since 1990, and decreased by less than 1 percent from 2020 levels.

Table 3-43: Total Greenhouse Gas Emissions (CO_2, CH_4, and N_2O) from Petroleum Systems (MMT CO_2 Eq.)

Activity	1990		2005		2017	2018	2019	2020	2021
Exploration	4.7		6.3		2.3	3.8	2.9	1.2	0.6
Production	52.0		50.1		79.3	88.2	97.6	77.1	68.9
Transportation	0.2		0.1		0.2	0.2	0.3	0.2	0.2
Crude Refining	4.0		4.6		4.5	4.6	6.0	5.1	5.1
Total	**60.8**		**61.2**		**86.4**	**96.8**	**106.8**	**83.6**	**74.8**

[73] Petroleum Systems includes fugitive emissions (leaks, venting, and flaring). In many industries, including petroleum refineries, the largest source of onsite CO_2 emissions is often fossil fuel combustion, which is covered in Section 3.1 of this chapter.

Table 3-44: CH_4 Emissions from Petroleum Systems (MMT CO_2 Eq.)

Activity	1990	2005	2017	2018	2019	2020	2021
Exploration	**4.3**	**5.9**	**0.5**	**0.6**	**0.5**	**0.3**	**0.2**
Production	**46.1**	**44.0**	**60.3**	**59.0**	**58.2**	**53.0**	**48.9**
Pneumatic Controllers	21.3	23.3	38.1	35.3	24.8	31.7	28.4
Offshore Production	9.9	7.2	5.7	5.5	5.5	5.3	5.5
Equipment Leaks	2.3	2.9	3.3	3.7	3.9	3.2	3.3
Gas Engines	2.3	2.0	2.5	2.6	2.6	2.5	2.5
Produced Water	2.6	1.7	2.4	2.6	2.7	2.5	2.5
Chemical Injection Pumps	1.3	3.0	3.4	3.9	10.8	3.3	3.2
Assoc Gas Flaring	0.5	0.4	1.1	1.9	2.3	1.2	0.8
Other Sources	5.9	3.5	3.7	3.6	5.5	3.5	2.8
Crude Oil Transportation	**0.2**	**0.1**	**0.2**	**0.2**	**0.3**	**0.2**	**0.2**
Refining	**0.7**	**0.8**	**0.9**	**0.8**	**1.0**	**0.9**	**0.8**
Total	**51.3**	**50.9**	**61.9**	**60.6**	**59.9**	**54.5**	**50.2**

Table 3-45: CH_4 Emissions from Petroleum Systems (kt CH_4)

Activity	1990	2005	2017	2018	2019	2020	2021
Exploration	**154**	**211**	**17**	**20**	**16**	**12**	**6**
Production	**1,646**	**1,573**	**2,152**	**2,107**	**2,078**	**1,894**	**1,748**
Pneumatic Controllers	760	833	1,362	1,260	886	1,131	1,015
Offshore Production	353	259	205	197	196	188	195
Equipment Leaks	82	102	120	132	138	115	117
Gas Engines	82	71	89	92	94	89	89
Produced Water	91	62	84	93	98	89	89
Chemical Injection Pumps	47	105	121	139	387	116	116
Assoc Gas Flaring	20	14	38	66	82	43	28
Other Sources	211	125	133	128	197	124	99
Crude Oil Transportation	**7**	**5**	**8**	**8**	**9**	**8**	**8**
Refining	**26**	**30**	**33**	**30**	**36**	**31**	**30**
Total	**1,833**	**1,891**	**2,209**	**2,165**	**2,138**	**1,945**	**1,791**

Table 3-46: CO_2 Emissions from Petroleum Systems (MMT CO_2)

Activity	1990	2005	2017	2018	2019	2020	2021
Exploration	0.4	0.4	1.9	3.2	2.4	0.8	0.5
Production	5.9	6.1	19.0	29.2	39.4	24.0	20.0
Transportation	+	+	+	+	+	+	+
Crude Refining	3.3	3.7	3.6	3.7	5.0	4.2	4.2
Total	**9.5**	**10.2**	**24.5**	**36.1**	**46.9**	**29.1**	**24.7**

+ Does not exceed 0.05 MMT CO_2 Eq.

Table 3-47: CO_2 Emissions from Petroleum Systems (kt CO_2)

Activity	1990	2005	2017	2018	2019	2020	2021
Exploration	364	395	1,853	3,208	2,434	838	467
Production	5,869	6,097	19,025	29,187	39,429	24,000	19,985
Transportation	0.9	0.7	1.1	1.2	1.3	1.2	1.1
Crude Refining	3,284	3,728	3,582	3,706	5,009	4,242	4,214
Total	**9,519**	**10,221**	**24,462**	**36,102**	**46,874**	**29,081**	**24,667**

Table 3-48: N_2O Emissions from Petroleum Systems (Metric Tons CO_2 Eq.)

Activity	1990	2005	2017	2018	2019	2020	2021
Exploration	161	174	722	1,338	894	361	219
Production	4,907	5,465	13,450	25,638	26,522	21,665	10,539
Transportation	NE	NE	NE	NE	NE	NE	NE
Crude Refining	8,096	9,189	9,286	9,351	13,127	11,149	11,083
Total	**13,164**	**14,827**	**23,458**	**36,327**	**40,542**	**33,175**	**21,841**

NE (Not Estimated)

Table 3-49: N_2O Emissions from Petroleum Systems (Metric Tons N_2O)

Activity	1990	2005	2017	2018	2019	2020	2021
Exploration	0.6	0.7	2.7	5.0	3.4	1.4	0.8
Production	18.5	20.6	50.8	96.7	100.1	81.8	39.8
Transportation	NE	NE	NE	NE	NE	NE	NE
Crude Refining	30.5	34.7	35.0	35.3	49.5	42.1	41.8
Total	**49.7**	**56.0**	**88.5**	**137.1**	**153.0**	**125.2**	**82.4**

NE (Not Estimated)

Methodology and Time-Series Consistency

See Annex 3.5 for the full time series of emissions data, activity data, emission factors, and additional information on methods and data sources.

Petroleum systems includes emission estimates for activities occurring in petroleum systems from the oil wellhead through crude oil refining, including activities for crude oil exploration, production field operations, crude oil transportation activities, and refining operations. Generally, emissions are estimated for each activity by multiplying emission factors (e.g., emission rate per equipment or per activity) by corresponding activity data (e.g., equipment count or frequency of activity). Certain sources within petroleum refineries are developed using an IPCC Tier 3 approach (i.e., all refineries in the nation report facility-level emissions data to the GHGRP, which are included directly in the national emissions estimates here). Other estimates are developed with a Tier 2 approach. Tier 1 approaches are not used.

EPA received stakeholder feedback on updates in the Inventory through EPA's stakeholder process on oil and gas in the Inventory. Stakeholder feedback is noted below in Recalculations Discussion and Planned Improvements. More information on the stakeholder process can be found online.[74]

Emission Factors. Key references for emission factors include *Methane Emissions from the Natural Gas Industry by the Gas Research Institute and EPA* (GRI/EPA 1996), *Estimates of Methane Emissions from the U.S. Oil Industry* (EPA 1999), *Compilation of Air Pollutant Emission Factors, AP-42* (EPA 1997), *Global Emissions of Methane from Petroleum Sources* (API 1992), consensus of industry peer review panels, Bureau of Ocean Energy Management (BOEM) reports, *Nonpoint Oil and Gas Emission Estimation Tool* (EPA 2017), and analysis of GHGRP data (EPA 2022).

Emission factors for hydraulically fractured (HF) oil well completions and workovers (in four control categories) were developed using EPA's GHGRP data; year-specific data were used to calculate emission factors from 2016-forward and the year 2016 emission factors were applied to all prior years in the time series. The emission factors

[74] See https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems.

for well testing and associated gas venting and flaring were developed using year-specific GHGRP data for years 2015 forward; earlier years in the time series use 2015 emission factors. For miscellaneous production flaring, year-specific emission factors were developed for years 2015 forward from GHGRP data, an emission factor of 0 (assumption of no flaring) was assumed for 1990 through 1992, and linear interpolation was applied to develop emission factors for 1993 through 2014. For more information, please see memoranda available online.[75] For offshore oil production, emission factors were calculated using BOEM data for offshore facilities in federal waters of the Gulf of Mexico (and these data were also applied to facilities located in state waters of the Gulf of Mexico) and GHGRP data for offshore facilities off the coasts of California and Alaska. For many other sources, emission factors were held constant for the period 1990 through 2021, and trends in emissions reflect changes in activity levels. Emission factors from EPA 1999 are used for all other production and transportation activities.

For associated gas venting and flaring and miscellaneous production flaring, emission factors were developed on a production basis (i.e., emissions per unit oil produced). Additionally, for these two sources, basin-specific activity and emission factors were developed for each basin that in any year from 2011 forward contributed at least 10 percent of total source emissions (on a CO_2 Eq. basis) in the GHGRP. For associated gas venting and flaring, basin-specific factors were developed for four basins: Williston, Permian, Gulf Coast, and Anadarko. For miscellaneous production flaring, basin-specific factors were developed for three basins: Williston, Permian, and Gulf Coast. For each source, data from all other basins were combined, and activity and emission factors were developed for the other basins as a single group.

For pneumatic controllers and tanks, basin-specific emission factors were calculated for all the basins reporting to the GHGRP. These emission factors were calculated for all the years with applicable GHGRP data (i.e., 2011 - 2021 or 2015 - 2021). For the remaining basins (i.e., basins not reporting to the GHGRP), subpart W average emission factors were used. For more information, please see memoranda available online.[3]

For the exploration and production segments, in general, CO_2 emissions for each source were estimated with GHGRP data or by multiplying CO_2 content factors by the corresponding CH_4 data, as the CO_2 content of gas relates to its CH_4 content. Sources with CO_2 emission estimates calculated using GHGRP data include HF completions and workovers, associated gas venting and flaring, tanks, well testing, pneumatic controllers, chemical injection pumps, miscellaneous production flaring, and certain offshore production facilities (those located off the coasts of California and Alaska). For these sources, CO_2 was calculated using the same methods as used for CH_4. Carbon dioxide emission factors for offshore oil production in the Gulf of Mexico were derived using data from BOEM, following the same methods as used for CH_4 estimates. For other sources, the production field operations emission factors for CO_2 are generally estimated by multiplying the CH_4 emission factors by a conversion factor, which is the ratio of CO_2 content and CH_4 content in produced associated gas.

For the exploration and production segments, N_2O emissions were estimated for flaring sources using GHGRP or BOEM OGOR-B data and the same method used for CO_2. Sources with N_2O emissions in the exploration segment include well testing and HF completions with flaring. Sources with N_2O emissions in the production segment include associated gas flaring, tank flaring, miscellaneous production flaring, HF workovers with flaring, and flaring from offshore production sources.

For crude oil transportation, emission factors for CH_4 were largely developed using data from EPA (1997), API (1992), and EPA (1999). Emission factors for CO_2 were estimated by multiplying the CH_4 emission factors by a conversion factor, which is the ratio of CO_2 content and CH_4 content in whole crude post-separator.

For petroleum refining activities, year-specific emissions from 2010 forward were directly obtained from EPA's GHGRP. All U.S. refineries have been required to report CH_4, CO_2, and N_2O emissions for all major activities starting with emissions that occurred in 2010. The reported total CH_4, CO_2, and N_2O emissions for each activity was used for the emissions in each year from 2010 forward. To estimate emissions for 1990 to 2009, the 2010 to 2013 emissions data from GHGRP along with the refinery feed data for 2010 to 2013 were used to derive CH_4 and CO_2 emission factors (i.e., sum of activity emissions/sum of refinery feed) and 2010 to 2017 data were used to derive

[75] See https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems.

N_2O emission factors; these emission factors were then applied to the annual refinery feed in years 1990 to 2009. GHGRP delayed coker CH_4 emissions for 2010 through 2017 were increased using the ratio of certain reported emissions for 2018 to 2017, to account for a more accurate GHGRP calculation methodology that was implemented starting in reporting year 2018.

A complete list of references for emission factors and activity data by emission source is provided in Annex 3.5.

Activity Data. References for activity data include Enverus data (Enverus 2021), Energy Information Administration (EIA) reports, *Methane Emissions from the Natural Gas Industry by the Gas Research Institute and EPA* (EPA/GRI 1996), *Estimates of Methane Emissions from the U.S. Oil Industry* (EPA 1999), consensus of industry peer review panels, BOEM reports, the Oil & Gas Journal, the Interstate Oil and Gas Compact Commission, the United States Army Corps of Engineers, and analysis of GHGRP data (EPA 2022).Enverus data for 2021 are not currently available; this version of the Inventory uses 2020 data as proxy for 2021.

For pneumatic controllers, equipment leaks, chemical injection pumps, and tanks, basin-specific activity factors were calculated for all the basins reporting to the GHGRP. These factors were calculated for all the years with applicable GHGRP data (i.e., 2011 through 2021 or 2015 through 2021). For the remaining basins (i.e., basins not reporting to the GHGRP), GHGRP average activity factors were used. For more information, please see memoranda available online.[76]

For many sources, complete activity data were not available for all years of the time series. In such cases, one of three approaches was employed to estimate values, consistent with IPCC good practice. Where appropriate, the activity data were calculated from related statistics using ratios developed based on EPA/GRI (1996) and/or GHGRP data. In some cases, activity data are developed by interpolating between recent data points (such as from GHGRP) and earlier data points, such as from EPA/GRI (1996). Lastly, in limited instances the previous year's data were used if current year data were not yet available.

A complete list of references for emission factors and activity data by emission source is provided in Annex 3.5. The United States reports data to the UNFCCC using this Inventory report along with Common Reporting Format (CRF) tables. This note is provided for those reviewing the CRF tables: The notation key "IE" is used for CO_2 and CH_4 emissions from venting and flaring in CRF table 1.B.2. Disaggregating flaring and venting estimates across the Inventory would involve the application of assumptions and could result in inconsistent reporting and, potentially, decreased transparency. Data availability varies across segments within oil and gas activities systems, and emission factor data available for activities that include flaring can include emissions from multiple sources (flaring, venting and leaks).

As noted above, EPA's GHGRP data, available starting in 2010 for refineries and in 2011 for other sources, have improved estimates of emissions from petroleum systems. Many of the previously available datasets were collected in the 1990s. To develop a consistent time series for sources with new data, EPA reviewed available information on factors that may have resulted in changes over the time series (e.g., regulations, voluntary actions) and requested stakeholder feedback on trends as well. For most sources, EPA developed annual data for 1993 through 2009 or 2014 by interpolating activity data or emission factors or both between 1992 (when GRI/EPA data are available) and 2010 or 2015 data points. Information on time-series consistency for sources updated in this year's Inventory can be found in the Recalculations Discussion below, with additional detail provided in supporting memos (relevant memos are cited in the Recalculations Discussion). For information on other sources, please see the Methodology Discussion above and Annex 3.5.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2021.

[76] See https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems.

Uncertainty— TO BE UPDATED FOR FINAL INVENTORY REPORT

EPA conducted a quantitative uncertainty analysis using the IPCC Approach 2 methodology (Monte Carlo Simulation technique) to characterize uncertainty for petroleum systems. For more information on the approach, please see the memoranda *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2016: Natural Gas and Petroleum Systems Uncertainty Estimates* and *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2019: Update for Natural Gas and Petroleum Systems CO_2 Uncertainty Estimates*.[77]

EPA used Microsoft Excel's @RISK add-in tool to estimate the 95 percent confidence bound around CH_4 and CO_2 emissions from petroleum systems for the current Inventory. For the CH_4 uncertainty analysis, EPA focused on the six highest methane-emitting sources for the year 2020, which together emitted 76 percent of methane from petroleum systems in 2020, and extrapolated the estimated uncertainty for the remaining sources For the CO_2 uncertainty analysis, EPA focused on the 3 highest-emitting sources for the year 2020 which together emitted 80 percent of CO_2 from petroleum systems in 2020, and extrapolated the estimated uncertainty for the remaining sources. The @RISK add-in provides for the specification of probability density functions (PDFs) for key variables within a computational structure that mirrors the calculation of the inventory estimate. The IPCC guidance notes that in using this method, "some uncertainties that are not addressed by statistical means may exist, including those arising from omissions or double counting, or other conceptual errors, or from incomplete understanding of the processes that may lead to inaccuracies in estimates developed from models." As a result, the understanding of the uncertainty of emission estimates for this category evolves and improves as the underlying methodologies and datasets improve. The uncertainty bounds reported below only reflect those uncertainties that EPA has been able to quantify and do not incorporate considerations such as modeling uncertainty, data representativeness, measurement errors, misreporting or misclassification. To estimate uncertainty for N_2O, EPA applied the uncertainty bounds calculated for CO_2. EPA will seek to refine this estimate in future Inventories.

The results presented below provide the 95 percent confidence bound within which actual emissions from this source category are likely to fall for the year 2020, using the recommended IPCC methodology. The results of the Approach 2 uncertainty analysis are summarized in Table 3-50. Petroleum systems CH_4 emissions in 2020 were estimated to be between 29.0 and 53.1 MMT CO_2 Eq., while CO_2 emissions were estimated to be between 23.5 and 38.0 MMT CO_2 Eq. at a 95 percent confidence level. Petroleum systems N_2O emissions in 2020 were estimated to be between 0.03 and 0.05 MMT CO_2 Eq. at a 95 percent confidence level.

Uncertainty bounds for other years of the time series have not been calculated, but uncertainty is expected to vary over the time series. For example, years where many emission sources are calculated with interpolated data would likely have higher uncertainty than years with predominantly year-specific data. In addition, the emission sources that contribute the most to CH_4 and CO_2 emissions are different over the time series, particularly when comparing recent years to early years in the time series. For example, associated gas venting emissions were higher and flaring emissions were lower in early years of the time series, compared to recent years. Technologies also changed over the time series (e.g., reduced emissions completions were not used early in the time series).

Table 3-50: Approach 2 Quantitative Uncertainty Estimates for CH_4 and CO_2 Emissions from Petroleum Systems (MMT CO_2 Eq. and Percent)

Source	Gas	2020 Emission Estimate (MMT CO_2 Eq.)[b]	Uncertainty Range Relative to Emission Estimate[a]			
			(MMT CO_2 Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Petroleum Systems	CH_4	40.2	29.0	53.1	-28%	+32%
Petroleum Systems	CO_2	30.2	23.5	38.0	-22%	+26%
Petroleum Systems	N_2O	0.04	0.03	0.05	-22%	+26%

[77] See https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems.

[a] Range of emission estimates estimated by applying the 95 percent confidence intervals obtained from the Monte Carlo Simulation analysis conducted for the year 2020 CH_4 and CO_2 emissions.
[b] All reported values are rounded after calculation. As a result, lower and upper bounds may not be duplicable from other rounded values as shown in table.

QA/QC and Verification Discussion

The petroleum systems emission estimates in the Inventory are continually being reviewed and assessed to determine whether emission factors and activity factors accurately reflect current industry practices. A QA/QC analysis was performed for data gathering and input, documentation, and calculation. QA/QC checks are consistently conducted to minimize human error in the emission calculations. EPA performs a thorough review of information associated with new studies, GHGRP data, regulations, public webcasts, and the Natural Gas STAR Program to assess whether the assumptions in the Inventory are consistent with current industry practices. EPA has a multi-step data verification process for GHGRP data, including automatic checks during data-entry, statistical analyses on completed reports, and staff review of the reported data. Based on the results of the verification process, EPA follows up with facilities to resolve mistakes that may have occurred.[78]

As in previous years, EPA conducted early engagement and communication with stakeholders on updates prior to public review of the current Inventory. EPA held stakeholder webinars on greenhouse gas data for oil and gas in September and November of 2022. EPA released memos detailing updates under consideration and requesting stakeholder feedback. Stakeholder feedback received through these processes is discussed in the Recalculations Discussion and Planned Improvements sections below.

In recent years, several studies have measured emissions at the source level and at the national or regional level and calculated emission estimates that may differ from the Inventory. There are a variety of potential uses of data from new studies, including replacing a previous estimate or factor, verifying or QA of an existing estimate or factor, and identifying areas for updates. In general, there are two major types of studies related to oil and gas greenhouse gas data: studies that focus on measurement or quantification of emissions from specific activities, processes, and equipment, and studies that use tools such as inverse modeling to estimate the level of overall emissions needed to account for measured atmospheric concentrations of greenhouse gases at various scales. The first type of study can lead to direct improvements to or verification of Inventory estimates. In the past few years, EPA has reviewed, and in many cases, incorporated data from these data sources. The second type of study can provide general indications on potential over- and under-estimates.

A key challenge in using these types of studies to assess Inventory results is having a relevant basis for comparison (e.g., the two data sets should have comparable time frames and geographic coverage, and the independent study should assess data from the Inventory and not another data set, such as the Emissions Database for Global Atmospheric Research, or “EDGAR”). In an effort to improve the ability to compare the national-level Inventory with measurement results that may be at other spatial and temporal scales, a team at Harvard University along with EPA and other coauthors developed a gridded inventory of U.S. anthropogenic methane emissions with 0.1 degree x 0.1 degree spatial resolution, monthly temporal resolution, and detailed scale-dependent error characterization.[79] The gridded methane inventory is designed to be consistent with the U.S. EPA’s *Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990-2014* estimates for the year 2012, which presents national totals.[80] An updated version of the gridded inventory is being developed and will improve efforts to compare results of the inventory with atmospheric studies.

[78] See https://www.epa.gov/sites/production/files/2015-07/documents/ghgrp_verification_factsheet.pdf.
[79] See https://www.epa.gov/ghgemissions/gridded-2012-methane-emissions.
[80] See https://www.epa.gov/ghgemissions/us-greenhouse-gas-inventory-report-1990-2014.

