



Environmental Assessment for Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category

April 2024

This page intentionally left blank.

U.S. Environmental Protection Agency
Office of Water (4303T)
1200 Pennsylvania Avenue, NW
Washington, DC 20460

EPA-821-R-24-005

This document was prepared by the Environmental Protection Agency. Neither the United States Government nor any of its employees, contractors, subcontractors, or their employees make any warrant, expressed or implied, or assume any legal liability or responsibility for any third party's use of or the results of such use of any information, apparatus, product, or process discussed in this report, or represents that its use by such party would not infringe on privately owned rights.

Questions regarding this document should be directed to:

U.S. EPA Engineering and Analysis Division (4303T)
1200 Pennsylvania Avenue NW
Washington, DC 20460
(202) 566-1000

Contents

1. Introduction	1
1.1 Background on Steam Electric Power Plant Wastewater Discharges	1
1.2 Scope of the EA	2
2. Literature Review of the Environmental and Human Health Concerns Associated with the Evaluated Wastestreams	5
2.1 Pollutants Discharged in the Evaluated Wastestreams	5
2.1.1 Metals and Toxic Bioaccumulative Pollutants	5
2.1.2 Nutrients	6
2.1.3 TDS and Salinity.....	6
2.1.4 Bromine/Bromide	7
2.1.5 Iodine/Iodide.....	8
2.2 Potential Impacts from the Evaluated Wastestreams	9
2.2.1 Ecological Impacts.....	10
2.2.2 Human Health Effects	12
2.2.3 Groundwater Impacts	13
2.2.4 CCR Surface Impoundments as Attractive Nuisances	14
3. Environmental Assessment Methodology	16
3.1 Pollutant Loadings for the Evaluated Wastestreams.....	16
3.2 Pollutant Exposure Pathways	18
3.3 Environmental Impacts Selected for Qualitative and Quantitative Assessments in the EA.....	20
3.4 Overview of the IRW Model	21
3.4.1 Structure of the IRW Model.....	21
3.4.2 Pollutants Evaluated by the IRW Model.....	23
3.5 Proximity Analysis.....	26
3.6 Downstream Analysis	26
3.7 Scope of the Evaluated Plants and Immediate Receiving Waters	27
4. Results of the Quantitative Environmental Assessment for the Final Supplemental Rule	31
4.1 Environmental Impacts Identified by the IRW Model	31
4.1.1 Water Quality Impacts	32
4.1.2 Wildlife Impacts	35
4.1.3 Human Health Impacts	37
4.2 Discharges to Sensitive Environments.....	43
4.2.1 Impaired Waters	44
4.2.2 Fish Consumption Advisories	47
4.2.3 Drinking Water Resources.....	49
4.3 Impacts in Downstream Surface Waters	49
4.4 Summary of Key Environmental and Human Health Improvements.....	51
5. References	52

Attachments

Attachment A . Additional IRW Model Results

List of Figures

Figure 1. Overview of the IRW Model	23
Figure 2. Locations of Immediate Receiving Waters Evaluated in the Environmental Assessment for the Final Supplemental Rule.....	30
Figure 3. Immediate Receiving Waters Impaired by Mercury	45
Figure 4. Immediate Receiving Waters Impaired by Metals Other Than Mercury	46
Figure 5. Immediate Receiving Waters Impaired by Nutrients.....	46
Figure 6. Immediate Receiving Waters with Fish Consumption Advisories for Mercury.....	48

