

APPENDIX A

LITERATURE SEARCH METHODS

A-1. Literature Search Methods

The literature review serves as the basis for this analysis. For the LCA portion of the review, the search queries (Table A-1) returned 3,159 peer-reviewed articles. Additional keyword searches (see Table A-5) were used to provide flexible searching for a variety of terms, as well as their frequency, in the titles and abstracts of the query results. We reviewed the titles and abstracts of all of the records in order to screen for relevance, selecting a subset for careful screening, and a final set of 247 articles for careful reading.

In addition to searching the peer-reviewed literature, we performed searches for non-peer-reviewed literature: reports and articles from governments, -governmental organization reports, and trade organizations. These searches were based on custom internet searches (e.g., restricting domains to “.edu” or “.gov”), finding non-peer-reviewed literature in the citations of the articles selected for review, and professional experience (i.e., contributors to this report drew upon their knowledge of the field to suggest additional sources). In all, 25 additional records were identified for careful reading.

Combining the results from the searches of peer-reviewed and non-peer-reviewed literature, all 272 records were reviewed. Twenty were excluded as being out of scope, leaving 252 articles and reports for inclusion in the study. All 252 references were screened for relevant content, which was used to provide pathway descriptions, discussions of drivers, and other qualitative text for this report. Ninety-two of these references provided data that could potentially be used for quantitative comparison; for these, the full text was reviewed, and documentation was developed describing their scope, results, and conclusions. Of these, 75 provided data based on LCA approaches. Of those, 47 provided useable data; data from 23 studies was used in both the inter- and intra-study analyses; 10 were used only in the interstudy, and 14 were used only in the intra-study. Table A-2 shows the disaggregation of these studies and data points by management pathway and environmental impact (note that individual studies can, and often do, address multiple pathways or impacts; therefore, the numbers in these tables cannot be summed). Forty-seven studies provide the ranking data summarized in Section 3.5.

Supplemental searches were performed on qualities of variety of soil amendments made by wasted food pathways. These searches (Table A-3) revealed 146 total results, of which 101 were unique. Screening resulted in 19 relevant articles for careful review.

A supplemental literature review was performed to the circularity assessment (Chapter 4). The search queries (Table A-4) returned 67 unique peer-reviewed articles. We reviewed the titles and abstracts of all the records to screen for relevance, selecting a subset for careful screening, and a final set of 20 articles for careful reading. Citations from within these articles were also scanned for relevance and reviewed if applicable. As with the initial literature review, we also performed searches for non-peer-reviewed literature, finding 6 additional references for careful reading.

A-1-1 Literature Queries and Keyword Searches

TABLE A-1. SUMMARY OF LCA LITERATURE SEARCH QUERIES, INCLUDING COUNTS OF UNIQUE RECORDS RETURNED (TOTAL UNIQUE = 3159)

| Name | Source | Date | Number Unique | Total Results | Description | Search Terms |
|-----------|----------------|-----------|---------------|---------------|--|---|
| Q1 | Web of Science | 3/31/2021 | 264 | 264 | Initial search: "food waste" as topic; pathway words in title, and LCA/environment words in title | TS="Food Waste" AND TI=(landfill OR donation OR upcycling OR "animal feed" OR unharvested OR compost* OR "anaerobic digestion" OR "aerobic digestion" OR co-digestion OR "land application" OR "wastewater treatment" OR incineration OR "controlled combustion" OR "waste-to-energy") AND TI=(environment* OR "soil health" OR LCA OR "life cycle assessment" OR "life cycle analysis", OR "water quality" OR eutrophication OR energy OR toxicity OR "greenhouse gas" OR "air emissions" OR "air pollution" OR "land use" OR "land occupation" OR "water use" OR "water consumption" OR "criteria air") |
| Q2 | Web of Science | 4/7/2021 | 393 | 503 | Basic search revised: require some form of food waste in title, require something environmental in abstract, require some pathway or indicator in abstract | TI=("food waste" OR "food loss" OR "wasted food") AND AB=(environment* OR LCA OR "life cycle" or "lifecycle" or "life-cycle") AND AB=(landfill OR donation OR upcycling OR "animal feed" OR unharvested OR compost* OR "anaerobic digestion" OR "aerobic digestion" OR co-digestion OR "land application" OR "wastewater treatment" OR incineration OR "controlled combustion" OR "waste-to-energy" OR "soil health" OR "water quality" OR eutrophication OR energy OR toxicity OR "greenhouse gas" OR "air emissions" OR "air pollution" OR "land use" OR "land occupation" OR "water use" OR "water consumption" OR "criteria air") |

| Name | Source | Date | Number Unique | Total Results | Description | Search Terms |
|------|----------------|----------|---------------|---------------|--|--|
| Q3 | Web of Science | 4/7/2021 | 346 | 891 | initial search, with improved TI (with more stemming) improved environmental words (sustain, etc.) improved general (added pollution, recycling, etc.) | <p>TI=("food waste*" OR "food loss*" OR "wasted food" OR "lost food") AND</p> <p>AB=(environment* OR LCA OR "life cycle" or "lifecycle" or "life-cycle" or resource* or sustain*) AND</p> <p>AB=("waste prevention" OR "waste management" OR pollut* OR recycl* OR landfill OR donation OR upcycling OR "animal feed" OR unharvested OR compost* OR "anaerobic digestion" OR "aerobic digestion" OR co-digestion OR "land application" OR "wastewater treatment" OR incineration OR "controlled combustion" OR "waste-to-energy" OR "soil health" OR biodivers* OR energy OR toxicity OR "land use" OR "land occupation" OR "water use" OR "water consumption" OR "water bod*" OR river OR lake OR "water quality" OR eutrophication OR "greenhouse gas" OR "air emissions" OR "air pollution" OR atmospher* OR "criteria air" OR climate)</p> |
| Q4 | Web of Science | 4/8/2021 | not used | 1659 | Basic search revised to a) exclude anaerobic digestion and compost (well-represented in results), and b) provide more terms that are higher specificity, c) split the impacts from the pathways, and d) use an OR for the two abstract searches. | <p>TI=("food waste*" OR "food loss*" OR "wasted food" OR "lost food" OR "organic waste" OR "biowaste" OR "organic fraction" OR "bio-waste" OR "food-based") AND</p> <p>(</p> <p>AB=(landfill OR donation OR "animal feed" OR "livestock feed" OR unharvested OR "soil incorporation" OR "land application" OR incineration OR combustion OR "waste-to-energy" OR "waste to energy" OR upcycling OR valorization OR valorisation OR carbonization OR carbonisation OR bioconversion OR pyrolysis OR gasification OR ferment)</p> <p>OR</p> <p>AB= ("soil health" OR biodivers* OR toxicity OR "land use" OR "land occupation" OR "water use" OR "water consumption" OR "water bod*" OR river OR lake OR "water quality" OR eutrophication OR "greenhouse gas" OR "air emissions" OR "air pollution" OR "criteria air" OR "climate change" OR "global warming")</p> |

| Name | Source | Date | Number Unique | Total Results | Description | Search Terms |
|------|----------------|-----------|---------------|---------------|--|---|
| Q5 | Web of Science | 4/8/2021 | 403 | 932 | Q4, but with Title more restrictive | TI=("food waste*" OR "food loss*" OR "wasted food" OR "lost food") AND (AB=(landfill OR donation OR "animal feed" OR "livestock feed" OR unharvested OR "soil incorporation" OR "land application" OR incineration OR combustion OR "waste-to-energy" OR "waste to energy" OR upcycling OR valorization OR valorisation OR carbonization OR carbonisation OR bioconversion OR pyrolysis OR gasification OR ferment) OR AB= ("soil health" OR biodivers* OR toxicity OR "land use" OR "land occupation" OR "water use" OR "water consumption" OR "water bod*" OR river OR lake OR "water quality" OR eutrophication OR "greenhouse gas" OR "air emissions" OR "air pollution" OR "criteria air" OR "climate change" OR "global warming") |
| PM4 | PubMed | 3/24/2021 | 3 | 26 | Initial pubmed search; combination of searches 1,2,3 (see terms) | Search 1: landfill[Title] OR donation[Title] OR upcycling[Title] OR "animal feed"[Title] OR unharvested[Title] OR compost[Title] OR "anaerobic digestion"[Title] OR "aerobic digestion"[Title] OR co-digestion OR "land application"[Title] OR "wastewater treatment"[Title] OR incineration[Title] OR "controlled combustion"[Title] OR "waste-to-energy"[Title] Search 2: environment[Title] OR "soil health"[Title] OR LCA[Title] OR "life cycle assessment"[Title] OR "life cycle analysis"[Title], OR "water quality"[Title] OR eutrophication[Title] OR energy[Title] OR toxicity[Title] OR "greenhouse gas"[Title] OR "air emissions"[Title] OR "air pollution"[Title] OR "land use"[Title] OR "land occupation"[Title] OR "water use"[Title] OR "water consumption"[Title] OR "criteria air"[Title] Search 3: "food waste"[Title] |
| PM5 | PubMed | 3/24/2021 | 33 | 59 | combination: searches 3 and 2 | (see PM4) |
| PM6 | PubMed | 3/24/2021 | 349 | 383 | combination: searches 3 and 1 | (see PM4) |

| Name | Source | Date | Number Unique | Total Results | Description | Search Terms |
|-------|----------------|-----------|---------------|---------------|---|--|
| PMQ2 | PubMed | 4/7/2021 | 23 | 189 | Basic search revised: require some form of food waste in title, require something environmental in abstract, require some pathway or indicator in abstract. Combination of searches 1,2,3 in "Search Terms" | <p>#3 Search: landfill[Title/Abstract] OR donation[Title/Abstract] OR upcycling[Title/Abstract] OR "animal feed"[Title/Abstract] OR unharvested[Title/Abstract] OR compost[Title/Abstract] OR "anaerobic digestion"[Title/Abstract] OR "aerobic digestion"[Title/Abstract] OR co-digestion OR "land application"[Title/Abstract] OR "wastewater treatment"[Title/Abstract] OR incineration[Title/Abstract] OR "controlled combustion"[Title/Abstract] OR "waste-to-energy"[Title/Abstract] OR "soil health"[Title/Abstract] OR "water quality"[Title/Abstract] OR eutrophication[Title/Abstract] OR energy[Title/Abstract] OR toxicity[Title/Abstract] OR "greenhouse gas"[Title/Abstract] OR "air emissions"[Title/Abstract] OR "air pollution"[Title/Abstract] OR "land use"[Title/Abstract] OR "land occupation"[Title/Abstract] OR "water use"[Title/Abstract] OR "water consumption"[Title/Abstract] OR "criteria air"[Title/Abstract]</p> <p>#1 Search: "food waste"[Title] OR "food loss"[Title] OR "wasted food"[Title]</p> <p>#2 Search: environment*[Title/Abstract] OR LCA[Title/Abstract] OR "life cycle"[Title/Abstract] OR lifecycle[Title/Abstract]</p> |
| E1 | Ebsco | 3/24/2021 | 11 | 62 | Basic search, with "food waste" as title, and pathway words in title, and LCA/environment words in title. Note quotes did not function well; application was a hit for 'landfill application' | <p>"Food Waste" AND (landfill OR donation OR upcycling OR "animal feed" OR unharvested OR compost* OR "anaerobic digestion" OR "aerobic digestion" OR co-digestion OR "land application" OR "wastewater treatment" OR incineration OR "controlled combustion" OR "waste-to-energy") AND (environment* OR "soil health" OR LCA OR "life cycle assessment" OR "life cycle analysis", OR "water quality" OR eutrophication OR energy OR toxicity OR "greenhouse gas" OR "air emissions" OR "air pollution" OR "land use" OR "land occupation" OR "water use" OR "water consumption" OR "criteria air")</p> |
| Q1_F1 | Web of Science | 3/29/2021 | 207 | 231 | forward searches for 5 articles from Q1 | |
| Q1_B1 | Web of Science | 3/29/2021 | 469 | 580 | backward searches for 5 articles from Q1 | |

| Name | Source | Date | Number Unique | Total Results | Description | Search Terms |
|---------|----------------|-----------|---------------|---------------|------------------------|------------------------|
| Q1_2022 | Web of Science | 9/25/2022 | 76 | 93 | Q1 + years = 2021-2023 | ... AND PY=(2021-2023) |
| Q2_2022 | Web of Science | 9/25/2022 | 219 | 263 | Q2 + years = 2021-2023 | ... AND PY=(2021-2023) |
| Q3_2022 | Web of Science | | 195 | 492 | Q3 + years = 2021-2023 | ... AND PY=(2021-2023) |
| Q5_2022 | Web of Science | | 168 | 461 | Q5 + years = 2021-2023 | ... AND PY=(2021-2023) |

**TABLE A-2. NUMBERS OF STUDIES AND DATA POINTS,
DISAGGREGATED BY MANAGEMENT PATHWAY AND ENVIRONMENTAL IMPACTS.**

| Management Pathway | Number of Studies | Number of Data Points | Environmental Impact | Number of Studies | Number of Data Points |
|--------------------------------------|-------------------|-----------------------|----------------------|-------------------|-----------------------|
| Source Reduction | 4 | 17 | Acidification | 18 | 57 |
| Donation | 5 | 17 | Energy Demand | 9 | 31 |
| Anaerobic Digestion | 23 | 144 | Eutrophication | 20 | 92 |
| Animal Feed | 5 | 21 | GWP | 35 | 168 |
| Compost | 18 | 68 | Land Occupation | 6 | 26 |
| Controlled Combustion / Incineration | 19 | 72 | Water Consumption | 9 | 26 |
| Land application | 1 | 5 | | | |
| Landfill | 14 | 48 | | | |
| Sewer / Wastewater Treatment | 3 | 7 | | | |
| Unharvested / Plowed In | 1 | 1 | | | |
| Upcycling | - | - | | | |

TABLE A-3. SUMMARY OF WASTED FOOD-RELATED SOIL AMENDMENT SEARCH QUERIES, INCLUDING COUNTS OF UNIQUE RECORDS RETURNED (TOTAL UNIQUE = 101)

| Source | Date | Number Unique | Total Results | Description | Search Terms |
|----------------|---------|---------------|---------------|--|---|
| Web of Science | 2/10/23 | 35 | 35 | Initial search on carbon sequestration by digestate | ((PY=(2010-2023)) AND AB=("carbon sequestration" OR "sequester carbon")) AND AB="digestate" |
| Science Direct | 2/10/23 | 2 | 2 | carbon sequestration by digestate | Title, abstract or key word: ("carbon sequestration" OR "sequester carbon") AND "digestate" Publication Year: 2010-2023 |
| PubMed | 2/10/23 | 1 | 11 | carbon sequestration by digestate | Title/abstract: ("carbon sequestration" OR "sequester carbon") AND "digestate" Publication Date: 1/1/2010 to present |
| Web of Science | 2/10/23 | 13 | 13 | Initial search on digestate and soil health | ((PY=(2010-2023)) AND AB=("soil health")) AND AB=(digestate) |
| Science Direct | 2/10/23 | 0 | 5 | digestate and soil health | Title, abstract or key word: "soil health" AND "digestate" Publication Year: 2010-2023 |
| PubMed | 2/10/23 | 0 | 5 | digestate and soil health | Title/abstract: "soil health" AND "digestate" Publication Date: 1/1/2010 to present |
| Science Direct | 2/15/23 | 12 | 12 | Initial search on comparison of nutrient leaching from soil amendments | Title, abstract, keywords: "nutrient" AND "leach" AND "compost" AND ("digestate" OR "biosolids") |
| PubMed | 2/15/23 | 4 | 6 | comparison of nutrient leaching from soil amendments | leach*[Title/Abstract] AND nutrients[Title/Abstract] AND compost[Title/Abstract] AND (biosolids[Title/Abstract] OR digestate[Title/Abstract]) |
| Web of Science | 2/15/23 | 30 | 37 | comparison of nutrient leaching from soil amendments | AB= leach* AND nutrients AND compost AND (biosolids OR digestate) |
| Science Direct | 2/15/23 | 4 | 20 | Initial search on biosolids* | Title, abstract, keywords: "soil" AND "health" AND "biosolids" AND "compost" |

* Similar searches on Web of Science and PubMed did not reveal any unique and relevant sources

**TABLE A-4. SUMMARY OF CIRCULARITY LITERATURE SEARCH QUERIES,
INCLUDING COUNTS OF UNIQUE RECORDS RETURNED (TOTAL UNIQUE = 67)**

| Source | Date | Number Unique | Total Results | Description | Search Terms |
|----------------|----------|---------------|---------------|--|--|
| Web of Science | 10/14/22 | 51 | 51 | Initial search for circularity literature | PY=(2014-2022) AND TI=circular* AND TI=(food OR organic) AND AB=food and AB=("animal feed" OR "compost" OR "anaerobic digestion" OR "land application" OR "wastewater treatment" OR incineration OR landfill OR "controlled combustion") NOT DT=patent |
| Science Direct | 10/14/22 | 8 | 32 | Initial Science Direct search for circularity literature | Years: 2014 – 2022; Title, abstract, or author-specified keywords: Food AND ("animal feed" OR compost OR "anaerobic digestion" OR "land application" OR "wastewater treatment" OR incineration OR landfill OR "controlled combustion"); and Title: (circular OR circularity) AND (food OR organic) |
| PubMed | 2/10/23 | 8 | 29 | Initial PubMed search for circularity literature | (((((("2014/01/01"[Date - Publication] : "3000"[Date - Publication])) AND ("animal feed"[Title/Abstract] OR "compost"[Title/Abstract] OR "anaerobic digestion"[Title/Abstract] OR "land application"[Title/Abstract] OR "wastewater treatment"[Title/Abstract] OR incineration[Title/Abstract] OR landfill[Title/Abstract] OR "controlled combustion"[Title/Abstract])) AND (circular*[Title])) AND (food[Title] OR organic[Title])) |

TABLE A-5. KEYWORDS USED IN SCREENING OF LCA LITERATURE QUERY RESULTS. STEM SEARCHES ALLOW FOR ADDITION OF LETTERS TO THE SEARCH TERMS, EXACT SEARCHES REQUIRE NO ADDITIONAL CHARACTERS, AND REGEX SEARCHES PROVIDE ADDITIONAL FLEXIBILITY TO DEFINE CONSTRAINTS.

| Full Name | Stem Search | Exact Search | Regex Search |
|---------------------------------|--|--------------|--|
| Food Waste | food waste, food - waste, food loss, wasted food, waste food, lost food, fw, flw | FW, FLW | food waste food-waste food loss wasted food waste food lost food fw flw |
| LCA | life cycle assessment, life cycle analysis, lifecycle assessment, lifecycle analysis, life - cycle assessment, life - cycle analysis | LCA | |
| Food Waste LCA | | | (food waste food - waste food loss wasted food waste food lost food bw bwb bflw b).{0,10}(life cycle assessment lifecycle assessment life - cycle assessment life cycle analysis lifecycle analysis life - cycle analysis bLCA b)(life cycle assessment lifecycle assessment life - cycle assessment life cycle analysis lifecycle analysis life - cycle analysis bLCA b).{0,10}(food waste food - waste food loss wasted food waste food lost food bw bwb bflw b) |
| Environment | environm | | |
| Sustain | sustain | | |
| Resources | resourc | | |
| Recycling | recycl | | |
| Pollut | pollut, contamin | | |
| Waste | waste prevention, waste management | | |
| Circular Economy | circular economy, circularity | | |
| TEA | techno- economic, technoeconomic | TEA | |
| Source Reduction | source reduction | | |
| Unharvest / soil incorp. | harvest, crop residue, stover, not harvest, plow in, plough in | | |
| Upcycling | upcycl, value - added surplus, value added surplus | | |
| Donation | donat | | |
| Animal feed | Animal feed, animal ration, fish feed, livestock feed, livestock ration | | |
| Anaerobic Digestion | anaerobic digest | AD | |
| Gasification | gasificat | | |
| Pyrolysis | pyrolysis | | |

| Full Name | Stem Search | Exact Search | Regex Search |
|---------------------------------|---|--------------|---|
| Hydrothermal Carbonization | | | (hydrothermal){0,5}(carbon) |
| Microbial Fuel Cell | microbial fuel cell | | |
| Other Valorization | valorization, bioplastic, platform chemical, biochar | | |
| Mechanical Biological Treatment | mechanical treatment, biological treatment | | |
| Compost | compost | | |
| Land Application | land application | | |
| Wastewater treat. | wastewater treat | WRRF | |
| Controlled Combustion | combust, inciner | | |
| Landfill | landfill, municipal solid waste | MSW | |
| Down the drain | disposer, kitchen sink, in - sink | | |
| Landfill | | | (organic){0,20}(solid waste landfill MSW) |
| Products | bioplastic, platform chemical | | |
| Waste-energy | waste - energy, waste - to - energy, waste to energy, energy from waste, combustion, incineration | | |
| Biochar | biochar | | |
| Agricultural Wastes | | | (agricultur){0,10}(waste) |
| Criteria air pollut. | air pollut | CAP, HAP | |
| Biodiversity | biodiver | | |
| Climate change | climate change, global warming | GWP | |
| Energy demand | energy demand | | |
| Eutroph | eutroph | | |
| Human health | human health, public health | | |
| Land use | land use, land occupation, arable land | | |
| Soil health | soil health | | |
| Tox | toxic, ecotox, human-tox | | |
| Water use | water use, water consumption, water demand | | |
| Water quality | water quality, water pollution, polluted water | | |
| Water | water bod, river, lake, ocean | | |

| Full Name | Stem Search | Exact Search | Regex Search |
|------------------------|--|--------------|--------------|
| Commercial | commerc, business | | |
| Residential | residen, homeowner | | |
| Production | food production, production of food | | |
| Processing | processing | | |
| China | China | | |
| South Korea | South Korea | | |
| European Union | European Union, Austria, Belgium, Bulgaria, Croatia, Cyprus, Czechia, Denmark, Estonia, Finland, France, Germany, Greece, Hungary, Ireland, Italy, Latvia, Lithuania, Luxembourg, Malta, Netherlands, Poland, Portugal, Romania, Slovakia, Slovenia, Spain, Sweden | EU | |
| Europe (not EU) | Albania, Andorra, Armenia, Azerbaijan, Belarus, Bosnia and Herzegovina, Georgia, Iceland, Kosovo, Liechtenstein, Moldova, Monaco, Montenegro, North Macedonia, Norway, Russia, San Marino, Serbia, Switzerland, Turkey, Ukraine, United Kingdom | UK | |
| United States | United States, America | U.S. | |