As discussed above, refinery emissions are quantified by using the total emissions reported to GHGRP for the refinery emission categories included in Petroleum Systems. Subpart Y has provisions that refineries are not required to report under Subpart Y if their emissions fall below certain thresholds. Each year, a review is conducted to determine whether an adjustment is needed to the Inventory emissions to include emissions from refineries that stopped reporting to the GHGRP. Based on the review of the most recent GHGRP data, EPA identified a refinery last reported annual emissions data to the GHGRP for reporting year 2012, due to meeting the criteria for cessation of reporting. EPA used the 2012 reported emissions for the refinery as proxy to gap fill annual emissions for 2013 through 2020 for this refinery.

Recalculations Discussion

EPA received information and data related to the emission estimates through GHGRP reporting and stakeholder feedback on updates under consideration. In October 2022, EPA released a draft memorandum that discussed changes under consideration and requested stakeholder feedback on those changes.[81] EPA did not receive written feedback on the memorandum. Memoranda cited in the Recalculations Discussion below are: Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2021: Updates Under Consideration for Incorporating Additional Geographically Disaggregated Data (*Disaggregation* memo) and Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2021: Updates Under Consideration for Incorporating Additional Geographically Disaggregated Data for the Production Segment (*Production Disaggregation* memo).

In this Inventory, an update that incorporates additional basin-level data from GHGRP subpart W was implemented for several emission sources in the onshore production segment. The update seeks to improve the ability of EPA's gridded and state inventories to reflect variation due to differences in formation types, technologies and practices, regulations, or voluntary initiatives, and not only the differences in key activity levels that are reflected in the current gridded and state inventories. This would allow EPA to use the gridded inventory for improved comparisons of the national Inventory with various atmospheric observation studies (since regions will better reflect the local differences in emissions rates as reported to GHGRP) and would allow the state-level inventory to reflect differences in state-level programs, formation type mixes, and varying technologies and practices. For many sources, an approach that develops estimates using geographically disaggregated data may not be possible or preferable to a national level approach based on the currently available data. For some emission sources in the Inventory, emission factor data come from research studies and are applied at the national level. For example, many of the emission factors used to quantify emissions in the Inventory for the gathering and boosting, transmission and storage, distribution, and post-meter segments are from research studies and do not have a level of detail or total population comparable to GHGRP. For petroleum refineries, because there is no reporting threshold for GHGRP Subpart Y, facility-level data are generally available for all refineries in the United States, and these site-specific data are already used to develop the gridded and state-level greenhouse gas estimates. Even in cases where geographically disaggregated data are available, such an approach may not always be preferable. In cases with limited variation between areas, such an approach would have limited impact on emissions estimates regionally or nationally. In cases with limited data in certain areas, disaggregated approaches might substantially increase the uncertainty of estimates and basin-specific calculations would not be an improvement over use of a national average. EPA continues to seek stakeholder feedback on the draft approach in this Inventory.

EPA evaluated relevant information available and made several updates to the Inventory, including for pneumatic controllers, equipment leaks, chemical injection pumps, and storage tanks. For each of these emission sources, EPA modified the calculation methodology to use GHGRP data to develop basin-specific activity factors and/or emission factors. General information for these source specific recalculations are presented below and details (including the basin-specific emissions estimates) are available in the *Disaggregation* memo and *Production Disaggregation* memo.

[81] Stakeholder materials including draft memoranda for the current (i.e., 1990 to 2021) Inventory are available at https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems.

In addition to the updates to production segment sources mentioned above, for certain sources, CH_4 and/or CO_2 emissions changed by greater than 0.05 MMT CO_2 Eq., comparing the previous estimate for 2020 to the current (recalculated) estimate for 2020. The emissions changes were mostly due to GHGRP data submission revisions. These sources are discussed below and include associated gas flaring, miscellaneous production flaring, offshore production, and refinery flaring.

In addition, for the current Inventory, CO_2-equivalent emissions totals have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth Assessment Report* (AR4) (IPCC 2007) used in the previous inventories. The AR5 GWPs have been applied across the entire time series for consistency. The GWP of CH_4 has increased from 25 to 28, leading to an increase in the calculated CO_2-equivalent emissions of CH_4, while the GWP of N_2O has decreased from 298 to 265, leading to a decrease in the calculated CO_2-equivalent emissions of N_2O. Further discussion on this update and the overall impacts of updating the Inventory GWP values to reflect the IPCC *Fifth Assessment Report* can be found in Chapter 9, Recalculations and Improvements.

The combined impact of revisions to 2020 petroleum systems CH_4 emission estimates on a CO_2-equivalent basis, compared to the previous Inventory, is an increase from 40.2 to 54.5 MMT CO_2 Eq. (14.2 MMT CO_2 Eq., or 35 percent). The recalculations resulted in higher CH_4 emission estimates on average across the 1990 through 2020 time series, compared to the previous Inventory, by 11.0 MMT CO_2 Eq., or 25 percent.

The combined impact of revisions to 2020 petroleum systems CO_2 emission estimates, compared to the previous Inventory, is a decrease from 30.2 to 29.1 MMT CO_2 (1.1 MMT CO_2, or 4 percent). The recalculations resulted in lower emission estimates on average across the 1990 through 2020 time series, compared to the previous Inventory, by 1.2 MMT CO_2 Eq., or 9 percent.

The combined impact of revisions to 2020 petroleum systems N_2O emission estimates on a CO_2-equivalent basis, compared to the previous Inventory, is a decrease of 0.004 MMT CO_2, Eq. or 12 percent. The recalculations resulted in an average decrease in emission estimates across the 1990 through 2020 time series, compared to the previous Inventory, of 0.002 MMT CO_2 Eq., or 11 percent.

In Table 3-51 and Table 3-52 below are categories in Petroleum Systems with updated methodologies or with recalculations resulting in a change of greater than 0.05 MMT CO_2 Eq., comparing the previous estimate for 2020 to the current (recalculated) estimate for 2020. For more information, please see the discussion below.

For certain sources, CH_4 emissions for 2020 changed by greater than 0.05 MMT CO_2 Eq., compared to the previous Inventory due to the use of an updated GWP value (AR5). These sources are not discussed below and include associated gas venting and flaring, produced water, gas engines, heaters, and refineries.

Table 3-51: Recalculations of CO_2 in Petroleum Systems (MMT CO_2)

Segment/Source	*Previous Estimate Year 2020, 2022 Inventory*	Current Estimate Year 2020, 2023 Inventory	Current Estimate Year 2021, 2023 Inventory
Exploration	***0.9***	**0.8**	**0.5**
Production	***25.0***	**24.0**	**20.0**
Tanks	*6.5*	5.3	5.4
Pneumatic Controllers	*0.1*	0.1	0.1
Equipment Leaks	*+*	+	+
Chemical Injection Pumps	*+*	+	+
Associated Gas Flaring	*13.0*	13.3	9.6
Miscellaneous Production Flaring	*4.6*	4.7	4.2
Transportation	***+***	**+**	**+**
Refining	***4.3***	**4.2**	**4.2**
Flares	4.3	4.2	4.2
Petroleum Systems Total	***30.2***	**29.1**	**24.7**

+ Does not exceed 0.05 MMT CO_2.

Table 3-52: Recalculations of CH_4 in Petroleum Systems (MMT CO_2 Eq.)

Segment/Source	*Previous Estimate Year 2020, 2022 Inventory*	Current Estimate Year *2020*, 2023 Inventory	Current Estimate Year 2021, 2023 Inventory
Exploration	***0.3***	**0.3**	**0.2**
Production	***38.9***	**53.0**	**48.9**
Tanks	*0.7*	0.8	0.6
Pneumatic Controllers	*21.3*	31.7	28.4
Equipment Leaks	*2.4*	3.2	3.3
Chemical Injection Pumps	*1.9*	3.3	3.2
Miscellaneous Production Flaring	*0.4*	0.6	0.5
Offshore Production	*4.8*	5.3	5.5
Transportation	***0.2***	**0.2**	**0.2**
Refining	***0.8***	**0.9**	**0.8**
Petroleum Systems Total	***40.2***	**54.5**	**50.2**

Exploration

Recalculations for the exploration segment have resulted in lower calculated CH_4 and CO_2 emissions over the time series (less than 0.1 percent), compared to the previous Inventory.

Production

Pneumatic Controllers (Methodological Update)

EPA updated the calculation methodology for pneumatic controllers to use basin-specific activity factors and emission factors calculated from subpart W data for each type of controller (i.e., high, intermittent, and low bleed). Previously, national average activity and emission factors calculated using subpart W data were applied to estimate pneumatic controller emissions. In this methodological update, EPA summed basin-level emissions together to develop national emissions. The *Disaggregation* memo and *Production Disaggregation* memo present additional information and considerations for this update.

EPA calculated basin-specific activity factors and CH_4 emission factors for all basins that reported subpart W data. The factors were year-specific for RY2011 through RY2021. EPA retained the previous Inventory's activity factor assumptions for 1990 through 1993 and applied linear interpolation between the 1993 and 2011 activity factors at the basin-level. Year 2011 emission factors were applied to all prior years for each basin. For basins without subpart W data available, EPA applied national average activity and emission factors.

The estimation methodology for CO_2 emissions was not updated to use the basin-specific approach for the public review version of the Inventory. CO_2 emissions were estimated by applying a CO_2 to CH_4 ratio to the estimated CH_4 emissions. EPA will calculate pneumatic controller CO_2 emissions in the same manner as CH_4 emissions for the final Inventory.

As a result of this methodological update, CH_4 emissions estimates are an average of 22 percent higher across the time-series and 32 percent higher in 2020, compared to the previous Inventory. The most significant changes are in recent years, 2013 through 2020, due specifically to changes in intermittent bleed controller emissions estimates. Certain basins (e.g., Anadarko Basin, Appalachian, Appalachian Basin (Eastern Overthrust), Bend Arch, Fort Worth Syncline, Gulf Coast, and Sedgwick) have higher activity factors (mainly the average number of controllers per well) and/or emission factors for intermittent bleed pneumatic controllers, compared to the national average. Some of these basins also exhibit large changes in emissions over these recent years.

Table 3-53: Pneumatic Controllers National CH_4 Emissions (Metric Tons CH_4)

Source	1990	2005	2017	2018	2019	2020	2021
High Bleed Controllers	708,800	493,011	89,472	73,438	73,278	87,884	48,202
Low Bleed Controllers	51,170	63,773	20,104	31,779	50,456	36,752	46,360
Intermittent Bleed Controllers	0	276,145	1,252,028	1,155,041	762,647	1,006,263	920,518
Total Emissions	**759,970**	**832,929**	**1,361,605**	**1,260,259**	**886,382**	**1,130,899**	**1,015,080**
Previous Estimate	*736,447*	*708,680*	*835,129*	*727,365*	*732,092*	*853,562*	*NA*

NA (Not Applicable)

Equipment Leaks (Methodological Update)

EPA updated the calculation methodology for onshore production equipment leaks to use basin-specific equipment-level activity factors (e.g., separators/well) from GHGRP data. Previously, national average equipment activity factors developed using RY2014 GHGRP data were used in the Inventory for all years. In this methodological update, EPA summed basin-level emissions together to develop national emissions. The *Disaggregation* memo and *Production Disaggregation* memo present additional information and considerations for this update.

EPA calculated basin-specific equipment-level activity factors for all basins that reported subpart W data. The factors were year-specific for RY2015 through RY2021. EPA retained the previous Inventory's activity factors for 1990 through 1993 and used linear interpolation between the 1993 and 2015 activity factors at the basin-level. For basins without subpart W data available, EPA applied national average activity factors using all subpart W data. This methodological update applies only for activity factors. The previous Inventory's CH_4 emission factors for onshore production segment equipment leaks (by equipment type) were retained and used to develop CH_4 estimates.

The calculation methodology for CO_2 emissions was not updated for the public review version of the Inventory. The previous Inventory's methodology was retained to develop CO_2 estimates. EPA will calculate equipment leak CO_2 emissions in the same manner as CH_4 emissions for the final Inventory.

This update resulted in CH_4 emissions an average of 18 percent higher across the time-series compared with the previous Inventory and a 21 percent higher estimate for 2020, compared to the previous Inventory. The emissions increase is due to certain basins having higher activity factors compared to the national average activity factors (e.g., Anadarko, Appalachian, Appalachian Basin (Eastern Overthrust), and Gulf Coast).

Table 3-54: Production Equipment Leaks National CH_4 Emissions (Metric Tons CH_4)

Source	1990	2005	2017	2018	2019	2020	2021
Oil Wellheads	56,524	51,563	60,557	59,195	60,877	58,632	60,029
Separators	10,970	17,514	30,021	42,001	38,510	29,356	27,107
Heater/Treaters	11,119	20,741	16,245	17,492	22,706	18,734	21,307
Headers	3,323	12,434	12,754	13,217	15,595	8,075	8,444
Total Emissions	**81,936**	**102,251**	**119,577**	**131,904**	**137,688**	**114,797**	**116,887**
Previous Estimate	*81,874*	*86,248*	*100,450*	*99,287*	*98,459*	*94,921*	*NA*

NA (Not Applicable)

Chemical injection Pumps (Methodological Update)

EPA updated the calculation methodology for chemical injection pumps to use basin-specific activity factors from GHGRP data. Previously, a national average activity factor developed using RY2014 GHGRP data was used in the Inventory for all years. In this methodological update, EPA summed basin-level emissions together to develop national emissions. The *Disaggregation* memo and *Production Disaggregation* memo present additional information and considerations for this update.

EPA calculated basin-specific activity factor for all basins that reported subpart W data. The factors were year-specific for RY2015 through RY2021. EPA also retained the previous Inventory's activity factor for 1990 through 1993 and used linear interpolation between the 1993 and 2015 activity factors at the basin-level. For basins without subpart W data available, EPA applied the national average unweighted activity factor from all subpart W data. This methodological update applies only to activity factors. The previous Inventory's CH_4 emission factor for chemical injection pumps was retained and used to develop CH_4 estimates.

The estimation methodology for CO_2 emissions was not updated for the public review version of the Inventory. The previous Inventory's methodology was retained to develop CO_2 estimates. EPA will calculate chemical injection pump CO_2 emissions in the same manner as CH_4 emissions for the final Inventory.

This update resulted in calculated CH_4 emissions an average of 63 percent higher across the time-series compared with the previous Inventory and 52 percent higher in 2020, compared to the previous Inventory. The emissions increase is due to certain basins having a higher activity factor compared to the national average activity factor (e.g., Anadarko Basin, Appalachian, Appalachian Basin (Eastern Overthrust), Bend Arch, Fort Worth Syncline, Green River, and Gulf Coast).

Table 3-55: Chemical Injection Pumps National CH_4 Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Chemical Injection Pumps	**47,401**		**105,458**		**121,469**	**138,866**	**387,416**	**116,080**	**115,678**
Previous Estimate	*46,758*		*67,685*		*80,728*	*79,793*	*79,128*	*76,284*	*NA*

NA (Not Applicable)

Storage Tanks (Methodological Update)

EPA updated the calculation methodology for production segment storage tanks to use basin-specific activity factors and emission factors, calculated from Ssubpart W data for each storage tank category. Previously, national annual average activity and emission factors calculated using subpart W data were applied to estimate storage tank emissions. In this update, EPA developed national emission estimates by summing calculated basin-level total emission estimates, using basin-level data emission and activity factors developed from Subpart W. The *Production Disaggregation* memo presents additional information and considerations for this update.

EPA calculated basin-specific activity factors and CH_4 and CO_2 emission factors for all basins that reported subpart W data. The factors were year-specific for reporting year (RY) 2015 through RY2021. EPA also retained the previous Inventory's activity factor assumptions (i.e., all oil tanks were uncontrolled in 1990) and used linear interpolation between the 1990 and 2015 activity factors at the basin-level. Year 2015 emission factors were applied to all prior years for each basin. For basins without Subpart W data available, EPA applied national average activity and emission factors (unweighted average of all Subpart W reported data).

This update resulted in oil tank CH_4 emission estimates that are on average 16 percent lower across the time series than in the previous Inventory. The CH_4 estimates for 2020 are 2 percent lower than in the previous Inventory. Oil tank CO_2 emissions are on average 55 percent lower across the time series than in the previous Inventory and 2020 emissions estimates are 20 percent lower than in the previous Inventory. The CH_4 emissions estimate decrease occurs mainly from 1990 through 2005, where there is an average decrease in calculated emissions of 39 percent, compared to the previous inventory. Oil tank CO_2 emissions have a similarly large decrease in that time frame.

The Arctic Coastal Plains Province Basin has a large impact on these earlier time series year emissions, when this basin accounts for a large percentage of total liquids production, but very little of the production in that basin is stored in tanks. Oil tank CO_2 emissions decreased in recent years of the time series due to certain basins with higher production (e.g., Denver Basin, Gulf Coast, Permian) having lower activity factors and emission factors than the national average.

Table 3-56: Storage Tanks National CH_4 Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Large Tanks w/Flares	0		993		5,142	6,330	4,226	3,715	3,108

Large Tanks w/VRU	0	721	9,334	2,410	2,320	1,026	513
Large Tanks w/o Control	105,668	40,150	42,112	42,679	26,491	21,294	12,290
Small Tanks w/Flares	0	15	45	16	23	29	68
Small Tanks w/o Flares	7,438	3,448	2,991	3,326	2,755	2,709	3,598
Malfunctioning Separator Dump Valves	2,397	1,472	4,247	785	428	338	320
Total Emissions	**115,503**	**46,799**	**63,871**	**55,546**	**36,243**	**29,112**	**19,896**
Previous Estimate	*218,419*	*60,186*	*61,098*	*57,412*	*35,266*	*29,613*	*NA*

NA (Not Applicable)

Table 3-57: Storage Tanks National CO_2 Emissions (kt CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
Large Tanks w/Flares	0	716	3,771	5,348	5,974	5,212	5,381
Large Tanks w/VRU	0	3	4	4	6	2	1
Large Tanks w/o Control	24	8	5	4	5	6	5
Small Tanks w/Flares	0	3	11	7	9	10	9
Small Tanks w/o Flares	12	5	4	5	4	4	5
Malfunctioning Separator Dump Valves	12	13	32	30	26	20	37
Total Emissions	**47**	**748**	**3,828**	**5,398**	**6,024**	**5,255**	**5,439**
Previous Estimate	*115*	*2,505*	*4,313*	*6,189*	*6,682*	*6,537*	*NA*

NA (Not Applicable)

Associated Gas Flaring (Recalculation with Updated Data)

Associated gas flaring CO_2 emission estimates are on average of 0.1 percent higher across the time series compared with the previous Inventory and in 2020 are 2 percent higher than in the previous Inventory. The emission changes were due to GHGRP data submission revisions.

Table 3-58: Associated Gas Flaring National CO_2 Emissions (kt CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
220 - Gulf Coast Basin (LA, TX)	225	124	749	645	712	801	410
360 - Anadarko Basin	102	63	62	79	18	10	8
395 - Williston Basin	969	1,243	6,954	10,698	15,334	8,257	6,772
430 - Permian Basin	2,844	1,971	3,141	6,700	7,333	3,605	1,942
"Other" Basins	944	507	384	633	1,006	619	486
Total Emissions	**5,084**	**3,908**	**11,291**	**18,756**	**24,403**	**13,293**	**9,619**
220 - Gulf Coast Basin (LA, TX)	*225*	*124*	*749*	*651*	*713*	*798*	*NA*
360 - Anadarko Basin	*102*	*63*	*62*	*79*	*18*	*10*	*NA*
395 - Williston Basin	*969*	*1,243*	*6,909*	*11,140*	*14,762*	*8,052*	*NA*
430 - Permian Basin	*2,844*	*1,971*	*3,141*	*6,711*	*7,227*	*3,558*	*NA*
"Other" Basins	*944*	*507*	*384*	*624*	*990*	*624*	*NA*
Previous Estimate	*5,084*	*3,908*	*11,245*	*19,206*	*23,710*	*13,041*	*NA*

NA (Not Applicable)

Miscellaneous Production Flaring

Miscellaneous production flaring CO_2 emission estimates are on average 0.3 percent higher across the time series than in the previous Inventory and in 2020 are 2 percent higher than in the previous Inventory. The emission estimate changes were due to GHGRP data submission revisions.

Table 3-59: Miscellaneous Production Flaring National CO_2 Emissions (kt CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
220 - Gulf Coast Basin (LA, TX)	0	105	509	584	616	651	787
395 - Williston Basin	0	72	537	1,701	2,643	852	882
430 - Permian Basin	0	209	1,465	1,406	4,320	2,798	2,216
"Other" Basins	0	400	551	615	646	378	270
Total Emissions	**0**	**786**	**3,063**	**4,307**	**8,225**	**4,679**	**4,154**
Previous Estimate	*0*	*786*	*3,031*	*4,166*	*7,989*	*4,589*	*NA*

NA (Not Applicable)

Miscellaneous production flaring CH_4 emission estimates are on average 2 percent higher across the time series compared with the previous inventory and in 2020 are 31 percent higher than calculated in the previous Inventory. The emission changes were due to GHGRP data submission revisions.

Table 3-60: Miscellaneous Production Flaring National CH_4 Emissions (Metric Tons CH_4)

Source	1990	2005	2017	2018	2019	2020	2021
220 - Gulf Coast Basin (LA, TX)	0	440	2,119	1,978	2,506	2,452	2,989
395 - Williston Basin	0	179	1,618	3,031	3,503	1,670	1,396
430 - Permian Basin	0	1,097	5,389	5,296	21,296	16,712	11,305
"Other" Basins	0	1,291	1,904	1,816	1,731	1,249	961
Total Emissions	**0**	**3,008**	**11,030**	**12,121**	**29,036**	**22,082**	**16,650**
Previous Estimate	*0*	*3,008*	*10,928*	*11,669*	*22,994*	*16,807*	*NA*

NA (Not Applicable)

Offshore Production (Recalculation with Updated Data)

Offshore production CH_4 emission estimates are on average less than 0.05 percent lower across the time series than in the previous Inventory. The 2020 value is 3 percent lower than in the previous Inventory. The emission changes were due to updated offshore complex counts.