List of Tables

Table 1. Wastestreams Evaluated in the EA.....	2
Table 2. Estimated Annual Baseline Mass Pollutant Loadings and Estimated Reduction in Loadings Under Regulatory Options for the Evaluated Wastestreams ^a	17
Table 3. Steam Electric Power Plant Wastewater Environmental Pathways and Routes of Exposure Evaluated in the Environmental Assessment for the Final Supplemental Rule	19
Table 4. Water Quality Benchmarks: NRWQC and MCLs.....	24
Table 5. Sediment Biota and Wildlife Benchmarks: TECs and NEHCs.....	25
Table 6. Human Health Benchmarks: Oral RfDs and CSFs	25
Table 7. Plants, Generating Units, and Immediate Receiving Waters Evaluated in the Environmental Assessment for the Final Supplemental Rule	28
Table 8. Plants, Generating Units, and Immediate Receiving Waters with Pollutant Loadings Under Baseline and Regulatory Options for the Final Supplemental Rule	29
Table 9. Modeled IRWs with Exceedances of NRWQC and MCLs Under Baseline and Regulatory Options.....	32
Table 10. Modeled IRWs with Exceedances of NRWQC and MCLs, by Pollutant, Under Baseline and Regulatory Options.....	34
Table 11. Modeled IRWs with Exceedances of TECs and NEHCs Under Baseline and Regulatory Options.....	36
Table 12. Modeled IRWs with Exceedances of Oral RfD (Noncancer Human Health Effects) Under Baseline and Regulatory Options.....	38
Table 13. Modeled IRWs with LECR Greater Than One-in-a-Million (Cancer Human Health Effects) Under Baseline and Regulatory Options.....	39
Table 14. Modeled IRWs with Exceedances of Oral RfDs by Race/Ethnicity Under Baseline and Regulatory Options.....	40
Table 15. Modeled IRWs with LECR Greater Than One-in-a-Million (Cancer Human Health Effects) Race/Ethnicity Under Baseline and Regulatory Options.....	41
Table 16. Comparison of Modeled T4 Fish Tissue Concentrations to Fish Advisory Screening Values Under Baseline and Regulatory Options.....	42

Table 17. Modeled IRWs Identified as CWA Section 303(d) Impaired Waters, Fish Consumption Advisory Waters, or Drinking Water Resources Under Baseline and Regulatory Options..... 43

Table 18. Modeled IRWs Identified as CWA Section 303(d) Impaired Waters for Pollutants Present in the Evaluated Wastestreams Under Baseline and Regulatory Options..... 45

Table 19. Modeled IRWs Identified as Fish Consumption Advisory Waters for Pollutants Present in the Evaluated Wastestreams Under Baseline and Regulatory Options..... 48

Table 20. Modeled IRWs Identified as Located Within 5 Miles of a Drinking Water Resource Under Baseline and Regulatory Options..... 49

List of Abbreviations

AETX	aetokthonotoxin
BA	bottom ash
BAT	best available technology economically achievable
BCA	benefit and cost analysis
Br-DBP	brominated disinfection byproduct
CCR	coal combustion residuals
CFR	Code of Federal Regulations
CRL	combustion residual leachate
CSF	cancer slope factor
CWA	Clean Water Act
DBP	disinfection byproduct
DCN	document control number
D-FATE	Downstream Fate and Transport Equations
DNA	deoxyribonucleic acid
DWTP	drinking water treatment plant
EA	environmental assessment
EGU	electric generating unit
EJ	environmental justice
ELGs	effluent limitations guidelines and standards
EPA	U.S. Environmental Protection Agency
FGD	flue gas desulfurization
FR	Federal Register
FW	freshwater
HAAs	haloacetic acids
HANs	haloacetonitriles
HH O	human health for the consumption of organism only
HH WO	human health for the consumption of water and organism
I-DBP	iodinated disinfection byproduct
IRIS	Integrated Risk Information System
IRW	immediate receiving water
lb/year	pounds per year
LC50	median lethal concentration
LECR	lifetime excess cancer risk
MCL	maximum contaminant level
MRL	minimal risk level
mg/kg	milligrams per kilogram

mg/kg-day	milligrams per kilogram body weight per day
mg/L	milligrams per liter
µg/g	micrograms per gram
N	nitrogen
NEHC	no effect hazard concentration
NHDPlus	National Hydrography Dataset Plus
NRWQC	National Recommended Water Quality Criteria
POTW	publicly owned treatment works
ppm	parts per million
PSES	pretreatment standards for existing sources
RfD	reference dose
RIA	regulatory impact analysis
SO ₂	sulfur dioxide
T3	trophic level 3
T4	trophic level 4
TDD	technical development document
TDS	total dissolved solids
TEC	threshold effect concentration
THMs	trihalomethanes
TKN	total Kjeldahl nitrogen
TSS	total suspended solids
UV	ultraviolet
VM	vacuolar myelinopathy
WHO	World Health Organization