APPENDIX B

ENERGY, FERTILIZER, AND CARBON SEQUESTRATION OFFSETS

B-1. Energy

The following table documents energy production for electricity and heat in AD, incineration, landfill and WWT. The analysis showed that all or most facilities (some are vague) apply credits only on net energy production, as they should. This can be done in two ways, as documented in the Energy Recovery Modeling column of Table B-1. All values in the electricity column represent total electricity production, so when a net credit is applied, only a partial credit is documented. All studies take some credit for produced, exported electricity. Some studies take a credit for a portion of the heat that is sold, used for district heating, etc. Across the reviewed studies, incineration has the highest total median energy production (280 kWh/Mg wasted food (WF)); however, values in the table do not reflect onsite use, so net energy recovery cannot be determined from the extracted data. Median energy recovery for the AD pathway is 177 kWh/Mg WF. Fewer data points were identified for landfills and the values range widely.

| | |
|---------------------------------------|---|
| Credit - Gross; Facility - Net | Full credit for all energy production and full impact for all energy consumption are included. Thus, there is a credit/impact only on the net difference. |
| Credit - Net; Facility - Net | Credit/impact based on net energy (production - consumption). |

TABLE B-1. SUMMARY OF ENERGY CREDIT MODELING

| Study | Pathway | Basis Unit (1 Mg) | Electricity Recovery | | Heat Recovery | | Energy Recovery Modeling | Note |
|----------------|----------|-------------------|----------------------|----------------|---------------|----------------|--------------------------------|--|
| | | | kWh | Credit Applied | kWh | Credit Applied | | |
| Parry (2012) | AD | WF | 421 | TRUE | 477 | FALSE | Credit - Gross; Facility - Net | recovered heat was used to operate AD. The rest went unused. |
| Parry (2012) | WWT | WF | 233 | TRUE | 271 | FALSE | Credit - Gross; Facility - Net | |
| Parry 2012 | Landfill | WF | 224 | TRUE | 0 | FALSE | Credit - Gross; Facility - Net | |
| Yoshida (2012) | AD | WF | 191 | Partial | 0 | FALSE | Credit - Gross; Facility - Net | Wet AD |
| Yoshida 2012 | AD | WF | 100 | Partial | Not specified | TRUE | Credit - Gross; Facility - Net | Dry AD |

| Study | Pathway | Basis Unit (1 Mg) | Electricity Recovery | | Heat Recovery | | Energy Recovery Modeling | Note |
|------------------------|--------------|-------------------|----------------------|----------------|---|----------------|--------------------------------|--|
| | | | kWh | Credit Applied | kWh | Credit Applied | | |
| Zhou (2022) | AD | WF | 177 | TRUE | 0 | FALSE | Credit - Gross; Facility - Net | "After purification, the biogas was sent to generating heat and electricity, and around 176.5 kWh/t WF was generated for energy recovery." Possible that some heat is included in recovered energy. Not clear. |
| González et al. (2020) | AD | WF | 133 | Partial | 199 | Partial | Credit - Net; Facility - Net | Values for CHP scenario; Assumes some heat is available for sale. This amount is credited. |
| Tong (2018) | Incineration | WF | 157 | Partial | 0 | FALSE | Credit - Gross; Facility - Net | Could not find values for AD. |
| Lee (2020) | AD | WF | 163 | TRUE | 210 | Not specified | Credit - Gross; Facility - Net | |
| Lee 2020 | Incineration | WF | 280 | TRUE | 601 | Not specified | Credit - Gross; Facility - Net | |
| Pace (2018) | AD | Organic Waste | 121 | Partial | Not specified | FALSE | Credit - Net; Facility - Net | |
| Slorach (2020) | AD | WF | 254 | TRUE | 0 for current scenario; 252 for future scenario | Partial | Credit - Gross; Facility - Net | |
| Mondello (2017) | Landfill | WF | 166 | TRUE | 0 | FALSE | Credit - Gross; Facility - Net | |
| Mondello 2017 | Incineration | WF | 564 | TRUE | 0 | FALSE | Credit - Gross; Facility - Net | |
| Mondello 2017 | AD | WF | 175 | TRUE | 0 | FALSE | Credit - Gross; Facility - Net | |
| Tian (2021) | AD | WF | 385 | TRUE | Not specified | FALSE | Credit - Gross; Facility - Net | centralized AD, scenario D. |
| Slorach et al. (2019b) | Incineration | WF | 174 | TRUE | 0 | FALSE | Credit - Net; Facility - Net | |
| Slorach et al. (2019b) | AD | WF | 254 | TRUE | 476 | FALSE | Credit - Net; Facility - Net | |

| Study | Pathway | Basis Unit (1 Mg) | Electricity Recovery | | Heat Recovery | | Energy Recovery Modeling | Note |
|--------------------------|--------------|-------------------|----------------------|----------------|---------------|----------------|--------------------------------------|------|
| | | | kWh | Credit Applied | kWh | Credit Applied | | |
| Albizzati et al. (2021b) | Incineration | WF | 500 | TRUE | 0 | FALSE | Credit - Gross; Facility - Net | |
| Hoehn (2021) | Incineration | WF | 138 | TRUE | 355 | TRUE | Credit-Gross; Facility-Not specified | |
| Huang et al. (2022) | Incineration | WF | 427 | TRUE | Not specified | FALSE | Credit - Net; Facility - Net | |
| Huang et al. (2022) | Landfill | WF | 33.9 | TRUE | 0 | FALSE | Credit - Net; Facility - Net | |

B-1-1 Fertilizer and CO₂ Sequestration

Looking across all studies (regardless of scope criteria), we identified 6 studies that report the magnitude of the GWP credit associated with carbon sequestration for the compost pathway and 8 studies for the AD pathway. At the same time, we reviewed studies to determine the range of magnitudes associated with avoided fertilizer production (for brevity and context we report them together here). Table B-2 and Table B-3 show the median, min and max credit values for each pathway for fertilizer and carbon sequestration credits. For the compost and AD pathways, respectively, we found 8 and 13 studies that report the magnitude of their modeled fertilizer credit. Table B-4 shows the carbon sequestration and fertilizer credits for individual studies-pathways. All studies that include both pathways attribute more carbon sequestration to compost than to AD. Among these studies, in absolute terms, the difference in magnitude between the two pathways is typically moderate (median difference = -10.3 kg CO₂eq difference/Mg wasted food). In relative terms, this difference amounts to 7% of median impact for the AD pathway and 23% of median impact for the compost pathway. This difference is consequential for the net impact of the compost pathway (i.e., it could switch from positive to negative). One study, Parry (2012), attributes a much greater estimate of carbon sequestration to the compost pathway as compared to AD.

TABLE B-2. GWP OF FERTILIZER AND CARBON SEQUESTRATION CREDITS (KG CO₂EQ/METRIC TON ORGANIC WASTE) FOR AD DIGESTATE

| | Fertilizer Offset (n=13) | Carbon Sequestration (n=8) |
|--------|--------------------------|----------------------------|
| Median | -26 | -22 |
| Min | -110 | -80 |
| Max | -3.2 | -6.0 |

TABLE B-3: GWP OF FERTILIZER AND CARBON SEQUESTRATION CREDITS (KG CO₂EQ/METRIC TON ORGANIC WASTE) FOR COMPOST

| | Fertilizer Offset (n=8) | Carbon Sequestration (n=6) |
|--------|-------------------------|----------------------------|
| Median | -28 | -94 |
| Min | -50 | -224 |
| Max | -2.8 | -10 |

Table 3-4 shows median values for pathway impacts with and without CO₂ sequestration (CO₂ Sequest = TRUE or FALSE) for AD, Compost, Land Application, Landfill, Unharvest, and WRRF. Note that these data overlap with, but do not match the data in Table B-2 and Table B-3, which required that offsets be explicitly reported. Table 3-4 simply requires that CO₂ sequestration be identified as either included or not. Nonetheless, all tables agree that the CO₂ sequestration offset reduces the GWP impact. Across pathways with data for both conditions, impacts with sequestration are higher than results without CO₂ sequestration. For AD and compost (looking across all tables), the shift is on the order of -50 to -200 kg CO₂eq / metric ton wasted food.

**TABLE B-4: MAGNITUDE OF CARBON SEQUESTRATION AND AVOIDED FERTILIZER CREDITS
(ALL RESULTS STANDARDIZED TO 1 METRIC TON OF THE DOCUMENTED BASIS UNIT)**

| Study | Pathway | Basis Unit | Fertilizer Nutrient Offset | Carbon Sequestration |
|---------------------------|---------|-----------------------------------|----------------------------|----------------------------|
| | | | GWP (kg CO ₂ e) | GWP (kg CO ₂ e) |
| Yoshida (2012) | Compost | mixed organic waste | -2.82 | -67.7 |
| Yoshida (2012) | AD | mixed organic waste | -3.21 | -60.9 |
| Yoshida (2012) | AD | mixed organic waste | -9.18 | -61.7 |
| Levis and Barlaz (2011) | Compost | mixed organic waste | -11 | -148 |
| Hodge (2016) | Compost | mixed commercial wasted food (WF) | -14 | -10 |
| Hodge (2016) | AD | mixed commercial WF | -14 | -6 |
| Morelli et al. (2019) | Compost | WF | -16 | -21.7 |
| Morelli et al. (2019) | AD | WF | -6.0 | -7.4 |
| Schott (2016) | AD | WF | -67 | Excluded |
| Morris et al. (2017) | Compost | WF | -50 | -120 |
| Morris et al. (2017) | AD | WF | -20 | -80 |
| Slorach et al. (2019a) | Compost | household WF | Minor | Excluded |
| Parry (2012) | Compost | WF | -48 | -224 |
| Parry (2012) | AD | WF | -53 | -28 |
| Ebner et al. (2015) | AD | industrial WF and manure | -8.2 | -12.8 |
| Tong et al. (2018) | AD | WF | -26 | -15.4 |
| Mondello (2017) | AD | WF | -42 | Excluded |
| Mondello (2017) | Compost | WF | -40 | Excluded |
| Tian et al. (2021) | AD | WF | -49 | Excluded |
| Khosnevisan et al. (2018) | AD | mixed organic waste | -110 | Excluded |

APPENDIX C

DESCRIPTION OF APPROACH TO COMPARE IMPACT METRICS

Consistency in reported impact category units enabled standardization of results across studies for the impact categories acidification, energy demand, eutrophication, GWP, land occupation, and water consumption. Variability in units prevented standardization of results for categories such as ecotoxicity, human toxicity, or smog formation. This section briefly describes the standardization approach.

C-1. Description of Approach to Compare Impact Metrics

Midpoint acidification is typically reported in units of Accumulated Exceedance (AE, i.e. mol H+ eq) or kg SO₂eq. Midpoint eutrophication is typically reported in either kg of nitrogen or phosphorus equivalents, but some methods, such as Impact 2002+ (Jolliet et al. 2003), do not include marine eutrophication, and other methods, such as ReCiPe 2008 (Goedkoop et al. 2013), classify freshwater and marine eutrophication separately. The EPA TRACI method (Bare 2011), however, combines the two types of eutrophication, reporting impacts in terms of kg nitrogen equivalents, based on stoichiometric conversions of algal biomass from the Redfield ratio (Redfield 1934). A broader discussion of acidification and eutrophication units across methods is provided in the LCIA Encyclopedia (Henderson 2015; van Zelm et al. 2015). In this study, we have harmonized acidification and eutrophication impacts to those used in the TRACI method, in order to account for the different units and the freshwater/marine modeling split. Table C-1, below, shows the conversion factors for acidification and eutrophication between a sampling of impact methods and TRACI. The impact methods shown are representative of the variety of units reported; Subsequent updates to some of the methods have been made (with little to no changes to global midpoint values, which are shown and used here), but many of the LCA studies reviewed have used older versions of methods.

Midpoint global warming potential is reported in terms of CO₂eq. However, as IPCC updates factors for GWP of substances, these factors can change in LCIA methods. No attempt was made to adjust for changes in these factors, as the relative magnitude of changes is restricted relative to scope, inventory data and other differences between studies.

Finally, midpoint land and water metrics are typically area and volume based, respectively, so no conversion between impact assessment methods is necessary.

TABLE C-1. SELECTED CHARACTERIZATION FACTORS FOR LCIA METHODS, ALONG WITH CONVERSION TO TRACI 2.1

| Impact Method | Impact Category | Emission | Compartment | CF | Unit | TRACI 2.1 Impact Category | TRACI 2.1 Units | Conversion to TRACI 2.1 |
|--------------------|---------------------------|---------------------------------|-------------|--------|--|---------------------------|-----------------------|-------------------------|
| CML 2001 | Acidification | SO ₂ | air | 1.2 | kg SO ₂ eq | Acidification | kg SO ₂ eq | 0.833 |
| CML 2001 | Eutrophication | N | water | 0.42 | kg PO ₄ ⁽³⁻⁾ eq | Eutrophication | kg Neq | 2.349 |
| CML 2001 | Eutrophication | P | water | 3.06 | kg PO ₄ ⁽³⁻⁾ eq | Eutrophication | kg Neq | 2.382 |
| CML 2001 | Eutrophication | PO ₄ ⁽³⁻⁾ | water | 1 | kg PO ₄ ⁽³⁻⁾ eq | Eutrophication | kg Neq | 2.380 |
| ILCD 2011 Midpoint | Acidification | SO ₂ | air | 1.31 | mol H ⁺ eq | Acidification | kg SO ₂ eq | 0.763 |
| ILCD 2011 Midpoint | Freshwater Eutrophication | P | water | 1 | kg Peq | Eutrophication | kg Neq | 7.290 |
| ILCD 2011 Midpoint | Freshwater Eutrophication | PO ₄ ⁽³⁻⁾ | water | 0.33 | kg Peq | Eutrophication | kg Neq | 7.212 |
| ILCD 2011 Midpoint | Marine Eutrophication | N | water | 1 | kg Neq | Eutrophication | kg Neq | 0.986 |
| Impact 2002+ | Acidification | SO ₂ | air | 1 | kg SO ₂ eq | Acidification | kg SO ₂ eq | 1.000 |
| Impact 2002+ | Eutrophication | N | water | 0 | kg PO ₄ ⁽³⁻⁾ P-lim | Eutrophication | kg Neq | 0.986 |
| Impact 2002+ | Eutrophication | P | water | 3.06 | kg PO ₄ ⁽³⁻⁾ P-lim | Eutrophication | kg Neq | 2.382 |
| Impact 2002+ | Eutrophication | PO ₄ ⁽³⁻⁾ | water | 1 | kg PO ₄ ⁽³⁻⁾ P-lim | Eutrophication | kg Neq | 2.380 |
| ReCiPe 2008 | Acidification | SO ₂ | air | 1 | kg SO ₂ eq | Acidification | kg SO ₂ eq | 1.000 |
| ReCiPe 2008 | Freshwater Eutrophication | P | water | 1 | kg Peq | Eutrophication | kg Neq | 7.290 |
| ReCiPe 2008 | Freshwater Eutrophication | PO ₄ ⁽³⁻⁾ | water | 0.33 | kg Peq | Eutrophication | kg Neq | 7.212 |
| ReCiPe 2008 | Marine Eutrophication | N | water | 1 | kg Neq | Eutrophication | kg Neq | 0.986 |
| TRACI 2.1 | Acidification | SO ₂ | air | 1 | kg SO ₂ eq | Acidification | kg SO ₂ eq | 1.000 |
| TRACI 2.1 | Eutrophication | N | water | 0.9864 | kg Neq | Eutrophication | kg Neq | 1.000 |
| TRACI 2.1 | Eutrophication | P | water | 7.29 | kg Neq | Eutrophication | kg Neq | 1.000 |
| TRACI 2.1 | Eutrophication | PO ₄ ⁽³⁻⁾ | water | 2.38 | kg Neq | Eutrophication | kg Neq | 1.000 |

References and Notes for Table C-1:

- CML 2001: (Guinée et al. 2002)
- ILCD 2011: (Wolf et al. 2012). ILCD uses Accumulated Exceedance (Seppälä et al. 2006; Posch et al. 2008), with units of mol H⁺ eq.
- TRACI 2.1: (Bare 2011)
- ReCiPe 2008: (Goedkoop et al. 2013)
- Impact 2002+: (Jolliet et al. 2003)

APPENDIX D

EXTRACTED DATA VALUES

This section provides a table of data used in calculation of statistics for ranking, and in preparation of Figure 3-3 to Figure 3-8.