Table 3-61: Offshore Production National CH_4 Emissions (Metric Tons CH_4)

Source	1990	2005	2017	2018	2019	2020	2021
GOM Federal Waters	302,936	219,285	187,433	183,236	181,488	173,336	179,891
GOM State Waters	5,657	665	96	60	71	60	59
Pacific Waters	22,609	17,659	5,052	3,794	3,370	4,262	4,554
Alaska State Waters	21,936	21,191	12,163	9,834	10,711	10,366	10,664
Total Emissions	**353,138**	**258,801**	**204,745**	**196,924**	**195,640**	**188,024**	**195,168**
Previous Estimate	*353,138*	*258,801*	*203,917*	*196,349*	*195,626*	*192,943*	*NA*

NA (Not Applicable)

Transportation

Recalculations for the transportation segment have resulted in calculated CH_4 and CO_2 emissions over the time series from this segment that are lower (by less than 0.2 percent) than in the previous Inventory.

Refining

Recalculations due to resubmitted GHGRP data in the refining segment have resulted in average calculated CH_4 emissions over the time series 3 percent lower than in the previous Inventory, and 2020 CH_4 emissions 0.9 lower than in the previous Inventory.

Refining CO_2 emission estimates are on average 0.3 percent lower across the time series than in the previous Inventory and 2 percent lower in 2020 than in the previous Inventory. This change is due to GHGRP resubmissions and was largely due to a change in reported flaring CO_2 emissions.

Table 3-62: Refining National CO_2 Emissions (kt CO_2)

Source	1990		2005		2017	2018	2019	2020	2021
Flares	3,134		3,557		3,509	3,643	4,961	4,208	4,182
Total Refining	3,284		3,728		3,582	3,706	5,009	4,242	4,214
Previous Estimate	*3,284*		*3,728*		*3,725*	*3,820*	*5,080*	*4,326*	*NA*

NA (Not Applicable)

Planned Improvements

Planned Improvements for 2023 Inventory

This draft of the Inventory does not yet incorporate updated activity data products for the following data inputs, due to a data base subscription lapse: oil well counts, wells drilled, wells completed, and production. For these inputs, year 2020 values for activity data are used in place of year 2021. The Final Inventory (to be published April 2023) will incorporate the latest activity data.

Basin-level approaches for pneumatic controllers, equipment leaks, and chemical injection pumps were applied to calculate CH_4 emissions for public review. For the final Inventory, EPA would apply consistent methods for both CO_2 and CH_4 emissions calculations.

Additional information on the update and specific requests for stakeholder feedback can be found in the *Disaggregation* memo and *Production Disaggregation* memos. Feedback EPA has received in response to the memo include that basin-level data from GHGRP can improve accuracy of estimates when applied appropriately, and that EPA should consider application of the approach to only basins with 50 percent coverage or more, EPA will consider this feedback and any additional feedback received and may revise the calculations in the Inventory based.

Upcoming Data, and Additional Data that Could Inform the Inventory

EPA will assess new data received by the Greenhouse Gas Reporting Program, the Methane Challenge Program and other relevant programs on an ongoing basis, which may be used to confirm or improve existing estimates and assumptions.

EPA continues to track studies that contain data that may be used to update the Inventory. EPA will also continue to assess studies that include and compare both top-down and bottom-up estimates, and which could lead to improved understanding of unassigned high emitters (e.g., identification of emission sources and information on frequency of high emitters) as recommended in previous stakeholder comments.

Box 3-6: Carbon Dioxide Transport, Injection, and Geological Storage

Carbon dioxide is produced, captured, transported, and used for Enhanced Oil Recovery (EOR) as well as commercial and non-EOR industrial applications, or is stored geologically. This CO_2 is produced from both naturally-occurring CO_2 reservoirs and from industrial sources such as natural gas processing plants and ammonia plants. In the Inventory, emissions of CO_2 from naturally-occurring CO_2 reservoirs are estimated based on the specific application.

In the Inventory, CO_2 that is used in non-EOR industrial and commercial applications (e.g., food processing, chemical production) is assumed to be emitted to the atmosphere during its industrial use. These emissions are discussed in the Carbon Dioxide Consumption section, 4.15.

For EOR CO_2, as noted in the *2006 IPCC Guidelines*, "At the Tier 1 or 2 methodology levels [EOR CO_2 is] indistinguishable from fugitive greenhouse gas emissions by the associated oil and gas activities." In the U.S. estimates for oil and gas fugitive emissions, the Tier 2 emission factors for CO_2 include CO_2 that was originally injected and is emitted along with other gas from leak, venting, and flaring pathways, as measurement data used to develop those factors would not be able to distinguish between CO_2 from EOR and CO_2 occurring in the produced natural gas. Therefore, EOR CO_2 emitted through those pathways is included in CO_2 estimates in 1B2.

IPCC includes methodological guidance to estimate emissions from the capture, transport, injection, and geological storage of CO_2. The methodology is based on the principle that the carbon capture and storage system should be handled in a complete and consistent manner across the entire Energy sector. The approach accounts for CO_2 captured at natural and industrial sites as well as emissions from capture, transport, and use. For storage specifically, a Tier 3 methodology is outlined for estimating and reporting emissions based on site-specific evaluations. However, IPCC (IPCC 2006) notes that if a national regulatory process exists, emissions information available through that process may support development of CO_2 emission estimates for geologic storage.

In the United States, facilities that produce CO_2 for various end-use applications (including capture facilities such as acid gas removal plants and ammonia plants), importers of CO_2, exporters of CO_2, facilities that conduct geologic sequestration of CO_2, and facilities that inject CO_2 underground, are required to report greenhouse gas data annually to EPA through its GHGRP. Facilities reporting geologic sequestration of CO_2 to the GHGRP develop and implement an EPA-approved site-specific monitoring, reporting and verification plan, and report the amount of CO_2 sequestered using a mass balance approach.

GHGRP data relevant for this Inventory estimate consists of national-level annual quantities of CO_2 captured and extracted for EOR applications for 2010 to 2021 and data reported for geologic sequestration from 2016 to 2021.

The amount of CO_2 captured and extracted from natural and industrial sites for EOR applications in 2020 is 35,090 kt (35.1 MMT CO_2 Eq.) (see 6). The quantity of CO_2 captured and extracted is noted here for information purposes only; CO_2 captured and extracted from industrial and commercial processes is generally assumed to be emitted and included in emissions totals from those processes.

Table 3-63: Quantity of CO_2 Captured and Extracted for EOR Operations (kt CO_2)

Stage	2017	2018	2019	2020	2021
Quantity of CO_2 Captured and Extracted for EOR Operations	49,600	48,400	52,100	35,210	35,090

Several facilities are reporting under GHGRP Subpart RR (Geologic Sequestration of Carbon Dioxide). See Table 3-64 for the number of facilities reporting under Subpart RR, the reported CO_2 sequestered in subsurface geologic formations in each year, and of the quantity of CO_2 emitted from equipment leaks in each year. The quantity of CO_2 sequestered and emitted is noted here for information purposes only; EPA is considering updates to its approach in the Inventory for this source for future Inventories.

Table 3-64: Geologic Sequestration Information Reported Under GHGRP Subpart RR

Stage	2017	2018	2019	2020	2021
Number of Reporting Facilities	3	5	5	6	9
Reported Annual CO_2 Sequestered (kt)	5,958	7,662	8,332	6,802	6,947

Reported Annual CO_2 Emissions from Equipment Leaks (kt)	10	11	16	13	37

3.7 Natural Gas Systems (CRF Source Category 1B2b)

Note that this draft of the Inventory does not yet incorporate updated activity data products for the following data inputs, due to a data base subscription lapse: gas well counts, wells drilled, wells completed, and production. Year 2020 values for activity data are used in place of year 2021. The Final Inventory (to be published April 2023) will incorporate the latest activity data.

The U.S. natural gas system encompasses hundreds of thousands of wells, hundreds of processing facilities, and over a million miles of transmission and distribution pipelines. This IPCC category (1B2b) is for fugitive emissions from natural gas systems, which per IPCC guidelines include emissions from leaks, venting, and flaring. Total greenhouse gas emissions (CH_4, CO_2, and N_2O) from natural gas systems in 2021 were 218.3 MMT CO_2 Eq., a decrease of 12 percent from 1990 and a decrease of 2 percent from 2020, both primarily due to decreases in CH_4 emissions. From 2010, emissions decreased by 3 percent, primarily due to decreases in CH_4 emissions. National total dry gas production in the United States increased by 94 percent from 1990 to 2021, increased by 3 percent from 2020 to 2021, and increased by 62 percent from 2010 to 2021. Of the overall greenhouse gas emissions (218.3 MMT CO_2 Eq.), 83 percent are CH_4 emissions (181.4 MMT CO_2 Eq.), 17 percent are CO_2 emissions (36.8 MMT), and less than 0.01 percent are N_2O emissions (0.01 MMT CO_2 Eq.).

Overall, natural gas systems emitted 181.4 MMT CO_2 Eq. (6,479 kt CH_4) of CH_4 in 2021, a 16 percent decrease compared to 1990 emissions, and 2 percent decrease compared to 2020 emissions (see Table 3-66 and Table 3-67). For non-combustion CO_2, a total of 36.8 MMT CO_2 Eq. (36,846 kt) was emitted in 2021, a 14 percent increase compared to 1990 emissions, and a 2 percent increase compared to 2020 levels. The 2021 N_2O emissions were estimated to be 0.01 MMT CO_2 Eq. (0.03 kt N_2O), a 75 percent increase compared to 1990 emissions, and an 8 percent decrease compared to 2020 levels.

The 1990 to 2021 emissions trend is not consistent across segments or gases. Overall, the 1990 to 2021 decrease in CH_4 emissions is due primarily to the decrease in emissions from the following segments: distribution (70 percent decrease), transmission and storage (30 percent decrease), processing (40 percent decrease), and exploration (94 percent decrease). Over the same time period, the production segment saw increased CH_4 emissions of 45 percent (with onshore production emissions increasing 27 percent, offshore production emissions decreasing 86 percent, and gathering and boosting [G&B] emissions increasing 110 percent), and post-meter emissions increasing by 60 percent. The 1990 to 2021 increase in CO_2 emissions is primarily due to an increase in CO_2 emissions in the production segment, where emissions from flaring have increased over time.

Methane and CO_2 emissions from natural gas systems include those resulting from normal operations, routine maintenance, and system upsets. Emissions from normal operations include natural gas engine and turbine uncombusted exhaust, flaring, and leak emissions from system components. Routine maintenance emissions originate from pipelines, equipment, and wells during repair and maintenance activities. Pressure surge relief systems and accidents can lead to system upset emissions. Emissions of N_2O from flaring activities are included in the Inventory, with most of the emissions occurring in the processing and production segments. Note, CO_2 emissions exclude all combustion emissions (e.g., engine combustion) except for flaring CO_2 emissions. All combustion CO_2 emissions (except for flaring) are accounted for in Section 3.1 CO_2 from Fossil Fuel Combustion.

Each year, some estimates in the Inventory are recalculated with improved methods and/or data. These improvements are implemented consistently across the previous Inventory's time series (i.e., 1990 to 2020) to ensure that the trend is representative of changes in emissions. Recalculations in natural gas systems in this year's Inventory include:

- Methodological updates to five onshore production segment sources - pneumatic controllers, equipment leaks, chemical injection pumps, storage tanks, and liquids unloading
- Recalculations due to Greenhouse Gas Reporting Program (GHGRP) submission revisions
- Recalculations due to updating the global warming potential (GWP) for CH_4 and N_2O to use AR5 values.

Updates to well counts and produced water volumes were not available for Public Review estimates, and 2021 data were set equal to 2020.

The Recalculations Discussion section below provides more details on the updated methods.

Below is a characterization of the six emission subcategories of natural gas systems: exploration, production (including gathering and boosting), processing, transmission and storage, distribution, and post-meter. Each of the segments is described and the different factors affecting CH_4, CO_2, and N_2O emissions are discussed.

Exploration. Exploration includes well drilling, testing, and completion. Emissions from exploration accounted for less than 0.2 percent of CH_4 emissions and of CO_2 emissions from natural gas systems in 2021. Well completions accounted for approximately 88 percent of CH_4 emissions from the exploration segment in 2021, with the rest resulting from well testing and drilling. Well completion flaring emissions account for most of the CO_2 emissions. Methane emissions from exploration decreased by 94 percent from 1990 to 2021, with the largest decreases coming from hydraulically fractured gas well completions without reduced emissions completions (RECs). Methane emissions decreased 17 percent from 2020 to 2021 due to decreases in emissions from non-hydraulically fractured well completions with venting. Methane emissions were highest from 2005 to 2008. Carbon dioxide emissions from exploration decreased by 94 percent from 1990 to 2021 primarily due to decreases in hydraulically fractured gas well completions. Carbon dioxide emissions from exploration decreased by 83 percent from 2020 to 2021 due to decreases in emissions from hydraulically fractured gas well completions with flaring. Carbon dioxide emissions were highest from 2006 to 2008. Nitrous oxide emissions decreased 98 percent from 1990 to 2021 and decreased 86 percent from 2020 to 2021.

Production (including gathering and boosting). In the production segment, wells are used to withdraw raw gas from underground formations. Emissions arise from the wells themselves, and from well-site equipment and activities such as pneumatic controllers, tanks and separators, and liquids unloading. Gathering and boosting emission sources are included within the production sector. The gathering and boosting sources include gathering and boosting stations (with multiple emission sources on site) and gathering pipelines. The gathering and boosting stations receive natural gas from production sites and transfer it, via gathering pipelines, to transmission pipelines or processing facilities (custody transfer points are typically used to segregate sources between each segment). Boosting processes include compression, dehydration, and transport of gas to a processing facility or pipeline. Emissions from production (including gathering and boosting) accounted for 52 percent of CH_4 emissions and 25 percent of CO_2 emissions from natural gas systems in 2021. Emissions from gathering and boosting and pneumatic controllers in onshore production accounted for most of the production segment CH_4 emissions in 2021. Within gathering and boosting, the largest sources of CH_4 are compressor exhaust slip, compressor venting and leaks, and tanks. Flaring emissions account for most of the CO_2 emissions from production, with the highest emissions coming from flare stacks at gathering stations, miscellaneous onshore production flaring, and tank flaring. Methane emissions from production increased by 45 percent from 1990 to 2021, due primarily to increases in emissions from pneumatic controllers (due to an increase in the number of controllers, particularly in the number of intermittent bleed controllers) and increases in emissions from compressor exhaust slip in gathering and boosting. Methane emissions decreased 3 percent from 2020 to 2021 due to decreases in emissions from pneumatic controllers and liquids unloading. Carbon dioxide emissions from production increased by approximately a factor of 2.7 from 1990 to 2021 due to increases in emissions at flare stacks in gathering and boosting and miscellaneous onshore production flaring and increased 3 percent from 2020 to 2021 due primarily to increases in emissions from tanks and acid gas removal units at gathering and boosting stations. Nitrous oxide emissions decreased by 28 percent from 1990 to 2021 and decreased 16 percent from 2020 to 2021. The decrease in N_2O emissions from 1990 to 2021 and from 2020 to 2021 is primarily due to decreases in emissions from flaring at gathering and boosting stations.

Processing. In the processing segment, natural gas liquids and various other constituents from the raw gas are removed, resulting in "pipeline quality" gas, which is injected into the transmission system. Methane emissions from compressors, including compressor seals, are the primary emission source from this stage. Most of the CO_2 emissions come from acid gas removal (AGR) units, which are designed to remove CO_2 from natural gas. Processing plants accounted for 8 percent of CH_4 emissions and 71 percent of CO_2 emissions from natural gas systems. Methane emissions from processing decreased by 40 percent from 1990 to 2021 as emissions from compressors (leaks and venting) and equipment leaks decreased; and increased 3 percent from 2020 to 2021 due to increased emissions from gas engines. Carbon dioxide emissions from processing decreased by 8 percent from 1990 to 2021, due to a decrease in AGR emissions, and increased 3 percent from 2020 to 2021 due to increased emissions from AGR. Nitrous oxide emissions decreased 1 percent from 2020 to 2021.

Transmission and Storage. Natural gas transmission involves high pressure, large diameter pipelines that transport gas long distances from field production and processing areas to distribution systems or large volume customers such as power plants or chemical plants. Compressor station facilities are used to move the gas throughout the U.S. transmission system. Leak CH_4 emissions from these compressor stations and venting from pneumatic controllers account for most of the emissions from this stage. Uncombusted compressor engine exhaust and pipeline venting are also sources of CH_4 emissions from transmission. Natural gas is also injected and stored in underground formations, or liquefied and stored in above ground tanks, during periods of low demand (e.g., summer), and withdrawn, processed, and distributed during periods of high demand (e.g., winter). Leak and venting emissions from compressors are the primary contributors to CH_4 emissions from storage. Emissions from liquefied natural gas (LNG) stations and terminals are also calculated under the transmission and storage segment. Methane emissions from the transmission and storage segment accounted for approximately 25 percent of emissions from natural gas systems, while CO_2 emissions from transmission and storage accounted for 4 percent of the CO_2 emissions from natural gas systems. CH_4 emissions from this source decreased by 30 percent from 1990 to 2021 due to reduced pneumatic device and compressor station emissions (including emissions from compressors and leaks) and decreased 2 percent from 2020 to 2021 due to decreased emissions from pipeline venting transmission compressors. CO_2 emissions from transmission and storage were 4.7 times higher in 2021 than in 1990, due to increased emissions from LNG export terminals, and decreased by 16 percent from 2020 to 2021, also due to LNG export terminals and flaring (both transmission and storage). The quantity of LNG exported from the United States increased by a factor of 68 from 1990 to 2021, and by 49 percent from 2020 to 2021. LNG emissions are about 1 percent of CH_4 and 89 percent of CO_2 emissions from transmission and storage in year 2021. Nitrous oxide emissions from transmission and storage increased by 165 percent from 1990 to 2021 and decreased 12 percent from 2020 to 2021.

Distribution. Distribution pipelines take the high-pressure gas from the transmission system at "city gate" stations, reduce the pressure and distribute the gas through primarily underground mains and service lines to individual end users. There were 1,337,012 miles of distribution mains in 2021, an increase of 392,855 miles since 1990 (PHMSA 2021). Distribution system emissions, which accounted for 8 percent of CH_4 emissions from natural gas systems and less than 1 percent of CO_2 emissions, result mainly from leak emissions from pipelines and stations. An increased use of plastic piping, which has lower emissions than other pipe materials, has reduced both CH_4 and CO_2 emissions from this stage, as have station upgrades at metering and regulating (M&R) stations. Distribution system CH_4 emissions in 2021 were 70 percent lower than 1990 levels and 1 percent lower than 2020 emissions. Distribution system CO_2 emissions in 2021 were 70 percent lower than 1990 levels and 1 percent lower than 2020 emissions. Annual CO_2 emissions from this segment are less than 0.1 MMT CO_2 Eq. across the time series.

Post-Meter. Post-meter includes leak emissions from residential and commercial appliances, industrial facilities and power plants, and natural gas fueled vehicles. Leak emissions from residential appliances and industrial facilities and power plants account for the majority of post-meter CH_4 emissions. Methane emissions from the post-meter segment accounted for approximately 7 percent of emissions from natural gas systems in 2021. Post-meter CH_4 emissions increased by 60 percent from 1990 to 2021 and increased by less than 1 percent from 2020 to 2021, due to increases in the number of residential houses using natural gas and increased natural gas consumption at industrial facilities and power plants. CO_2 emissions from post-meter account for less than 0.01 percent of total CO_2 emissions from natural gas systems.

Total greenhouse gas emissions from the six subcategories within natural gas systems are shown in MMT CO_2 Eq. in Table 3-65. Total CH_4 emissions for these same segments of natural gas systems are shown in MMT CO_2 Eq. (Table 3-66) and kt (Table 3-67). Most emission estimates are calculated using a net emission approach. However, a few sources are still calculated with a potential emission approach. Reductions data are applied to those sources. In 2021, 2.6 MMT CO_2 Eq. CH_4 is subtracted from production segment emissions, 4.3 MMT CO_2 Eq. CH_4 is subtracted from the transmission and storage segment, and 0.1 MMT CO_2 Eq. CH_4 is subtracted from the distribution segment to calculate net emissions. More disaggregated information on potential emissions, net emissions, and reductions data is available in Annex 3.6, Methodology for Estimating CH_4 and CO_2 Emissions from Natural Gas Systems.

Table 3-65: Total Greenhouse Gas Emissions (CH_4, CO_2, and N_2O) from Natural Gas Systems (MMT CO_2 Eq.)

Segment	1990		2005		2017	2018	2019	2020	2021
Exploration	3.6		11.5		1.8	3.0	2.3	0.3	0.2
Production	68.1		102.4		111.5	116.2	115.5	106.2	103.2
Processing	52.2		31.8		35.8	36.3	40.4	39.3	40.4
Transmission and Storage	64.4		44.7		41.4	43.9	45.7	47.4	46.2
Distribution	51.0		28.5		15.7	15.6	15.5	15.5	15.3
Post-Meter	8.1		9.6		11.9	12.5	12.8	13.0	13.0
Total	**247.5**		**228.6**		**218.2**	**227.4**	**232.3**	**221.7**	**218.3**

Note: Totals may not sum due to independent rounding.