1. Introduction

The U.S. Environmental Protection Agency (EPA) promulgated revised effluent limitations guidelines and standards (ELGs) for the Steam Electric Power Generating Point Source Category (40 CFR 423) on November 3, 2015 (80 FR 67838), referred to hereinafter as the “2015 rule.” Following promulgation, the EPA received seven petitions for review of the 2015 rule and the Administrator announced his decision to reconsider the 2015 rule. The EPA finalized a revision to the regulations for the Steam Electric Power Generating category (85 FR 64650, October 13, 2020), referred to as the “2020 rule,” which established revised ELGs for flue gas desulfurization (FGD) wastewater and bottom ash (BA) transport water discharged from steam electric power plants. See the *Technical Development Document for Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*, or TDD (EPA-821-R-24-004) for more background and information on the rulemaking history.

This 2024 supplemental rulemaking is based on a review of the ELGs promulgated in 2020 under Executive Order 13990. The supplement rule covers best available technology economically achievable (BAT) and pretreatment standards for existing sources (PSES) requirements for FGD wastewater, BA transport water, combustion residual leachate (CRL), and legacy wastewater from steam electric power plants. It also establishes new source performance standards (NSPS) and pretreatment standards for new sources (PSNS) for CRL.

In support of the development of the 2015 rule and the 2020 rule, the EPA conducted an environmental assessment (EA) to evaluate the environmental impact of pollutant loadings discharged by steam electric power plants and assess the potential environmental improvement from pollutant loading changes under the rules. The EPA documented the EA in the September 2015 report *Environmental Assessment for the Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category* (EPA-821-R-15-006) (U.S. EPA, 2015a), referred to hereinafter as the “2015 EA,” and the *Supplemental Environmental Assessment for Revisions to the Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category* (EPA-821-R-20-002) (U.S. EPA, 2020a), referred to hereinafter as the “2020 EA.” To support the 2024 final rule, the EPA updated its EA for the 2015 rule and 2020 rule to include the steam electric power plants discharging one or more of the four wastestreams. In addition, the EPA evaluated potential cumulative impacts from multiple pollutants (Joint Toxic Action analysis) in support of the proposed rulemaking.

The Clean Water Act does not require that the EPA assess the water-quality-related environmental impacts, or the benefits, of its ELGs, and the Agency did not make its decisions in the final rule based on the expected benefits of the rule. The EPA does, however, inform itself and the public of the benefits of its proposed and final rules, as required by Executive Order 12866. See the *Benefit and Cost Analysis for Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*, or BCA Report (EPA-821-R-24-006). This EA report presents the EPA’s evaluation of the potential environmental impacts due to pollutant loadings under baseline discharge practices (*i.e.*, following full implementation of the requirements under the 2015 rule and 2020 rule and any known retirements, fuel conversions, and treatment technologies in place at in-scope steam electric power plants) and the improvements to those impacts under the evaluated regulatory options.

1.1 Background on Steam Electric Power Plant Wastewater Discharges

Based on demonstrated impacts documented in literature and modeled receiving water pollutant concentrations, discharges of steam electric power plant wastewater can affect the water quality in receiving waters, affect the wildlife in the surrounding environments, and pose a human health risk to nearby communities. There is substantial evidence that certain pollutants found in these wastewater discharges, such as mercury and selenium, propagate from the aquatic environment to terrestrial food webs, indicating a potential for broader impacts on surrounding ecological systems by diminishing

population diversity and disrupting community dynamics. Ecosystem recovery from exposure to these pollutants can be extremely slow, and even short periods of exposure (*e.g.*, less than a year) can cause observable ecological impacts that last for years.

Steam electric power plants often discharge wastewater into waterbodies used for fishing, for recreation, and/or as sources of drinking water. Many studies have raised concerns about the toxicity of these wastestreams and their impacts on downstream drinking water treatment systems. For example, these discharges can elevate halogen levels in surface water, which may contribute to disinfection byproduct formation at downstream drinking water treatment plants. Leaching of pollutants from surface impoundments and landfills containing combustion residuals is known to affect off-site groundwater and drinking water wells at concentrations above maximum contaminant level drinking water standards, posing a threat to human health.