TABLE D-1. EXTRACTED DATA VALUES USED FOR INTER-STUDY COMPARISONS

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|-------------|--|---------|-----------------------------------|-----------|-------------------------------|-----------------|
| Acidification | Combust | incineration with electricity recovery | 0.5 | kg SO ₂ eq | Singapore | N.A. | Ahamed 2016 |
| CED | Combust | incineration with electricity recovery | 3530 | MJ | Singapore | N.A. | Ahamed 2016 |
| Eutrophication | Combust | incineration with electricity recovery | 0.197 | kg Neq | Singapore | N.A. | Ahamed 2016 |
| GWP | Combust | incineration with electricity recovery | 235 | kg CO ₂ eq | Singapore | N.A. | Ahamed 2016 |
| Eutrophication | Animal Feed | Wet | 0.671 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Animal Feed | BSF | -0.0385 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Animal Feed | PC | 0.148 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Animal Feed | Wet | 3.43 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Animal Feed | BSF | 0.0372 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|-------------|-------------|--------|-----------------------------------|-----------|-------------------------------|-----------------|
| Eutrophication | Animal Feed | PC | 2.19 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| GWP | Animal Feed | Wet | -76 | kg CO ₂ eq | EU-27+1 | N.A. | Albizzati 2021a |
| GWP | Animal Feed | BSF | 17.1 | kg CO ₂ eq | EU-27+1 | N.A. | Albizzati 2021a |
| GWP | Animal Feed | PC | -210 | kg CO ₂ eq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Compost | Centralized | 0.158 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Compost | Home | 1.09 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Compost | Centralized | 4.88 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Compost | Home | 7.29 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| GWP | Compost | Centralized | 51 | kg CO ₂ eq | EU-27+1 | TRUE | Albizzati 2021a |
| GWP | Compost | Home | 270 | kg CO ₂ eq | EU-27+1 | TRUE | Albizzati 2021a |
| Eutrophication | Donate | Base | -3.26 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| Eutrophication | Donate | Base | -1.31 | kg Neq | EU-27+1 | N.A. | Albizzati 2021a |
| GWP | Donate | Base | -1300 | kg CO ₂ eq | EU-27+1 | N.A. | Albizzati 2021a |
| GWP | AD | Baseline | 123 | kg CO ₂ eq | Europe | FALSE | Albizzati 2021b |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|-------------|------------------------------------|--------|-----------------------------------|-----------|-------------------------------|--|
| GWP | Animal Feed | With avoided products (net result) | -329 | kg CO ₂ eq | Europe | N.A. | Albizzati 2021b |
| GWP | Animal Feed | With avoided products (net result) | -385 | kg CO ₂ eq | Europe | N.A. | Albizzati 2021b |
| GWP | Combust | Baseline | 41.8 | kg CO ₂ eq | Europe | N.A. | Albizzati 2021b |
| GWP | Compost | Baseline | 152 | kg CO ₂ eq | Europe | FALSE | Albizzati 2021b |
| GWP | Landfill | Baseline | 2410 | kg CO ₂ eq | Europe | FALSE | Albizzati 2021b |
| GWP | AD | Baseline | 26.5 | kg CO ₂ eq | Qatar | FALSE | Al-Rumaihi 2020 |
| GWP | Compost | Baseline | 128 | kg CO ₂ eq | Qatar | FALSE | Al-Rumaihi 2020 |
| GWP | AD | total food waste | -194 | kg CO ₂ eq | Sweden | FALSE | Bernstad Saraiva Schott and Andersson 2015 |
| GWP | Combust | total food waste | -82.8 | kg CO ₂ eq | Sweden | N.A. | Bernstad Saraiva Schott and Andersson 2015 |
| CED | Donate | Base | -6510 | MJ | Sweden | N.A. | Eriksson 2017 |
| GWP | Donate | Base | -574 | kg CO ₂ eq | Sweden | N.A. | Eriksson 2017 |
| CED | Upcycle | Base | -8240 | MJ | Sweden | N.A. | Eriksson 2017 |
| GWP | Upcycle | Base | -600 | kg CO ₂ eq | Sweden | N.A. | Eriksson 2017 |
| Acidification | Upcycle | Bread Additive | 0.725 | kg SO ₂ eq | Sweden | N.A. | Eriksson 2021 |
| Acidification | Upcycle | Soup Additive | -2.29 | kg SO ₂ eq | Sweden | N.A. | Eriksson 2021 |
| Eutrophication | Upcycle | Bread Additive | 0.948 | kg Neq | Sweden | N.A. | Eriksson 2021 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|----------|--|-----------|-----------------------------------|---------------|-------------------------------|---------------|
| Eutrophication | Upcycle | Bread Additive | -0.109 | kg Neq | Sweden | N.A. | Eriksson 2021 |
| Eutrophication | Upcycle | Soup Additive | -0.496 | kg Neq | Sweden | N.A. | Eriksson 2021 |
| Eutrophication | Upcycle | Soup Additive | -4.14 | kg Neq | Sweden | N.A. | Eriksson 2021 |
| GWP | Upcycle | Bread Additive | 68 | kg CO ₂ eq | Sweden | N.A. | Eriksson 2021 |
| GWP | Upcycle | Soup Additive | -450 | kg CO ₂ eq | Sweden | N.A. | Eriksson 2021 |
| Water Consumption | Upcycle | Bread Additive | -0.16 | m ³ water | Sweden | N.A. | Eriksson 2021 |
| Water Consumption | Upcycle | Soup Additive | -3 | m ³ water | Sweden | N.A. | Eriksson 2021 |
| GWP | AD | AD-LF -BU ; residuals to landfill; with beneficial units | -179 | kg CO ₂ eq | United States | TRUE | Hodge 2016 |
| GWP | AD | AD-WTE - BUresiduals to waste to energy incineration; with beneficial units | -130 | kg CO ₂ eq | United States | TRUE | Hodge 2016 |
| GWP | Combust | WTE - BU; beneficial units | -106 | kg CO ₂ eq | United States | N.A. | Hodge 2016 |
| GWP | Compost | AC-LF - BU; residuals to landfill; with beneficial units | -78 | kg CO ₂ eq | United States | TRUE | Hodge 2016 |
| GWP | Compost | AC-WTE; BU; residuals to waste to energy incineration; with beneficial units | -25 | kg CO ₂ eq | United States | TRUE | Hodge 2016 |
| GWP | Landfill | LF - BU; energy recovery | -10 | kg CO ₂ eq | United States | TRUE | Hodge 2016 |
| Eutrophication | Combust | Base | 0.0000456 | kg Neq | Spain | N.A. | Hoehn 2021 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|----------|---|----------|-----------------------------------|-----------|-------------------------------|-------------------|
| Eutrophication | Combust | Base | 0.000986 | kg Neq | Spain | N.A. | Hoehn 2021 |
| Eutrophication | Compost | Base | 0.000346 | kg Neq | Spain | N.A. | Hoehn 2021 |
| Eutrophication | Compost | Base | 0.000197 | kg Neq | Spain | N.A. | Hoehn 2021 |
| Eutrophication | Landfill | Base | 0.0219 | kg Neq | Spain | N.A. | Hoehn 2021 |
| Eutrophication | Landfill | Base | 0.0076 | kg Neq | Spain | N.A. | Hoehn 2021 |
| Acidification | Combust | Base | -1 | kg SO ₂ eq | China | N.A. | Huang 2022 |
| Eutrophication | Combust | Base | 0.365 | kg Neq | China | N.A. | Huang 2022 |
| GWP | Combust | Base | 30 | kg CO ₂ eq | China | N.A. | Huang 2022 |
| Water Consumption | Combust | Base | -1.5 | m ³ water | China | N.A. | Huang 2022 |
| Acidification | Landfill | Base | 12 | kg SO ₂ eq | China | N.A. | Huang 2022 |
| Eutrophication | Landfill | Base | 0.219 | kg Neq | China | N.A. | Huang 2022 |
| GWP | Landfill | Base | 590 | kg CO ₂ eq | China | FALSE | Huang 2022 |
| Water Consumption | Landfill | Base | 0.5 | m ³ water | China | N.A. | Huang 2022 |
| CED | AD | Bio-1, Pre-treatment: biopulp AD with CHP | -13000 | MJ | Denmark | N.A. | Khoshnevisan 2018 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|--------|---------|--|--------|-----------------------------------|-----------|-------------------------------|-------------------|
| CED | AD | Bio-2, Pre-treatment: biopulp AD with electricity recovery and nutrient | -101 | MJ | Denmark | N.A. | Khoshnevisan 2018 |
| CED | AD | Bio-3, Pre-treatment: biopulp AD with biogasoline and nutrient recovery | -9640 | MJ | Denmark | N.A. | Khoshnevisan 2018 |
| CED | AD | Sp-1, Pre-treatment: Screw press AD with CHP | -9530 | MJ | Denmark | N.A. | Khoshnevisan 2018 |
| CED | AD | Sp-2, Pre-treatment: Screw press AD with electricity recovery and nutrient | -7470 | MJ | Denmark | N.A. | Khoshnevisan 2018 |
| CED | AD | Sp-3, Pre-treatment: Screw press AD with biogasoline and nutrient recovery | -7180 | MJ | Denmark | N.A. | Khoshnevisan 2018 |
| CED | AD | Ds-1, Pre-treatment: Disc screen AD with CHP | -9320 | MJ | Denmark | N.A. | Khoshnevisan 2018 |
| CED | AD | Ds-2, Pre-treatment: Disc screen AD with electricity recovery and nutrient | -6820 | MJ | Denmark | N.A. | Khoshnevisan 2018 |
| CED | AD | Ds-3, Pre-treatment: Disc screen AD with biogasoline and nutrient recovery | -6470 | MJ | Denmark | N.A. | Khoshnevisan 2018 |
| GWP | AD | Bio-1, Pre-treatment: biopulp AD with CHP | -941 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |
| GWP | AD | Bio-2, Pre-treatment: biopulp AD with electricity recovery and nutrient | -672 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |
| GWP | AD | Bio-3, Pre-treatment: biopulp AD with biogasoline and nutrient recovery | -333 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |
| GWP | AD | Sp-1, Pre-treatment: Screw press AD with CHP | -750 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |
| GWP | AD | Sp-2, Pre-treatment: Screw press AD with electricity recovery and nutrient | -562 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|---------|--|--------|-----------------------------------|---------------|-------------------------------|-------------------|
| GWP | AD | Sp-3, Pre-treatment: Screw press AD with biogasoline and nutrient recovery | -327 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |
| GWP | AD | Ds-1, Pre-treatment: Disc screen AD with CHP | -758 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |
| GWP | AD | Ds-2, Pre-treatment: Disc screen AD with electricity recovery and nutrient | -532 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |
| GWP | AD | Ds-3, Pre-treatment: Disc screen AD with biogasoline and nutrient recovery | -248 | kg CO ₂ eq | Denmark | FALSE | Khoshnevisan 2018 |
| GWP | Compost | Windrow compost | -148 | kg CO ₂ eq | United States | TRUE | Levis 2011 |
| GWP | Compost | ASP compost | -73 | kg CO ₂ eq | United States | TRUE | Levis 2011 |
| GWP | Compost | Gore compost | -102 | kg CO ₂ eq | United States | TRUE | Levis 2011 |
| GWP | Compost | In-vessel compost | -64 | kg CO ₂ eq | United States | TRUE | Levis 2011 |
| Acidification | Compost | S2- Aerated windrow composting | 1.01 | kg SO ₂ eq | Malaysia | N.A. | Lin 2022 |
| Eutrophication | Compost | S2- Aerated windrow composting | -0.122 | kg Neq | Malaysia | N.A. | Lin 2022 |
| GWP | Compost | S2- Aerated windrow composting | 463 | kg CO ₂ eq | Malaysia | FALSE | Lin 2022 |
| Acidification | AD | Anaerobic Digestion | 0.203 | kg SO ₂ eq | Germany | N.A. | Mayer 2020 |
| Acidification | AD | AD + Incineration | 0.0844 | kg SO ₂ eq | Germany | N.A. | Mayer 2020 |
| Eutrophication | AD | Anaerobic Digestion | -1.14 | kg Neq | Germany | N.A. | Mayer 2020 |
| Eutrophication | AD | Anaerobic Digestion | 0.307 | kg Neq | Germany | N.A. | Mayer 2020 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|---------|-----------------------|---------|-----------------------------------|-----------|-------------------------------|---------------|
| Eutrophication | AD | AD + Incineration | -0.712 | kg Neq | Germany | N.A. | Mayer 2020 |
| Eutrophication | AD | AD + Incineration | 0.109 | kg Neq | Germany | N.A. | Mayer 2020 |
| GWP | AD | Anaerobic Digestion | -45.3 | kg CO ₂ eq | Germany | FALSE | Mayer 2020 |
| GWP | AD | AD + Incineration | -0.298 | kg CO ₂ eq | Germany | FALSE | Mayer 2020 |
| Acidification | Combust | Incineration | 0.0427 | kg SO ₂ eq | Germany | N.A. | Mayer 2020 |
| Acidification | Combust | Incineration + drying | 0.0894 | kg SO ₂ eq | Germany | N.A. | Mayer 2020 |
| Eutrophication | Combust | Incineration | -0.547 | kg Neq | Germany | N.A. | Mayer 2020 |
| Eutrophication | Combust | Incineration | 0.153 | kg Neq | Germany | N.A. | Mayer 2020 |
| Eutrophication | Combust | Incineration + drying | -0.618 | kg Neq | Germany | N.A. | Mayer 2020 |
| Eutrophication | Combust | Incineration + drying | 0.165 | kg Neq | Germany | N.A. | Mayer 2020 |
| GWP | Combust | Incineration | -55 | kg CO ₂ eq | Germany | N.A. | Mayer 2020 |
| GWP | Combust | Incineration + drying | -2.97 | kg CO ₂ eq | Germany | N.A. | Mayer 2020 |
| Acidification | AD | With avoided products | -0.23 | kg SO ₂ eq | Italy | N.A. | Mondello 2017 |
| CED | AD | With avoided products | -1660 | MJ | Italy | N.A. | Mondello 2017 |
| Eutrophication | AD | With avoided products | -0.0948 | kg Neq | Italy | N.A. | Mondello 2017 |
| GWP | AD | With avoided products | -299 | kg CO ₂ eq | Italy | FALSE | Mondello 2017 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-----------------|-------------|--------------------------------|--------|-----------------------------------|-------------------------------|-------------------------------|---------------|
| Land Occupation | AD | With avoided products | -0.14 | m ² .yr | Italy | N.A. | Mondello 2017 |
| Acidification | Animal Feed | With avoided products | -1.28 | kg SO ₂ eq | Italy | N.A. | Mondello 2017 |
| Eutrophication | Animal Feed | With avoided products | -0.45 | kg Neq | Italy | N.A. | Mondello 2017 |
| GWP | Animal Feed | With avoided products | -420 | kg CO ₂ eq | Italy | N.A. | Mondello 2017 |
| Acidification | Compost | With avoided products | 0.38 | kg SO ₂ eq | Italy | N.A. | Mondello 2017 |
| Eutrophication | Compost | With avoided products | 0.166 | kg Neq | Italy | N.A. | Mondello 2017 |
| GWP | Compost | With avoided products | 59.3 | kg CO ₂ eq | Italy | FALSE | Mondello 2017 |
| Acidification | AD | Full-capacity base performance | -0.11 | kg SO ₂ eq | United States (Massachusetts) | N.A. | Morelli 2019 |
| Acidification | AD | Full-capacity low performance | 0.21 | kg SO ₂ eq | United States (Massachusetts) | N.A. | Morelli 2019 |
| CED | AD | Full-capacity base performance | -7200 | MJ | United States (Massachusetts) | N.A. | Morelli 2019 |
| CED | AD | Full-capacity low performance | -3200 | MJ | United States (Massachusetts) | N.A. | Morelli 2019 |
| Eutrophication | AD | Full-capacity base performance | 2.37 | kg Neq | United States (Massachusetts) | N.A. | Morelli 2019 |
| Eutrophication | AD | Full-capacity low performance | 3.06 | kg Neq | United States (Massachusetts) | N.A. | Morelli 2019 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|---------|--------------------------------|---------|-----------------------------------|-------------------------------|-------------------------------|--------------|
| GWP | AD | Full-capacity base performance | -140 | kg CO ₂ eq | United States (Massachusetts) | TRUE | Morelli 2019 |
| GWP | AD | Full-capacity low performance | -30 | kg CO ₂ eq | United States (Massachusetts) | TRUE | Morelli 2019 |
| Acidification | Combust | Baseline | 0.081 | kg SO ₂ eq | United States (Massachusetts) | N.A. | Morelli 2019 |
| CED | Combust | Baseline | -960 | MJ | United States (Massachusetts) | N.A. | Morelli 2019 |
| Eutrophication | Combust | Baseline | 0.00612 | kg Neq | United States (Massachusetts) | N.A. | Morelli 2019 |
| GWP | Combust | Baseline | -20 | kg CO ₂ eq | United States (Massachusetts) | N.A. | Morelli 2019 |
| Acidification | Compost | Windrow base performance | 1.2 | kg SO ₂ eq | United States (Massachusetts) | N.A. | Morelli 2019 |
| Acidification | Compost | Windrow improved performance | 0.5 | kg SO ₂ eq | United States (Massachusetts) | N.A. | Morelli 2019 |
| Acidification | Compost | ASP improved performance | 0.49 | kg SO ₂ eq | United States (Massachusetts) | N.A. | Morelli 2019 |
| CED | Compost | Windrow base performance | 290 | MJ | United States (Massachusetts) | N.A. | Morelli 2019 |
| CED | Compost | ASP base performance | 540 | MJ | United States (Massachusetts) | N.A. | Morelli 2019 |
| CED | Compost | Windrow improved performance | 220 | MJ | United States (Massachusetts) | N.A. | Morelli 2019 |
| CED | Compost | ASP improved performance | 390 | MJ | United States (Massachusetts) | N.A. | Morelli 2019 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|----------|------------------------------|---------|-----------------------------------|-------------------------------|-------------------------------|---------------|
| Eutrophication | Compost | Windrow base performance | 0.937 | kg Neq | United States (Massachusetts) | N.A. | Morelli 2019 |
| Eutrophication | Compost | Windrow improved performance | 0.661 | kg Neq | United States (Massachusetts) | N.A. | Morelli 2019 |
| GWP | Compost | Windrow base performance | 100 | kg CO ₂ eq | United States (Massachusetts) | TRUE | Morelli 2019 |
| GWP | Compost | ASP base performance | 70 | kg CO ₂ eq | United States (Massachusetts) | TRUE | Morelli 2019 |
| GWP | Compost | Windrow improved performance | -10 | kg CO ₂ eq | United States (Massachusetts) | TRUE | Morelli 2019 |
| Acidification | Landfill | Baseline | 0.14 | kg SO ₂ eq | United States (Massachusetts) | N.A. | Morelli 2019 |
| CED | Landfill | Baseline | -200 | MJ | United States (Massachusetts) | N.A. | Morelli 2019 |
| Eutrophication | Landfill | Baseline | 0.00848 | kg Neq | United States (Massachusetts) | N.A. | Morelli 2019 |
| GWP | Landfill | Baseline | 320 | kg CO ₂ eq | United States (Massachusetts) | FALSE | Morelli 2019 |
| GWP | AD | Baseline | -170 | kg CO ₂ eq | United States | TRUE | Morris 2017 |
| GWP | Compost | Baseline | -50 | kg CO ₂ eq | United States | TRUE | Morris 2017 |
| GWP | Landfill | Baseline | 30 | kg CO ₂ eq | United States | TRUE | Morris 2017 |
| GWP | WWTP | Baseline | 100 | kg CO ₂ eq | United States | TRUE | Morris 2017 |
| Acidification | AD | AD (plus ~10% compost) | 0 | kg SO ₂ eq | Ireland | N.A. | Oldfield 2016 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|-------------|--|--------|-----------------------------------|---------------|-------------------------------|---------------|
| Eutrophication | AD | AD (plus ~10% compost) | 0 | kg Neq | Ireland | N.A. | Oldfield 2016 |
| GWP | AD | AD (plus ~10% compost) | -108 | kg CO ₂ eq | Ireland | FALSE | Oldfield 2016 |
| Acidification | Combust | Incineration | 0.313 | kg SO ₂ eq | Ireland | N.A. | Oldfield 2016 |
| Eutrophication | Combust | Incineration | 0.272 | kg Neq | Ireland | N.A. | Oldfield 2016 |
| GWP | Combust | Incineration | -18.7 | kg CO ₂ eq | Ireland | N.A. | Oldfield 2016 |
| Acidification | Compost | Composting | 1.36 | kg SO ₂ eq | Ireland | N.A. | Oldfield 2016 |
| Eutrophication | Compost | Composting | 0.884 | kg Neq | Ireland | N.A. | Oldfield 2016 |
| GWP | Compost | Composting | 15 | kg CO ₂ eq | Ireland | TRUE | Oldfield 2016 |
| Acidification | Source Red. | Source reduction (plus compost, AD, or incineration) | -34.6 | kg SO ₂ eq | Ireland | N.A. | Oldfield 2016 |
| Eutrophication | Source Red. | Source reduction (plus compost, AD, or incineration) | -21.2 | kg Neq | Ireland | N.A. | Oldfield 2016 |
| GWP | Source Red. | Source reduction (plus compost, AD, or incineration) | -3550 | kg CO ₂ eq | Ireland | N.A. | Oldfield 2016 |
| Acidification | Donate | Average value of all scenarios | -57 | kg SO ₂ eq | United States | N.A. | OR DEQ 2019 |
| CED | Donate | Average value of all scenarios | -27300 | MJ | United States | N.A. | OR DEQ 2019 |
| Eutrophication | Donate | Average value of all scenarios | -36.9 | kg Neq | United States | N.A. | OR DEQ 2019 |
| GWP | Donate | Average value of all scenarios | -2760 | kg CO ₂ eq | United States | N.A. | OR DEQ 2019 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|----------|--|--------|-----------------------------------|-------------------------------------|-------------------------------|--------------|
| Water Consumption | Donate | Average value of all scenarios | -398 | m ³ water | United States | N.A. | OR DEQ 2019 |
| GWP | AD | Energy and fertilizer displacement, CT = 200 | 26.4 | kg CO ₂ eq | United States (Northern California) | FALSE | Pace 2018 |
| GWP | AD | Energy and fertilizer displacement, CT = 350 | 18.9 | kg CO ₂ eq | United States (Northern California) | FALSE | Pace 2018 |
| GWP | AD | Energy and fertilizer displacement, CT = 500 | 15.3 | kg CO ₂ eq | United States (Northern California) | FALSE | Pace 2018 |
| GWP | AD | Energy and fertilizer displacement, CT = 650 | 11.7 | kg CO ₂ eq | United States (Northern California) | FALSE | Pace 2018 |
| GWP | AD | Energy and fertilizer displacement, CT = 800 | 8.1 | kg CO ₂ eq | United States (Northern California) | FALSE | Pace 2018 |
| GWP | AD | nonbiogenic only | -217 | kg CO ₂ eq | United States | TRUE | Parry 2012 |
| GWP | Compost | nonbiogenic only | 83.1 | kg CO ₂ eq | United States | TRUE | Parry 2012 |
| GWP | Landfill | nonbiogenic only | 508 | kg CO ₂ eq | United States | TRUE | Parry 2012 |
| GWP | WWTP | nonbiogenic only | 279 | kg CO ₂ eq | United States | TRUE | Parry 2012 |
| Acidification | AD | Food residue | 0.03 | kg SO ₂ eq | Germany | N.A. | Poeschl 2012 |
| Acidification | AD | Pomace | 0.06 | kg SO ₂ eq | Germany | N.A. | Poeschl 2012 |
| Acidification | AD | Slaughterhouse waste | 0.01 | kg SO ₂ eq | Germany | N.A. | Poeschl 2012 |
| Acidification | AD | Grease separator sludge | 0.05 | kg SO ₂ eq | Germany | N.A. | Poeschl 2012 |
| Eutrophication | AD | Pomace | 0 | kg Neq | Germany | N.A. | Poeschl 2012 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-----------------|---------|-------------------------|--------|-----------------------------------|-----------|-------------------------------|--------------|
| Eutrophication | AD | Food residue | 0.0888 | kg Neq | Germany | N.A. | Poeschl 2012 |
| Eutrophication | AD | Pomace | 0.109 | kg Neq | Germany | N.A. | Poeschl 2012 |
| Eutrophication | AD | Slaughterhouse waste | 0.0592 | kg Neq | Germany | N.A. | Poeschl 2012 |
| Eutrophication | AD | Grease separator sludge | 0.0493 | kg Neq | Germany | N.A. | Poeschl 2012 |
| GWP | AD | Food residue | -51.7 | kg CO ₂ eq | Germany | FALSE | Poeschl 2012 |
| GWP | AD | Pomace | -85.5 | kg CO ₂ eq | Germany | FALSE | Poeschl 2012 |
| GWP | AD | Slaughterhouse waste | -50.6 | kg CO ₂ eq | Germany | FALSE | Poeschl 2012 |
| GWP | AD | Grease separator sludge | -26.4 | kg CO ₂ eq | Germany | FALSE | Poeschl 2012 |
| Land Occupation | AD | Pomace | 0.66 | m ² .yr | Germany | N.A. | Poeschl 2012 |
| Land Occupation | AD | Food residue | 3.8 | m ² .yr | Germany | N.A. | Poeschl 2012 |
| Land Occupation | AD | Pomace | 3.58 | m ² .yr | Germany | N.A. | Poeschl 2012 |
| Land Occupation | AD | Slaughterhouse waste | 2.12 | m ² .yr | Germany | N.A. | Poeschl 2012 |
| Land Occupation | AD | Slaughterhouse waste | 0.42 | m ² .yr | Germany | N.A. | Poeschl 2012 |
| Land Occupation | AD | Grease separator sludge | 2.05 | m ² .yr | Germany | N.A. | Poeschl 2012 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|-------------|-------------------------|--------|-----------------------------------|-----------|-------------------------------|--------------|
| Land Occupation | AD | Grease separator sludge | 0.4 | m ² .yr | Germany | N.A. | Poeschl 2012 |
| Water Consumption | AD | Food residue | -0.73 | m ³ water | Germany | N.A. | Poeschl 2012 |
| Water Consumption | AD | Pomace | -0.89 | m ³ water | Germany | N.A. | Poeschl 2012 |
| Water Consumption | AD | Slaughterhouse waste | -0.48 | m ³ water | Germany | N.A. | Poeschl 2012 |
| Water Consumption | AD | Grease separator sludge | -0.31 | m ³ water | Germany | N.A. | Poeschl 2012 |
| GWP | AD | Farm | -76 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | AD | Foodservice | -243 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | AD | Manufacturing | -206 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | AD | Residential | -197 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | AD | Retail | -204 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | Animal Feed | Farm | -100 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Animal Feed | Foodservice | -222 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Animal Feed | Manufacturing | -176 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Animal Feed | Residential | -191 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|--------|-------------|---------------|--------|-----------------------------------|-----------|-------------------------------|------------|
| GWP | Animal Feed | Retail | -213 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Combust | Farm | 996 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Combust | Foodservice | 642 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Combust | Manufacturing | 717 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Combust | Residential | 737 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Combust | Retail | 725 | kg CO ₂ eq | Various | N.A. | ReFED 2023 |
| GWP | Compost | Farm | 37 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Compost | Foodservice | 58.1 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Compost | Manufacturing | 53.3 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Compost | Residential | 52.2 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Compost | Retail | 53.1 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Land App. | Farm | 12.7 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | Land App. | Foodservice | 29.6 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | Land App. | Manufacturing | 25.1 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | Land App. | Residential | 22.5 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | Land App. | Retail | 21.8 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-----------------|-----------|---------------|--------|-----------------------------------|-----------|-------------------------------|------------|
| GWP | Landfill | Farm | 328 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Landfill | Foodservice | 834 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Landfill | Manufacturing | 722 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Landfill | Residential | 693 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Landfill | Retail | 715 | kg CO ₂ eq | Various | TRUE | ReFED 2023 |
| GWP | Unharvest | Farm | 47.4 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | WWTP | Farm | 215 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | WWTP | Foodservice | 572 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | WWTP | Manufacturing | 493 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | WWTP | Residential | 473 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| GWP | WWTP | Retail | 488 | kg CO ₂ eq | Various | FALSE | ReFED 2023 |
| Acidification | Combust | Baseline | -0.64 | kg SO ₂ eq | Taiwan | N.A. | Shih 2021 |
| CED | Combust | Baseline | -2070 | MJ | Taiwan | N.A. | Shih 2021 |
| Eutrophication | Combust | Baseline | -0.284 | kg Neq | Taiwan | N.A. | Shih 2021 |
| GWP | Combust | Baseline | 493 | kg CO ₂ eq | Taiwan | N.A. | Shih 2021 |
| Land Occupation | Combust | Baseline | -1.04 | m ² .yr | Taiwan | N.A. | Shih 2021 |
| Acidification | Compost | Baseline | 0.47 | kg SO ₂ eq | Taiwan | N.A. | Shih 2021 |
| CED | Compost | Baseline | 894 | MJ | Taiwan | N.A. | Shih 2021 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|---------|---|---------|-----------------------------------|----------------|-------------------------------|---------------|
| Eutrophication | Compost | Baseline | 0.261 | kg Neq | Taiwan | N.A. | Shih 2021 |
| GWP | Compost | Baseline | 63.2 | kg CO ₂ eq | Taiwan | FALSE | Shih 2021 |
| Land Occupation | Compost | Baseline | 0.27 | m ² .yr | Taiwan | N.A. | Shih 2021 |
| Acidification | AD | average value (10th and 90th also provided) | 7.63 | kg SO ₂ eq | United Kingdom | N.A. | Slorach 2019a |
| CED | AD | average value (10th and 90th also provided) | -1930 | MJ | United Kingdom | N.A. | Slorach 2019a |
| Eutrophication | AD | average value (10th and 90th also provided) | -0.0846 | kg Neq | United Kingdom | N.A. | Slorach 2019a |
| Eutrophication | AD | average value (10th and 90th also provided) | 0.71 | kg Neq | United Kingdom | N.A. | Slorach 2019a |
| GWP | AD | average value (10th and 90th also provided) | -31.6 | kg CO ₂ eq | United Kingdom | FALSE | Slorach 2019a |
| Land Occupation | AD | average value (10th and 90th also provided) | 0.6 | m ² .yr | United Kingdom | N.A. | Slorach 2019a |
| Land Occupation | AD | average value (10th and 90th also provided) | -0.21 | m ² .yr | United Kingdom | N.A. | Slorach 2019a |
| Water Consumption | AD | average value (10th and 90th also provided) | -274 | m ³ water | United Kingdom | N.A. | Slorach 2019a |
| Acidification | Combust | average value (10th and 90th also provided) | 0.39 | kg SO ₂ eq | United Kingdom | N.A. | Slorach 2019a |
| CED | Combust | average value (10th and 90th also provided) | -900 | MJ | United Kingdom | N.A. | Slorach 2019a |
| Eutrophication | Combust | average value (10th and 90th also provided) | 0.0348 | kg Neq | United Kingdom | N.A. | Slorach 2019a |
| Eutrophication | Combust | average value (10th and 90th also provided) | 0.118 | kg Neq | United Kingdom | N.A. | Slorach 2019a |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|----------|---|----------|-----------------------------------|----------------|-------------------------------|---------------|
| GWP | Combust | average value (10th and 90th also provided) | -4.97 | kg CO ₂ eq | United Kingdom | N.A. | Slorach 2019a |
| Land Occupation | Combust | average value (10th and 90th also provided) | -0.37 | m ² .yr | United Kingdom | N.A. | Slorach 2019a |
| Land Occupation | Combust | average value (10th and 90th also provided) | 0.01 | m ² .yr | United Kingdom | N.A. | Slorach 2019a |
| Water Consumption | Combust | average value (10th and 90th also provided) | -149 | m ³ water | United Kingdom | N.A. | Slorach 2019a |
| Acidification | Compost | average value (10th and 90th also provided) | 10.1 | kg SO ₂ eq | United Kingdom | N.A. | Slorach 2019a |
| CED | Compost | average value (10th and 90th also provided) | 1310 | MJ | United Kingdom | N.A. | Slorach 2019a |
| Eutrophication | Compost | average value (10th and 90th also provided) | -0.00642 | kg Neq | United Kingdom | N.A. | Slorach 2019a |
| Eutrophication | Compost | average value (10th and 90th also provided) | 0.385 | kg Neq | United Kingdom | N.A. | Slorach 2019a |
| GWP | Compost | average value (10th and 90th also provided) | 77.5 | kg CO ₂ eq | United Kingdom | FALSE | Slorach 2019a |
| Land Occupation | Compost | average value (10th and 90th also provided) | 4.08 | m ² .yr | United Kingdom | N.A. | Slorach 2019a |
| Land Occupation | Compost | average value (10th and 90th also provided) | -0.67 | m ² .yr | United Kingdom | N.A. | Slorach 2019a |
| Water Consumption | Compost | average value (10th and 90th also provided) | 97.3 | m ³ water | United Kingdom | N.A. | Slorach 2019a |
| Acidification | Landfill | average value (10th and 90th also provided) | 0.24 | kg SO ₂ eq | United Kingdom | N.A. | Slorach 2019a |
| CED | Landfill | average value (10th and 90th also provided) | 140 | MJ | United Kingdom | N.A. | Slorach 2019a |
| Eutrophication | Landfill | average value (10th and 90th also provided) | 0.0204 | kg Neq | United Kingdom | N.A. | Slorach 2019a |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|----------|---|--------|-----------------------------------|----------------|-------------------------------|---------------|
| Eutrophication | Landfill | average value (10th and 90th also provided) | 7.32 | kg Neq | United Kingdom | N.A. | Slorach 2019a |
| GWP | Landfill | average value (10th and 90th also provided) | 195 | kg CO ₂ eq | United Kingdom | TRUE | Slorach 2019a |
| Land Occupation | Landfill | average value (10th and 90th also provided) | 0.79 | m ² .yr | United Kingdom | N.A. | Slorach 2019a |
| Land Occupation | Landfill | average value (10th and 90th also provided) | 3.85 | m ² .yr | United Kingdom | N.A. | Slorach 2019a |
| Water Consumption | Landfill | average value (10th and 90th also provided) | -39.4 | m ³ water | United Kingdom | N.A. | Slorach 2019a |
| Acidification | AD | Baseline | 7.6 | kg SO ₂ eq | United Kingdom | N.A. | Slorach 2019b |
| CED | AD | Baseline | -1990 | MJ | United Kingdom | N.A. | Slorach 2019b |
| Eutrophication | AD | Baseline | -0.101 | kg Neq | United Kingdom | N.A. | Slorach 2019b |
| Eutrophication | AD | Baseline | 0.69 | kg Neq | United Kingdom | N.A. | Slorach 2019b |
| GWP | AD | Baseline | -39 | kg CO ₂ eq | United Kingdom | FALSE | Slorach 2019b |
| Land Occupation | AD | Baseline | 0.6 | m ² .yr | United Kingdom | N.A. | Slorach 2019b |
| Land Occupation | AD | Baseline | -0.23 | m ² .yr | United Kingdom | N.A. | Slorach 2019b |
| Water Consumption | AD | Baseline | -249 | m ³ water | United Kingdom | N.A. | Slorach 2019b |
| Acidification | Combust | Baseline | 0.4 | kg SO ₂ eq | United Kingdom | N.A. | Slorach 2019b |
| CED | Combust | Baseline | -940 | MJ | United Kingdom | N.A. | Slorach 2019b |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|----------|----------|--------|-----------------------------------|----------------|-------------------------------|---------------|
| Eutrophication | Combust | Baseline | 0.0241 | kg Neq | United Kingdom | N.A. | Slorach 2019b |
| Eutrophication | Combust | Baseline | 0.0986 | kg Neq | United Kingdom | N.A. | Slorach 2019b |
| GWP | Combust | Baseline | -10 | kg CO ₂ eq | United Kingdom | N.A. | Slorach 2019b |
| Land Occupation | Combust | Baseline | -0.04 | m ² .yr | United Kingdom | N.A. | Slorach 2019b |
| Land Occupation | Combust | Baseline | -0.005 | m ² .yr | United Kingdom | N.A. | Slorach 2019b |
| Water Consumption | Combust | Baseline | -133 | m ³ water | United Kingdom | N.A. | Slorach 2019b |
| Acidification | Landfill | Baseline | 0.2 | kg SO ₂ eq | United Kingdom | N.A. | Slorach 2019b |
| CED | Landfill | Baseline | 120 | MJ | United Kingdom | N.A. | Slorach 2019b |
| Eutrophication | Landfill | Baseline | 0.0168 | kg Neq | United Kingdom | N.A. | Slorach 2019b |
| Eutrophication | Landfill | Baseline | 7.3 | kg Neq | United Kingdom | N.A. | Slorach 2019b |
| GWP | Landfill | Baseline | 193 | kg CO ₂ eq | United Kingdom | TRUE | Slorach 2019b |
| Land Occupation | Landfill | Baseline | 0.1 | m ² .yr | United Kingdom | N.A. | Slorach 2019b |
| Land Occupation | Landfill | Baseline | 3.84 | m ² .yr | United Kingdom | N.A. | Slorach 2019b |
| Water Consumption | Landfill | Baseline | -33 | m ³ water | United Kingdom | N.A. | Slorach 2019b |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|----------|--|---------|-----------------------------------|-----------|-------------------------------|-------------|
| Eutrophication | Donate | 80% donation with avoided emissions and 20% landfilling with energy recovery | -0.892 | kg Neq | Brazil | N.A. | Sulis 2021 |
| GWP | Donate | 80% donation with avoided emissions and 20% landfilling with energy recovery | -325 | kg CO ₂ eq | Brazil | N.A. | Sulis 2021 |
| Water Consumption | Donate | 80% donation with avoided emissions and 20% landfilling with energy recovery | -69 | m ³ water | Brazil | N.A. | Sulis 2021 |
| Acidification | Landfill | With energy recovery | 1.2 | kg SO ₂ eq | Brazil | N.A. | Sulis 2021 |
| Eutrophication | Landfill | With energy recovery | 0.0219 | kg Neq | Brazil | N.A. | Sulis 2021 |
| GWP | Landfill | With energy recovery | 175 | kg CO ₂ eq | Brazil | FALSE | Sulis 2021 |
| Water Consumption | Landfill | With energy recovery | -3.5 | m ³ water | Brazil | N.A. | Sulis 2021 |
| GWP | Donate | base | -400 | kg CO ₂ eq | Sweden | N.A. | Sundin 2022 |
| Acidification | AD | Decentralized AD, biogas to electricity | 0.275 | kg SO ₂ eq | Singapore | N.A. | Tian 2021 |
| Acidification | AD | Decentralized AD, biogas to cook fuel | 0.119 | kg SO ₂ eq | Singapore | N.A. | Tian 2021 |
| Acidification | AD | Centralized AD, biogas to electricity | 0.309 | kg SO ₂ eq | Singapore | N.A. | Tian 2021 |
| Acidification | AD | Centralized AD, biogas to transport fuel | -0.0989 | kg SO ₂ eq | Singapore | N.A. | Tian 2021 |
| Eutrophication | AD | Decentralized AD, biogas to electricity | 0.0698 | kg Neq | Singapore | N.A. | Tian 2021 |
| Eutrophication | AD | Decentralized AD, biogas to cook fuel | 0.0625 | kg Neq | Singapore | N.A. | Tian 2021 |
| Eutrophication | AD | Centralized AD, biogas to electricity | 0.0955 | kg Neq | Singapore | N.A. | Tian 2021 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|---------|--|---------|-----------------------------------|-----------|-------------------------------|-----------|
| Eutrophication | AD | Centralized AD, biogas to transport fuel | 0.0853 | kg Neq | Singapore | N.A. | Tian 2021 |
| Eutrophication | AD | Decentralized AD, biogas to electricity | 0.953 | kg Neq | Singapore | N.A. | Tian 2021 |
| GWP | AD | Decentralized AD, biogas to electricity | -72.8 | kg CO ₂ eq | Singapore | FALSE | Tian 2021 |
| GWP | AD | Decentralized AD, biogas to cook fuel | -238 | kg CO ₂ eq | Singapore | FALSE | Tian 2021 |
| GWP | AD | Centralized AD, biogas to electricity | -81.4 | kg CO ₂ eq | Singapore | FALSE | Tian 2021 |
| GWP | AD | Centralized AD, biogas to transport fuel | -137 | kg CO ₂ eq | Singapore | FALSE | Tian 2021 |
| Water Consumption | AD | Decentralized AD, biogas to electricity | 0.0508 | m ³ water | Singapore | N.A. | Tian 2021 |
| Water Consumption | AD | Decentralized AD, biogas to cook fuel | 0.311 | m ³ water | Singapore | N.A. | Tian 2021 |
| Water Consumption | AD | Centralized AD, biogas to electricity | 0.0662 | m ³ water | Singapore | N.A. | Tian 2021 |
| Water Consumption | AD | Centralized AD, biogas to transport fuel | 0.282 | m ³ water | Singapore | N.A. | Tian 2021 |
| Acidification | Combust | Base | 0.0873 | kg SO ₂ eq | Singapore | N.A. | Tian 2021 |
| Eutrophication | Combust | Base | 0.0153 | kg Neq | Singapore | N.A. | Tian 2021 |
| Eutrophication | Combust | Base | 0.00155 | kg Neq | Singapore | N.A. | Tian 2021 |
| GWP | Combust | Base | -29.2 | kg CO ₂ eq | Singapore | N.A. | Tian 2021 |
| Water Consumption | Combust | Base | 0.0213 | m ³ water | Singapore | N.A. | Tian 2021 |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|----------------|-------------|--|---------|-----------------------------------|-----------|-------------------------------|----------------|
| Acidification | AD | Baseline | -0.169 | kg SO ₂ eq | Singapore | N.A. | Tong 2018 |
| Acidification | AD | Baseline | -0.0901 | kg SO ₂ eq | Singapore | N.A. | Tong 2018 |
| Eutrophication | AD | Baseline | 0.0948 | kg Neq | Singapore | N.A. | Tong 2018 |
| Eutrophication | AD | Baseline | 0.163 | kg Neq | Singapore | N.A. | Tong 2018 |
| GWP | AD | Baseline | 94.1 | kg CO ₂ eq | Singapore | TRUE | Tong 2018 |
| GWP | AD | Baseline | 38.5 | kg CO ₂ eq | Singapore | TRUE | Tong 2018 |
| Acidification | Combust | Baseline | -0.0864 | kg SO ₂ eq | Singapore | N.A. | Tong 2018 |
| Eutrophication | Combust | Baseline | 0.0837 | kg Neq | Singapore | N.A. | Tong 2018 |
| GWP | Combust | Baseline | 107 | kg CO ₂ eq | Singapore | N.A. | Tong 2018 |
| GWP | AD | Net emissions; with curing | -66.1 | kg CO ₂ eq | Various | TRUE | U.S. EPA 2020b |
| GWP | AD | Net emissions; with direct application | -154 | kg CO ₂ eq | Various | TRUE | U.S. EPA 2020b |
| GWP | AD | Net emissions; with curing | -44.1 | kg CO ₂ eq | Various | TRUE | U.S. EPA 2020b |
| GWP | AD | Net emissions; with direct application | -110 | kg CO ₂ eq | Various | TRUE | U.S. EPA 2020b |
| GWP | Combust | Net emissions | -143 | kg CO ₂ eq | Various | N.A. | U.S. EPA 2020b |
| GWP | Compost | Net emissions | -132 | kg CO ₂ eq | Various | TRUE | U.S. EPA 2020b |
| GWP | Landfill | Net emissions | 551 | kg CO ₂ eq | Various | TRUE | U.S. EPA 2020b |
| GWP | Source Red. | Net emissions | -4030 | kg CO ₂ eq | Various | N.A. | U.S. EPA 2020c |