Table 3-66: CH_4 Emissions from Natural Gas Systems (MMT CO_2 Eq.)

Segment	1990		2005		2017	2018	2019	2020	2021
Exploration	3.3		10.0		1.4	2.6	2.1	0.2	0.2
Production	64.7		97.9		103.5	107.0	104.7	97.3	94.0
Onshore Production	39.3		69.0		59.9	62.9	59.4	53.8	50.0
Gathering and Boosting	20.7		26.8		42.9	43.3	44.6	42.6	43.4
Offshore Production	4.8		2.0		0.7	0.8	0.7	0.9	0.7
Processing	23.9		13.0		12.9	13.5	14.2	13.9	14.3
Transmission and Storage	64.1		44.3		41.0	43.2	44.3	45.5	44.6
Distribution	50.9		28.5		15.7	15.6	15.5	15.5	15.3
Post-Meter	8.1		9.6		11.9	12.5	12.8	13.0	13.0
Total	**215.1**		**203.4**		**186.4**	**194.4**	**193.6**	**185.4**	**181.4**

Note: Totals may not sum due to independent rounding.

Table 3-67: CH_4 Emissions from Natural Gas Systems (kt)

Segment	1990		2005		2017	2018	2019	2020	2021
Exploration	119		358		49	94	75	9	7
Production	2,311		3,495		3,697	3,823	3,739	3,475	3,359
Onshore Production	1,403		2,464		2,139	2,246	2,122	1,923	1,786
Gathering and Boosting	739		958		1,533	1,547	1,591	1,520	1,548
Offshore Production	170		73		26	30	25	32	24
Processing	853		463		460	483	506	495	510
Transmission and Storage	2,289		1,584		1,465	1,542	1,584	1,625	1,592

Distribution	1,819	1,018	561	557	554	553	548
Post-Meter	290	344	424	445	457	463	463
Total	**7,682**	**7,263**	**6,657**	**6,943**	**6,915**	**6,620**	**6,479**

Note: Totals may not sum due to independent rounding.

Table 3-68: CO_2 Emissions from Natural Gas Systems (MMT)

Segment	1990	2005	2017	2018	2019	2020	2021
Exploration	0.3	1.4	0.4	0.3	0.2	0.1	+
Production	3.3	4.6	8.0	9.1	10.9	8.9	9.1
Processing	28.3	18.8	22.9	22.8	26.2	25.4	26.1
Transmission and Storage	0.3	0.3	0.4	0.7	1.4	1.9	1.6
Distribution	0.1	+	+	+	+	+	+
Post-Meter	+	+	+	+	+	+	+
Total	**32.4**	**25.2**	**31.8**	**33.0**	**38.7**	**36.3**	**36.8**

+ Does not exceed 0.05 MMT CO_2 Eq.
Note: Totals may not sum due to independent rounding.

Table 3-69: CO_2 Emissions from Natural Gas Systems (kt)

Segment	1990	2005	2017	2018	2019	2020	2021
Exploration	297	1,434	444	336	220	96	17
Production	3,337	4,556	7,967	9,147	10,857	8,878	9,141
Processing	28,338	18,836	22,935	22,766	26,225	25,419	26,096
Transmission and Storage	336	349	405	707	1,384	1,884	1,574
Distribution	54	30	17	17	16	16	16
Post-Meter	1	1	2	2	2	2	2
Total	**32,363**	**25,206**	**31,770**	**32,974**	**38,705**	**36,296**	**36,846**

NE (Not Estimated)
Note: Totals may not sum due to independent rounding.

Table 3-70: N_2O Emissions from Natural Gas Systems (Metric Tons CO_2 Eq.)

Segment	1990	2005	2017	2018	2019	2020	2021
Exploration	355	1,090	217	156	103	45	6
Production	3,840	5,153	3,730	4,061	4,774	3,310	2,779
Processing	NO	2,977	2,643	2,998	5,081	4,349	4,300
Transmission and Storage	298	351	364	290	636	903	791
Distribution	NO	NO	NO	NO	NO	NO	NO
Post-Meter	NO	NO	NO	NO	NO	NO	NO
Total	**4,494**	**9,572**	**6,953**	**7,506**	**10,594**	**8,608**	**7,877**

NO (Not Occurring)
Note: Totals may not sum due to independent rounding.

Table 3-71: N_2O Emissions from Natural Gas Systems (Metric Tons N_2O)

Segment	1990	2005	2017	2018	2019	2020	2021
Exploration	1.3	4.1	0.8	0.6	0.4	0.2	0.0
Production	14.5	19.4	14.1	15.3	18.0	12.5	10.5
Processing	NO	11.2	10.0	11.3	19.2	16.4	16.2
Transmission and Storage	1.1	1.3	1.4	1.1	2.4	3.4	3.0
Distribution	NO	NO	NO	NO	NO	NO	NO
Post-Meter	NO	NO	NO	NO	NO	NO	NO
Total	**17.0**	**36.1**	**26.2**	**28.3**	**40.0**	**32.5**	**29.7**

NO (Not Occurring)
Note: Totals may not sum due to independent rounding.

Methodology and Time-Series Consistency

See Annex 3.6 for the full time series of emissions data, activity data, and emission factors, and additional information on methods and data sources—for example, the specific years of reporting data from EPA's GHGRP that are used to develop certain factors.

This section provides a general overview of the methodology for natural gas system emission estimates in the Inventory, which involves the calculation of CH_4, CO_2, and N_2O emissions for over 100 emissions sources (i.e., equipment types or processes), and then the summation of emissions for each natural gas segment.

The approach for calculating emissions for natural gas systems generally involves the application of emission factors to activity data. For most sources, the approach uses technology-specific emission factors or emission factors that vary over time and take into account changes to technologies and practices, which are used to calculate net emissions directly. For others, the approach uses what are considered "potential methane factors" and emission reduction data to calculate net emissions. The estimates are developed with an IPCC Tier 2 approach. Tier 1 approaches are not used.

Emission Factors. Key references for emission factors for CH_4 and CO_2 emissions from the U.S. natural gas industry include a 1996 study published by the Gas Research Institute (GRI) and EPA (GRI/EPA 1996), EPA's GHGRP (EPA 2022), and others.

The 1996 GRI/EPA study developed over 80 CH_4 emission factors to characterize emissions from the various components within the operating segments of the U.S. natural gas system. The GRI/EPA study was based on a combination of process engineering studies, collection of activity data, and measurements at representative natural gas facilities conducted in the early 1990s. Year-specific natural gas CH_4 compositions are calculated using U.S. Department of Energy's Energy Information Administration (EIA) annual gross production data for National Energy Modeling System (NEMS) oil and gas supply module regions in conjunction with data from the Gas Technology Institute (GTI, formerly GRI) Unconventional Natural Gas and Gas Composition Databases (GTI 2001). These year-specific CH_4 compositions are applied to emission factors, which therefore may vary from year to year due to slight changes in the CH_4 composition of natural gas for each NEMS region.

GHGRP Subpart W data were used to develop CH_4, CO_2, and N_2O emission factors for many sources in the Inventory. In the exploration and production segments, GHGRP data were used to develop emission factors used for all years of the time series for well testing, gas well completions and workovers with and without hydraulic fracturing, pneumatic controllers and chemical injection pumps, condensate tanks, liquids unloading, miscellaneous flaring, gathering and boosting pipelines, and certain sources at gathering and boosting stations. In the processing segment, for recent years of the times series, GHGRP data were used to develop emission factors for leaks, compressors, flares, dehydrators, and blowdowns/venting. In the transmission and storage segment, GHGRP data were used to develop factors for all years of the time series for LNG stations and terminals and transmission pipeline blowdowns, and for pneumatic controllers for recent years of the times series.

Other data sources used for CH_4 emission factors include Zimmerle et al. (2015) for transmission and storage station leaks and compressors, GTI (2009 and 2019) for commercial and industrial meters, Lamb et al. (2015) for recent years for distribution pipelines and meter/regulator stations, Zimmerle et al. (2019) for gathering and boosting stations, Bureau of Ocean Energy Management (BOEM) reports, and Fischer et al. (2019) and IPCC (2019) for post-meter emissions.

For CO_2 emissions from sources in the exploration, production and processing segments that use emission factors not directly calculated from GHGRP data, data from the 1996 GRI/EPA study and a 2001 GTI publication were used to adapt the CH_4 emission factors into related CO_2 emission factors. For sources in the transmission and storage segment that use emission factors not directly calculated from GHGRP data, and for sources in the distribution segment, data from the 1996 GRI/EPA study and a 1993 GTI publication were used to adapt the CH_4 emission factors into non-combustion related CO_2 emission factors. CO_2 emissions from post-meter sources (commercial, industrial and vehicles) were estimated using default emission factors from IPCC (2019). Carbon dioxide emissions from post-meter residential sources are included in fossil fuel combustion data.

Flaring N_2O emissions were estimated for flaring sources using GHGRP data.

See Annex 3.6 for more detailed information on the methodology and data used to calculate CH_4, CO_2, and N_2O emissions from natural gas systems.

Activity Data. Activity data were taken from various published data sets, as detailed in Annex 3.6. Key activity data sources include data sets developed and maintained by EPA's GHGRP (EPA 2022); Enverus (Enverus 2021); BOEM; Federal Energy Regulatory Commission (FERC); EIA; the Natural Gas STAR and Methane Challenge Programs annual data; Oil and Gas Journal; and PHMSA. Enverus data for 2021 are not currently available; this public review version of the Inventory uses 2020 data as proxy for 2021.

For a few sources, recent direct activity data are not available. For these sources, either 2020 data were used as a proxy for 2021 data, or a set of industry activity data drivers was developed and used to calculate activity data over the time series. Drivers include statistics on gas production, number of wells, system throughput, miles of various kinds of pipe, and other statistics that characterize the changes in the U.S. natural gas system infrastructure and operations. More information on activity data and drivers is available in Annex 3.6.

A complete list of references for emission factors and activity data by emission source is provided in Annex 3.6.

Calculating Net Emissions. For most sources, net emissions are calculated directly by applying emission factors to activity data. Emission factors used in net emission approaches reflect technology-specific information, and take into account regulatory and voluntary reductions. However, for production, transmission and storage, and distribution, some sources are calculated using potential emission factors, and CH_4 that is not emitted is deducted from the total CH_4 potential estimates. To take into account use of such technologies and practices that result in lower emissions but are not reflected in "potential" emission factors, data are collected on both regulatory and voluntary reductions. Regulatory actions addressed using this method include EPA National Emission Standards for Hazardous Air Pollutants (NESHAP) regulations for dehydrator vents. Voluntary reductions included in the Inventory are those reported to Natural Gas STAR and Methane Challenge for certain sources. Natural Gas STAR and Methane Challenge reductions were reassessed for this Inventory, see the Recalculations Discussion for more information.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2020. GHGRP data available (starting in 2011) and other recent data sources have improved estimates of emissions from natural gas systems. To develop a consistent time series, for sources with new data, EPA reviewed available information on factors that may have resulted in changes over the time series (e.g., regulations, voluntary actions) and requested stakeholder feedback on trends as well. For most sources, EPA developed annual data for 1993 through 2010 by interpolating activity data or emission factors or both between 1992 and 2011 data points. Information on time-series consistency for sources updated in this year's Inventory can be found in the Recalculations Discussion below, with additional detail provided in supporting memos (relevant memos are cited in the Recalculations Discussion). For detailed documentation of methodologies, please see Annex 3.5.

Through EPA's stakeholder process on oil and gas in the Inventory, EPA received stakeholder feedback on updates under consideration for the Inventory. Stakeholder feedback is noted below in Recalculations Discussion and Planned Improvements.

The United States reports data to the UNFCCC using this Inventory report along with Common Reporting Format (CRF) tables. This note is provided for those reviewing the CRF tables: The notation key "IE" is used for CO_2 and CH_4 emissions from venting and flaring in CRF table 1.B.2. Disaggregating flaring and venting estimates across the Inventory would involve the application of assumptions and could result in inconsistent reporting and, potentially, decreased transparency. Data availability varies across segments within oil and gas activities systems, and emission factor data available for activities that include flaring can include emissions from multiple sources (flaring, venting and leaks).

Uncertainty – TO BE UPDATED FOR FINAL INVENTORY REPORT

EPA has conducted a quantitative uncertainty analysis using the IPCC Approach 2 methodology (Monte Carlo

Simulation technique) to characterize the uncertainty for natural gas systems. For more information on the approach, please see the memoranda *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2016: Natural Gas and Petroleum Systems Uncertainty Estimates* and *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2019: Update for Natural Gas and Petroleum Systems CO_2 Uncertainty Estimates.*[82]

EPA used Microsoft Excel's @RISK add-in tool to estimate the 95 percent confidence bound around CH_4 and CO_2 emissions from natural gas systems for the current Inventory. For the CH_4 uncertainty analysis, EPA focused on the 16 highest-emitting sources for the year 2020, which together emitted 76 percent of methane from natural gas systems in 2020, and extrapolated the estimated uncertainty for the remaining sources. For the CO_2 uncertainty analysis, EPA focused on the 3 highest-emitting sources for the year 2020, which together emitted 80 percent of CO_2 from natural gas systems in 2020, and extrapolated the estimated uncertainty for the remaining sources. To estimate uncertainty for N_2O, EPA applied the uncertainty bounds calculated for CO_2. EPA will seek to refine this estimate in future Inventories. The @RISK add-in provides for the specification of probability density functions (PDFs) for key variables within a computational structure that mirrors the calculation of the inventory estimate. The IPCC guidance notes that in using this method, "some uncertainties that are not addressed by statistical means may exist, including those arising from omissions or double counting, or other conceptual errors, or from incomplete understanding of the processes that may lead to inaccuracies in estimates developed from models." The uncertainty bounds reported below only reflect those uncertainties that EPA has been able to quantify and do not incorporate considerations such as modeling uncertainty, data representativeness, measurement errors, misreporting or misclassification. The understanding of the uncertainty of emission estimates for this category evolves and improves as the underlying methodologies and datasets improve.

The results presented below provide the 95 percent confidence bound within which actual emissions from this source category are likely to fall for the year 2020, using the IPCC methodology. The results of the Approach 2 uncertainty analysis are summarized in Table 3-72. Natural gas systems CH_4 emissions in 2020 were estimated to be between 135.2 and 194.6 MMT CO_2 Eq. at a 95 percent confidence level. Natural gas systems CO_2 emissions in 2020 were estimated to be between 29.7 and 42.2 MMT CO_2 Eq. at a 95 percent confidence level. Natural gas systems N_2O emissions in 2020 were estimated to be between 0.009 and 0.012 MMT CO_2 Eq. at a 95 percent confidence level.

Uncertainty bounds for other years of the time series have not been calculated, but uncertainty is expected to vary over the time series. For example, years where many emission sources are calculated with interpolated data would likely have higher uncertainty than years with predominantly year-specific data. In addition, the emission sources that contribute the most to CH_4 and CO_2 emissions are different over the time series, particularly when comparing recent years to early years in the time series. For example, venting emissions were higher and flaring emissions were lower in early years of the time series, compared to recent years. Technologies also changed over the time series (e.g., liquids unloading with plunger lifts and reduced emissions completions were not used early in the time series and cast iron distribution mains were more prevalent than plastic mains in early years). Transmission and gas processing compressor leak and vent emissions were also higher in the early years of the time series.

Table 3-72: Approach 2 Quantitative Uncertainty Estimates for CH_4 and Non-combustion CO_2 Emissions from Natural Gas Systems (MMT CO_2 Eq. and Percent)

Source	Gas	2020 Emission Estimate (MMT CO_2 Eq.)[b]	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(%)	
			Lower Bound[b]	Upper Bound[b]	Lower Bound[b]	Upper Bound[b]
Natural Gas Systems	CH_4	164.9	135.2	194.6	-18%	+18%
Natural Gas Systems	CO_2	35.4	29.7	42.3	-16%	+19%
Natural Gas Systems	N_2O	+	+	+	-16%	+19%

+ Less than 0.05 MMT CO_2 Eq.

[82] See https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems.

[a] Range of emission estimates estimated by applying the 95 percent confidence intervals obtained from the Monte Carlo Simulation analysis conducted for the year 2020 CH_4 and CO_2 emissions.
[b] All reported values are rounded after calculation. As a result, lower and upper bounds may not be duplicable from other rounded values as shown in Table 3-66 and Table 3-67.

QA/QC and Verification Discussion

The natural gas systems emission estimates in the Inventory are continually being reviewed and assessed to determine whether emission factors and activity factors accurately reflect current industry practices. A QA/QC analysis was performed for data gathering and input, documentation, and calculation. QA/QC checks are consistently conducted to minimize human error in the model calculations. EPA performs a thorough review of information associated with new studies, GHGRP data, regulations, public webcasts, and the Natural Gas STAR Program to assess whether the assumptions in the Inventory are consistent with current industry practices. The EPA has a multi-step data verification process for GHGRP data, including automatic checks during data-entry, statistical analyses on completed reports, and staff review of the reported data. Based on the results of the verification process, the EPA follows up with facilities to resolve mistakes that may have occurred.[83]

As in previous years, EPA conducted early engagement and communication with stakeholders on updates prior to public review of the current Inventory. EPA held stakeholder webinars in September and November of 2022. EPA released memos detailing updates under consideration and requesting stakeholder feedback.

In recent years, several studies have measured emissions at the source level and at the national or regional level and calculated emission estimates that may differ from the Inventory. There are a variety of potential uses of data from new studies, including replacing a previous estimate or factor, verifying or QA of an existing estimate or factor, and identifying areas for updates. In general, there are two major types of studies related to oil and gas greenhouse gas data: studies that focus on measurement or quantification of emissions from specific activities, processes and equipment, and studies that use tools such as inverse modeling to estimate the level of overall emissions needed to account for measured atmospheric concentrations of greenhouse gases at various scales. The first type of study can lead to direct improvements to or verification of Inventory estimates. In the past few years, EPA has reviewed and in many cases, incorporated data from these data sources. The second type of study can provide general indications of potential over- and under-estimates. In addition, in recent years information from top-down studies has been directly incorporated to quantify emissions from well blowouts.

A key challenge in using these types of studies to assess Inventory results is having a relevant basis for comparison (e.g., the two data sets should have comparable time frames and geographic coverage, and the independent study should assess data from the Inventory and not another data set, such as the Emissions Database for Global Atmospheric Research, or "EDGAR"). In an effort to improve the ability to compare the national-level Inventory with measurement results that may be at other spatial or temporal scales, a team at Harvard University along with EPA and other coauthors developed a gridded inventory of U.S. anthropogenic methane emissions with 0.1 degree x 0.1 degree spatial resolution, monthly temporal resolution, and detailed scale-dependent error characterization.[84] The gridded methane inventory is designed to be consistent with the U.S. EPA's *Inventory of U.S. Greenhouse Gas Emissions and Sinks: 1990-2014* estimates for the year 2012, which presents national totals.[85] An updated version of the gridded inventory is being developed and will improve efforts to compare results of the Inventory with atmospheric studies.

[83] See https://www.epa.gov/sites/production/files/2015-07/documents/ghgrp_verification_factsheet.pdf.
[84] See https://www.epa.gov/ghgemissions/gridded-2012-methane-emissions.
[85] See https://www.epa.gov/ghgemissions/us-greenhouse-gas-inventory-report-1990-2014.

Recalculations Discussion

EPA received information and data related to the emission estimates through GHGRP reporting and stakeholder feedback on updates under consideration. In October 2022, EPA released a draft memorandum that discussed changes under consideration and requested stakeholder feedback on those changes.[86] EPA did not receive written feedback on the memorandum. Memoranda cited in the Recalculations Discussion below are: Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2021: Updates Under Consideration for Incorporating Additional Geographically Disaggregated Data (*Disaggregation* memo) and Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2021: Updates Under Consideration for Incorporating Additional Geographically Disaggregated Data for the Production Segment (*Production Disaggregation* memo).

In this Inventory, an update that incorporates additional basin-level data from GHGRP subpart W was implemented for several emission sources in the onshore production segment. The update seeks to improve the ability of EPA's gridded and state inventories to reflect variation due to differences in formation types, technologies and practices, regulations, or voluntary initiatives, and not only the differences in key activity levels that are reflected in the current gridded and state inventories. This would allow EPA to use the gridded inventory for improved comparisons of the national Inventory with various atmospheric observation studies (since regions will better reflect the local differences in emissions rates as reported to GHGRP) and would allow the state-level inventory to reflect differences in state-level programs, formation type mixes, and varying technologies and practices. For many sources, an approach that develops estimates using geographically disaggregated data may not be possible or preferable to a national level approach based on the currently available data. For some emission sources in the Inventory, emission factor data come from research studies and are applied at the national level. For example, many of the emission factors used to quantify emissions in the Inventory for the gathering and boosting, transmission and storage, distribution, and post-meter segments are from research studies and do not have a level of detail or total population comparable to GHGRP. Even in cases where geographically disaggregated data are available, such an approach may not always be preferable. In cases with limited variation between areas, such an approach would have limited impact on emissions estimates regionally or nationally. In cases with limited data in certain areas, disaggregated approaches might substantially increase the uncertainty of estimates and basin-specific calculations would not be an improvement over use of a national average. EPA continues to seek stakeholder feedback on the draft approach in this Inventory.

EPA evaluated relevant information available and made several updates to the Inventory, including for pneumatic controllers, equipment leaks, chemical injection pumps, storage tanks, and liquids unloading. For each of these emission sources, EPA modified the calculation methodology to use GHGRP data to develop basin-specific activity factors and/or emission factors. General information for these source specific recalculations are presented below and details are available in the *Disaggregation* memo and *Production Disaggregation* memo, including additional considerations for the updates.