1.2 Scope of the EA

The Steam Electric Power Generating Point Source Category ELGs apply to establishments whose generation of electricity is the predominant source of revenue or principal reason for operation, and whose generation results primarily from a process using fossil-type fuels (coal, oil, or gas), fuel derived from fossil fuel (*e.g.*, petroleum coke, synthesis gas), or nuclear fuel in conjunction with a thermal cycle using the steam water system as the thermodynamic medium. The EPA evaluated four wastestreams from steam electric power plants whose limitations and standards would be revised under the new rulemaking: FGD wastewater, BA transport water, CRL, and legacy wastewater, as described in Table 1.

Table 1. Wastestreams Evaluated in the EA

Evaluated Wastestream	Description
FGD wastewater	<p>Wastewater generated from a wet FGD scrubber system. Wet FGD systems are used to control sulfur dioxide (SO₂) and mercury emissions from the flue gas generated in the plant's electric generating unit (EGU).</p> <p>The pollutant concentrations in FGD wastewater vary from plant to plant depending on the coal type, the burning of refined coal, the sorbents and additives used, the materials used to construct the FGD system, the FGD system operation, the level of recycle within the absorber, and the air pollution control systems operated upstream of the FGD system. FGD wastewater contains total dissolved solids (TDS), total suspended solids (TSS), nutrients, halogens, metals, and other toxic and bioaccumulative pollutants, such as arsenic and selenium (see the TDD [U.S. EPA, 2024a] for further details).</p>
BA transport water	<p>Water used to convey the BA particles collected at the bottom of the EGU.</p> <p>BA transport waters contain halogens, TDS, TSS, metals, and other toxic and bioaccumulative pollutants, such as arsenic and selenium (see the TDD [U.S. EPA, 2024a] for details). The effluent from BA surface impoundments typically contains low concentrations of TSS; however, arsenic, bromide, selenium, and metals are still present in the wastewater, predominantly in dissolved form.</p>
CRL	<p>Leachate is composed of liquid, including any suspended or dissolved constituents in the liquid, that has percolated through waste or other materials emplaced in a landfill, or that passes through the surface impoundment's containment structure (<i>e.g.</i>, bottom, dikes, berms). CRL includes seepage and/or leakage from a combustion residual landfill or impoundment unit.</p> <p>CRL contains pollutants similar to those in FGD wastewater.</p>

Table 1. Wastestreams Evaluated in the EA

Evaluated Wastestream	Description
Legacy wastewater	As described in the preamble to the final rule, legacy wastewater is comprised of FGD wastewater, BA transport water, fly ash transport water, CRL, gasification wastewater, and/or flue gas mercury control wastewater generated before the “as soon as possible” date that more stringent effluent limitations from the 2015 or 2020 rules would apply. Legacy wastewater contains pollutants similar to those in the other wastestreams described in this table.

The goal of the EA is to answer the following questions about pollutant loadings from the four evaluated wastestreams:

- What are the environmental concerns?
- What are baseline environmental impacts to water quality and wildlife and impacts to human health?
- What are the potential improvements to water quality, wildlife, and human health under the regulatory options?

This EA report presents the EPA’s evaluation of environmental concerns and potential exposures (ecological and human) to pollutants commonly found in wastewater discharges from steam electric power plants. The EPA carried out both qualitative and quantitative analyses. Qualitative analyses included reviewing additional literature documenting site impacts and pollutant-specific research. Quantitative analyses included assessing the pollutant loadings to receiving waters—including those designated as impaired or with a fish consumption advisory—under baseline and the evaluated regulatory options and reviewing the effects of pollutant exposure on ecological and human receptors. To quantify impacts associated with these discharges, the EPA used a computer model to estimate pollutant concentrations in the immediate receiving waters, pollutant concentrations in fish tissue, and potential exposure doses to ecological and human receptors from fish consumption. The EPA compared the values calculated by the model to benchmark values to assess the extent of the environmental impacts nationwide. The EPA evaluated the impacts of FGD wastewater, BA transport water, CRL,¹ and legacy wastewater discharges.