| Impact | Pathway | Scenario | Result | Unit (per metric ton wasted food) | Geography | CO ₂ Sequestration | Citation |
|-------------------|-------------|----------|--------|-----------------------------------|---------------|-------------------------------|----------------|
| CED | Source Red. | Baseline | -14700 | MJ | United States | N.A. | U.S. EPA 2021c |
| CED | Source Red. | Baseline | -33100 | MJ | United States | N.A. | U.S. EPA 2021c |
| GWP | Source Red. | Baseline | -1250 | kg CO ₂ eq | United States | N.A. | U.S. EPA 2021c |
| GWP | Source Red. | Baseline | -3040 | kg CO ₂ eq | United States | N.A. | U.S. EPA 2021c |
| Land Occupation | Source Red. | Baseline | -1560 | m ² .yr | United States | N.A. | U.S. EPA 2021c |
| Land Occupation | Source Red. | Baseline | -7310 | m ² .yr | United States | N.A. | U.S. EPA 2021c |
| Water Consumption | Source Red. | Baseline | -113 | m ³ water | United States | N.A. | U.S. EPA 2021c |
| Water Consumption | Source Red. | Baseline | -298 | m ³ water | United States | N.A. | U.S. EPA 2021c |

TABLE D-2. EXTRACTED DATA VALUES USED FOR INTRA-STUDY RANKINGS

Non-standardized impact results are presented per the listed functional unit and units. In cells with multiple values listed, multiple pathway scenario impact results are averaged to establish the pathway ranking. Rank 1 = Best performance. Higher numbers are associated with higher impact. Numbers do not always start at 1 as additional pathways (not included in the report) are sometimes ranked as well. The first set of pathway columns are the data used to establish rankings; the second set are the rankings. WF = wasted food.