In addition to the production segment sources mentioned above, for certain sources, CH_4 and/or CO_2 emissions changed by greater than 0.05 MMT CO_2 Eq., comparing the previous estimate for 2020 to the current (recalculated) estimate for 2020. The emissions changes were mostly due to GHGRP data submission revisions. These sources are discussed below and include miscellaneous production flaring, offshore production, distribution pipelines, and post-meter emissions.

In addition, for the current Inventory, CO_2-equivalent emissions totals have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth Assessment Report* (AR4) (IPCC 2007) used in the previous inventories. The AR5 GWPs have been applied across the entire time series for consistency. The GWP of CH_4 has increased from 25 to 28, leading to an increase in the calculated CO_2-equivalent emissions of CH_4, while the GWP of N_2O has decreased from 298 to 265, leading to a decrease in the calculated CO_2-equivalent emissions

[86] Stakeholder materials including draft memoranda for the current (i.e., 1990 to 2021) Inventory are available at https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems.

of N_2O. Further discussion on this update and the overall impacts of updating the Inventory GWP values to reflect the IPCC *Fifth Assessment Report* can be found in Chapter 9, Recalculations and Improvements.

The combined impact of revisions to 2020 natural gas systems CH_4 emissions, compared to the previous Inventory, is an increase from 164.9 to 185.4 MMT CO_2 Eq. (20.5 MMT CO_2 Eq., or 12 percent). The recalculations resulted in an average increase in the annual CH_4 emission estimates across the 1990 through 2020 time series, compared to the previous Inventory, of 24.1 MMT CO_2 Eq., or 14 percent.

The combined impact of revisions to 2020 natural gas systems CO_2 emissions, compared to the previous Inventory, is an increase from 35.4 MMT to 36.3 MMT, or 2.7 percent. The recalculations resulted in an average increase in emission estimates across the 1990 through 2020 time series, compared to the previous Inventory, of 0.4 MMT CO_2 Eq., or 1.3 percent.

The combined impact of revisions to 2020 natural gas systems N_2O emissions, compared to the previous Inventory, is a decrease from 10.2 kt CO_2 Eq. to 8.6 kt CO_2 Eq., or 15 percent. The recalculations resulted in an average decrease in emission estimates across the 1990 through 2020 time series, compared to the previous Inventory, of 11 percent.

In Table 3-73 and Table 3-74 below are categories in Natural Gas Systems with recalculations resulting in a change of greater than 0.05 MMT CO_2 Eq., comparing the previous estimate for 2019 to the current (recalculated) estimate for 2019. No changes made to N_2O estimates resulted in a change greater than 0.05 MMT CO_2 Eq. For more information, please see the Recalculations Discussion below.

For certain sources, the change in GWP for CH_4 alone (i.e., not the results of other recalculations) resulted in calculated CH_4 CO_2-equivalent emissions for 2020 changing by greater than 0.05 MMT CO_2 Eq., compared to the previous Inventory. These sources are not discussed below. The production segment sources impacted by the GWP update are: wellhead leaks, produced water, dehydrator kimray pumps, gas engine exhaust, G&B compressors, G&B pneumatic controllers, G&B pneumatic pumps, G&B combustion slip, G&B yard piping, and G&B pipeline leaks. The natural gas processing sources impacted by the GWP update are: reciprocating compressors, gas engine exhaust, and blowdowns. The transmission and storage sources impacted by the GWP update are: compressor station leaks, reciprocating compressors, centrifugal compressors, M&R, gas engine exhaust, pneumatic controllers, pipeline venting, and compressor station venting. The distribution sources impacted by the GWP update are distribution main and service leaks, customer meters, and mishaps.

Table 3-73: Recalculations of CO_2 in Natural Gas Systems (MMT CO_2)

Segment and Emission Sources with Changes of Greater than 0.05 MMT CO_2 due to Recalculations	*Previous Estimate Year 2020, 2022 Inventory*	Current Estimate Year 2020, 2023 Inventory	Current Estimate Year 2021, 2023 Inventory
Exploration	***0.1***	**0.1**	**+**
Production	***7.7***	**8.9**	**9.1**
Misc. Onshore Production Flaring	*1.1*	1.3	1.0
Large Tanks with Flares	*0.6*	0.8	0.8
Liquids Unloading	*+*	+	+
G&B Station Sources	*5.8*	6.5	7.1
Processing	***25.5***	**25.4**	**26.1**
Flares	*7.9*	8.1	7.4
Transmission and Storage	***2.0***	**1.9**	**1.6**
Distribution	*+*	+	+
Post-Meter	*+*	+	+
Total	***35.4***	**36.3**	**36.8**

+ Does not exceed 0.05 MMT CO_2.

Table 3-74: Recalculations of CH_4 in Natural Gas Systems (MMT CO_2 Eq.)

Segment and Emission Sources with Changes of Greater than 0.05 MMT CO_2 due to Recalculations	*Previous Estimate Year 2020, 2022 Inventory*	Current Estimate Year 2020, 2023 Inventory	Current Estimate Year 2021, 2023 Inventory
Exploration	***0.2***	**0.2**	**0.2**
Production	***86.4***	**97.3**	**94.0**
Well pad Equipment Leaks	*6.6*	10.3	9.6
Chemical Injection Pumps	*2.8*	2.4	2.1
Pneumatic Controllers	*23.8*	22.8	21.3
Tanks	*0.4*	1.5	1.2
Liquids Unloading	*3.2*	4.5	3.4
G&B Station Sources	*34.1*	38.7	39.8
Processing	***12.4***	**13.9**	**14.3**
Transmission and Storage	***40.6***	**45.5**	**44.6**
Distribution	***13.9***	**15.5**	**15.3**
Pipeline Mains – Unprotected Steel	*1.0*	1.1	1.0
Post-Meter	***11.5***	**13.0**	**13.0**
Total	***164.9***	**185.4**	**181.4**

Exploration

There were no methodological updates to the exploration segment, and recalculations due to updated data resulted in average decreases in calculated CH_4 and CO_2 emissions over the time series of less than 1 percent.

Production

Pneumatic Controllers (Methodological Update)

EPA updated the calculation methodology for pneumatic controllers to use basin-specific activity factors and emission factors calculated from subpart W data for each type of controller (i.e., high, intermittent, and low bleed). Previously, national average activity and emission factors calculated using subpart W data were applied to estimate pneumatic controller emissions. In this methodological update, EPA summed basin-level emissions together to develop national emissions. The *Disaggregation* memo and *Production Disaggregation* memo present additional information and considerations for this update.

EPA calculated basin-specific activity factors and CH_4 emission factors were calculated for all basins that reported subpart W data. The factors were year-specific for RY2011 through RY2021. EPA retained the previous Inventory's activity factor assumptions for 1990 through 1992 and applied linear interpolation between the 1992 and 2011 activity factors at the basin-level. Year 2011 emission factors were applied to all prior years for each basin. For basins without subpart W data available, EPA applied national average activity and emission factors.

The estimation methodology for CO_2 emissions was not updated to use the basin-specific approach for the public review version of the Inventory. CO_2 emissions were estimated by applying a CO_2 to CH_4 ratio to the estimated CH_4 emissions. EPA will calculate pneumatic controller CO_2 emissions in the same manner as CH_4 emissions for the final Inventory.

As a result of this methodological update, CH_4 emissions estimates are on average 4 percent higher across the time-series than in the previous Inventory. The estimate for 2020 is 14 percent lower than in the previous Inventory. Pneumatic controller CH_4 emissions were higher for all years between 1990 through 2011 by an average of 8 percent and CH_4 emissions were lower for 2011 through 2020 by an average of 6 percent, compared to the previous inventory. Emissions were lower in recent years due to some basins having slightly lower activity factors and/or emission factors for intermittent bleed pneumatic controllers, compared to the national average. Emissions were higher in early years of the time series due to basins having higher emission factors than the national average. Multiple basins impact the emissions changes for pneumatic controllers at gas wells.

Table 3-75: Pneumatic Controllers National CH_4 Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Low Bleed Controllers	0		22,745		32,360	33,805	31,475	27,364	25,609
High Bleed Controllers	350,535		483,375		108,533	87,071	53,233	42,332	42,828
Intermittent Bleed Controllers	230,504		569,592		873,015	835,249	874,372	744,622	692,097
Total Emissions	**581,039**		**1,075,712**		**1,013,908**	**956,125**	**959,080**	**814,318**	**760,534**
Previous Estimate	*510,354*		*1,041,503*		*1,104,896*	*1,072,874*	*1,024,678*	*950,718*	*NA*

NA (Not Applicable)

Storage Tanks (Methodological Update)

EPA updated the calculation methodology for production segment storage tanks to use basin-specific activity factors and emission factors calculated from Subpart W data for each storage tank category. Previously, national annual average activity and emission factors calculated using Subpart W data were applied to estimate storage tank emissions. In this methodological update, EPA summed basin-level emissions together to develop national emissions. The calculation methodology was updated to estimate CH_4 and CO_2 emissions using basin-level data from subpart W. The *Production Disaggregation* memo presents additional information and considerations for this update.

EPA calculated basin-specific activity factors and CH_4 and CO_2 emission factors for all basins that reported subpart W data. The factors were year-specific for reporting year (RY) 2015 through RY2021. EPA also retained the previous Inventory's activity factor assumptions for 1990 and used linear interpolation between the 1990 and 2015 activity factors at the basin-level. Year 2015 emission factors were applied to all prior years for each basin. For basins without Subpart W data available, EPA applied national average activity and emission factors.

This update resulted in CH_4 emission estimates an average of 276 percent higher across the time series compared with the previous Inventory. The estimate for 2020 is 210 percent higher than in the previous Inventory. Storage tank CO_2 emissions are an average of 43 percent higher across the time series compared to the previous Inventory. The 2020 emission estimate is 50 percent higher than in the previous Inventory.

The basin-level approach's emissions increased because certain basins with high liquids production and storage tank throughput had higher emission factors and/or activity factors than the national average. The time-series is also impacted as the basin-level approach reflects changing levels of liquids production, and hence storage tank throughput, for basins across the time-series; basins with more production and storage tank throughput in the early 90s also corresponded to basins with higher emission factors and/or activity factors than the national average. For CH_4, this is particularly noticeable for basins with small tanks without flares (e.g., Arkoma Basin, Bend Arch, Central Western Overthrust, East Texas, Piceance) and for CO_2 emissions this is noticeable for basins using large tanks with flares (e.g., Anadarko Basin, Appalachian, Chautauqua Platform, Denver, Gulf Coast, Permian, South Oklahoma Folded Belt).

Table 3-76: Storage Tanks National CH_4 Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Large Tanks w/Flares	505		336		1,016	1,273	789	600	606
Large Tanks w/VRU	0		27		205	143	905	525	371
Large Tanks w/o Control	16,161		6,867		6,622	15,416	2,446	4,284	4,916
Small Tanks w/Flares	0		51		249	237	208	201	168
Small Tanks w/o Flares	89,757		31,176		40,152	43,448	63,168	47,749	37,959
Malfunctioning Separator Dump Valves	7		4		648	40	80	254	197
Total Emissions	**106,429**		**38,461**		**48,892**	**60,556**	**67,595**	**53,613**	**44,217**
Previous Estimate	*16,421*		*11,331*		*21,493*	*24,435*	*21,194*	*17,294*	*NA*

NA (Not Applicable)

Table 3-77: Storage Tanks National CO_2 Emissions (kt CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
Large Tanks w/Flares	579	422	1,804	1,356	840	795	825
Large Tanks w/VRU	0	2	0	0	1	1	1
Large Tanks w/o Control	2	1	1	37	1	1	1
Small Tanks w/Flares	0	13	72	87	82	41	28
Small Tanks w/o Flares	47	18	23	26	33	24	18
Malfunctioning Separator Dump Valves	0	0	2	0	0	1	0
Total Emissions	**628**	**456**	**1,902**	**1,507**	**956**	**862**	**873**
Previous Estimate	*298*	*380*	*1,131*	*844*	*634*	*574*	*NA*

NA (Not Applicable)

Equipment Leaks (Methodological Update)

EPA updated the calculation methodology for onshore production equipment leaks to use basin-specific equipment-level activity factors (e.g., separators per well) from GHGRP data. Previously, national average equipment activity factors developed using RY2014 GHGRP data were used in the Inventory for all years. In this methodological update, EPA summed basin-level emissions together to develop national emissions. The *Disaggregation* memo and *Production Disaggregation* memo present additional information and considerations for this update.

EPA calculated basin-specific equipment-level activity factors for all basins that reported Subpart W data. The factors were year-specific for RY2015 through RY2021. EPA also retained the previous Inventory's activity factors for 1990 through 1992 and used linear interpolation between the 1992 and 2015 activity factors at the basin-level. For basins without subpart W data available, EPA applied national average activity factors. This methodological update applies only for activity factors. The previous Inventory's CH_4 emission factors for onshore production segment equipment leaks (by equipment type) were retained and used to develop CH_4 estimates. Since the CH_4 emission factors were not updated, EPA also retained the Gas STAR reductions that are applicable to equipment leaks.

The calculation methodology for CO_2 emissions was not updated for the public review version of the Inventory. The previous Inventory's methodology was retained to develop CO_2 estimates. EPA will calculate equipment leak CO_2 emissions in the same manner as CH_4 emissions for the final Inventory.

This update resulted in CH_4 emission estimates an average of 8 percent higher across the time series compared to the previous Inventory. The 2020 emission estimate is 39 percent higher than in the previous Inventory. The early years of the time series are minimally impacted by the update, with average CH_4 emissions 1 percent lower for years 1990 through 2002, compared to the previous Inventory. Methane emissions are an average of 14 percent higher for 2002 through 2020, compared to the previous Inventory. These recent years of the time series relied on the basin-specific activity factors and certain basins had higher activity factors compared to the national average factors (e.g., Anadarko Basin, Arkla, Fort Worth Syncline, Gulf Coast, Powder River, San Juan, Strawn).

Table 3-78: Production Equipment Leaks National CH_4 Emissions (Metric Tons CH_4)

Source	1990	2005	2017	2018	2019	2020	2021
Heaters	12,116	20,307	20,068	80,312	16,421	19,223	17,694
Separators	40,746	92,060	129,978	124,339	128,675	132,409	112,425
Dehydrators	12,722	12,796	4,485	5,552	3,739	3,133	4,128
Meters/Piping	42,205	72,148	78,403	81,139	85,625	154,544	135,476
Compressors	29,858	64,877	73,000	72,026	64,471	60,157	73,963
Gas STAR Reductions for Leaks	0	20,908	2,748	71	133	133	133
Total Emissions	**137,647**	**239,280**	**303,187**	**363,296**	**298,797**	**369,333**	**343,553**

Previous Estimate	*138,844*	*220,489*	*273,028*	*274,664*	*270,662*	*265,657*	*NA*

NA (Not Applicable)

Chemical injection Pumps (Methodological Update)

EPA updated the calculation methodology for chemical injection pumps to use basin-specific activity factors from GHGRP data. Previously, national average activity factors developed using RY2014 GHGRP data were used in the Inventory for all years. In this methodological update, EPA summed basin-level emissions together to develop national emissions. The *Disaggregation* memo and *Production Disaggregation* memo present additional information and considerations for this update.

EPA calculated basin-specific activity factors for all basins that reported subpart W data. The factors were year-specific for RY2015 through RY2021. EPA also retained the previous Inventory's activity factors for 1990 through 1992 and applied linear interpolation between the 1992 and 2015 activity factors at the basin-level. For basins without subpart W data available, EPA applied national average activity factors. This methodological update applies only to activity factors. The previous Inventory's CH_4 emission factor for chemical injection pumps was retained and used to develop CH_4 estimates.

The estimation methodology for CO_2 emissions was not updated for the public review version of the Inventory. The previous Inventory's methodology was retained to develop CO_2 estimates. EPA will calculate chemical injection pump CO_2 emissions in the same manner as CH_4 emissions for the final Inventory.

This update resulted in CH_4 emission estimates an average of 86 percent higher across the time-series. The 2020 emission estimate is 24 percent lower than in the previous Inventory. The emissions increase across the time-series is predominantly due to the Bend Arch, which has a very high RY2015 activity factor (chemical injection pumps per well), which then impacts prior years because it's used in the linear interpolation back to the 1992 activity factor.

Table 3-79: Chemical Injection Pumps National CH_4 Emissions (Metric Tons CH_4)

Source	1990	2005	2017	2018	2019	2020	2021
Chemical Injection Pumps	25,345	183,832	113,726	120,984	108,546	84,002	76,315
Previous Estimate	*27,158*	*84,573*	*116,107*	*115,140*	*113,538*	*110,785*	*NA*

NA (Not Applicable)

Liquids Unloading (Methodological Update)

EPA updated the calculation methodology for liquids unloading to use basin-specific activity factors and emission factors calculated from subpart W data for each type of liquids unloading (i.e., with and without plunger lifts). Previously, national average activity and emission factors calculated using Subpart W data were applied to estimate liquids unloading emissions. In this methodological update, EPA summed basin-level emissions together to develop national emissions. The *Disaggregation* memo and *Production Disaggregation* memo present additional information and considerations for this update.

EPA calculated basin-specific activity factors, and CH_4 and CO_2 emission factors for all basins that reported subpart W data. The factors were also year-specific for RY2011 through RY2021. EPA also revised the previous Inventory's activity factor and emission factor assumptions for 1990 through 1992. Previously, Year 2011 emission factors were applied to all prior years of the time series and activity factors were derived by linear interpolation between Year 2011 data and API/ANGA data (collected in 2011) for 1990. In the current Inventory, EPA used activity and emission factors developed using GRI data for 1990 through 1992 (GRI/EPA 1996). The 1996 GRI study did not include CO_2 data for liquids unloading. EPA used RY2011 CO_2 emission factors for the earlier years in the time series (i.e., 1990 through 2010). The same activity and emission factors derived from the GRI data were used for all basins for 1990 through 1992. For the remaining time series years (i.e., 1993-2010), EPA applied linear interpolation between the 1992 and 2011 factors at the basin-level. For basins without subpart W data available, EPA applied national average activity and emission factors.

This update resulted in CH_4 and CO_2 emission estimates an average of 15 percent lower across the time series than in the previous Inventory. In the earlier years of the time series (i.e., 1990 through 2006), CH_4 emissions are lower than in the previous Inventory by an average of 43 percent. CO_2 emissions over the same time period are lower than in the previous Inventory by an average of 38 percent. For the time series years with reported GHGRP data (i.e., 2011 through 2020), CH_4 emissions increased by an average of 21 percent, compared to the previous Inventory. Similarly, CO_2 emissions also increased by an average of 17 percent during 2011 through 2020. The basin-level approach's emissions were higher than the previous Inventory's because certain basins with high gas well counts (e.g., Appalachian and Anadarko basins) had higher emission factors than the national average.

Table 3-80: Liquids Unloading National CH_4 Emissions (Metric Tons CH_4)

Source	1990	2005	2017	2018	2019	2020	2021
Liquids Unloading With Plunger Lifts	0	144,856	68,633	99,159	85,536	60,280	39,456
Liquids Unloading Without Plunger Lifts	76,815	214,070	116,012	166,014	124,428	98,687	80,690
Total Emissions	**76,815**	**358,925**	**184,645**	**265,173**	**209,964**	**158,968**	**120,145**
Previous Estimate	*373,528*	*379,184*	*155,178*	*207,603*	*175,156*	*129,831*	*NA*

NA (Not Applicable)

Table 3-81: Liquids Unloading National CO_2 Emissions (Metric Tons CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
Liquids Unloading With Plunger Lifts	0	11,926	3,376	4,212	2,864	2,606	1,967
Liquids Unloading Without Plunger Lifts	44,810	40,806	5,390	7,227	7,270	3,562	3,733
Total Emissions	**44,810**	**52,733**	**8,767**	**11,439**	**10,134**	**6,168**	**5,700**
Previous Estimate	*83,155*	*67,087*	*7,487*	*9,181*	*8,284*	*5,491*	*NA*

NA (Not Applicable)

Miscellaneous Production Flaring (Recalculation with Updated Data)

Miscellaneous production flaring CO_2 emissions estimates are on average 0.2 percent higher across the 1990 to 2020 time series compared with the previous Inventory and the 2020 estimate is 23 percent higher, compared to the previous Inventory. These changes were due to GHGRP submission revisions.

Table 3-82: Miscellaneous Production Flaring National Emissions (kt CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
Miscellaneous Flaring-Gulf Coast Basin	NO	166	209	137	398	250	267
Miscellaneous Flaring-Williston Basin	NO	+	10	6	3	4	4
Miscellaneous Flaring-Permian Basin	NO	260	622	707	889	831	483
Miscellaneous Flaring-Other Basins	NO	117	306	476	305	213	236
Total Emissions	**NO**	**543**	**1,148**	**1,326**	**1,595**	**1,298**	**991**
Previous Estimate	*NO*	*543*	*1,145*	*1,344*	*1,904*	*1,060*	*NA*

+ Does not exceed 0.5 kt.
NO (Not Occurring)
NA (Not Applicable)

Gathering and Boosting – Tanks (Recalculation with Updated Data)

Methane emission estimates for gathering and boosting tanks are on average 0.1 percent lower across the 1990 to 2020 time series than in the previous Inventory. The 2020 estimate is 2 percent lower than in the previous Inventory. These changes were due to GHGRP submission revisions.

Table 3-83: Tanks National Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Tanks	129,829		165,236		255,244	249,489	295,914	239,623	276,748
Previous Estimate	*129,829*		*165,236*		*255,244*	*249,489*	*300,169*	*244,257*	*NA*

NA (Not Applicable)

Gathering and Boosting – Station Blowdowns

Methane emissions estimates for gathering and boosting station blowdowns are on average 0.7 percent lower across the 1990 to 2020 time series than in the previous Inventory. The 2020 estimate is 10 percent lower than in the previous Inventory. These changes were due to GHGRP submission revisions.

Table 3-84: Station Blowdowns National Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Station Blowdowns	20,517		26,113		63,852	78,548	38,412	40,468	42,231
Previous Estimate	*20,517*		*26,113*		*63,852*	*78,548*	*43,865*	*44,881*	*NA*

NA (Not Applicable)

Gathering and Boosting – Dehydrator Vents (Large Units)

Methane emissions for dehydrator vents at large units are on average of 4 percent higher across the 1990 to 2020 time series compared with the previous Inventory. The 2020 estimate is 115 percent higher compared to the previous Inventory. The dehydrator vents at large units CO_2 emissions estimate increased by an average of 10 percent across the time series and by 292 percent in 2020, compared to the previous Inventory. These changes were due to GHGRP submission revisions.

Table 3-85: Dehydrator Vents National Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Dehydrator Vents	35,716		45,457		61,754	56,543	56,405	52,323	59,207
Previous Estimate	*35,716*		*45,457*		*61,386*	*56,381*	*55,967*	*24,345*	*NA*

NA (Not Applicable)

Table 3-86: Dehydrator Vents National Emissions (kt CO_2)

Source	1990		2005		2017	2018	2019	2020	2021
Dehydrator Vents	371		472		771	820	1,039	1,048	995
Previous Estimate	*371*		*472*		*772*	*820*	*907*	*267*	*NA*

NA (Not Applicable)

Gathering and Boosting – Flare Stacks (Recalculation with Updated Data)

The flare stacks CO_2 emissions estimate are an average of 0.3 percent lower across the time series compared with the previous Inventory. The 2020 estimate is 4 percent lower, compared to the previous Inventory. These changes were due to GHGRP submission revisions.

Table 3-87: Production Storage Tanks National Emissions (kt CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
Flare Stacks	1,355	1,725	2,256	3,696	4,777	2,822	2,631
Previous Estimate	*1,355*	*1,725*	*2,256*	*3,695*	*5,028*	*2,926*	*NA*

NA (Not Applicable)

Processing

Flares (Recalculation with Updated Data)

Processing segment flare CO_2 emission estimates are on average of less than 1 percent higher across the 1993 to 2020 time series than in the previous Inventory. The estimate for 2020 is 3 percent higher than in the previous Inventory. These changes were due to GHGRP submission revisions.

Table 3-88: Processing Segment Flares National CO_2 Emissions (kt CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
Flares	NO	3,517	5,587	5,945	9,859	8,120	7,381
Previous Estimate	*NO*	*3,517*	*5,590*	*6,176*	*9,837*	*7,879*	*NA*

NA (Not Applicable)
NO (Not Occurring)

AGR Vents (Recalculation with Updated Data)

AGR vents CO_2 emission estimates are on average lower than the previous Inventory by less than 1 percent across the 1990 to 2020 time series. Emission estimates for 2020 are 2 percent lower than in the previous Inventory. These changes were due to GHGRP submission revisions.

Table 3-89: AGR Vents National CO_2 Emissions (kt CO_2)

Source	1990	2005	2017	2018	2019	2020	2021
AGR Vents	28,282	15,281	17,313	16,788	16,325	17,258	18,658
Previous Estimate	*28,282*	*15,281*	*17,364*	*16,792*	*16,505*	*17,559*	*NA*

NA (Not Applicable)
NO (Not Occurring)

Transmission and Storage

There were no methodological updates to the transmission and storage segment, and recalculations resulted in an average increase in calculated CH_4 emissions over the time series of 0.2 percent. CO_2 emissions will be updated for the Final Inventory; see Planned Improvements.

Distribution

Mains – Unprotected Steel (Recalculation with Updated Data)

Methane emissions estimates for unprotected steel distribution mains are on average 0.6 percent lower across the 1990 to 2020 time series compared to the previous Inventory and 6 percent lower in 2020, compared to the previous Inventory. The emission changes were due to updated PHMSA pipeline mileage data.

Table 3-90: Mains – Unprotected Steel National CH_4 Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Mains – Unprotected Steel	231,201		91,262		44,574	42,581	40,732	39.261	37,488
Previous Estimate	*231,201*		*91,262*		*47,236*	*45,213*	*43,369*	*41,554*	*NA*

NA (Not Applicable)

Post-Meter

Post-Meter (Recalculation with Updated Data)

Post-Meter CH_4 emissions estimates are higher by an average of 0.1 percent across the 1990 to 2020 time series compared with the previous Inventory, and 1 percent higher in 2020, compared to the previous Inventory. The emission changes were due to changes in residential and industrial natural gas consumption data.

Table 3-91: Post-Meter National CH_4 Emissions (Metric Tons CH_4)

Source	1990		2005		2017	2018	2019	2020	2021
Post-Meter	289,951		344,464		424,492	445,323	456,679	462,751	463,072
Previous Estimate	*289,951*		*344,464*		*424,492*	*445,220*	*456,551*	*459,072*	*NA*

NA (Not Applicable)

Planned Improvements

Planned Improvements for 2023 Inventory

This draft of the Inventory does not yet incorporate updated activity data products for the following data inputs, due to a data base subscription lapse: gas well counts, wells drilled, wells completed, and production. For these inputs, year 2020 values for activity data are used in place of year 2021. The Final Inventory (to be published April 2023) will incorporate the latest activity data.

The CO_2 emissions estimates for LNG export terminals will be updated for the Final Inventory to correct an error in the emission factor calculations in this draft Inventory. The recalculation will result in average annual CO_2 emissions estimates for 1990 through 2015 decreasing from 122 kt to 23 kt, consistent with the prior Inventory, and annual average CO_2 emissions for 2016 through 2021 will increase by 69 kt.

Basin-level approaches for pneumatic controllers, equipment leaks, and chemical injection pumps were applied to calculate CH_4 emissions for public review. For the final Inventory, EPA would apply consistent methods for both CO_2 and CH_4 emissions calculations.

Additional information on the update and specific requests for stakeholder feedback can be found in the *Disaggregation* memo and *Production Disaggregation* memos. Feedback EPA has received in response to the memo include that basin-level data from GHGRP can improve accuracy of estimates when applied appropriately, that EPA should consider application of the approach to only basins with 50 percent coverage or more, and that liquids unloading is a source that may be well-suited to a basin-level approach, EPA will consider this feedback and any additional feedback received and may revise the calculations in the Inventory.

Upcoming Data, and Additional Data that Could Inform the Inventory

EPA will assess new data received by EPA's Greenhouse Gas Reporting Program, Methane Challenge Program on an ongoing basis, which may be used to validate or improve existing estimates and assumptions.

EPA continues to track studies that contain data that may be used to update the Inventory. EPA will also continue to assess studies that include and compare both top-down and bottom-up emission estimates, which could lead to improved understanding of unassigned high emitters (e.g., identification of emission sources and information on frequency of high emitters) as recommended in previous stakeholder comments.

3.8 Abandoned Oil and Gas Wells (CRF Source Categories 1B2a and 1B2b)

Note that this draft of the Inventory does not yet incorporate updated activity data for the following data inputs, due to a data base subscription lapse: abandoned well counts, and fractions of plugged and unplugged abandoned wells. Year 2020 values for activity data are used in place of year 2021. The Final Inventory (to be published April 2023) will incorporate the latest activity data.

The term "abandoned wells", as used in the Inventory, encompasses various types of oil and gas wells, including orphaned wells and other non-producing wells:

- Wells with no recent production, and not plugged. Common terms (such as those used in state databases) might include: inactive, temporarily abandoned, shut-in, dormant, and idle.
- Wells with no recent production and no responsible operator. Common terms might include: orphaned, deserted, long-term idle, and abandoned.
- Wells that have been plugged to prevent migration of gas or fluids.

The U.S. population of abandoned oil and gas wells (including orphaned wells and other non-producing wells) is around 3.7 million (with around 2.9 million abandoned oil wells and 0.8 million abandoned gas wells). The methods to calculate emissions from abandoned wells involve calculating the total populations of plugged and unplugged abandoned oil and gas wells in the United States and the application of emission factors. An estimate of the number of orphaned wells within this population is not developed as part of the methodology. Wells that are plugged have much lower average emissions than wells that are unplugged (less than 1 kg CH_4 per well per year, versus over 100 kg CH_4 per well per year). Around 42 percent of the abandoned well population in the United States is plugged. This fraction has increased over the Inventory time series (from around 22 percent in 1990) as more wells fall under regulations and programs requiring or promoting plugging of abandoned wells.

Abandoned oil wells. Abandoned oil wells emitted 231 kt CH_4 and 5 kt CO_2 in 2021. Emissions of both gases increased by 3 percent from 1990, while the total population of abandoned oil wells increased 37 percent.

Abandoned gas wells. Abandoned gas wells emitted 63 kt CH_4 and 3 kt CO_2 in 2021. Emissions of both gases increased by 25 percent from 1990, while the total population of abandoned gas wells increased 75 percent.

Table 3-92: CH_4 Emissions from Abandoned Oil and Gas Wells (MMT CO_2 Eq.)

Activity	1990	2005	2017	2018	2019	2020	2021
Abandoned Oil Wells	6.3	6.5	6.5	6.5	6.5	6.5	6.5
Abandoned Gas Wells	1.4	1.6	1.8	1.8	1.8	1.8	1.8
Total	**7.7**	**8.1**	**8.3**	**8.3**	**8.3**	**8.2**	**8.2**

Table 3-93: CH_4 Emissions from Abandoned Oil and Gas Wells (kt)

Activity	1990	2005	2017	2018	2019	2020	2021
Abandoned Oil Wells	223	232	232	232	233	231	231
Abandoned Gas Wells	51	57	63	63	64	63	63
Total	**274**	**289**	**295**	**296**	**297**	**295**	**295**

Table 3-94: CO_2 Emissions from Abandoned Oil and Gas Wells (MMT CO_2)

Activity	1990	2005	2017	2018	2019	2020	2021
Abandoned Oil Wells	+	+	+	+	+	+	+
Abandoned Gas Wells	+	+	+	+	+	+	+
Total	**+**	**+**	**+**	**+**	**+**	**+**	**+**

+ Does not exceed 0.05 MMT CO_2 Eq.

Table 3-95: CO_2 Emissions from Abandoned Oil and Gas Wells (kt)

Activity	1990	2005	2017	2018	2019	2020	2021
Abandoned Oil Wells	5	5	5	5	5	5	5
Abandoned Gas Wells	2	2	3	3	3	3	3
Total	**7**	**7**	**7**	**7**	**7**	**7**	**8**

Note: Totals may not sum due to independent rounding.

Methodology and Time-Series Consistency

EPA uses a Tier 2 method from IPCC (2019) to quantify emissions from abandoned oil and gas wells. EPA's approach is based on the number of plugged and unplugged abandoned wells in the Appalachian region and in the rest of the U.S., and emission factors for plugged and unplugged abandoned wells in Appalachia and the rest of the U.S. Methods for abandoned wells are unavailable in IPCC (2006). The details of this approach and of the data sources used are described in the memorandum *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2016: Abandoned Wells in Natural Gas and Petroleum Systems (2018 Abandoned Wells Memo).*

EPA developed abandoned well CH_4 emission factors using data from Kang et al. (2016) and Townsend-Small et al. (2016). Plugged and unplugged abandoned well CH_4 emission factors were developed at the national-level (using emission data from Townsend-Small et al.) and for the Appalachia region (using emission data from measurements in Pennsylvania and Ohio conducted by Kang et al. and Townsend-Small et al., respectively). The Appalachia region emissions factors were applied to abandoned wells in states in the Appalachian basin region, and the national-level emission factors were applied to abandoned wells in all other states. EPA developed abandoned well CO_2 emission factors using the CH_4 emission factors and an assumed ratio of CO_2-to-CH_4 gas content, similar to the approach used to calculate CO_2 emissions for many sources in Petroleum Systems and Natural Gas Systems. For abandoned oil wells, EPA used the Petroleum Systems default production segment associated gas ratio of 0.020 MT CO_2/MT CH_4, which was derived through API TankCalc modeling runs. For abandoned gas wells, EPA used the Natural Gas Systems default production segment CH_4 and CO_2 gas content values (GRI/EPA 1996, GTI 2001) to develop a ratio of 0.044 MT CO_2/MT CH_4. The same respective emission factors are applied for each year of the time series.

EPA developed state-level annual counts of abandoned wells for 1990 through 2020 by summing together an annual estimate of abandoned wells in the Enverus data set (Enverus 2021), and an estimate of total abandoned wells not included the Enverus dataset (see *2018 Abandoned Wells Memo* for additional information on how the value was calculated) for each state. References reviewed to develop the number of abandoned wells not included in the Enverus dataset include historical records collected by state agencies and by USGS.

The total abandoned well population was then split into plugged and unplugged wells by applying an assumption that all abandoned wells were unplugged in 1950 and using Enverus data to calculate the fraction of plugged abandoned wells in 2020 in that data set, which was then applied to the total population of abandoned wells for 2020 and 2021. Linear interpolation was applied between the 1950 value and 2020 value to calculate the plugged fraction for intermediate years. See the memorandum *Inventory of U.S. Greenhouse Gas Emissions and Sinks 1990-2016: Abandoned Wells in Natural Gas and Petroleum Systems (2018 Abandoned Wells Memo)* for details.[87] State-level plugged and unplugged fractions were developed for the time-series using state-level Enverus data for 2020

[87] See https://www.epa.gov/ghgemissions/natural-gas-and-petroleum-systems.

and linear interpolation between 1950 and 2020 plugged and unplugged fractions. Abandoned wells in all states were assumed to be unplugged in 1950.

Abandoned Oil Wells

Table 3-96: Abandoned Oil Wells Activity Data, CH_4 and CO_2 Emissions (kt)

Source	1990	2005	2017	2018	2019	2020	2021
Plugged abandoned oil wells	474,432	799,331	1,105,366	1,139,476	1,175,867	1,192,907	1,192,907
Unplugged abandoned oil wells	1,664,717	1,749,329	1,749,813	1,751,999	1,756,573	1,739,533	1,739,533
Total Abandoned Oil Wells	2,139,149	2,548,660	2,855,179	2,891,475	2,932,440	2,932,440	2,932,440
Abandoned oil wells in Appalachia	23%	21%	19%	19%	19%	19%	19%
Abandoned oil wells outside of Appalachia	77%	79%	81%	81%	81%	81%	81%
CH_4 from plugged abandoned oil wells (kt)	0.20	0.30	0.39	0.40	0.41	0.42	0.42
CH_4 from unplugged abandoned oil wells(kt)	223.1	231.3	231.5	231.8	232.5	230.7	230.7
Total CH_4 from Abandoned oil wells (kt)	**223.3**	**231.6**	**231.9**	**232.2**	**232.9**	**231.1**	**231.1**
Total CO_2 from Abandoned oil wells (kt)	**4.5**	**4.7**	**4.7**	**4.7**	**4.7**	**4.7**	**4.7**

Abandoned Gas Wells

Table 3-97: Abandoned Gas Wells Activity Data, CH_4 and CO_2 Emissions (kt)

Source	1990	2005	2017	2018	2019	2020	2021
Plugged abandoned gas wells	107,292	206,413	332,743	342,495	353,746	358,871	358,871
Unplugged abandoned gas wells	349,041	397,844	440,367	442,014	444,532	439,407	439,407
Total Abandoned Gas Wells	456,333	604,257	773,110	784,509	798,278	798,278	798,278
Abandoned gas wells in Appalachia	29%	26%	24%	24%	25%	25%	25%
Abandoned gas wells outside of Appalachia	71%	74%	76%	76%	75%	75%	75%
CH_4 from plugged abandoned gas wells (kt)	0.07	0.12	0.17	0.18	0.19	0.19	0.19
CH_4 from unplugged abandoned gas wells (kt)	50.9	56.8	62.8	63.2	63.8	63.2	63.2
Total CH_4 from Abandoned gas wells (kt)	**50.9**	**56.9**	**63.0**	**63.4**	**64.0**	**63.4**	**63.4**
Total CO_2 from Abandoned gas wells (kt)	**2.2**	**2.5**	**2.8**	**2.8**	**2.8**	**2.8**	**2.8**

Uncertainty—TO BE UPDATED FOR FINAL INVENTORY REPORT

To characterize uncertainty surrounding estimates of abandoned well emissions, EPA conducted a quantitative uncertainty analysis using the IPCC Approach 2 methodology (Monte Carlo simulation technique). See the *2018 Abandoned Wells Memo* for details of the uncertainty analysis methods. EPA used Microsoft Excel's @RISK add-in tool to estimate the 95 percent confidence bound around total methane emissions from abandoned oil and gas wells in year 2019, then applied the calculated bounds to both CH_4 and CO_2 emissions estimates for each population. The @RISK add-in provides for the specification of probability density functions (PDFs) for key variables within a computational structure that mirrors the calculation of the inventory estimate. EPA used measurement data from the Kang et al. (2016) and Townsend-Small et al. (2016) studies to characterize the CH_4 emission factor PDFs. For activity data inputs (e.g., total count of abandoned wells, split between plugged and unplugged), EPA assigned default uncertainty bounds of ± 10 percent based on expert judgment.

The IPCC guidance notes that in using this method, "some uncertainties that are not addressed by statistical means may exist, including those arising from omissions or double counting, or other conceptual errors, or from incomplete understanding of the processes that may lead to inaccuracies in estimates developed from models." As a result, the understanding of the uncertainty of emission estimates for this category evolves and improves as the underlying methodologies and datasets improve. The uncertainty bounds reported below only reflect those uncertainties that EPA has been able to quantify and do not incorporate considerations such as modeling uncertainty, data representativeness, measurement errors, misreporting or misclassification.

The results presented below in Table 3-98 provide the 95 percent confidence bound within which actual emissions from abandoned oil and gas wells are likely to fall for the year 2019, using the recommended IPCC methodology. Abandoned oil well CH_4 emissions in 2019 were estimated to be between 0.9 and 16.5 MMT CO_2 Eq., while abandoned gas well CH_4 emissions were estimated to be between 0.2 and 4.3 MMT CO_2 Eq. at a 95 percent confidence level. Uncertainty bounds for other years of the time series have not been calculated, but uncertainty is expected to vary over the time series.

Table 3-98: Approach 2 Quantitative Uncertainty Estimates for CH_4 and CO_2 Emissions from Petroleum and Natural Gas Systems (MMT CO_2 Eq. and Percent)

Source	Gas	2019 Emission Estimate (MMT CO_2 Eq.)[b]	Uncertainty Range Relative to Emission Estimate[a] (MMT CO_2 Eq.)		(%)	
			Lower Bound	Upper Bound	Lower Bound	Upper Bound
Abandoned Oil Wells	CH_4	5.2	0.9	16.5	-83%	+217%
Abandoned Gas Wells	CH_4	1.4	0.2	4.3	-83%	+217%
Abandoned Oil Wells	CO_2	0.004	0.001	0.013	-83%	+217%
Abandoned Gas Wells	CO_2	0.002	0.0004	0.008	-83%	+217%

[a] Range of emission estimates estimated by applying the 95 percent confidence intervals obtained from the Monte Carlo Simulation analysis conducted for total abandoned oil and gas well CH_4 emissions in year 2019.

[b] All reported values are rounded after calculation. As a result, lower and upper bounds may not be duplicable from other rounded values as shown in table.

QA/QC and Verification Discussion

The emission estimates in the Inventory are continually reviewed and assessed to determine whether emission factors and activity factors accurately reflect current industry practices. A QA/QC analysis was performed for data gathering and input, documentation, and calculation. QA/QC checks are consistently conducted to minimize human error in the model calculations. EPA performs a thorough review of information associated with new studies to assess whether the assumptions in the Inventory are consistent with industry practices and whether new data is available that could be considered for updates to the estimates. As in previous years, EPA conducted early engagement and communication with stakeholders on updates prior to public review. EPA held stakeholder webinars on greenhouse gas data for oil and gas in September and November of 2022.

Recalculations Discussion

EPA updated the Inventory methodology to estimate abandoned well emissions at the state-level as an intermediate step to calculating national emissions. Previously, well counts were developed for the Appalachian region and for all other regions as a total, and plugged and unplugged fractions were developed at the national-level. In the current Inventory, EPA used abandoned well counts and plugged and unplugged fractions at the state-level to estimate emissions. The incorporation of disaggregated, state-level data will improve future versions of both the gridded and state-level greenhouse gas inventories as geographic differences in plugging rates can now be reflected. This will allow EPA to use the gridded greenhouse gas inventory for improved comparisons with atmospheric observation studies, because regions will reflect local differences. In addition, this update will improve the ability of the state-level Inventory to reflect impacts of state-level programs.

The emission factors from the previous Inventory were retained and used to estimate state-level emissions, with Appalachia-specific factors applied to states in Appalachia. The state-level emissions were then summed up to the national level. As an outcome of these revisions, total calculated abandoned well CH_4 emissions across the time series are an average of 6 percent higher than in the previous Inventory. The calculated value for 2020 is 7 percent higher than in the previous Inventory.

The main cause of increased emission estimate across the time series is the application of state-specific fractions of plugged wells, which resulted in a larger fraction of unplugged wells in Appalachia (which has a higher unplugged well emission factor than other regions) than in the previous inventory, which applied a national average plugging fraction to the entire U.S. abandoned well population.