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original Impact category |
|--------------------------|---------------------|----------|----------|---------|-----------------------|----------|-----------------|------------------------------|-----------------------|-----------|----------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|--------------------------|
| Eutrophication | 0.02 | | | | 0.08 | | Ahamed 2016 | 1 tonne WF | kg PO ₄ eq | Singapore | | 1 | | | | 2 | | Eutrophication Potential |
| Acidification | 0.04 | | | | 0.50 | | Ahamed 2016 | 1 tonne WF | kg SO ₂ eq | Singapore | | 1 | | | | 2 | | Acidification |
| GWP | -1.5 | | | | 2.4E+2 | | Ahamed 2016 | 1 tonne WF | kg CO ₂ eq | Singapore | | 1 | | | | 2 | | Global Warming Potential |
| GWP | | -3.3E+2 | | | | 2.4E+3 | Albizzati 2021b | 1 metric ton of WF processed | kg CO ₂ eq | Europe | Wet Animal Feed | | 1 | | | | 2 | Global Warming Potential |
| GWP | | -3.3E+2 | | | 42 | | Albizzati 2021b | 1 metric ton of WF processed | kg CO ₂ eq | Europe | Wet Animal Feed | | 1 | | | 2 | | Global Warming Potential |
| GWP | | -3.3E+2 | | 1.5E+2 | | | Albizzati 2021b | 1 metric ton of WF processed | kg CO ₂ eq | Europe | Wet Animal Feed | | 1 | | 2 | | | Global Warming Potential |
| GWP | 1.2E+2 | -3.3E+2 | | | | | Albizzati 2021b | 1 metric ton of WF processed | kg CO ₂ eq | Europe | Wet Animal Feed | 2 | 1 | | | | | Global Warming Potential |
| GWP | | | 1 | | | 2 | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | | 2 | Global Warming Potential |
| GWP | | | 1 | | 2 | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | 2 | | Global Warming Potential |
| GWP | | | 1 | 2 | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | 2 | | | Global Warming Potential |
| GWP | 2 | | 1 | | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | 2 | | 1 | | | | Global Warming Potential |
| GWP | 1.2E+2 | | | 1.5E+2 | 42 | 2.4E+3 | Albizzati 2021b | WF processing | kg CO ₂ eq | Europe | Traditional Pathways | 1 | | | 1 | 1 | 2 | Global Warming Potential |
| Human Toxicity | | 1 | | | | 2 | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | | | 2 | Human toxicity, cancer |
| Human Toxicity | | 1 | | | 2 | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | | 2 | | Human toxicity, cancer |
| Human Toxicity | | 1 | | 1 | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | 1 | | | Human toxicity, cancer |
| Human Toxicity | 1 | 1 | | | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | 1 | 1 | | | | | Human toxicity, cancer |
| Human Toxicity | | | 1 | | | 2 | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | | 2 | Human toxicity, cancer |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|-----------------|-----------------|------------------|-----------|----------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|------------------------------|
| Human Toxicity | | | 1 | | 2 | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | 2 | | Human toxicity, cancer |
| Human Toxicity | 1 | | | 1 | 2 | 2 | Albizzati 2021b | WF processing | relative ranking | Europe | Traditional Pathways | 1 | | | 1 | 2 | 2 | Human toxicity, cancer |
| Human Toxicity | | 2 | | | | 1 | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 2 | | | | 1 | Human Toxicity, non-cancer |
| Human Toxicity | | 2 | | | 1 | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 2 | | | 1 | | Human Toxicity, non-cancer |
| Human Toxicity | | 1 | | 2 | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | 2 | | | Human Toxicity, non-cancer |
| Human Toxicity | 1 | 1 | | | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | 1 | 1 | | | | | Human Toxicity, non-cancer |
| Human Toxicity | | | 2 | | | 1 | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 2 | | | 1 | Human Toxicity, non-cancer |
| Human Toxicity | | | 2 | | 1 | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 2 | | 1 | | Human Toxicity, non-cancer |
| Human Toxicity | 3 | | | 4 | 1 | 2 | Albizzati 2021b | WF processing | relative ranking | Europe | Traditional Pathways | 3 | | | 4 | 1 | 2 | Human Toxicity, non-cancer |
| Particulate matter formation | | 1 | | | | 2 | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | | | 2 | Particulate matter formation |
| Particulate matter formation | | 1 | | | 2 | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | | 2 | | Particulate matter formation |
| Particulate matter formation | | 1 | | 2 | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | 2 | | | Particulate matter formation |
| Particulate matter formation | 2 | 1 | | | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | 2 | 1 | | | | | Particulate matter formation |
| Particulate matter formation | | | 1 | | | 2 | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | | 2 | Particulate matter formation |
| Particulate matter formation | | | 1 | | 2 | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | 2 | | Particulate matter formation |
| Particulate matter formation | 2 | | 1 | | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | 2 | | 1 | | | | Particulate matter formation |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|-----------------|-----------------|------------------|-----------|----------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|------------------------------|
| Particulate matter formation | 1 | | | 2 | 1 | 1 | Albizzati 2021b | WF processing | relative ranking | Europe | Traditional Pathways | 1 | | | 2 | 1 | 1 | Particulate matter formation |
| Acidification | | 2 | | | | 1 | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 2 | | | | 1 | Acidification, terrestrial |
| Acidification | | 2 | | | 1 | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 2 | | | 1 | | Acidification, terrestrial |
| Acidification | | 1 | | 2 | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | 2 | | | Acidification, terrestrial |
| Acidification | 2 | 1 | | | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | 2 | 1 | | | | | Acidification, terrestrial |
| Acidification | | | 2 | | | 1 | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 2 | | | 1 | Acidification, terrestrial |
| Acidification | | | 2 | | 1 | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 2 | | 1 | | Acidification, terrestrial |
| Acidification | | | 1 | 2 | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | 2 | | | Acidification, terrestrial |
| Acidification | 2 | | 1 | | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | 2 | | 1 | | | | Acidification, terrestrial |
| Acidification | 3 | | | 4 | 1 | 2 | Albizzati 2021b | WF processing | relative ranking | Europe | Traditional Pathways | 3 | | | 4 | 1 | 2 | Acidification, terrestrial |
| Eutrophication | | 2 | | | | 1 | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 2 | | | | 1 | Eutrophication, terrestrial |
| Eutrophication | | 2 | | | 1 | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 2 | | | 1 | | Eutrophication, terrestrial |
| Eutrophication | | 1 | | 2 | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | 2 | | | Eutrophication, terrestrial |
| Eutrophication | 2 | 1 | | | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | 2 | 1 | | | | | Eutrophication, terrestrial |
| Eutrophication | | | 1 | 2 | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | 2 | | | Eutrophication, terrestrial |
| Eutrophication | 2 | | 1 | | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | 2 | | 1 | | | | Eutrophication, terrestrial |
| Eutrophication | 2 | | | 3 | 1 | 1 | Albizzati 2021b | WF processing | relative ranking | Europe | Traditional Pathways | 2 | | | 3 | 1 | 1 | Eutrophication, terrestrial |
| Eutrophication | | 2 | | | | 1 | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 2 | | | | 1 | Eutrophication, marine |
| Eutrophication | | 2 | | | 1 | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 2 | | | 1 | | Eutrophication, marine |
| Eutrophication | | 1 | | 2 | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | 2 | | | Eutrophication, marine |
| Eutrophication | 2 | 1 | | | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | 2 | 1 | | | | | Eutrophication, marine |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|--------------------------|---------------------|----------|----------|---------|-----------------------|----------|-----------------|-----------------|------------------|-----------|----------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|----------------------------|
| Eutrophication | | | 1 | | | 2 | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | | 2 | Eutrophication, marine |
| Eutrophication | | | 1 | | 2 | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | 2 | | Eutrophication, marine |
| Eutrophication | | | 1 | 2 | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | 2 | | | Eutrophication, marine |
| Eutrophication | 2 | | 1 | | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | 2 | | 1 | | | | Eutrophication, marine |
| Eutrophication | 3 | | | 2 | 1 | 1 | Albizzati 2021b | WF processing | relative ranking | Europe | Traditional Pathways | 3 | | | 2 | 1 | 1 | Eutrophication, marine |
| Eutrophication | | 1 | | | | 2 | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | | | 2 | Eutrophication, freshwater |
| Eutrophication | | 1 | | | 2 | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | | 2 | | Eutrophication, freshwater |
| Eutrophication | | 1 | | 2 | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | 2 | | | Eutrophication, freshwater |
| Eutrophication | 2 | 1 | | | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | 2 | 1 | | | | | Eutrophication, freshwater |
| Eutrophication | | | 1 | | | 2 | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | | 2 | Eutrophication, freshwater |
| Eutrophication | | | 1 | | 2 | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | | 2 | | Eutrophication, freshwater |
| Eutrophication | 2 | | 1 | 2 | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | 2 | | 1 | | | | Eutrophication, freshwater |
| Eutrophication | 1 | | | 2 | 1 | 1 | Albizzati 2021b | WF processing | relative ranking | Europe | Traditional Pathways | 1 | | | 2 | 1 | 1 | Eutrophication, freshwater |
| Ecotoxicity | | 1 | | | | 2 | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | | | 2 | Ecotoxicity, freshwater |
| Ecotoxicity | | 1 | | | 2 | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | | 2 | | Ecotoxicity, freshwater |
| Ecotoxicity | | 1 | | 2 | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | | 1 | | 2 | | | Ecotoxicity, freshwater |
| Ecotoxicity | 2 | 1 | | | | | Albizzati 2021b | 7.9 kg WF | relative ranking | Europe | Wet Animal Feed | 2 | 1 | | | | | Ecotoxicity, freshwater |
| Ecotoxicity | | | 1 | 2 | | | Albizzati 2021b | 10.4 kg WF | relative ranking | Europe | Dry Animal Feed | | | 1 | 2 | | | Ecotoxicity, freshwater |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|--------------------------|---------------------|----------|-------------|-----------|------------------------------|----------|-----------------|---------------------------------------|------------------------|-----------|---|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|-----------------------------|
| Ecotoxicity | 1 | | | 2 | 1 | 1 | Albizzati 2021b | WF processing | relative ranking | Europe | Traditional Pathways | 1 | | | 2 | 1 | 1 | Ecotoxicity, freshwater |
| GWP | 81, 22 | -76 | -210, 17 | 51, 270 | | | Albizzati 2021a | 1 ton post-process WF | kg CO ₂ eq | Europe | | 3 | 2 | 1 | 4 | | | GWP |
| Eutrophication | 1, 1 | 0.68 | -0.04, 0.15 | 0.16, 1.1 | | | Albizzati 2021a | 1 ton post-process WF | kg Neq | Europe | | 4 | 3 | 1 | 2 | | | Marine Eutrophication |
| Eutrophication | 0.7, 0.71 | 0.47 | 0.01, 0.3 | 0.67, 1 | | | Albizzati 2021a | 1 ton post-process WF | kg Peq | Europe | | 3 | 2 | 1 | 4 | | | Freshwater eutrophication |
| GWP | 27 | | | 1.3E+2 | | | Al-Rumaihi 2020 | metric ton WF | kg CO ₂ eq | Qatar | | 1 | | | 2 | | | Global Warming Potential |
| Human Toxicity | 2 | | | 1 | | | Al-Rumaihi 2020 | metric ton WF | relative ranking | Qatar | | 2 | | | 1 | | | Human Toxicity |
| Eutrophication | 2 | | | 1 | | | Al-Rumaihi 2020 | metric ton WF | relative ranking | Qatar | Moderate difference | 2 | | | 1 | | | Eutrophication |
| Acidification | 2 | | | 1 | | | Al-Rumaihi 2020 | metric ton WF | relative ranking | Qatar | Moderate difference | 2 | | | 1 | | | Acidification |
| GWP | 0.36, 0.28 | | | 0.40 | -0.15, -0.2 | | Benavente 2017 | kg two-phase olive mill waste (TPOMW) | kg CO ₂ eq | Spain | B1 = composting B2 = Anaerobic digestion B3 = Anaerobic digestion T1 = combustion T2 = combustion | 2 | | | 3 | 1 | | Global Warming Potential |
| Acidification | 4.2E-3, 1.8E-03 | | | 4.2E-3 | -3.0E-03, -2.2E-03 | | Benavente 2017 | kg two-phase olive mill waste (TPOMW) | Accumulated Exceedance | Spain | | 2 | | | 3 | 1 | | Acidification |
| Eutrophication | 6.6E-03, 3.0E-03 | | | 7.4E-3 | -3.7E-03, -3.0E-03, -2.2E-03 | | Benavente 2017 | kg two-phase olive mill waste (TPOMW) | Accumulated Exceedance | Spain | | 2 | | | 3 | 1 | | Eutrophication, terrestrial |
| Eutrophication | -0.4E-04, -0.4E-04 | | | -3.0E-4 | 0, 0 | | Benavente 2017 | kg two-phase olive mill waste (TPOMW) | kg Peq | Spain | | 2 | | | 1 | 3 | | Freshwater Eutrophication |
| Eutrophication | 1.5E-03, 1.2E-03 | | | 3.5E-3 | -0.3E-03, -0.2E-03 | | Benavente 2017 | kg two-phase olive mill waste (TPOMW) | kg Neq | Spain | | 2 | | | 3 | 1 | | Eutrophication, marine |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|--------------------------|---------------------|----------|----------|---------|-----------------------|----------|----------------|--|-----------------------|---------------|-----------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|-------------------------------------|
| Ecotoxicity | 4.2E-01, 4.2E-01 | | | 4.0 | 0, 0 | | Benavente 2017 | kg two-phase olive mill waste (TPOMW) | CTUe | Spain | | 2 | | | 3 | 1 | | Ecotoxicity |
| Human Toxicity | 3.2E-10, 3E-10 | | | 5.5E-9 | -1E-10, -0.8E-10 | | Benavente 2017 | kg two-phase olive mill waste (TPOMW) | CTUh | Spain | | 2 | | | 3 | 1 | | Human toxicity, cancer |
| Human Toxicity | 5E-7, 5E-7 | | | 7.6E-6 | 0, 0 | | Benavente 2017 | kg two-phase olive mill waste (TPOMW) | CTUh | Spain | | 2 | | | 3 | 1 | | Human Toxicity, non-cancer |
| GWP | -3.0E+2 | | | | -3.6E+2 | | Clavreul 2012 | 1 tonne of organic kitchen waste | kg CO ₂ eq | Denmark | | 2 | | | | 1 | | Global Warming Potential |
| GWP | | 61 | 2.0E+2 | 1.2E+2 | | 1.0E+3 | Dou 2018 | 1 tonne food waste | kg CO ₂ eq | South Korea | From Kim and Kim 2010 | | 1 | 3 | 2 | | 4 | Global Warming Potential |
| GWP | -1.8E+2 | | | -78 | -1.1E+2 | -25 | Hodge 2016 | 1000 kg mixed waste (58% food waste, 42% non-food waste) | kg CO ₂ eq | United States | | 1 | | | 3 | 2 | 4 | Global Warming Potential |
| Eutrophication | 2.0E-4 | | | 4.8E-5 | 6.3E-6 | 3.0E-3 | Hoehn 2021 | 1 metric ton of WF treated | kg Peq | Spain | | 3 | | | 2 | 1 | 4 | Freshwater eutrophication |
| Eutrophication | 1.8E-5 | | | 2.0E-4 | 1.0E-3 | 7.7E-3 | Hoehn 2021 | 1 metric ton of WF treated | kg Neq | Spain | | 1 | | | 2 | 3 | 4 | Marine eutrophication |
| GWP | -10, 120 | | | | 30 | 5.9E+2 | Huang 2022 | 1 metric ton (wet basis) of collected and treated food waste | kg CO ₂ eq | China | | 2 | | | | 1 | 3 | Climate change |
| Acidification | 0.5, -0.5 | | | | -1.0 | 12 | Huang 2022 | 1 metric ton (wet basis) of collected and treated food waste | kg SO ₂ eq | China | | 2 | | | | 1 | 3 | Terrestrial acidification potential |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|---------------|--|-------------------------|-----------|---------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|--|
| Eutrophication | 0, 0.01 | | | | 0.05 | 0.03 | Huang 2022 | 1 metric ton (wet basis) of collected and treated food waste | kg Peq | China | | 1 | | | | 3 | 2 | Freshwater eutrophication potential |
| Particulate matter formation | 0, 0.05 | | | | -0.50 | 0.05 | Huang 2022 | 1 metric ton (wet basis) of collected and treated food waste | kg PM _{2.5} eq | China | | 2 | | | | 1 | 3 | Particulate matter formation potential |
| GWP | | | | 88 | | 4.9E+2 | Keng 2020 | 1 metric ton organic waste (food + landscape) | kg CO ₂ eq | Malaysia | minor difference | | | | 1 | | 2 | Global Warming Potential |
| Acidification | | | | 0.75 | | 0.07 | Keng 2020 | 1 metric ton organic waste (food + landscape) | kg SO ₂ eq | Malaysia | moderate difference | | | | 2 | | 1 | Acidification |
| Eutrophication | | | | 0.05 | | 6.7 | Keng 2020 | 1 metric ton organic waste (food + landscape) | kg Neq | Malaysia | moderate difference | | | | 1 | | 2 | Eutrophication |
| Human Toxicity | | | | 8.0E-10 | | 2.2E-5 | Keng 2020 | 1 metric ton organic waste (food + landscape) | CTUh | Malaysia | moderate difference | | | | 1 | | 2 | Human Toxicity, cancer |
| Human Toxicity | | | | 1.5E-4 | | 1.5E-3 | Keng 2020 | 1 metric ton organic waste (food + landscape) | CTUh | Malaysia | moderate difference | | | | 1 | | 2 | Human Toxicity, non-cancer |
| Particulate matter formation | | | | 0.05 | | 5.1E-3 | Keng 2020 | 1 metric ton organic waste (food + landscape) | kg PM _{2.5} eq | Malaysia | moderate difference | | | | 2 | | 1 | Particulate Matter Formation Potential |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|-----------------------------|----------|----------|----------------------|-----------------------|---------------------|---------------|---|-----------------------|-----------|------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|-----------------------------------|
| Ecotoxicity | | | | 1.5E+2 | | 1.1E+5 | Keng 2020 | 1 metric ton organic waste (food + landscape) | CTUe | Malaysia | major difference | | | | 1 | | 2 | Ecotoxicity |
| GWP | -4.0E+2 | | | -148, -73, -102, -64 | | 1140, -25, -235, -5 | Levis 2011 | 1000 kg of food waste plus 550 kg of branches | kg CO ₂ eq | | | 1 | | | 3,4,5,6 | | 2,7,8,9 | Global Warming Potential |
| GWP | -2.9E-3, -1.2E-4, -6.5E-6 | | -3.3E-3 | 4.3E-4 | | 1.1E-3, -6.7E-4 | Lin 2022 | 1 tonne of WF (wet basis) | DALY | Malaysia | | 2,4,5 | | 1 | 6 | | 3,7 | Global Warming Potential |
| Particulate matter formation | -4.3E-4, 8.5E-7, -1E-5 | | -2.3E-4 | 7.0E-5 | | 5.4E-5, -9.8E-5 | Lin 2022 | 1 tonne of WF (wet basis) | DALY | Malaysia | | 1,4,5 | | 2 | 7 | | 3,6 | Fine particulate matter formation |
| Human Toxicity | -2.2E-3, 1.4E-5, -4E-5 | | -1.3E-4 | -2.5E-4 | | 1.9E-6, -6.8E-4 | Lin 2022 | 1 tonne of WF (wet basis) | DALY | Malaysia | | 1,5,7 | | 4 | 3 | | 2,6 | Human carcinogenic toxicity |
| Human Toxicity | -7.3E-3, 1.4E-5, -2.2E-4 | | -1.2E-3 | -3.7E-3 | | 2.7E2, -2.1E-3 | Lin 2022 | 1 tonne of WF (wet basis) | DALY | Malaysia | | 1,5,6 | | 4 | 2 | | 3,7 | Human non-carcinogenic toxicity |
| Acidification | -1.9E-7, -1.6E-8, -2.1E-8 | | -7.0E-8 | 2.2E-7 | | 1.5E-7, 4.5E-8 | Lin 2022 | 1 tonne of WF (wet basis) | species.yr | Malaysia | | 1,3,4 | | 2 | 7 | | 5,6 | Terrestrial acidification |
| Eutrophication | -7.7E-8, 4.0E-9, 9.6E-10 | | -1.7E-8 | -7.6E-9 | | 3.1E-9, -2.5E-8 | Lin 2022 | 1 tonne of WF (wet basis) | species.yr | Malaysia | | 1,5,7 | | 3 | 4 | | 2,6 | Freshwater eutrophication |
| Eutrophication | -1.3E-11, 4.9E-14, -1.5E-13 | | -4.1E-10 | -9.3E-12 | | 0, -4.1E-12 | Lin 2022 | 1 tonne of WF (wet basis) | species.yr | Malaysia | | 2,5,7 | | 1 | 3 | | 4,6 | Marine eutrophication |
| Ecotoxicity | -1.3E-9, -7E-12, -8.4E-11 | | -2.9E-9 | -2.1E-9 | | 1.2E-24, -4.2E-10 | Lin 2022 | 1 tonne of WF (wet basis) | species.yr | Malaysia | | 3,5,6 | | 1 | 2 | | 4,6 | Terrestrial ecotoxicity |
| Ecotoxicity | -2.6E-9, -8.8E-11, -1.0E-10 | | -2.5E-9 | -1.4E-9 | | 6.6E-9, -7.7E-10 | Lin 2022 | 1 tonne of WF (wet basis) | species.yr | Malaysia | | 1,5,6 | | 2 | 3 | | 4,7 | Freshwater ecotoxicity |
| Ecotoxicity | -4.2E-6, 3.84E-8, -1.1E-7 | | -7.2E-7 | -2.0E-6 | | 1.5E-5, -1.2E-6 | Lin 2022 | 1 tonne of WF (wet basis) | species.yr | Malaysia | | 1,5,6 | | 4 | 2 | | 3,7 | Marine ecotoxicity |
| GWP | -045, -2.9E-5 | | | | | | Mayer 2020 | kg OF-MSW | kg CO ₂ eq | Germany | | 3 | | | | 2 | | Global Warming Potential |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|--------------|-----------------------|----------|---------------|-----------------------------------|------------------------|----------------|---|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|--|
| Ecotoxicity | -0.02 | | | | | - | Opatokun 2017 | 1 kg food waste | g 1,4 DB eq | Australia | | 1 | | | | | 2 | Terrestrial ecotoxicity |
| Eutrophication | -1.14, -0.71 | | | | -0.55, -0.06 | | Mayer 2020 | kg OF-MSW | kg Peq | Germany | | 1,2 | | | | 3,4 | | Freshwater eutrophication |
| Human Toxicity | -0.034, -0.039 | | | | -0.01, -0.016 | | Mayer 2020 | kg OF-MSW | kg 1,4 DCBeq | Germany | | 1,2 | | | | 3,4 | | Human toxicity |
| Ecotoxicity | -0.17 | | | | | 0.12 | Opatokun 2017 | 1 kg food waste | g 1,4 DB eq | Australia | | 1 | | | | | 2 | Freshwater ecotoxicity |
| Eutrophication | 3.1E-4 1.1E-4 | | | | 1.6E-4 1.7E-4 | | Mayer 2020 | kg OF-MSW | kg Neq | Germany | | 1,4 | | | | 2,3 | | Marine eutrophication |
| Particulate matter formation | 4.8E-5 5.8E-5 | | | | 5.5E-5 7.6E-5 | | Mayer 2020 | kg OF-MSW | kg PM ₁₀ eq | Germany | | 1,3 | | | | 2,4 | | Particulate Matter Formation Potential |
| Acidification | 2.0E-4 8.4E-5 | | | | 4.3E-5 8.9E-5 | | Mayer 2020 | kg OF-MSW | kg SO ₂ eq | Germany | | 2,4 | | | | 1,3 | | Terrestrial acidification |
| Ecotoxicity | -0.20 | | | | | 0.07 | Opatokun 2017 | 1 kg food waste | g 1,4 DB eq | Australia | | 1 | | | | | 2 | Marine ecotoxicity |
| Acidification | -0.23 | | -1.3 | 0.38 | -0.74 | 1.5 | Mondello 2017 | 1 tonne of food waste | kg SO ₂ eq | Italy | With avoided products | 3 | | 1 | 4 | 2 | 5 | Acidification |
| Eutrophication | -0.04 | | -0.19 | 0.07 | -0.14 | 0.75 | Mondello 2017 | 1 tonne of food waste | kg PO ₄ eq | Italy | With avoided products | 3 | | 1 | 4 | 2 | 5 | Eutrophication |
| GWP | -3.0E+2 | | -4.2E+2 | 59 | 4.9E+2 | 1.1E+3 | Mondello 2017 | 1 tonne of food waste | kg CO ₂ eq | Italy | With avoided products | 2 | | 1 | 3 | 4 | 5 | Global Warming Potential |
| Ecotoxicity | -4.8 | | | | 1.4 | 3.0 | Slorach 2019b | 1 tonne WF | g 1,4 DB eq | United Kingdom | | 1 | | | | 2 | 3 | Terrestrial ecotoxicity |
| Ecotoxicity | -5.2 | | | 3.8 | 1.2 | 2.9 | Slorach 2019a | 1 metric ton household food waste | g 1,4 DB eq | United Kingdom | | 1 | | | 4 | 2 | 3 | Terrestrial ecotoxicity |
| Ecotoxicity | -4.2 | | | -0.27 | -0.34 | | Slorach 2020 | 1 tonne of household food waste | g 1,4 DB eq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 1 | | | 3 | 2 | | Terrestrial ecotoxicity |
| Human Toxicity | -11 | | | | | 3.4 | Opatokun 2017 | 1 kg food waste | g 1,4 DBeq | Australia | | 1 | | | | | 2 | Human toxicity |
| Eutrophication | 0.002, 0.003 | | | 0.001, 0.001 | 6.2E-6 | 8.6E-6 | Morelli 2019 | 1 kg food waste | kg Neq | United States | | 4 | | | 3 | 1 | 2 | Eutrophication |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|--------------------------------|-----------------------|----------|---------------|------------------------------------|-------------------------|---------------|------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|------------------------------|
| GWP | -0.14, -0.03 | | | 0.10, -0.01, 0.07 | -0.02 | 0.32 | Morelli 2019 | 1 kg food waste | kg CO ₂ eq | United States | | 1 | | | 3 | 2 | 4 | Global Warming Potential |
| Acidification | -1.1E-4, 2.1E-4 | | | 1.2E-3, 5E-4, 4.9E-4 | 8.1E-5 | 1.4E-4 | Morelli 2019 | 1 kg food waste | kg SO ₂ eq | United States | | 1 | | | 4 | 2 | 3 | Acidification Potential |
| Particulate matter formation | 2.5E-5, 4.0E-7 | | | 2.8E-5, 2.8E-5, 7.4E-6, 7.4E-6 | 2.9E-6 | 7.5E-6 | Morelli 2019 | 1 kg food waste | kg PM _{2.5} eq | United States | | 1 | | | 4 | 2 | 3 | Particulate Matter Formation |
| GWP | -0.17 | | | -0.05 | | 0.52 | Morris 2014 | 1 kg food waste | kg CO ₂ eq | United States | | 1 | | | 2 | | 3 | Global Warming Potential |
| GWP | -0.17 | | | -0.05 | | 0.03 | Morris 2017 | 1 kg food waste | kg CO ₂ eq | United States | | 1 | | | 2 | | 3 | Global Warming Potential |
| GWP | | | | 4.5E+2 | | 6.5E+2 | Mu 2017 | 1 tonne fresh matter in food waste | kg CO ₂ eq | United States | | | | | 1 | | 2 | Global Warming Potential |
| Acidification | | | | 2.4 | | 0.47 | Mu 2017 | 1 tonne fresh matter in food waste | kg SO ₂ eq | United States | | | | | 2 | | 1 | Acidification |
| Eutrophication | | | | -4.4 | | 1.3 | Mu 2017 | 1 tonne fresh matter in food waste | kg Neq | United States | | | | | 1 | | 2 | Eutrophication Potential |
| Human Toxicity | | | | 4.0E-4 | | 1.2E-6 | Mu 2017 | 1 tonne fresh matter in food waste | CTUh | United States | | | | | 2 | | 1 | Human Toxicity, non-cancer |
| Human Toxicity | | | | 2.3E-5 | | 1.6E-7 | Mu 2017 | 1 tonne fresh matter in food waste | CTUh | United States | | | | | 2 | | 1 | Human Toxicity, cancer |
| Particulate matter formation | | | | 0.02 | | 0.12 | Mu 2017 | 1 tonne fresh matter in food waste | kg PM _{2.5} eq | United States | | | | | 1 | | 2 | Human Health (respiratory) |
| Ecotoxicity | | | | 2.5E+3 | | 13 | Mu 2017 | 1 tonne fresh matter in food waste | CTUe | United States | | | | | 2 | | 1 | Ecotoxicity |
| GWP | -70, -215, -260 | | | 1.1E+2 | | 3.5E+2 | Murphy 2006 | 1 metric ton food scraps | kg CO ₂ eq | Ireland | | 1,2,3 | | | 4 | | 5 | Global Warming Potential |
| GWP | -1.4E+8 | | | 1.9E+7 | -2.4E+7 | 5.7E+8 | Oldfield 2016 | 1,267,749 tonnes WF | kg CO ₂ eq | Ireland | | 1 | | | 3 | 2 | 4 | Global Warming Potential |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|------------------------------|-----------------------------------|------------------------------|----------------|-----------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|----------------------------|
| Acidification | 0 | | | 1.7E+6 | 4.0E+5 | 5.3E+5 | Oldfield 2016 | 1,267,749 tonnes WF | kg SO ₂ eq | Ireland | | 1 | | | 4 | 2 | 3 | Acidification Potential |
| Eutrophication | 0 | | | 4.7E+5 | 1.5E+5 | 1.5E+6 | Oldfield 2016 | 1,267,749 tonnes WF | kg PO ₄ eq | Ireland | | 1 | | | 3 | 2 | 4 | Eutrophication Potential |
| GWP | | | | 2 | 1 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Manfredi et al. 2011 | | | | 2 | 1 | | Global Warming Potential |
| GWP | | | | | 1 | 2 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Andersen et al 2012 | | | | | 1 | 2 | Global Warming Potential |
| GWP | | | | 1 | | 2 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Lundie et al. | | | | 1 | | 2 | Global Warming Potential |
| GWP | -7.6E+2 | | | | | 5.0E+2 | Opatokun 2017 | 1 kg food waste | g CO ₂ eq | Australia | | 1 | | | | | 2 | Global Warming Potential |
| Acidification | -1.3 | | | | | 0.08 | Opatokun 2017 | 1 kg food waste | g SO ₂ eq | Australia | | 1 | | | | | 2 | Terrestrial acidification |
| Eutrophication | -0.21 | | | | | 0.01 | Opatokun 2017 | 1 kg food waste | g Peq | Australia | | 1 | | | | | 2 | Freshwater eutrophication |
| Eutrophication | -2.9 | | | | | 2.8 | Opatokun 2017 | 1 kg food waste | g Neq | Australia | | 1 | | | | | 2 | Marine eutrophication |
| Human Toxicity | -33 | | 5.3 | 2.1 | -42 | 69 | Mondello 2017 | 1 tonne of food waste | kg 1,4 DBeq | Italy | With avoided products | 2 | | 4 | 3 | 1 | 5 | Human toxicity |
| Particulate matter formation | -0.39 | | | | | 0.03 | Opatokun 2017 | 1 kg food waste | g PM ₁₀ eq | Australia | | 1 | | | | | 2 | Particle matter formation |
| Ecotoxicity | 0.68 | | 3.8 | 7.4 | -27 | 42 | Mondello 2017 | 1 tonne of food waste | kg 1,4 DBeq | Italy | With avoided products | 2 | | 3 | 4 | 1 | 5 | Freshwater ecotoxicity |
| Ecotoxicity | -1.7E+3 | | 1.2E+4 | 1.6E+4 | -6.6E+4 | 1.0E+5 | Mondello 2017 | 1 tonne of food waste | kg 1,4 DBeq | Italy | With avoided products | 2 | | 3 | 4 | 1 | 5 | Marine ecotoxicity |
| Ecotoxicity | -0.15 | | 0.06 | - | -0.98 | 0.52 | Mondello 2017 | 1 tonne of food waste | kg 1,4 DBeq | Italy | With avoided products | 2 | | 4 | 3 | 1 | 5 | Terrestrial ecotoxicity |
| GWP | -9.4E+2 | | | 3.6E+2 | | 2.2E+3 | Parry 2012 | 3930 tons food waste | short ton CO ₂ eq | United States | only non-biogenic | 1 | | | 2 | | 3 | Global Warming Potential |
| GWP | 38 | 2.1 | 40 | 2.8E+2 | | | Salemdeeb 2017 | 1 metric ton municipal food waste | kg CO ₂ eq | United Kingdom | | 2 | 1 | 3 | 4 | | | Global Warming Potential |
| Human Toxicity | 1.2E-4 | -1.0E-4 | -9.4E-5 | 1.2E-4 | | | Salemdeeb 2017 | 1 metric ton municipal food waste | CTU | United Kingdom | | 3 | 1 | 2 | 3 | | | Human Toxicity, non-cancer |
| Eutrophication | 1.9 | -1.7 | -1.4 | 1.9 | | | Salemdeeb 2017 | 1 metric ton | kg Neq | United Kingdom | | 3 | 1 | 2 | 3 | | | Eutrophication, marine |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|--|---|-------------------------|----------------|-------------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|--|
| | | | | | | | | municipal food waste | | | | | | | | | | |
| Ecotoxicity | 3.0E+2 | -2.8E+2 | -2.2E+2 | 3.0E+2 | | | Salemdeeb 2017 | 1 metric ton municipal food waste | CTU | United Kingdom | | 4 | 1 | 2 | 3 | | | Ecotoxicity |
| Acidification | 2.1 | -1.0 | -0.65 | 1.6 | | | Salemdeeb 2017 | 1 metric ton municipal food waste | Accumulated Exceedance | United Kingdom | | 4 | 1 | 2 | 3 | | | Acidification |
| Human Toxicity | 9.3E-7 | -2.8E-7 | -1.3E-7 | 1.1E-6 | | | Salemdeeb 2017 | 1 metric ton municipal food waste | CTU | United Kingdom | | 3 | 1 | 2 | 4 | | | Human Toxicity, cancer |
| Eutrophication | 0.03 | -0.03 | -0.02 | 0.03 | | | Salemdeeb 2017 | 1 metric ton municipal food waste | kg Peq | United Kingdom | | 3 | 1 | 2 | 4 | | | Eutrophication, freshwater |
| Particulate matter formation | 0.09 | -0.05 | -0.02 | 0.08 | | | Salemdeeb 2017 | 1 metric ton municipal food waste | kg PM _{2.5} eq | United Kingdom | | 4 | 1 | 2 | 3 | | | Particulate Matter Formation Potential |
| Eutrophication | 9.5 | -4.4 | -3.3 | 7.0 | | | Salemdeeb 2017 | 1 metric ton municipal food waste | Accumulated Exceedance | United Kingdom | | 4 | 1 | 2 | 3 | | | Eutrophication, terrestrial |
| GWP | -1.9E+2 | | | | -83 | | Bernstad Saraiva Schott and Andersson 2015 | metric ton currently generated food waste | kg CO ₂ eq | Sweden | | 1 | | | | 2 | | Global Warming Potential |
| GWP | 1 | | | | | 2 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Sanscartier et al. 2012 | 1 | | | | | 2 | Global Warming Potential |
| GWP | 1 | | | 2 | | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Colon et al. 2012 | 1 | | | 2 | | | Global Warming Potential |
| GWP | | | | | 1 | 2 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Manfredi et al. 2011 | | | | | 1 | 2 | Global Warming Potential |
| GWP | | | | 2 | 1 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Andersen et al 2012 | | | | 2 | 1 | | Global Warming Potential |
| GWP | 2 | | | 1 | | | Bernstad Saraiva Schott 2016 | Food waste | relative ranking | Various | Lundie et al. | 2 | | | 1 | | | Global Warming Potential |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|--------------------------|---------------------|----------|----------|---------|-----------------------|----------|------------------------------|-----------------------|------------------|-----------|-----------------------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|--------------------------|
| | | | | | | | | management | | | | | | | | | | |
| GWP | 2 | | | | 3 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Assefa et al | 1 | | | | 2 | | Global Warming Potential |
| GWP | 1 | | | 2 | | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Boldrin et al 2011 | 1 | | | 2 | | | Global Warming Potential |
| GWP | 2 | | | 3 | | 1 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Kong et al. 2012 | 2 | | | 3 | | 1 | Global Warming Potential |
| GWP | 2 | | | | 1 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Kirkeby et al. 2006 | 2 | | | | 1 | | Global Warming Potential |
| GWP | | | | 1 | | 2 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Kim and Kim, 2010 | | | | 1 | | 2 | Global Warming Potential |
| GWP | 1 | | | 2 | | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Bernstad and la Cour Jansen, 2001 | 1 | | | 2 | | | Global Warming Potential |
| GWP | 2 | | | | 1 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Fruergaard and Astrup 2010 | 2 | | | | 1 | | Global Warming Potential |
| GWP | 1 | | | 2 | | 3 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Aye and Widjaya, 2005 | 1 | | | 2 | | 3 | Global Warming Potential |
| GWP | 2 | | | 1 | | 3 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Blengini et al. 2008 | 2 | | | 1 | | 3 | Global Warming Potential |
| GWP | 1 | | | 2 | 3 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Diggelman and Ham 2003 | 1 | | | 2 | 3 | | Global Warming Potential |
| GWP | 1 | | | 2 | 3 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Khoo et al. 2010 | 1 | | | 2 | 3 | | Global Warming Potential |
| GWP | | | | 1 | 2 | | Bernstad Saraiva Schott 2016 | Food waste | relative ranking | Various | Lee et al. 2007; if | | | | 1 | 2 | | Global Warming Potential |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|--------------------------|---------------------|----------|----------|---------|-----------------------|----------|------------------------------|-----------------------------------|-----------------------|----------------|--|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|---------------------------------|
| | | | | | | | | management | | | biogenic CO ₂ is included | | | | | | | |
| GWP | | | | 2 | 1 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Lee et al. 2007; if biogenic CO ₂ is included | | | | 2 | 1 | | Global Warming Potential |
| GWP | 1 | | | | | 2 | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Hamelin et al. 2013 | 1 | | | | | 2 | Global Warming Potential |
| Acidification | -0.13 | | | 0.47 | -0.64 | 1.5 | Shih 2021 | 1 metric ton kitchen waste | kg SO ₂ eq | Taiwan | | 2 | | | 3 | 1 | 4 | Acidification |
| Eutrophication | -0.03 | | | 0.11 | -0.12 | 0.83 | Shih 2021 | 1 metric ton kitchen waste | kg PO ₄ eq | Taiwan | | 2 | | | 3 | 1 | 4 | Eutrophication |
| GWP | -2.9E+2 | | | 63 | 4.9E+2 | 1.2E+3 | Shih 2021 | 1 metric ton kitchen waste | kg CO ₂ eq | Taiwan | | 1 | | | 2 | 3 | 4 | Global Warming |
| Human Toxicity | -29 | | | 2.8 | -41 | 71 | Shih 2021 | 1 metric ton kitchen waste | kg 1,4 DBeq | Taiwan | | 2 | | | 3 | 1 | 4 | Human Toxicity |
| Ecotoxicity | 0.82 | | | 7.9 | -27 | 43 | Shih 2021 | 1 metric ton kitchen waste | kg 1,4 DBeq | Taiwan | | 2 | | | 3 | 1 | 4 | Fresh Water Aquatic Ecotoxicity |
| Ecotoxicity | -1.7E+3 | | | 1.6E+4 | -6.6E+4 | 1.0E+5 | Shih 2021 | 1 metric ton kitchen waste | kg 1,4 DBeq | Taiwan | | 2 | | | 3 | 1 | 4 | Marine Aquatic Ecotoxicity |
| Ecotoxicity | -0.08 | | | 0.07 | -0.87 | 0.64 | Shih 2021 | 1 metric ton kitchen waste | kg 1,4 DBeq | Taiwan | | 2 | | | 3 | 1 | 4 | Terrestrial Ecotoxicity |
| GWP | -39 | | | | -10 | 1.9E+2 | Slorach 2019b | 1 tonne WF | kg CO ₂ eq | United Kingdom | | 1 | | | | 2 | 3 | Global Warming Potential |
| Human Toxicity | -22 | | | | -0.70 | 4.3 | Slorach 2019b | 1 tonne WF | kg 1,4 DBeq | United Kingdom | | 1 | | | | 2 | 3 | Human toxicity |
| Eutrophication | -14 | | | | 2.3 | 3.3 | Slorach 2019b | 1 tonne WF | g Peq | United Kingdom | | 1 | | | | 3 | 2 | Freshwater eutrophication |
| Acidification | 7.6 | | | | 0.40 | 0.20 | Slorach 2019b | 1 tonne WF | kg SO ₂ eq | United Kingdom | | 3 | | | | 2 | 1 | Terrestrial acidification |
| Ecotoxicity | -3.3 | | | | -1.5 | -0.04 | Slorach 2019b | 1 tonne WF | kg 1,4 DBeq | United Kingdom | | 1 | | | | 2 | 3 | Freshwater ecotoxicity |
| Ecotoxicity | -3.0 | | | | -1.4 | -0.10 | Slorach 2019b | 1 tonne WF | kg 1,4 DBeq | United Kingdom | | 1 | | | | 2 | 3 | Marine ecotoxicity |
| Ecotoxicity | -3.9 | | | 1.4 | -1.9 | -0.18 | Slorach 2019a | 1 metric ton household food waste | kg 1,4 DBeq | United Kingdom | | 1 | | | 4 | 2 | 3 | Freshwater ecotoxicity |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|---------------|-----------------------------------|------------------------|----------------|---|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|------------------------------|
| Eutrophication | 0.70 | | | | 0.10 | 7.4 | Slorach 2019b | 1 tonne WF | kg Neq | United Kingdom | | 2 | | | | 1 | 3 | Marine eutrophication |
| Particulate matter formation | 0.99 | | | | 0.17 | 0.10 | Slorach 2019b | 1 tonne WF | kg PM ₁₀ eq | United Kingdom | | 3 | | | | 2 | 1 | Particulate matter formation |
| GWP | -32 | | | 78 | -5.0 | 2.0E+2 | Slorach 2019a | 1 metric ton household food waste | kg CO ₂ eq | United Kingdom | | 1 | | | 3 | 2 | 4 | Global Warming Potential |
| Ecotoxicity | -3.5 | | | 1.3 | -1.7 | -0.23 | Slorach 2019a | 1 metric ton household food waste | kg 1,4 DBeq | United Kingdom | | 1 | | | 4 | 2 | 3 | Marine ecotoxicity |
| Human Toxicity | -22 | | | 4.5 | -0.19 | 4.5 | Slorach 2019a | 1 metric ton household food waste | kg 1,4 DBeq | United Kingdom | | 1 | | | 3 | 2 | 4 | Human toxicity |
| Ecotoxicity | -0.58 | | | -0.13 | -0.25 | | Slorach 2020 | 1 tonne of household food waste | kg 1,4 DBeq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 1 | | | 3 | 2 | | Freshwater ecotoxicity |
| Ecotoxicity | -0.55 | | | -0.12 | -0.23 | | Slorach 2020 | 1 tonne of household food waste | kg 1,4 DBeq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 1 | | | 3 | 2 | | Marine ecotoxicity |
| Eutrophication | -12 | | | -0.88 | 4.8 | 2.8 | Slorach 2019a | 1 metric ton household food waste | g Peq | United Kingdom | | 1 | | | 2 | 4 | 3 | Freshwater Eutrophication |
| Eutrophication | 0.72 | | | 0.39 | 0.12 | 7.4 | Slorach 2019a | 1 metric ton household food waste | kg Neq | United Kingdom | | 3 | | | 2 | 1 | 4 | Marine Eutrophication |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|---------------|---|------------------------|----------------|---|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|------------------------------|
| Acidification | 7.6 | | | 10 | 0.39 | 0.24 | Slorach 2019a | 1 metric ton household food waste | kg SO ₂ eq | United Kingdom | | 3 | | | 4 | 2 | 1 | Terrestrial Acidification |
| Particulate matter formation | 1.0 | | | 1.4 | 0.18 | 0.11 | Slorach 2019a | 1 metric ton household food waste | kg PM ₁₀ eq | United Kingdom | | 3 | | | 4 | 2 | 1 | Particulate Matter Formation |
| GWP | 9.6 | | | 44 | 42 | | Slorach 2020 | 1 tonne of household food waste | kg CO ₂ eq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 1 | | | 3 | 2 | | Global warming potential |
| Human Toxicity | -8.1 | | | -0.09 | 0.33 | | Slorach 2020 | 1 tonne of household food waste | kg 1,4 DBeq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 1 | | | 2 | 2 | | Human toxicity |
| Human Toxicity | -21 | | | | | -21 | Xu 2015 | Management of 1 tonne of Food Waste Volatile Solids | kg 1,4 DBeq | China | | 1 | | | | | 2 | Human toxicity |
| Ecotoxicity | -0.42 | | | | | -0.37 | Xu 2015 | Management of 1 tonne of Food Waste Volatile Solids | kg 1,4 DBeq | China | | 1 | | | | | 2 | Freshwater ecotoxicity |
| Ecotoxicity | -0.43 | | | | | -0.37 | Xu 2015 | Management of 1 tonne of Food Waste Volatile Solids | kg 1,4 DBeq | China | | 1 | | | | | 2 | Marine ecotoxicity |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|---------------|---------------------------------|------------------------|----------------|---|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|------------------------------|
| Eutrophication | -4.3 | | | 2.1 | 6.1 | | Slorach 2020 | 1 tonne of household food waste | g Peq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 1 | | | 2 | 3 | | Freshwater eutrophication |
| Eutrophication | 0.73 | | | 0.64 | 0.92 | | Slorach 2020 | 1 tonne of household food waste | kg Neq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 2 | | | 1 | 3 | | Marine eutrophication |
| Acidification | 7.5 | | | 5.1 | 1.9 | | Slorach 2020 | 1 tonne of household food waste | kg SO ₂ eq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 3 | | | 2 | 1 | | Terrestrial acidification |
| Particulate matter formation | 0.99 | | | 0.51 | 0.36 | 1.0 | Slorach 2020 | 1 tonne of household food waste | kg PM ₁₀ eq | United Kingdom | Scenarios include mix of management practices. AD = Scenario 4 (92%), Incineration = Scenario 1 (74%), Compost = scenario 3, Landfill = Current/BAU | 3 | | | 2 | 1 | | Particulate matter formation |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|--------------------------|--------------------------|----------|----------|---------|-----------------------|----------|------------------------------|---|-----------------------|-----------|-----------------------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|---------------------------|
| GWP | 1 | | | | 2 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Bernstad and la Cour Jansen, 2001 | 1 | | | | 2 | | Global Warming Potential |
| GWP | 1 | | | | 2 | | Bernstad Saraiva Schott 2016 | Food waste management | relative ranking | Various | Hamelin et al. 2013 | 1 | | | | 2 | | Global Warming Potential |
| GWP | -72.8, -238, -81.4, -137 | | | | -29 | | Tian 2021 | Disposal of 1-ton WF generated from local food centre | kg CO ₂ eq | Singapore | | 1 | | | | 2 | | GWP |
| Eutrophication | 0.01 | | | | 2.0E-3 | | Tian 2021 | Disposal of 1-ton WF generated from local food centre | kg Peq | Singapore | | 2 | | | | 1 | | Freshwater eutrophication |
| Eutrophication | 0.97 | | | | 1.5E-3 | | Tian 2021 | Disposal of 1-ton WF generated from local food centre | kg Neq | Singapore | | 2 | | | | 1 | | Marine eutrophication |
| Acidification | 0.28, 0.12, 0.31, -0.1 | | | | 0.09 | | Tian 2021 | Disposal of 1-ton WF generated from local food centre | kg SO ₂ eq | Singapore | | 2 | | | | 1 | | Terrestrial acidification |
| Acidification | -169, -26 | | | | -86 | | Tong 2018 | 1000 tonnes food waste | kg SO ₂ eq | Singapore | | 1,3 | | | | 2 | | Acidification |
| Eutrophication | 40, 73 | | | | 35 | | Tong 2018 | 1000 tonnes food waste | kg PO ₄ eq | Singapore | | 2,3 | | | | 1 | | Eutrophication |
| Ecotoxicity | 7.2E-3, 3.9E-3 | | | | 1.0E-2, 9.7E-3 | | Mayer 2020 | kg OF-MSW | kg 1,4 DCBeq | Germany | | 1,2 | | | | 3,4 | | Freshwater ecotoxicity |
| GWP | 94100, 92500 | | | | 1.1E+5 | | Tong 2018 | 1000 tonnes food waste | kg CO ₂ eq | Singapore | | 1,2 | | | | 2 | | Global Warming Potential |
| Ecotoxicity | 0.007, 0.004 | | | | 0.01, 0.009 | | Mayer 2020 | kg OF-MSW | kg 1,4 DCBeq | Germany | | 1,2 | | | | 3,4 | | Marine ecotoxicity |
| Ecotoxicity | 6.7E-5, 6.2E-6 | | | | 6.4E-6, 8.2E-6 | | Mayer 2020 | kg OF-MSW | kg 1,4 DCBeq | Germany | | 1,4 | | | | 2,3 | | Terrestrial ecotoxicity |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|------------------------------|---------------------|----------|----------|---------|-----------------------|----------|----------------|---|------------------------|---------------|---------------------------------------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|------------------------------|
| Ecotoxicity | -11300, -1960 | | | | -8.9E+2 | | Tong 2018 | 1000 tonnes food waste | kg DCBeq | Singapore | | 1,2 | | | | 3 | | Freshwater ecotoxicity |
| GWP | -0.14 | | | -0.12 | -0.13 | 0.50 | U.S. EPA 2020b | short ton of food waste | MT CO ₂ eq | United States | wet AD, direct application | 1 | | | 3 | 2 | 4 | Global Warming Potential |
| GWP | -0.06 | | | -0.12 | -0.13 | 0.50 | U.S. EPA 2020b | short ton of food waste | MT CO ₂ eq | United States | alternate ranking. Wet AD with curing | 3 | | | 2 | 1 | 4 | Global Warming Potential |
| GWP | -0.10 | | | -0.12 | -0.13 | 0.50 | U.S. EPA 2020b | short ton of food waste | MT CO ₂ eq | United States | Dry AD Direct application | 3 | | | 2 | 1 | 4 | Global Warming Potential |
| GWP | -0.04 | | | -0.12 | -0.13 | 0.50 | U.S. EPA 2020b | short ton of food waste | MT CO ₂ eq | United States | Dry AD with Curing | 3 | | | 2 | 1 | 4 | Global Warming Potential |
| GWP | -3.7E+2 | | | | | -3.3E+2 | Xu 2015 | Management of 1 tonne of Food Waste Volatile Solids | kg CO ₂ eq | China | | 1 | | | | | 2 | Global Warming Potential |
| Human Toxicity | -25500, -4470 | | | | -1.3E+2 | | Tong 2018 | 1000 tonnes food waste | kg DCBeq | Singapore | | 1,2 | | | | 3 | | Human toxicity |
| Particulate matter formation | -0.66 | | | | | -0.48 | Xu 2015 | Management of 1 tonne of Food Waste Volatile Solids | kg PM ₁₀ eq | China | | 1 | | | | | 2 | Particulate matter formation |
| Acidification | -1.8 | | | | | -1.4 | Xu 2015 | Management of 1 tonne of Food Waste Volatile Solids | kg SO ₂ eq | China | | 1 | | | | | 2 | Terrestrial acidification |
| Eutrophication | -0.01 | | | | | -9.4E-3 | Xu 2015 | Management of 1 tonne of Food Waste Volatile Solids | kg Peq | China | | 1 | | | | | 2 | Freshwater eutrophication |
| Eutrophication | -0.36 | | | | | -0.19 | Xu 2015 | Management of 1 tonne of Food Waste Volatile Solids | kg Neq | China | | 1 | | | | | 2 | Marine eutrophication |