In the previous Inventory, abandoned dry wells were proportionally allocated between abandoned oil and gas wells at the national level. In the current Inventory, dry wells are proportionally allocated to abandoned oil and gas wells at the state level. The total counts of abandoned wells changed by 0.02 percent (decrease), compared with the previous inventory. The counts of abandoned oil wells are about 1.6 percent lower across the time series compared to the previous Inventory and gas wells are about 7 percent higher.

In addition, for the current Inventory, CO_2-equivalent emissions totals have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth Assessment Report* (AR4) (IPCC 2007) used in the previous inventories. The AR5 GWPs have been applied across the entire time series for consistency. The GWP of CH_4 has increased from 25 to 28, leading to an overall increase in the calculated CO_2-equivalent emissions of CH_4.

Compared to the previous Inventory which applied 100-year GWP values from AR4, in the current Inventory (including other recalculations noted above), CO_2-equivalent CH_4 emissions increased by 16 percent on average over the time series. Further discussion on this update and the overall impacts of updating the Inventory GWP values to reflect the IPCC AR5 can be found in Chapter 9, Recalculations and Improvements.

Planned Improvements

This draft of the Inventory does not yet incorporate updated activity data for the following data inputs, due to a data base subscription lapse: abandoned well counts, and fractions of plugged and unplugged abandoned wells. Year 2020 values for activity data are used in place of year 2021. The Final Inventory (to be published April 2023) will incorporate the latest activity data.

EPA will continue to assess new data and stakeholder feedback on considerations (such as potential use of emission factor data from regions not included in the measurement studies on which current emission factors are based) to improve the abandoned well count estimates and emission factors. In future Inventories, EPA will assess data that become available from Department of Interior and Department of Energy orphan well plugging programs.

3.9 International Bunker Fuels (CRF Source Category 1: Memo Items)

Emissions resulting from the combustion of fuels used for international transport activities, termed international bunker fuels under the UNFCCC, are not included in national emission totals, but are reported separately based upon location of fuel sales. The decision to report emissions from international bunker fuels separately, instead of allocating them to a particular country, was made by the Intergovernmental Negotiating Committee in establishing the Framework Convention on Climate Change.[88] These decisions are reflected in the IPCC methodological guidance, including IPCC (2006), in which countries are requested to report emissions from ships or aircraft that depart from their ports with fuel purchased within national boundaries and are engaged in international transport separately from national totals (IPCC 2006).[89]

Two transport modes are addressed under the IPCC definition of international bunker fuels: aviation and marine.[90] Greenhouse gases emitted from the combustion of international bunker fuels, like other fossil fuels, include CO_2, CH_4 and N_2O for marine transport modes, and CO_2 and N_2O for aviation transport modes. Emissions from ground transport activities—by road vehicles and trains—even when crossing international borders are allocated to the country where the fuel was loaded into the vehicle and, therefore, are not counted as bunker fuel emissions.

The *2006 IPCC Guidelines* distinguish between three different modes of air traffic: civil aviation, military aviation, and general aviation. Civil aviation comprises aircraft used for the commercial transport of passengers and freight, military aviation comprises aircraft under the control of national armed forces, and general aviation applies to recreational and small corporate aircraft. The *2006 IPCC Guidelines* further define international bunker fuel use from civil aviation as the fuel combusted for civil (e.g., commercial) aviation purposes by aircraft arriving or departing on international flight segments. However, as mentioned above, and in keeping with the *2006 IPCC Guidelines*, only the fuel purchased in the United States and used by aircraft taking-off (i.e., departing) from the United States are reported here. The standard fuel used for civil and military aviation is kerosene-type jet fuel, while the typical fuel used for general aviation is aviation gasoline.[91]

Emissions of CO_2 from aircraft are essentially a function of fuel consumption. Nitrous oxide emissions also depend upon engine characteristics, flight conditions, and flight phase (i.e., take-off, climb, cruise, decent, and landing). Recent data suggest that little or no CH_4 is emitted by modern engines (Anderson et al. 2011), and as a result, CH_4 emissions from this category are reported as zero. In jet engines, N_2O is primarily produced by the oxidation of atmospheric nitrogen, and the majority of emissions occur during the cruise phase.

International marine bunkers comprise emissions from fuels burned by ocean-going ships of all flags that are engaged in international transport. Ocean-going ships are generally classified as cargo and passenger carrying, military (i.e., U.S. Navy), fishing, and miscellaneous support ships (e.g., tugboats). For the purpose of estimating greenhouse gas emissions, international bunker fuels are solely related to cargo and passenger carrying vessels, which is the largest of the four categories, and military vessels. Two main types of fuels are used on sea-going vessels: distillate diesel fuel and residual fuel oil. Carbon dioxide is the primary greenhouse gas emitted from marine shipping.

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[88] See report of the Intergovernmental Negotiating Committee for a Framework Convention on Climate Change on the work of its ninth session, held at Geneva from 7 to 18 February 1994 (A/AC.237/55, annex I, para. 1c).

[89] Note that the definition of international bunker fuels used by the UNFCCC differs from that used by the International Civil Aviation Organization.

[90] Most emission related international aviation and marine regulations are under the rubric of the International Civil Aviation Organization (ICAO) or the International Maritime Organization (IMO), which develop international codes, recommendations, and conventions, such as the International Convention of the Prevention of Pollution from Ships (MARPOL).

[91] Naphtha-type jet fuel was used in the past by the military in turbojet and turboprop aircraft engines.

Overall, aggregate greenhouse gas emissions in 2021 from the combustion of international bunker fuels from both aviation and marine activities were 69.9 MMT CO_2 Eq., or 33.2 percent below emissions in 1990 (see Table 3-99 and Table 3-100). Emissions from international flights and international shipping voyages departing from the United States have increased by 4.5 percent and decreased by 55.1 percent, respectively, since 1990. The majority of these emissions were in the form of CO_2; however, small amounts of CH_4 (from marine transport modes) and N_2O were also emitted. Commercial aviation bunker fuel data for 2021 were not yet available and were proxied based on 2020 data.

Table 3-99: CO_2, CH_4, and N_2O Emissions from International Bunker Fuels (MMT CO_2 Eq.)

Gas/Mode	1990		2005		2017	2018	2019	2020	2021
CO_2	**103.6**		**113.3**		**120.2**	**122.2**	**116.1**	**69.6**	**69.3**
Aviation	38.2		60.2		77.8	80.9	80.8	39.8	39.9
Commercial	*30.0*		*55.6*		*74.5*	*77.7*	*77.6*	*36.7*	*36.7*
Military	*8.2*		*4.6*		*3.3*	*3.2*	*3.2*	*3.1*	*3.2*
Marine	65.4		53.1		42.4	41.3	35.4	29.9	29.4
CH_4	**0.2**		**0.1**		**0.1**	**0.1**	**0.1**	**0.1**	**0.1**
Aviation	NO		NO		NO	NO	NO	NO	NO
Marine	0.2		0.1		0.1	0.1	0.1	0.1	0.1
N_2O	**0.8**		**0.9**		**0.9**	**1.0**	**0.9**	**0.5**	**0.5**
Aviation	0.3		0.5		0.7	0.7	0.7	0.3	0.3
Marine	0.4		0.4		0.3	0.3	0.2	0.2	0.2
Total	**104.6**		**114.3**		**121.2**	**123.2**	**117.1**	**70.3**	**69.9**

NO (Not Occurring)

Notes: Totals may not sum due to independent rounding. Includes aircraft cruise altitude emissions. 2021 commercial aviation data were not yet available and were proxied based on 2020 data.

Table 3-100: CO_2, CH_4, and N_2O Emissions from International Bunker Fuels (kt)

Gas/Mode	1990		2005		2017	2018	2019	2020	2021
CO_2	**103,634**		**113,328**		**120,192**	**122,179**	**116,132**	**69,638**	**69,280**
Aviation	38,205		60,221		77,764	80,853	80,780	39,781	39,912
Marine	65,429		53,107		42,428	41,325	35,351	29,857	29,369
CH_4	**7**		**5**		**4**	**4**	**4**	**3**	**3**
Aviation	NO		NO		NO	NO	NO	NO	NO
Marine	7		5		4	4	4	3	3
N_2O	**3**		**3**		**4**	**4**	**3**	**2**	**2**
Aviation	1		2		2	3	3	1	1
Marine	2		1		1	1	1	1	1

NO (Not Occurring)

Notes: Totals by gas may not sum due to independent rounding. Includes aircraft cruise altitude emissions. 2021 commercial aviation data were not yet available and were proxied based on 2020 data.

Methodology and Time-Series Consistency

Emissions of CO_2 were for the most part estimated by applying C content and fraction oxidized factors to fuel consumption activity data. This approach is analogous to that described under Section 3.1 – CO_2 from Fossil Fuel Combustion. Carbon content and fraction oxidized factors for jet fuel (except for commercial aviation as per below), distillate fuel oil, and residual fuel oil are the same as used for CO_2 from Fossil Fuel Combustion and are presented in Annex 2.1, Annex 2.2, and Annex 3.8 of this Inventory. Density conversions were taken from ASTM (1989) and USAF (1998). Heat content for distillate fuel oil and residual fuel oil were taken from EIA (2022) and USAF (1998), and heat content for jet fuel was taken from EIA (2022). See below for details on how emission estimates for commercial aviation were determined.

A complete description of the methodology and a listing of the various factors employed can be found in Annex 2.1. See Annex 3.8 for a specific discussion on the methodology used for estimating emissions from international bunker fuel use by the U.S. military.

Emission estimates for CH_4 and N_2O were calculated by multiplying emission factors by measures of fuel consumption by fuel type and mode. Emission factors used in the calculations of CH_4 and N_2O emissions were obtained from the *Revised 1996 IPCC Guidelines* (IPCC/UNEP/OECD/IEA 1997), which is also referenced in the *2006 IPCC Guidelines* (IPCC 2006). For aircraft emissions, the following value, in units of grams of pollutant per kilogram of fuel consumed (g/kg), was employed: 0.1 for N_2O (IPCC 2006). For marine vessels consuming either distillate diesel or residual fuel oil the following values (g/MJ), were employed: 0.315 for CH_4 and 0.08 for N_2O. Activity data for aviation included solely jet fuel consumption statistics, while the marine mode included both distillate diesel and residual fuel oil.

Activity data on domestic and international aircraft fuel consumption were developed by the U.S. Federal Aviation Administration (FAA) using radar-informed data from the FAA Enhanced Traffic Management System (ETMS) for 1990 and 2000 through 2020 as modeled with the Aviation Environmental Design Tool (AEDT). This bottom-up approach is built from modeling dynamic aircraft performance for each flight occurring within an individual calendar year. The analysis incorporates data on the aircraft type, date, flight identifier, departure time, arrival time, departure airport, arrival airport, ground delay at each airport, and real-world flight trajectories. To generate results for a given flight within AEDT, the radar-informed aircraft data is correlated with engine and aircraft performance data to calculate fuel burn and exhaust emissions. Information on exhaust emissions for in-production aircraft engines comes from the International Civil Aviation Organization (ICAO) Aircraft Engine Emissions Databank (EDB). This bottom-up approach is in accordance with the Tier 3B method from the *2006 IPCC Guidelines* (IPCC 2006).

International aviation CO_2 estimates for 1990 and 2000 through 2020 were obtained directly from FAA's AEDT model (FAA 2022). The radar-informed method that was used to estimate CO_2 emissions for commercial aircraft for 1990 and 2000 through 2020 was not possible for 1991 through 1999 because the radar dataset was not available for years prior to 2000. FAA developed Official Airline Guide (OAG) schedule-informed inventories modeled with AEDT and great circle trajectories for 1990, 2000, and 2010. Because fuel consumption and CO_2 emission estimates for years 1991 through 1999 are unavailable, consumption estimates for these years were calculated using fuel consumption estimates from the Bureau of Transportation Statistics (DOT 1991 through 2013), adjusted based on 2000 through 2005 data. See Annex 3.3 for more information on the methodology for estimating emissions from commercial aircraft jet fuel consumption. Data for 2021 are not yet available so 2021 data were proxied based on 2020 data and scaled by the percent difference of 2020 and 2021 jet fuel consumption for commercial aviation reported by the Bureau of Transportation (DOT 1991 through 2021).

Data on U.S. Department of Defense (DoD) aviation bunker fuels and total jet fuel consumed by the U.S. military was supplied by the Office of the Under Secretary of Defense (Installations and Environment), DoD. Estimates of the percentage of each Service's total operations that were international operations were developed by DoD. Military aviation bunkers included international operations, operations conducted from naval vessels at sea, and operations conducted from U.S. installations principally over international water in direct support of military operations at sea. Military aviation bunker fuel emissions were estimated using military fuel and operations data synthesized from unpublished data from DoD's Defense Logistics Agency Energy (DLA Energy 2022). Together, the data allow the quantity of fuel used in military international operations to be estimated. Densities for each jet fuel type were obtained from a report from the U.S. Air Force (USAF 1998). Final jet fuel consumption estimates are presented in Table 3-101. See Annex 3.8 for additional discussion of military data.

Table 3-101: Aviation Jet Fuel Consumption for International Transport (Million Gallons)

Nationality	1990		2005		2017	2018	2019	2020	2021
U.S. and Foreign Carriers	3,155		5,858		7,844	8,178	8,170	3,859	3,859
U.S. Military	862		462		326	315	318	308	321
Total	**4,017**		**6,321**		**8,171**	**8,493**	**8,488**	**4,167**	**4,180**

Note: Totals may not sum due to independent rounding. U.S. and Foreign Carriers 2021 data are not available, so data were proxied based on 2020 data.

In order to quantify the civilian international component of marine bunker fuels, activity data on distillate diesel and residual fuel oil consumption by cargo or passenger carrying marine vessels departing from U.S. ports were collected for individual shipping agents on a monthly basis by the U.S. Customs and Border Protection. This information was then reported in unpublished data collected by the Foreign Trade Division of the U.S. Department of Commerce's Bureau of the Census (DOC 1991 through 2022) for 1990 through 2001, 2007 through 2021, and the Department of Homeland Security's Bunker Report for 2003 through 2006 (DHS 2008). Fuel consumption data for 2002 was interpolated due to inconsistencies in reported fuel consumption data. Activity data on distillate diesel consumption by military vessels departing from U.S. ports were provided by DLA Energy (2022). The total amount of fuel provided to naval vessels was reduced by 21 percent to account for fuel used while the vessels were not-underway (i.e., in port). Data on the percentage of steaming hours underway versus not underway were provided by the U.S. Navy. These fuel consumption estimates are presented in Table 3-102.

Table 3-102: Marine Fuel Consumption for International Transport (Million Gallons)

Fuel Type	1990		2005		2017	2018	2019	2020	2021
Residual Fuel Oil	4,781		3,881		2,975	2,790	2,246	1,964	1,953
Distillate Diesel Fuel & Other	617		444		568	684	702	461	437
U.S. Military Naval Fuels	522		471		307	285	281	296	285
Total	**5,920**		**4,796**		**3,850**	**3,759**	**3,229**	**2,721**	**2,674**

Note: Totals may not sum due to independent rounding.

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2021.

Uncertainty

Emission estimates related to the consumption of international bunker fuels are subject to the same uncertainties as those from domestic aviation and marine mobile combustion emissions; however, additional uncertainties result from the difficulty in collecting accurate fuel consumption activity data for international transport activities separate from domestic transport activities.[92] For example, smaller aircraft on shorter routes often carry sufficient fuel to complete several flight segments without refueling in order to minimize time spent at the airport gate or take advantage of lower fuel prices at particular airports. This practice, called tankering, when done on international flights, complicates the use of fuel sales data for estimating bunker fuel emissions. Tankering is less common with the type of large, long-range aircraft that make many international flights from the United States, however. Similar practices occur in the marine shipping industry where fuel costs represent a significant portion of overall operating costs and fuel prices vary from port to port, leading to some tankering from ports with low fuel costs.

Uncertainties exist with regard to the total fuel used by military aircraft and ships. Total aircraft and ship fuel use estimates were developed from DoD records, which document fuel sold to the DoD Components (e.g., Army, Department of Navy and Air Force) from the Defense Logistics Agency Energy. These data may not include fuel used in aircraft and ships as a result of a Service procuring fuel from, selling fuel to, trading fuel with, or giving fuel to other ships, aircraft, governments, or other entities.

Additionally, there are uncertainties in historical aircraft operations and training activity data. Estimates for the quantity of fuel actually used in Navy and Air Force flying activities reported as bunker fuel emissions had to be estimated based on a combination of available data and expert judgment. Estimates of marine bunker fuel emissions were based on Navy vessel steaming hour data, which reports fuel used while underway and fuel used while not underway. This approach does not capture some voyages that would be classified as domestic for a commercial vessel. Conversely, emissions from fuel used while not underway preceding an international voyage are reported as domestic rather than international as would be done for a commercial vessel. There is uncertainty

[92] See uncertainty discussions under section 3.1 CO_2 from Fossil Fuel Combustion.

associated with ground fuel estimates for 1997 through 2021, including estimates for the quantity of jet fuel allocated to ground transportation. Small fuel quantities may have been used in vehicles or equipment other than that which was assumed for each fuel type.

There are also uncertainties in fuel end-uses by fuel type, emissions factors, fuel densities, diesel fuel sulfur content, aircraft and vessel engine characteristics and fuel efficiencies, and the methodology used to back-calculate the data set to 1990 using the original set from 1995. The data were adjusted for trends in fuel use based on a closely correlating, but not matching, data set. All assumptions used to develop the estimate were based on process knowledge, DoD data, and expert judgments. The magnitude of the potential errors related to the various uncertainties has not been calculated but is believed to be small. The uncertainties associated with future military bunker fuel emission estimates could be reduced through revalidation of assumptions based on data regarding current equipment and operational tempo, however, it is doubtful data with more fidelity exist at this time.

Although aggregate fuel consumption data have been used to estimate emissions from aviation, the recommended method for estimating emissions of gases other than CO_2 in the *2006 IPCC Guidelines* (IPCC 2006) is to use data by specific aircraft type, number of individual flights and, ideally, movement data to better differentiate between domestic and international aviation and to facilitate estimating the effects of changes in technologies. The IPCC also recommends that cruise altitude emissions be estimated separately using fuel consumption data, while landing and take-off (LTO) cycle data be used to estimate near-ground level emissions of gases other than CO_2.[93]

There is also concern regarding the reliability of the existing DOC (1991 through 2022) data on marine vessel fuel consumption reported at U.S. customs stations due to the significant degree of inter-annual variation.

QA/QC and Verification

In order to ensure the quality of the emission estimates from international bunker fuels, General (IPCC Tier 1) and category-specific (Tier 2) Quality Assurance/Quality Control (QA/QC) procedures were implemented consistent with the U.S. Inventory QA/QC plan outlined in Annex 8. The Tier 2 procedures that were implemented involved checks specifically focusing on the activity data and emission factor sources and methodology used for estimating CO_2, CH_4, and N_2O emissions from international bunker fuels in the United States. Emission totals for the different sectors and fuels were compared and trends were investigated. No corrective actions were necessary.

Recalculations Discussion

For the current Inventory, CO_2-equivalent emissions of CH_4 and N_2O from international bunker fuels have been revised to reflect the 100-year global warming potentials (GWPs) provided in the IPCC *Fifth Assessment Report* (AR5) (IPCC 2013). AR5 GWP values differ slightly from those presented in the IPCC *Fourth Assessment Report* (AR4), which was used in the previous inventories (IPCC 2007). The AR5 GWPs have been applied across the entire time series for consistency. Prior inventories used GWPs of 25 and 298 for CH_4 and N_2O, respectively. These values have been updated to 28 and 265, respectively. Compared to the previous Inventory which applied 100-year GWP values from AR4, the average annual change in CO_2-equivalent CH_4 emissions was a 12 percent increase and the average annual change in CO_2-equivalent N_2O emissions was an 11 percent decrease for the time series. As a result of the change in methodology, total emissions across the time series changed by an average annual decrease of 0.1 MMT CO_2 Eq. (less than half a percent) relative to emissions results calculated using the prior GWPs. Further

[93] U.S. aviation emission estimates for CO, NO_x, and NMVOCs are reported by EPA's National Emission Inventory (NEI) Air Pollutant Emission Trends website, and reported under the Mobile Combustion section. It should be noted that these estimates are based solely upon LTO cycles and consequently only capture near ground-level emissions, which are more relevant for air quality evaluations. These estimates also include both domestic and international flights. Therefore, estimates reported under the Mobile Combustion section overestimate IPCC-defined domestic CO, NO_x, and NMVOC emissions by including landing and take-off (LTO) cycles by aircraft on international flights, but underestimate because they do not include emissions from aircraft on domestic flight segments at cruising altitudes.

discussion on this update and the overall impacts of updating the Inventory GWP values to reflect the IPCC AR5 can be found in Chapter 9, Recalculations and Improvements.

Planned Improvements

EPA will evaluate data availability to update the sources for densities, energy contents, and emission factors applied to estimate emissions from aviation and marine fuels. Many are from sources from the late 1990s, such as IPCC/UNEP/OECD/IEA (1997). Potential sources with more recent data include the International Maritime Organization (IMO) greenhouse gas emission inventory, International Air Transport Association (IATA)/ICAO greenhouse gas reporting system (CORSIA), and the EPA Greenhouse Gas Reporting Program (GHGRP) Technical Support Document for Petroleum Products. Specifically, EPA will evaluate data availability to support updating the heat contents and carbon contents of jet fuel with input from EIA.

A longer-term effort is underway to consider the feasibility of including data from a broader range of domestic and international sources for bunker fuels. Potential sources include the IMO greenhouse gas emission inventory, data from the U.S. Coast Guard on vehicle operation currently used in criteria pollutant modeling, data from the International Energy Agency (IEA), relevant updated FAA models to improve aviation bunker fuel estimates, and researching newly available marine bunker data.