| Standard Impact Category | Anaerobic Digestion | Wet Feed | Dry Feed | Compost | Controlled combustion | Landfill | Source Report | Functional Unit | Units | Geography | Note | Anaerobic Digestion, Rank | Wet Feed, Rank | Dry Feed, Rank | Compost, Rank | Controlled combustion, Rank | Landfill, Rank | Original impact category |
|--------------------------|---------------------|----------|----------|---------|-----------------------|----------|---------------|----------------------------|-----------------------|-----------|------|---------------------------|----------------|----------------|---------------|-----------------------------|----------------|--------------------------|
| Ecotoxicity | -4.1E7 -9.4E6 | | | | -4.0E+6 | | Tong 2018 | 1000 tonnes food waste | kg DCBeq | Singapore | | 1,2 | | | | 3 | | Marine ecotoxicity |
| Ecotoxicity | -336, 44.8 | | | | 65 | | Tong 2018 | 1000 tonnes food waste | kg DCBeq | Singapore | | 1,2 | | | | 3 | | Terrestrial ecotoxicity |
| Acidification | 0.20 | | | | 0.50 | 0.34 | Zhang 2019 | 1 metric ton of food waste | kg SO ₂ eq | China | | 1 | | | | 3 | 2 | Acidification |
| Eutrophication | 0.11 | | | | 0.10 | 0.30 | Zhang 2019 | 1 metric ton of food waste | kg Peq | China | | 2 | | | | 1 | 3 | Eutrophication |
| Ecotoxicity | 0.34 | | | | 0.03 | 0.02 | Zhang 2019 | 1 metric ton of food waste | kg DCBeq | China | | 3 | | | | 2 | 1 | Freshwater ecotoxicity |
| GWP | 2.7E+2 | | | | 7.6E+2 | 5.9E+2 | Zhang 2019 | 1 metric ton of food waste | kg CO ₂ eq | China | | 1 | | | | 3 | 2 | Global Warming Potential |
| Human Toxicity | 0.60 | | | | 1.6 | 0.32 | Zhang 2019 | 1 metric ton of food waste | kg DCBeq | China | | 2 | | | | 3 | 1 | Human toxicity |
| Ecotoxicity | 53 | | | | 67 | 4.9 | Zhang 2019 | 1 metric ton of food waste | kg DCBeq | China | | 2 | | | | 3 | 1 | Marine ecotoxicity |
| Ecotoxicity | 0.10 | | | | 0.22 | 0.01 | Zhang 2019 | 1 metric ton of food waste | kg DCBeq | China | | 2 | | | | 3 | 1 | Terrestrial ecotoxicity |
| GWP | -37 | | | 89 | | | Zhou 2022 | one ton of WF | kg CO ₂ eq | China | | 1 | | | 2 | | | GWP |
| Acidification | 0.10 | | | 0.61 | | | Zhou 2022 | one ton of WF | kg SO ₂ eq | China | | 1 | | | 2 | | | Acidification |
| Eutrophication | 0.24 | | | 1.6 | | | Zhou 2022 | one ton of WF | kg NO ₃ eq | China | | 1 | | | 2 | | | Nutrient Enrichment |
| GWP | -0.18 | -0.17 | | 0.05 | 0.67 | 0.63 | ReFED 2023 | Ton of food waste | MT CO ₂ eq | Various | | 1 | 2 | | 3 | 5 | 4 | GWP |
| GWP | -0.07 | -0.09 | | 0.03 | 0.90 | 0.30 | ReFED 2023 | Ton of food waste | MT CO ₂ eq | Various | | 2 | 1 | | 3 | 5 | 4 | GWP |
| GWP | -0.22 | -0.20 | | 0.05 | 0.58 | 0.76 | ReFED 2023 | Ton of food waste | MT CO ₂ eq | Various | | 1 | 2 | | 3 | 4 | 5 | GWP |
| GWP | -0.18 | -0.19 | | 0.05 | 0.66 | 0.65 | ReFED 2023 | Ton of food waste | MT CO ₂ eq | Various | | 2 | 1 | | 3 | 5 | 4 | GWP |