3.10 Biomass and Biofuels Consumption (CRF Source Category 1A)

The combustion of biomass fuels—such as wood, charcoal, the biogenic portions of MSW, and wood waste and biomass-based fuels such as ethanol, biogas, and biodiesel—generates CO_2 in addition to CH_4 and N_2O already covered in this chapter. In line with the reporting requirements for inventories submitted under the UNFCCC, CO_2 emissions from biomass combustion have been estimated separately from fossil fuel CO_2 emissions and are not directly included in the energy sector contributions to U.S. totals. In accordance with IPCC methodological guidelines, any such emissions are calculated by accounting for net carbon fluxes from changes in biogenic C reservoirs in wooded or crop lands. For a more complete description of this methodological approach, see the Land Use, Land-Use Change, and Forestry chapter (Chapter 6), which accounts for the contribution of any resulting CO_2 emissions to U.S. totals within the Land Use, Land-Use Change, and Forestry sector's approach.

Therefore, CO_2 emissions from biomass and biofuel consumption are not included specifically in summing energy sector totals. However, they are presented here for informational purposes and to provide detail on biomass and biofuels consumption.

In 2021, total CO_2 emissions from the burning of woody biomass in the industrial, residential, commercial, and electric power sectors were approximately 202.8 MMT CO_2 Eq. (202,841 kt) (see Table 3-103 and Table 3-104). As the largest consumer of woody biomass, the industrial sector was responsible for 62.1 percent of the CO_2 emissions from this source. The residential sector was the second largest emitter, constituting 23.6 percent of the total, while the electric power and commercial sectors accounted for the remainder.

Table 3-103: CO_2 Emissions from Wood Consumption by End-Use Sector (MMT CO_2 Eq.)

End-Use Sector	1990		2005		2017	2018	2019	2020	2021
Industrial	135.3		136.3		135.4	134.4	132.1	127.3	126.0
Residential	59.8		44.3		44.3	54.1	56.3	45.5	47.8
Commercial	6.8		7.2		8.6	8.7	8.7	8.6	8.5
Electric Power	13.3		19.1		23.6	22.8	20.7	19.1	20.5
Total	**215.2**		**206.9**		**212.0**	**220.0**	**217.7**	**200.4**	**202.8**

Table 3-104: CO_2 Emissions from Wood Consumption by End-Use Sector (kt)

End-Use Sector	1990	2005	2017	2018	2019	2020	2021
Industrial	135,348	136,269	135,386	134,417	132,069	127,301	125,970
Residential	59,808	44,340	44,298	54,124	56,253	45,452	47,823
Commercial	6,779	7,218	8,634	8,669	8,693	8,554	8,528
Electric Power	13,252	19,074	23,647	22,795	20,677	19,115	20,519
Total	**215,186**	**206,901**	**211,965**	**220,005**	**217,692**	**200,421**	**202,841**

Note: Totals may not sum due to independent rounding.

Carbon dioxide emissions from combustion of the biogenic components of MSW by the electric power sector were an estimated 15.3 MMT CO_2 (15,329 kt) in 2021. Emissions across the time series are shown in Table 3-105 and Table 3-106. As discussed in Section 3.3, MSW is combusted to produce electricity and the CO_2 emissions from the fossil portion of the MSW (e.g., plastics, textiles, etc.) are included in the energy sector FFC estimates. The MSW also includes biogenic components (e.g., food waste, yard trimmings, natural fibers) and the CO_2 emissions associated with that biogenic portion is included here.

Table 3-105: CO_2 Emissions from Biogenic Components of MSW (MMT CO_2 Eq.)

End-Use Sector	1990	2005	2017	2018	2019	2020	2021
Electric Power	18.5	14.7	16.1	16.1	15.7	15.6	15.3

Table 3-106: CO_2 Emissions from Biogenic Components of MSW (kt)

End-Use Sector	1990	2005	2017	2018	2019	2020	2021
Electric Power	18,534	14,722	16,130	16,115	15,709	15,614	15,329

The transportation sector is responsible for most of the fuel ethanol consumption in the United States. Ethanol used for fuel is currently produced primarily from corn grown in the Midwest, but it can be produced from a variety of biomass feedstocks. Most ethanol for transportation use is blended with gasoline to create a 90 percent gasoline, 10 percent by volume ethanol blend known as E-10 or gasohol.

In 2021, the United States transportation sector consumed an estimated 1,114.3 trillion Btu of ethanol (96 percent of total), and as a result, produced approximately 76.3 MMT CO_2 Eq. (76,279 kt) (see Table 3-107 and Table 3-108) of CO_2 emissions. Smaller quantities of ethanol were also used in the industrial and commercial sectors. Ethanol fuel production and consumption has grown significantly since 1990 due to the favorable economics of blending ethanol into gasoline and federal policies that have encouraged use of renewable fuels.

Table 3-107: CO_2 Emissions from Ethanol Consumption (MMT CO_2 Eq.)

End-Use Sector	1990	2005	2017	2018	2019	2020	2021
Transportation[a]	4.1	21.6	77.7	78.6	78.7	68.1	76.3
Industrial	0.1	1.2	1.9	1.4	1.6	1.6	1.2
Commercial	0.1	0.2	2.5	1.9	2.2	2.2	1.6
Total	**4.2**	**22.9**	**82.1**	**81.9**	**82.6**	**71.8**	**79.1**

[a] See Annex 3.2, Table A-76 for additional information on transportation consumption of these fuels.
Note: Totals may not sum due to independent rounding.

Table 3-108: CO_2 Emissions from Ethanol Consumption (kt)

End-Use Sector	1990	2005	2017	2018	2019	2020	2021
Transportation[a]	4,059	21,616	77,671	78,603	78,739	68,085	76,279
Industrial	105	1,176	1,868	1,404	1,610	1,582	1,171
Commercial	63	151	2,550	1,910	2,229	2,182	1,615
Total	**4,227**	**22,943**	**82,088**	**81,917**	**82,578**	**71,848**	**79,064**

[a] See Annex 3.2, Table A-76 for additional information on transportation consumption of these fuels.
Note: Totals may not sum due to independent rounding.

The transportation sector is assumed to be responsible for all of the biodiesel consumption in the United States (EIA 2022a). Biodiesel is currently produced primarily from soybean oil, but it can be produced from a variety of biomass feedstocks including waste oils, fats, and greases. Biodiesel for transportation use appears in low-level blends (less than 5 percent) with diesel fuel, high-level blends (between 6 and 20 percent) with diesel fuel, and 100 percent biodiesel (EIA 2022b).

In 2021, the United States consumed an estimated 218.2 trillion Btu of biodiesel, and as a result, produced approximately 16.1 MMT CO_2 Eq. (16,112 kt) (see Table 3-109 and Table 3-110) of CO_2 emissions. Biodiesel production and consumption has grown significantly since 2001 due to the favorable economics of blending biodiesel into diesel and federal policies that have encouraged use of renewable fuels (EIA 2022b). There was no measured biodiesel consumption prior to 2001 EIA (2022a).

Table 3-109: CO_2 Emissions from Biodiesel Consumption (MMT CO_2 Eq.)

End-Use Sector	1990		2005		2017	2018	2019	2020	2021
Transportation[a]	NO		0.9		18.7	17.9	17.1	17.7	16.1

NO (Not Occurring)
[a] See Annex 3.2, Table A-76 for additional information on transportation consumption of these fuels.

Table 3-110: CO_2 Emissions from Biodiesel Consumption (kt)

End-Use Sector	1990		2005		2017	2018	2019	2020	2021
Transportation[a]	NO		856		18,705	17,936	17,080	17,678	16,112

NO (Not Occurring)
[a] See Annex 3.2, Table A-76 for additional information on transportation consumption of these fuels.

Methodology and Time-Series Consistency

Woody biomass emissions were estimated by applying two gross heat contents from EIA (Lindstrom 2006) to U.S. consumption data (EIA 2022a) (see Table 3-112), provided in energy units for the industrial, residential, commercial, and electric power sectors. One heat content (16.95 MMBtu/MT wood and wood waste) was applied to the industrial sector's consumption, while the other heat content (15.43 MMBtu/MT wood and wood waste) was applied to the consumption data for the other sectors. An EIA emission factor of 0.434 MT C/MT wood (Lindstrom 2006) was then applied to the resulting quantities of woody biomass to obtain CO_2 emission estimates. The woody biomass is assumed to contain black liquor and other wood wastes, have a moisture content of 12 percent, and undergo complete combustion to be converted into CO_2.

Data for total waste incinerated, excluding tires, from 1990 to 2021 was derived following the methodology described in Section 3.3. Biogenic CO_2 emissions associated with MSW combustion were obtained from EPA's GHGRP FLIGHT data for MSW combustion sources (EPA 2022b). Dividing biogenic CO_2 emissions from GHGRP FLIGHT data for MSW combustors by estimated MSW tonnage combusted yielded an annual biogenic CO_2 emission factor. This approach follows the same approach used to develop the fossil CO_2 emissions from MSW combustion as discussed in Section 3.3. As this data was only available following 2011, all years prior use an average of the emission factors from 2011 through 2015.

Biogenic CO_2 emissions from MSW combustion were calculated by multiplying the annual tonnage estimates, excluding tires, by the calculated emissions factor. Calculated biogenic CO_2 emission factors are shown in Table 3-111.

Table 3-111: Calculated Biogenic CO_2 Content per Ton Waste (kg CO_2/Short Ton Combusted)

	1990		2005		2017	2018	2019	2020	2021
CO_2 Emission Factors	556		556		564	553	558	566	550

The amount of ethanol allocated across the transportation, industrial, and commercial sectors was based on the sector allocations of ethanol-blended motor gasoline. The sector allocations of ethanol-blended motor gasoline were determined using a bottom-up analysis conducted by EPA, as described in the Methodology section of Fossil Fuel Combustion. Total U.S. ethanol consumption from EIA (2022a) was allocated to individual sectors using the same sector allocations as ethanol-blended motor gasoline. The emissions from ethanol consumption were calculated by applying an emission factor of 18.67 MMT C/Qbtu (EPA 2010) to adjusted ethanol consumption estimates (see Table 3-113). The emissions from biodiesel consumption were calculated by applying an emission factor of 20.1 MMT C/Qbtu (EPA 2010) to U.S. biodiesel consumption estimates that were provided in energy units (EIA 2022a) (see Table 3-114).[94]

Table 3-112: Woody Biomass Consumption by Sector (Trillion Btu)

End-Use Sector	1990		2005		2017	2018	2019	2020	2021
Industrial	1,441.9		1,451.7		1,442.3	1,432.0	1,407.0	1,356.2	1,342.0
Residential	580.0		430.0		429.6	524.9	545.5	440.8	463.8
Commercial	65.7		70.0		83.7	84.1	84.3	83.0	82.7
Electric Power	128.5		185.0		229.3	221.1	200.5	185.4	199.0
Total	**2,216.2**		**2,136.7**		**2,185.0**	**2,262.0**	**2,237.3**	**2,065.3**	**2,087.5**

Note: Totals may not sum due to independent rounding.

Table 3-113: Ethanol Consumption by Sector (Trillion Btu)

End-Use Sector	1990		2005		2017	2018	2019	2020	2021
Transportation	59.3		315.8		1,134.6	1,148.2	1,150.2	994.6	1,114.3
Industrial	1.5		17.2		27.3	20.5	23.5	23.1	17.1
Commercial	0.9		2.2		37.2	27.9	32.6	31.9	23.6
Total	**61.7**		**335.1**		**1,199.1**	**1,196.6**	**1,206.3**	**1,049.5**	**1,155.0**

Note: Totals may not sum due to independent rounding.

Table 3-114: Biodiesel Consumption by Sector (Trillion Btu)

End-Use Sector	1990		2005		2017	2018	2019	2020	2021
Transportation	NO		11.6		253.3	242.9	231.3	239.4	218.2

NO (Not Occurring)

Methodological recalculations were applied to the entire time series to ensure time-series consistency from 1990 through 2021.

Uncertainty

It is assumed that the combustion efficiency for biomass is 100 percent, which is believed to be an overestimate of the efficiency of biomass combustion processes in the United States. Decreasing the combustion efficiency would decrease emission estimates for CO_2. Additionally, the heat content applied to the consumption of woody biomass in the residential, commercial, and electric power sectors is unlikely to be a completely accurate representation of the heat content for all the different types of woody biomass consumed within these sectors. Emission estimates from ethanol and biodiesel production are more certain than estimates from woody biomass consumption due to better activity data collection methods and uniform combustion techniques.

[94] CO_2 emissions from biodiesel do not include emissions associated with the C in the fuel that is from the methanol used in the process. Emissions from methanol use and combustion are assumed to be accounted for under Non-Energy Use of Fuels. See Annex 2.3 – Methodology for Estimating Carbon Emitted from Non-Energy Uses of Fossil Fuels.

Recalculations Discussion

The CO_2 emissions associated with the biogenic components of MSW were added to this year's report. The emissions were calculated based on the same approach used to develop fossil CO_2 emissions from the fossil components of MSW as described in Section 3.3.

Planned Improvements

Future research will investigate the availability of data on woody biomass heat contents and carbon emission factors to see if there are newer, improved data sources available for these factors.

Currently, emission estimates from biomass and biomass-based fuels included in this Inventory are limited to woody biomass, biogenic components of MSW, ethanol, and biodiesel. Additional forms of biomass-based fuel consumption include biogas, and other renewable diesel fuels. EPA will investigate additional forms of biomass-based fuel consumption, research the availability of relevant emissions factors, and integrate these into the Inventory as feasible. EPA will examine EIA data on biogas and other renewable diesel fuels to see if these fuel types can be included in future Inventories. EIA (2022a) natural gas data already deducts biogas used in the natural gas supply, so no adjustments are needed to the natural gas fuel consumption data to account for biogas. Distillate fuel statistics are adjusted in this Inventory to remove other renewable diesel fuels as well as biodiesel. Additionally, options for including "Other Renewable Fuels," as defined by EIA, will be evaluated.

The availability of facility-level combustion emissions through EPA's GHGRP will be examined to help better characterize the industrial sector's energy consumption in the United States and further classify woody biomass consumption by business establishments according to industrial economic activity type. Most methodologies used in EPA's GHGRP are consistent with IPCC, although for EPA's GHGRP, facilities collect detailed information specific to their operations according to detailed measurement standards, which may differ with the more aggregated data collected for the Inventory to estimate total, national U.S. emissions. In addition, and unlike the reporting requirements for this chapter under the UNFCCC reporting guidelines, some facility-level fuel combustion emissions reported under EPA's GHGRP may also include industrial process emissions.[95]

In line with UNFCCC reporting guidelines, fuel combustion emissions are included in this chapter, while process emissions are included in the Industrial Processes and Product Use chapter of this report. In examining data from EPA's GHGRP that would be useful to improve the emission estimates for the CO_2 from biomass combustion category, particular attention will also be made to ensure time-series consistency, as the facility-level reporting data from EPA's GHGRP are not available for all inventory years as reported in this Inventory. Additionally, analyses will focus on aligning reported facility-level fuel types and IPCC fuel types per the national energy statistics, ensuring CO_2 emissions from biomass are separated in the facility-level reported data, and maintaining consistency with national energy statistics provided by EIA. In implementing improvements and integration of data from EPA's GHGRP, the latest guidance from the IPCC on the use of facility-level data in national inventories will be relied upon.[96]

[95] See https://unfccc.int/resource/docs/2013/cop19/eng/10a03.pdf#page=2.

[96] See http://www.ipcc-nggip.iges.or.jp/public/tb/TFI_Technical_Bulletin_1.pdf.

3.11 Energy Sources of Precursor Greenhouse Gases – TO BE UPDATED FOR FINAL INVENTORY REPORT

In addition to the main greenhouse gases addressed above, energy-related activities are also sources of greenhouse gas precursors. The reporting requirements of the UNFCCC[97] request that information be provided on precursor emissions, which include carbon monoxide (CO), nitrogen oxides (NO_x), non-methane volatile organic compounds (NMVOCs), and sulfur dioxide (SO_2). These gases are not direct greenhouse gases, but indirectly impact Earth's radiative balance by altering the concentrations of greenhouse gases (e.g., tropospheric ozone) and atmospheric aerosol (e.g., particulate sulfate). Total emissions of NO_x, CO, NMVOCs, and SO_2 from energy-related activities from 1990 to 2020 are reported in Table 3-115.

Table 3-115: NO_x, CO, NMVOC, and SO_2 Emissions from Energy-Related Activities (kt)

Gas/Activity	1990	2005	2017	2018	2019	2020	2021
NO_x	**21,106**	**16,602**	**7,883**	**7,318**	**6,792**	**6,334**	**6,039**
Fossil Fuel Combustion	20,885	16,153	7,246	6,622	6,225	5,768	5,473
Transportation	10,862	10,295	4,519	3,903	3,790	3,502	3,214
Industrial	2,559	1,515	859	898	864	864	864
Electric Power Sector	6,045	3,434	1,049	1,025	886	717	710
Commercial	671	490	537	512	402	402	402
Residential	749	418	283	283	284	284	284
Petroleum and Natural Gas Systems	137	301	530	586	465	465	465
Incineration of Waste	82	128	71	71	71	71	71
Other Energy	2	20	35	39	31	31	31
International Bunker Fuels[a]	1,953	1,699	1,475	1,456	1,290	977	965
CO	**125,640**	**64,985**	**33,401**	**31,455**	**30,959**	**30,177**	**29,394**
Fossil Fuel Combustion	124,360	63,263	31,634	29,639	29,176	28,393	27,611
Transportation	119,360	58,615	27,942	25,957	25,580	24,798	24,015
Residential	3,668	2,856	2,291	2,286	2,286	2,286	2,286
Industrial	797	1,045	736	758	753	753	753
Electric Power Sector	329	582	532	505	424	424	424
Commercial	205	166	133	133	133	133	133
Petroleum and Natural Gas Systems	299	294	546	592	561	561	561
Incineration of Waste	978	1,403	1,175	1,175	1,175	1,175	1,175
Other Energy	3	24	46	49	47	47	47
International Bunker Fuels[a]	102	131	153	158	154	83	83
NMVOCs	**12,612**	**7,345**	**5,664**	**5,395**	**5,168**	**5,067**	**4,966**
Fossil Fuel Combustion	11,836	6,594	3,293	2,686	2,635	2,534	2,433
Transportation	10,932	5,724	2,728	2,114	2,064	1,963	1,862
Residential	686	518	319	319	319	319	319
Commercial	10	188	116	116	116	116	116
Industrial	165	120	101	106	107	107	107
Electric Power Sector	43	44	29	30	28	28	28
Petroleum and Natural Gas Systems	552	497	2,205	2,534	2,362	2,362	2,362
Incineration of Waste	222	241	109	109	109	109	109
Other Energy	2	13	57	66	62	62	62
International Bunker Fuels[a]	57	54	50	50	46	32	31

[97] See http://unfccc.int/resource/docs/2013/cop19/eng/10a03.pdf.

SO_2	**19,628**	**12,364**	**1,793**	**1,724**	**1,405**	**1,222**	**1,313**
Fossil Fuel Combustion	19,200	12,159	1,686	1,606	1,299	1,116	1,206
Electric Power Sector	14,433	9,439	1,257	1,193	921	739	831
Industrial	3,221	1,574	342	330	301	301	301
Transportation	793	619	48	45	40	38	37
Commercial	589	370	28	26	26	26	26
Residential	165	158	12	11	11	11	11
Petroleum and Natural Gas Systems	387	174	83	93	79	79	79
Incineration of Waste	38	25	22	22	24	24	24
Other Energy	3	5	2	3	2	2	2
International Bunker Fuels[a]	*NA*	*NA*	*NA*	*NA*	*NA*	*NA*	*NA*

NA (Not Applicable)
[a] These values are presented for informational purposes only and are not included in totals.
Note: Totals may not sum due to independent rounding.

Methodology and Time-Series Consistency

Emission estimates for 1990 through 2020 were obtained from data published on the National Emissions Inventory (NEI) Air Pollutant Emissions Trends Data website (EPA 2021a). For Table 3-117, NEI reported emissions of CO, NO_x, NMVOCs, and SO_2 are recategorized from NEI Tier 1/Tier 2 source categories to those more closely aligned with IPCC categories, based on EPA (2022).[98] NEI Tier 1 emission categories related to the energy sector categories in this report include: fuel combustion for electric utilities, industrial, and other; petroleum and related industries; highway vehicles; off-highway; and waste disposal and recycling (incineration, open burning). As described in detail in the NEI Technical Support Documentation (TSD) (EPA 2021b), NEI emissions are estimated through a combination of emissions data submitted directly to the EPA by state, local, and tribal air agencies, as well as additional information added by the Agency from EPA emissions programs, such as the emission trading program, Toxics Release Inventory (TRI), and data collected during rule development or compliance testing.

Methodological approaches were applied to the entire time series to ensure time-series consistency from 1990 through 2020, which are described in detail in the NEI's TSD and on EPA's Air Pollutant Emission Trends website (EPA ; EPA 2021). Updates to historical activity data are documented in NEI's TSD (EPA 2021). No quantitative estimates of uncertainty were calculated for this source category.

[98] The NEI estimates and reports emissions from six criteria air pollutants (CAPs) and 187 hazardous air pollutants (HAPs) in support of National Ambient Air Quality Standards. Reported NEI emission estimates are grouped into 60 sectors and 15 Tier 1 source categories, which broadly cover similar source categories to those presented in this chapter. For this report, EPA has mapped and regrouped emissions of greenhouse gas precursors (CO, NO_x, SO_2, and NMVOCs) from NEI Tier 1/Tier 2 categories to better align with NIR source categories, and to ensure consistency and completeness to the extent possible. See Annex 6.6 for more information on this mapping.