APPENDIX E

OTHER EMERGING PATHWAYS

The following sections provide a discussion of emerging pathways for wasted food management. Given that these technologies are in development, they were not included in the main analysis.

E-1. Hydrothermal Carbonization

Hydrothermal carbonization (HTC) is the conversion of watery slurry of organic material into carbon-rich hydrochar under moderate temperature and pressure (Mayer et al. 2019). The HTC process can operate with a variety of wet organic feedstocks including wasted food, green waste, digestate and the organic fraction of MSW (OFMSW) (Owsianiak et al. 2016). In the HTC process wasted food can be co-treated with non-wasted food (e.g., packaging), but the presence of packaging material in the waste stream has been found to drive environmental impacts (Berge et al. 2015).

Figure E-1 shows a system diagram for the HTC wasted food management pathway. Pre-treatment steps include moisture adjustment, sorting, mixing/shredding and chemical addition. Hydrochar is cooled, dewatered and combusted in a boiler or CHP system to recover heat and electrical energy. Liquid effluent resulting from the dewatering process may require advanced wastewater treatment due to high nutrient content or the presence of contaminants (Mayer et al. 2019).

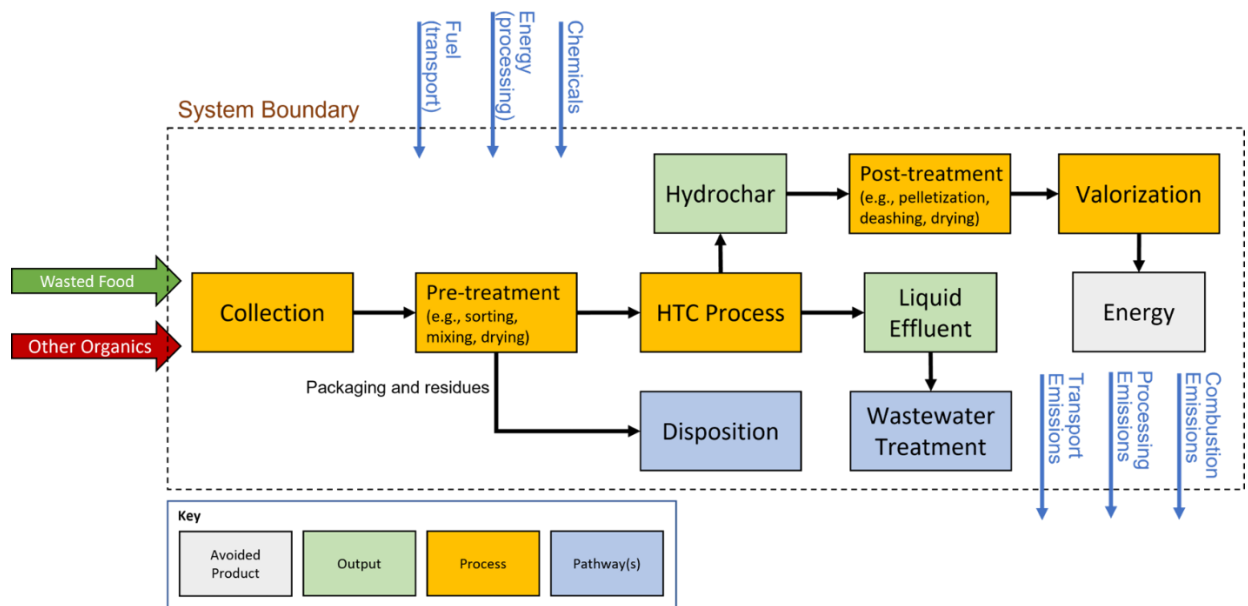


FIGURE E-1. DIAGRAM OF WASTED FOOD HYDROTHERMAL CARBONIZATION

E-1-1 Process Operation

During HTC, food waste is fed into a reactor where heat (around 180–250°C) and pressure (around 10-20 bar) remove moisture from the biomass (Owsianiak et al. 2016). Despite the temperature and pressure requirements, Berge et al. (2015) found that energy demand for the provision of heat had a “fairly

negligible impact on the system". An LCA of wasted food HTC found that operation of the HTC process contributed minimally to gross environmental impact across 15 impact categories (Owsianiak et al. 2016). Pumping of wet biowaste and drying and pelletization of hydrochar was analyzed separately from reactor operation and was a larger source of environmental impact (Owsianiak et al. 2016).

E-1-2 Process Emissions

The main products of HTC are hydrochar, non condensible gases leading to process air emissions (SO₂, NO_x, CO, CO₂, etc.) and water containing inorganic and organic compounds including trace metals such as chromium, arsenic, nickel, mercury and cadmium (Owsianiak et al. 2016). The HTC process can accept wasted food with moderate amounts of packaging contamination. Global warming impact of the process was found to increase significantly when 40% of incoming feedstock mass was packaging, due to the presence of fossil carbon in these materials (e.g., plastic packaging) (Berge et al. 2015).

E-1-3 Liquid Effluent Treatment

Hydrochar dewatering leads to the production of liquid effluent. This liquid may require treatment prior to discharge. Advanced treatment processes, such as reverse osmosis can be used to concentrate the effluent, which is then diluted to reduce its metal concentration for safe use as a fertilizer (Owsianiak et al. 2016).

E-1-4 Hydrochar Valorization

Hydrochar can be combusted to produce energy after pre-processing. The first stage of the hydrochar preparation process involves the separation of liquids and solids from the carbonization process in HTC. Afterwards, the solids are dried and any water remaining in the hydrochar must be evaporated before the hydrochar can be burned for energy. Global warming potential was found to be sensitive to the electricity demand of hydrochar drying (Berge et al. 2015). Hydrochar is a coal-like material that is typically transported to coal-fired power plants to substitute coal-based electricity. Hydrochar combustion has been shown to be a large contributor to smog and terrestrial eutrophication impact, but it was found that avoided energy production produced environmental benefits for the GWP and acidification impact categories. Acidification and GWP impact were found to be sensitive to the electrical efficiency of hydrochar energy recovery (Berge et al. 2015).

TABLE E-1. HTC: ENVIRONMENTAL DRIVERS

| | |
|--|---|
| Source of Environmental Impact | Environmental impact may be driven by packaging materials in wasted food. |
| Source of Environmental Impact | Hydrochar drying has high energy demand. |
| Source of Environmental Impact | Hydrochar combustion emits CAPs. |
| Source of Environmental Impact | Wastewater treatment of liquid effluent. |
| Source of Environmental Benefit | Hydrochar combustion displaces production and use of fossil energy. |
| Source of Environmental Benefit | Use of liquid effluent as fertilizer displaces conventional fertilizer production. |
| Options to Reduce Impact | Optimize hydrochar energy recovery. |
| Options to Reduce Impact | Additional pre-treatment sorting to remove packaging material from incoming wasted food and reduce global warming impact. |
| Additional Considerations | Environmental impacts associated with liquid-phase emissions decrease with the increasing presence of packaging material. ⁶⁵ |

E-2. Pyrolysis, Torrefaction and Gasification

Pyrolysis, torrefaction and gasification are all thermochemical processes where organic materials are converted in a controlled, low-oxygen environment to biochar (torrefaction), bio-oil (pyrolysis), synthesis gas (gasification), and associated gases and tar. The resulting biofuels can be further processed to produce gasoline, diesel, naphtha and refined industrial chemicals (U. Lee et al. 2017). In the case of gasification, the initial pyrolytic gases are further reacted with a gasifying agent to produce syngas (Mayer et al. 2019). Torrefaction is a mild pyrolytic process in which biochar is the primary energy product.

Commercial-scale application of pyrolysis, torrefaction, and gasification for wasted food management is nascent. Globally, there are facilities specializing in MSW gasification (Seo et al. 2018, 2), which is expected to include wasted food.

| Pyrolysis, Torrefaction, and Gasification: Associated Process Steps | | |
|---|-------------|--|
| Types | Allothermal | Use of outside fuel sources to meet the heat requirement. |
| | Autothermal | Combustion of internal feedstock to meet the heat requirement. |

Figure E-2 shows a system diagram for the wasted food pathways centered on pyrolysis, torrefaction and gasification. Potential pre-treatment steps include drying, shredding and chemical addition (Mayer et al. 2019). Bio-oil and synthesis gas (syngas) are the primary outputs of the pyrolysis and gasification management pathways and can be used either for energy production or undergo further processing to replace fossil-based feedstocks in the chemical industry. Produced fuel gas can be flared or used to provide process heat or other energy products. Biochar is a high carbon solid byproduct and can be used as a soil amendment (replacing peat), combusted to produce energy or disposed of in a landfill. Biochar can also be used as an additive in cement mortar to improve strength.

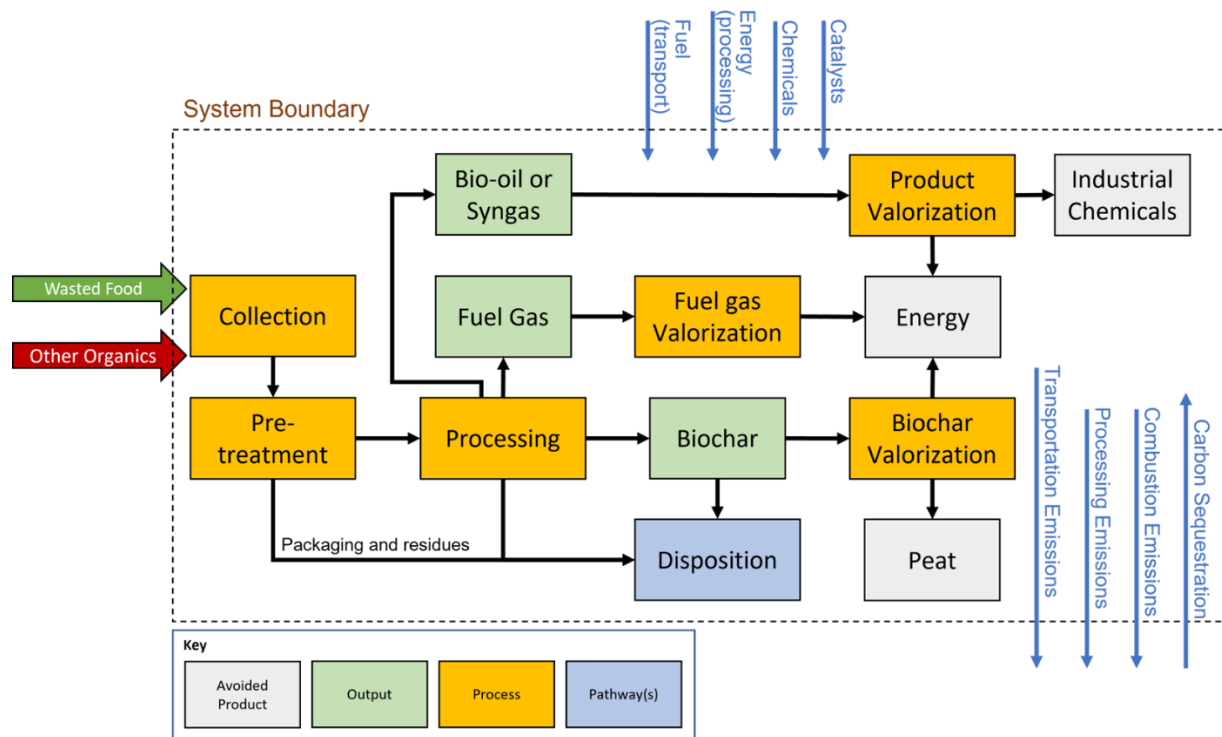


FIGURE E-2. DIAGRAM OF WASTED FOOD PYROLYSIS, TORREFACTION OR GASIFICATION

E-2-1 Process Operation

Torrefaction, gasification and pyrolysis are similar thermochemical processes that produce bio-fuels in low oxygen environments. During torrefaction, wasted food is introduced into a furnace and the heating temperature and nitrogen flowrate are adjusted to maximize biochar and energy output (Pahla et al. 2017).

Gasification involves the partial oxidation of food waste at high temperatures in gasifiers with a gasifying agent such as air, steam or oxygen until the organic matter is converted into syngas, a mixture of H₂, CO, CO₂, water, and hydrocarbons (Trabold and Babbitt 2018). Prior to the gasification process, Wasted Food requires processing to the designated particle size through methods such as sorting, shredding, grinding, drying and pelletization. These processes require additional energy use which contribute to the overall environmental impact of gasification (Nayak and Bhushan 2019).

Pyrolysis occurs at lower temperatures than gasification and requires heating wasted food without an oxidizing agent. This process creates biochar, bio-oils and gaseous products (Trabold and Babbitt 2018).⁶⁶

E-2-2 Process Emissions

Pyrolysis produces non-condensable gaseous products such as H₂, CO, CO₂, methane, and C₂H₆. These emissions contribute to the environmental impact of the pyrolysis process. Gasification emissions include ash, tars, unburnt char, alkali, metal compounds, S, N, carbonyl sulfide, ammonia and hydrogen cyanide, which contribute to the overall life cycle impact of gasification (Trabold and Babbitt 2018).

E-2-3 Bio-oil and Syngas Valorization

Pyrolysis bio-oil and gasification syngas can be transformed into biofuels by removing oxygen through catalytic treatment and hydrogenation. Syngas produced by gasification is frequently used for heat generation and stationary power. The produced syngas from gasification must be cleaned before being used for heat generation or power. Syngas can undergo further processing (e.g., catalytic conversion) to manufacture liquid fuels or chemical intermediates. Bio-oil (e.g., naphtha) can also be used as a feedstock or intermediary in the chemical and flavor industry (Trabold and Babbitt 2018).

E-2-4 Biochar Use and Disposal

Biochar produced by pyrolysis, torrefaction and gasification is an excellent soil amendment providing a variety of benefits such as acidity mediation, cation exchange capacity and water and nutrient retention. Biochar improves soil and crop health and boosts crop yields by immobilizing some organic pollutants and trace metals (Xiong et al. 2019). Biochar can also be ground into powder and used in cement mortar to adsorb water and improve strength (Xiong et al. 2019).

**TABLE E-2. PYROLYSIS, TORREFACTION AND GASIFICATION:
ENVIRONMENTAL DRIVERS**

| | |
|--|--|
| Source of Environmental Impact | Air emissions from processing (e.g., fuel gas). |
| Source of Environmental Impact | Process energy demand. |
| Source of Environmental Impact | Landfill or incineration of biochar. |
| Source of Environmental Impact | Waste management (e.g. ash from gasification). |
| Source of Environmental Benefit | Use of biochar for beneficial purposes such as a soil amendment or cement mixture additive. |
| | Fuel gas can be recovered and used for internal energy purposes. Excess fuel gas can also be recovered and sold for energy purposes. |

| | |
|---------------------------------|--|
| Options to Reduce Impact | Optimize recovery of biochar for beneficial use. |
| Other Considerations | Processing conditions (e.g., temperature, residence time) impact bio-product yields. |

E-3. Other Industrial Uses

For the understudied industrial use pathways (pyrolysis, gasification, HTC, etc.), it is not currently possible to draw robust conclusions about their environmental performance relative to other wasted food pathways. This is due both to the limited amount of research that considers these pathways and the wide range of technologies that fall in this broad category.

- For the pyrolysis and gasification pathways, no impact data were identified that met the basic scope criteria.
- Berge et al. (2015) is the only source of impact data for HTC that met basic scope criteria. Data from this study indicates acidification and GWP benefits that place HTC in the middle of other end-of-life pathways.
- More research is needed for these pathways as they are brought to commercial scale and pursue operational efficiency to optimize environmental performance in the lower end of the identified range.

Given the wide breadth of bio-based products that could ultimately be produced from the industrial use pathways, a wide range of performance outcomes are to be expected. The primary benefit of the industrial use pathways is their potential to avoid the production of non-renewable and resource intensive products such as electricity, diesel and virgin plastic resins. However, these pathways are also resource intensive and the realized net benefit depends on the balance of production and avoided product impacts.

E-3-1 Rendering

Rendering is the process of processing and heat treating unused portions of livestock carcasses and used cooking oil to produce fat and a protein-rich solids. Thus, this pathway applies to specific wasted food streams.

Figure E-3 shows a system diagram for rendering. The central component of the rendering process is heat treatment (e.g. at temperatures $>115^{\circ}\text{C}$ for >40 minutes). Heat treatment yields two main streams, one of proteinaceous solids, and the other of fatty liquid. Following additional processing of these two streams, major destinations for rendered material are livestock/aquaculture/pet feed, fuel, fertilizer, and industrial chemicals, although the output of rendering can be used in other products such as personal care products (Wilkinson and Meeker 2021). Rendered fat can be used directly in a variety of products (e.g., detergents, lubricants) or can be further processed.

Some of the local environmental concerns specific to rendering include microbiological safety, odors, and liquid effluents. With respect to the livestock carcass waste stream, microbiological safety is of concern: heat treatment of rendering does inactivate bacteria, viruses, and protozoa, but there has been concern about transmissible spongiform encephalopathy (TSE) in animal feed products, and the European Union requires that portions of the livestock carcass (brain, spinal cord) be incinerated (Gwyther et al. 2011). Rendering does produce odors, though a large fraction of these can be controlled with washing, with a higher degree of reduction from air cleaning technologies such as scrubbers. Liquid wastes such as oils and greases require treatment at on-site or local wastewater treatment facilities (Gwyther et al. 2011).

Our review found only one LCA study that included rendering (Corona et al. 2020), and the review of (Shurson 2020) identified very limited environmental data. Therefore, this pathway requires further study before it can be assessed quantitatively or compared to other pathways. For ReFED, Corona et al. (2020) model protein displacing soybean for feed production, biodiesel displacing conventional diesel production (and combustion), and bioglycerin displacing conventional glycerin. As documented in ReFED, the

performance of rendering appears similar to that of AD. However, Gwyther et al. (2011) suggest that anaerobic digestion or other technologies may have advantages, among them environmental, over rendering.

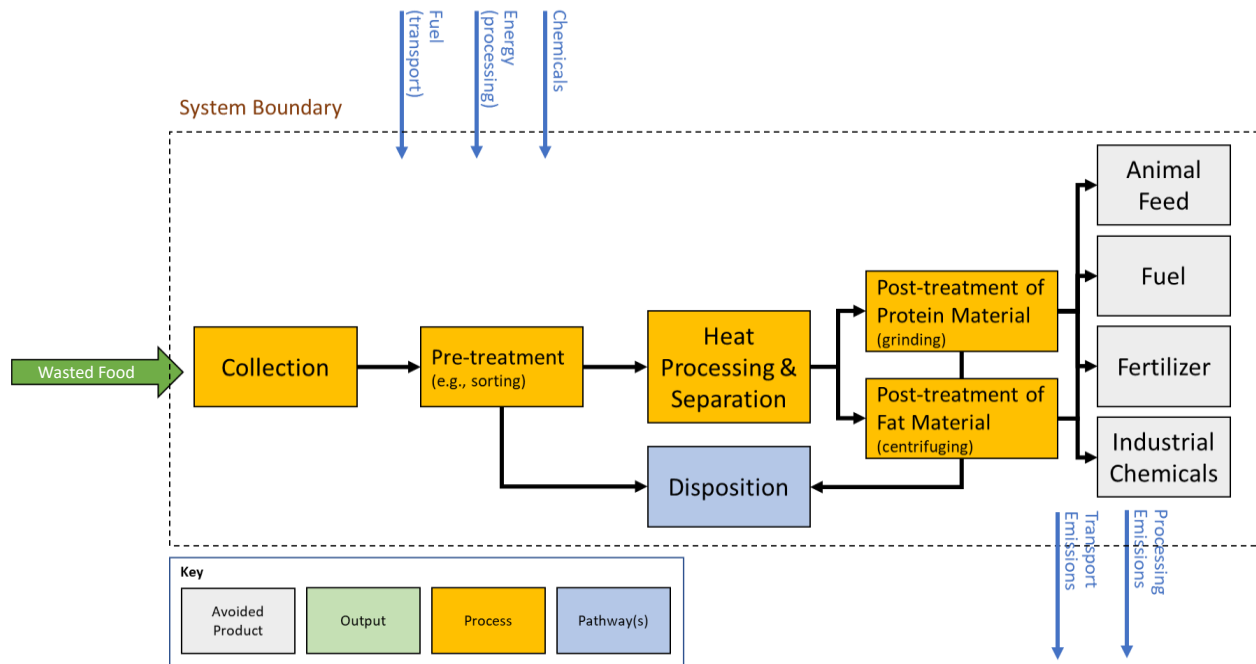


FIGURE E-3. DIAGRAM OF WASTED FOOD RENDERING

| | |
|--|---|
| Source of Environmental Impact | Energy for heat processing; transport. |
| Source of Environmental Benefit | Displacement of production of livestock/aquaculture/pet feed, fuel, fertilizer, and industrial chemicals. |
| Options to Reduce Impact | Optimize recovery of protein and fat. |

APPENDIX F

ENDNOTES

¹ World Resources Institute (WRI), Consumer Goods Forum (CGF), Food and Agriculture Organization of the United Nations (FAO), United Nations Environment Programme (UNEP), Waste and Resources Action Programme (WRAP), World Business Council for Sustainable Development (WBCSD).

² The term 'water resource recovery facility', WRRF, is used in this report to refer to wastewater treatment plants to highlight focuses on the resources that could be recovered. The term 'wastewater treatment' is used in this report to refer to the process of treating wastewater at the WRRF. For example, within the sewer/WWT pathway not all facilities receiving wasted food will have anaerobic digestion, so the more general term is preferred. Other forms of energy or resource recovery can occur at WRRFs, that may qualify them as WRRFs, but these practices are outside the scope of this document.

³ Damiani (2021) places the impact of beef production at more than 28 kg CO₂e/kg, on average, whereas rice production results in one-tenth the GWP impacts of beef (2.7 kg CO₂e/kg).

⁴ Eriksson and Spångberg (2017) considered the conversion of nonmarketable fruits and vegetables into chutney. They looked at five different fruits and vegetables (apples, bananas, oranges, peppers, and tomatoes) and evaluated only GWP (kg CO₂e/kg wasted food) and energy demand (MJ/kg wasted food). The analysis considered avoided production of fresh fruits and vegetables. Production intensities varied from 0.4 kg CO₂e and 5 MJ for an apple to 1 kg CO₂e and 15 MJ for a pepper, and thus the net impact was between -0.4 and -1 kg CO₂e and between -5 and -15 MJ for the upcycling of the apple and pepper. Additionally, several processes were essentially cancelled by their opposites: processing and avoided processing, glass jar production and avoided glass jar production, transportation to supermarket and avoided transportation to supermarket.

⁵ Eriksson et al. (2021) investigated changing broccoli harvesting practices to collect leaves and broccoli heads that fall outside the quality norms and are typically left behind. The leaves were considered for use in bread (displacing wheat flour) and broccoli soup (displacing imported broccoli powder), while the heads were considered for sliced broccoli. For all three scenarios, additional harvesting steps were negligible. Transportation was similar to that required for regular broccoli. The displacement of other products was a major contributor to overall impacts. In the case of bread, the displacement of wheat flour did not offset drying, so the net impact was 0.075 kg CO₂e/kg broccoli. However, the soup and sliced broccoli displaced products with higher impacts, leading to net negative impact (-0.45 to -0.35 kg CO₂e/kg broccoli).

⁶ For example, cows are often raised for milk and meat, but the leather produced from their hides is not treated as impact-free. In the case of upcycling, broccoli is generally cultivated for its florets, and the leaves are wasted food. As wasted food, broccoli leaf production is impact-free. However, if broccoli leaves were to become a regular substitute for wheat flour, they might cease to be classified as upcycled, and then they would no longer be considered impact-free. A study of broccoli leaves found that the leaves had higher environmental production impacts than wheat flour (Eriksson et al. 2021).

⁷ Co-digestion of significant quantities of wasted food with sludge from wastewater treatment operations may materially affect the composition and associated benefits and impacts of biosolids processing and use. The Biosolids Emissions Assessment Model (BEAM), was developed by the Canadian Council of

Ministers of the Environment, to estimate GHG emissions associated with biosolids management . Applicability of this model to biosolids generated from a co-digestion process is unknown.

⁸ About 1,265 of the roughly 16,000 WRRFs in the United States process biosolids with AD, and roughly 10% of those facilities co-digest source-separated wasted food with biosolids (ASCE 2021; U.S. EPA 2023b). Co-digestion of food waste along with wastewater solids has been shown to increase biogas production, with fats, oils, and grease providing some of the greatest biogas benefits (Parry 2014).

⁹ A U.S. EPA report of case studies of WRRFs adding wasted food (FOG, residential, commercial, and industrial) as co-digestate showed increases in biogas production ranging from 40% to 300% (U.S. EPA 2014b).

¹⁰ The ratio of produced energy to input energy (termed the energy ratio) is sometimes used to quantify net energy production. Research cited by Khoshnevisan et al. (2018) reports that biogas plants typically have an energy ratio of between 2 and 6. In their own analysis, energy ratios varied from 3.1 to 5.2.

¹¹ A review of sustainable approaches to wasted food management summarizes nine studies that provide evidence of improved methane yields, enhanced process stability, or increased volatile solids reduction when wasted food is co-digested with green waste, biosolids, animal manure, and other organic materials (Paritosh et al. 2017).

¹² Many reviewed studies specify either single- or two-stage digestion processes, but do not go into detail on why a specific reactor type was chosen. Ahamed et al. (2016) state clearly that “conventional single-stage reactors are not suitable for food waste,” and the study relies instead on a two-stage technology that is expected to increase process stability and yield more methane. Ahamed’s analysis looks at digestion of pure wasted food, while many other studies co-digest wasted food with other organic materials. Tiwary et al. (2015) go into greater detail, describing that methanogenic organisms struggle to deal with the rapid drop in pH that accompanies hydrolysis when highly degradable materials such as wasted food are introduced directly into single-stage reactors. Separation of these reactions within multi-stage reactors can increase biological stability and process performance. Alternatively, co-digestion of highly degradable materials with more stable organic materials can circumvent this issue but limits the quantity of wasted food that can be processed in each facility.

¹³ Instances in which it may not be appropriate to assign an avoided burden credit for the full quantity of energy produced include, for example, situations where there is no market for produced heat or where biogas has historically been used to provide heat for AD (i.e., baseline is not natural gas combustion).

¹⁴ Al-Rumaihi et al. (2020) showed that GWP results are sensitive to changes in CHP output, with a 30% increase in heat and electricity production leading to a greater than 100% decrease in impact potential (i.e., a net benefit).

¹⁵ A few studies examine disposal of digestate by incineration (Chiu and Lo 2018; Tong et al. 2018) or in a landfill (Righi et al. 2013).

¹⁶ One study states that 30% to 50% of total nitrogen can be lost as ammonia within six hours of application, though incorporation into soil may reduce losses by 85%. Nitrate leaching is also of concern, with losses of up to 15% of applied nitrogen from autumn applications (Nicholson et al. 2017).

¹⁷ Slorach et al. (2019a) report that there is not a significant difference in the production of biogas between mesophilic and thermophilic reactors, but that thermophilic reactors can have twice the energy consumption. Tiwary et al. (2015) come to a different conclusion, finding that thermophilic reactors do have greater biogas production.

¹⁸ A consequential analysis of animal feed, composting, and AD found that the lower eutrophication, toxicity, and acidification impacts of wet and dry feed are largely due to avoiding conventional animal feed production (Salemdeeb et al. 2017).

¹⁹ Albizzati et al. (2021b) consider avoided reference products for their wet and protein-concentrated animal feed pathways, which are maize grain and soybean-based, respectively. When compared to the gross impact of animal feed production and alternative management pathways, the reference product avoids considerable environmental burdens for several impact categories including human toxicity, cancer, particulate matter formation, global warming, ionizing radiation, photochemical ozone formation, eutrophication, and freshwater ecotoxicity.

²⁰ Thakali et al. (2022) studied the effect of source and regulatory requirements on the type and amount of contamination in food wastes intended for composting. Source separated food wastes generated from locations that were required by regulation to separate food waste from packaging and other non-food contaminants had lower levels of physical contamination than food wastes from locations that voluntarily separated. Chemical contaminants across both regulated and unregulated sources include zinc, copper, halogenated organics, and PFAS.

²¹ A review of methane emissions found that typically between 1% and 4% of carbon entering a compost pile is released as methane (Ermolaev et al. 2014). A separate review of laboratory and field studies found that 2.7% of carbon is released as methane (Pardo et al. 2015). A review by Cerda et al. (2018) identified that between 0.03 and 71.4 kg methane are released per ton wasted food during composting. Accurate measurement of compost gaseous emissions can be difficult, given emission gradients at different locations in the pile (Büyüksönmez 2012).

²² Levis and Barlaz (2011) performed a sensitivity analysis that looked at the fraction of carbon in compost feedstock that was liberated during biodegradation. The analysis considered loss of between 40% and 83% of initial carbon, with strong linear increases in GWP across all considered composting technologies as the fraction of carbon loss increased. Increasing carbon loss during composting increases process methane emissions while decreasing long-term carbon storage.

²³ An analysis of high and low emission scenarios for home composting systems (Quirós et al. 2014) illustrates the variable impacts of compost process emissions. In the high emission scenario, process emissions dominate all other operational impacts (energy, tools, and collection) for the acidification, eutrophication, and GWP categories. Energy-related inventory categories and ozone depletion are not influenced by compost process emissions. In the low emissions scenarios, process emissions still dominate for GWP, but the magnitude of impact is reduced. The contribution of process emissions to acidification and eutrophication is reduced dramatically in the low emission scenario and is similar in magnitude to impacts associated with infrastructure and energy consumption.

²⁴ The interplay of the various bacterial communities that compost organic waste is complex, and thus it can be straightforward to optimize for a single emission but difficult to optimize across emissions. For example, a review of strategies to minimize NH₃ emissions identified several approaches: 1) adjusting pH, 2) adjusting C:N ratios (e.g., through initial adjustment or co-composting a green waste that has a high C:N ratio), 3) adding absorbents (e.g., zeolite or biochar), 4) struvite precipitation (with additional adjustment of salinity), 5) adjusting nitrification, or 6) adjusting aeration rates and pressures (Wang et al. 2018; Wang and Zeng 2018).

²⁵ The studies that explicitly state their assumptions related to carbon sequestration assume that between 10% and 15% of land-applied carbon is sequestered (Levis and Barlaz 2011; Yoshida et al. 2012).

Whether or not the carbon in compost is assumed to be sequestered has a strong impact on results, potentially shifting net GWP emissions from positive to negative (Levis and Barlaz 2011; Silver et al. 2018).

²⁶ Wasted food is a good source of nitrogen and phosphorus in composting systems, can help establish ideal C/N, ratios in compost systems and leads to finished compost products with benefits both as a soil conditioner and as a replacement for chemical fertilizers.

²⁷ Energy use has been found to be a major contributor of gross emissions of nitrogen oxides and sulfur oxides in ASP and bioreactor compost systems (Levis and Barlaz 2011). The analysis shows considerably less gross impact associated with windrow composting energy demand, which is dominated by windrow turning.

²⁸ Carbon dioxide, ammonia, methane, nitrous oxide, and VOCs constitute 99% of total emissions; further detail is provided in Quirós et al. (2014).

²⁹ Increasing aeration through pile turning and bulking agents decreased emissions of both methane and nitrogen dioxide, according to a review of management approaches (Pardo et al. 2015) and laboratory-scale experiments (Yang et al. 2013). Experiments by the California Natural Resources Agency also confirm the importance of maintaining aeration to limit methanogenesis (Silver et al. 2018). A study of GHG emissions from home-scale composting systems came to an alternate conclusion: that fewer mixing events led to reduced methane emissions, while increased temperature led to increased emissions (Ermolaev et al. 2014). Other authors, cited in the Ermolaev study, had a similar finding and attributed their result partly to the release during turning events of methane trapped within the pile. Oxidation of methane to carbon dioxide near the pile's surface, by methanogenic bacteria, could also explain part of the finding that reduced mixing leads to less GHG emissions (Jäckel et al. 2005).

³⁰ An analysis of a community-scale composting system showed a moderate to major (10% to >30%) reduction in gross impact for GWP, acidification, smog formation, and ozone depletion as a result of avoided nitrogen fertilizer production. Carcinogenic toxicity impact most strongly benefited from the avoided fertilizer production, with the avoided impact of fertilizer production leading to nearly net zero impact (Keng et al. 2020). Results for a wasted food aerobic composting system showed moderate to major reductions in gross GWP and minor to moderate reductions in gross energy demand due to avoided fertilizer production (Morris et al. 2017). An analysis of wasted food composting that included windrow, ASP, Gore and bioreactor compost systems identified moderate reductions in gross GWP for all systems as a result of avoided fertilizer production (Levis and Barlaz 2011). Other authors have identified only minor environmental benefits associated with avoided fertilizer production for GWP (Righi et al. 2013).

³¹ The rate of compost application should be calibrated to consider all other nutrient sources such as legumes and chemical fertilizers, to avoid over-fertilization. Spreading equipment should be capable of metering compost application to avoid over-applying. Compost should not be applied when soil is frozen, and other best management practices related to organic amendment applications, such as manure, should be followed (NRCS and Agricultural BMP Task Force 2017).

³² Food waste is a high-moisture waste product. It should be mixed with a dry, high-carbon material to absorb moisture and provide structure to the compost pile necessary for aeration and odor control. The ideal C:N ratio at the time of pile formation is 30:1 with a moisture content of 50% to 70% (Risse and Faucette 2017).

³³ See Trabold and Babbitt (2018) or Mayer et al.(2019)for more information.

³⁴ Some food seasonings have high chlorine levels, which are related to dioxin production potential.

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- ³⁵ This statement refers to net impacts, which include olive production and olive oil processing in this analysis.
- ³⁶ Oily food waste is high in carbon (C:N ratio of 90:1) and can immobilize nitrogen in the soil, affecting crop yields if sufficient nitrogen is not available.
- ³⁷ The high carbon content of oily food waste can immobilize nitrate in the soil, prevent leaching, and make that nitrogen available for subsequent crops through re-mineralization, thus avoiding the need for additional fertilization.
- ³⁸ Of the reviewed studies that investigate land-applied wasted food, most indicate that direct application can benefit soil health, provide nutrients, and boost crop yields, but only if carefully managed temporally and co-applied with other organic wastes or inorganic fertilizers (Kumar et al. 2009; San Miguel et al. 2012).
- ³⁹ wasted food “must be applied at or below agronomic rates, and nuisance conditions (such as odors and flies) and negative impacts on surface and ground waters must be minimized” (Belcher and Aldrich 2006).
- ⁴⁰ Batuecas et al. (2019) commented on the risk that groundwater and surface water pollution could result from application of olive waste to agricultural fields, and a separate study on applying palm oil processing residues raised similar concerns (Embrandiri et al. 2012).
- ⁴¹ “Landfill gas contains many different gases. Methane and carbon dioxide make up 90% to 98% of landfill gas. The remaining 2% to 10% includes nitrogen, oxygen, ammonia, sulfides, hydrogen and various other gases” (NYSDOH 2019).
- ⁴² U.S. EPA’s Waste Reduction Model (WARM) suggests that under “moderate” moisture conditions, an appropriate decay rate for food waste is 0.14 yr⁻¹. For paper and wood products, under similar conditions, decay rates vary between 0.02 and 0.12 yr⁻¹, with higher values indicating more rapid decomposition. Yard trimmings have an average decay rate under moderate moisture conditions of 0.2 yr⁻¹.
- ⁴³ For example, lipids were calculated to contribute 59% to 70% of methane potential for a variety of food wastes in anaerobic environments, though there is a potential for lipids to inhibit methanogens (Lopez et al. 2016). The authors note the need for further research to clarify the role of lipids in wasted food decomposition.
- ⁴⁴ Reasonable ranges for decay rate and oxidation factor selection may lead to wasted food emissions varying from 0.23 to 0.71 MTCO₂e (metric tons of CO₂ equivalent) per ton of food waste (CARB 2017).
- ⁴⁵ Some studies suggest that typical EPA decay rates are too low (Wang et al. 2013), which would lead to underestimation of the portion of methane produced early in the landfill’s life, before the installation of robust LFG capture systems.
- ⁴⁶ The carbon storage potentials of other commonly landfilled organic materials such as cardboard, newspaper, grass, and dimensional lumber were estimated to be 55%, 84%, 53%, and 88%, respectively (U.S. EPA 2020b).
- ⁴⁷ One study that modeled the landfill as “a ‘dry tomb’ where microbial activity is suppressed” assumed that 43% of landfilled carbon was sequestered (Yoshida et al. 2012). This value is considerably higher than the 16% estimate provided by WARM and is unlikely to apply in moist regions. Additional research has shown that landfill degradation is more limited in arid regions (Jain et al. 2021).
- ⁴⁸ Landfill leachate is the predominant contributor to eutrophication potential in several studies (Edwards et al. 2017; Lee et al. 2020; Slorach et al. 2020).

⁴⁹ Lee et al. (2017) found that when their LCA model was altered to increase LFG collection in the early stages of cell development, GHG emissions could be reduced by 27% compared to a less aggressive gas collection strategy.

⁵⁰ The authors noted that use of these factors was uncertain but was based on the best available information at the time.

⁵¹ Zan et al. (2020) suggested that just 6% of the total COD in wasted food is degraded over a 4-hour sewer residence time. For comparison, direct GHG emissions from landfilling wasted food are approximately 16–30+ times greater than these marginal additional sewer emissions from added wasted food (Zan et al. 2020). In the U.S. context, this 6% figure may overestimate actual sewer emissions, given that a GIS-based analysis simulating the residence time of 3,422 U.S. sewer systems found that the median residence time of wastewater is approximately 3.3 hours (as opposed to Zan et al.'s 4-hour basis), with larger municipalities having longer residence times (Kapo et al. 2017).

⁵² About 1,265 of the roughly 16,000 WRRFs in the United States process biosolids with AD (ASCE 2021; U.S. EPA 2023b).

⁵³ Concern that additional wasted food in sewers will drive solids deposition and clogging has not been universally borne out (Bolzonella et al. 2003). Further study and decision support tools are needed to navigate potential trade-offs between wasted food conveyance and sewer performance. For instance, various studies of long-term impacts have shown that, when high-specific-gravity wasted food (e.g., bones and egg shells) and/or FOG-rich wasted food are excluded, marginal additional sewer deposits from wasted food are small and have a minor impact on sewer performance (Mattsson et al. 2015). Mattsson et al. (2014) used cameras to monitor the effects of food waste disposers on small-diameter residential pipes across three Swedish municipalities. Despite disposer usage correlating with more deposits, especially in pipe segments with sags or gentle slopes, these deposits were small and appeared to have a minor impact on sewer performance (Mattsson et al. 2014).

⁵⁴ Source separation requires separate collection and transportation infrastructure that is not currently available in all communities, and sewage collection avoids hygiene and odor problems that can arise during other forms of collection and transportation (Kaur et al. 2019).

⁵⁵ A survey of local municipal authorities and waste management operators revealed that organizations providing organic feedstocks to AD operators can expect to meet contamination requirements of less than 5% on a total weight basis (Ogden et al. 2022).

⁵⁶ The animal feed pathway can address contamination concerns through preprocessing. Vermicomposting can recover the value of the food waste (in this case nutritional value) by transferring it to a different carrier (like fly larvae), avoiding the issue of feeding animals food waste with physical contaminants (Nick Hacheny and Sally Brown 2017).

⁵⁷ Heterogenous wasted food feedstocks can be difficult to formulate for appropriate animal nutrition.

⁵⁸ In a 2019 survey of composters in California, 48% reported that contaminated waste streams limit their ability to accept wasted food (CalRecycle 2019). Regulations governing commercial composting facilities in Washington State limit physical contamination in compost piles to less than 1% of wet weight with a plastic content of not more than 0.25% (State of Washington 2019).

⁵⁹ <https://www.epa.gov/air-emissions-factors-and-quantification/emissions-estimation-tools>

⁶⁰ Their reported values range from -4300, -3090, -162 and +536 kg CO₂e when preventing (source reduction), donating, digesting and landfilling 1 metric ton of wasted food, respectively. The source

reduction pathway receives a credit for avoiding food production, as does food donation. The AD and landfilling options do not receive this credit. Stated differently, the wasted food collected for donation comes with a negative carbon footprint, so that additional transportation and other processing may produce some GHG emissions but leaves the overall pathway with a large, net negative carbon footprint. Food waste enters the AD and landfilling pathways with a neutral (zero) carbon footprint, given its status as a waste product. These modeling options describe the convention, used in this report, to compare wasted food pathways to source reduction, wherein source reduction/donation have a virtual net negative impact while wasted food enters other pathways burden free (i.e. neutral environmental impact).

⁶¹ An analysis of AD, composting, controlled combustion, and landfilling of wasted food found that management pathways have small GWP impacts (and potentials to reduce impact) compared to food production (Slorach et al. 2020).

⁶² In 2021, over 65% of landfilled waste was sent to facilities with active energy recovery projects, and nearly 50% of collected LFG was sent for energy recovery (U.S. EPA 2022a).

⁶³ Including 4% used as alternative daily cover for landfills

⁶⁴ Excluding scenarios where cover crops are harvested for bio-energy.

⁶⁵ Toxicity and eutrophication impacts were found to decrease with the increasing presence of packaging material as these impacts are dominated by liquid emissions, produced during hydrochar dewatering, which are not produced by low-moisture, packaging materials (Berge et al. 2015).

⁶⁶ Research has shown that slow pyrolysis, with its lower temperatures and slower heating rates, tends to produce more biochar and less bio-oil because of longer residence times with slower heating delays decomposition (Trabold and Babbitt 2018).