The EPA evaluated three regulatory options, summarized in Table VII-1 of the preamble to the final rule. The EPA evaluated 112 plants that discharge FGD wastewater, BA transport water, CRL, and/or legacy wastewater directly or indirectly to surface waters under baseline and/or the regulatory options and performed the quantitative modeling of pollutants in the immediate receiving water on a subset of 100 of these plants. The analyses presented in this report account for notice of planned participation as described in Section VI of the preamble to the final rule. See Section 3.7 of this report for additional details on the scope of this EA.

The assessments described in this EA report focus on environmental impacts caused by exposure to pollutants in the evaluated wastestreams through the surface water exposure pathway. However, the final rule may have other environmental impacts unrelated to exposure to pollutants in wastewater discharges. Examples include changes in groundwater and surface water withdrawals by plants and

¹ The EPA is establishing a new subcategory for discharges of unmanaged CRL, which the EPA is defining in this rule to mean the following: (1) discharges of CRL that the permitting authority determines are the functional equivalent of a direct discharge to a waters of the United States (WOTUS) through groundwater or (2) discharges of CRL that has leached from a waste management unit into the subsurface and mixed with groundwater prior to being captured and pumped to the surface for discharge directly to a WOTUS (see Section VII.C.5 of the preamble to the final rule). This subcategory of CRL is not evaluated in the EA.

changes in air emissions due to changes in electricity use, transportation requirements, and the profile of electricity generation. These impacts are discussed in the EPA's BCA Report (U.S. EPA, 2024b).

This EA report does not discuss impacts caused by pollutants in unmanaged CRL. See Section VII.C.5 of the preamble to the final rule.

This report presents the methodology and results of the qualitative and quantitative analyses performed for the EA to support the supplemental rule. In addition to this EA, the final rule is supported by several reports:

- *Technical Development Document for Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category* (TDD), Document No. EPA-821-R-24-004 (U.S. EPA, 2024a). This report includes background on the final rule, the industry, and treatment technologies and pollution prevention techniques; it also documents the EPA's engineering analyses to support the supplemental rule, including cost estimates, wastewater characterization and pollutant loadings, and a non-water-quality environmental impact assessment.
- *Benefit and Cost Analysis for Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category* (BCA Report), Document No. EPA-821-R-24-006 (U.S. EPA, 2024b). This report summarizes the monetary benefits and societal costs of implementing the regulatory options.
- *Regulatory Impact Analysis for Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category* (RIA), Document No. EPA-821-R-24-007 (U.S. EPA, 2024c). This report presents a profile of the steam electric power generating industry, a summary of the costs and impacts associated with the regulatory options, and an assessment of the supplemental rule's impact on employment and small businesses.
- *Environmental Justice Analysis for Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category* (EJ Report). Document No. EPA-821-R-24-008 (U.S. EPA, 2024d). This report presents the environmental justice (EJ) analysis to support the supplemental rule, including screening analysis to identify communities with potential EJ concerns, community outreach, literature review, and risk analysis.

The ELGs for the Steam Electric Power Generating Category are based on data generated or obtained in accordance with the EPA's Quality System and Information Quality Guidelines. The EPA's quality assurance and quality control activities for this rulemaking include developing, approving, and implementing quality assurance project plans for the use of environmental data generated or collected from sampling and analyses, existing databases, and literature searches, and for developing any models that used environmental data.

2. Literature Review of the Environmental and Human Health Concerns Associated with the Evaluated Wastestreams

Discharges of the evaluated wastestreams from steam electric power plants—flue gas desulfurization (FGD) wastewater, bottom ash (BA) transport water, combustion residual leachate (CRL), and legacy wastewater—contain toxic and bioaccumulative pollutants (*e.g.*, selenium, mercury, arsenic, nickel), halogens (containing bromides, chlorides, or iodides), nutrients, and total dissolved solids (TDS), which can cause environmental harm through the contamination of surface waters. Certain pollutants in the discharges pose a danger to ecological communities due to their persistence in the environment and bioaccumulation in organisms. These factors can slow ecological recovery and can have long-term impacts on aquatic organisms, wildlife, and human health. Many studies document ecological impacts such as fish mortality, genotoxicity, and lower fish survival and reproduction rates resulting from exposure to pollutants in steam electric power plant discharges (Brandt et al., 2017 and 2019; Carlson and Adriano, 1993; Hopkins et al., 2000; Javed et al., 2016; Lemly, 1997b and 2018; Rowe et al., 1996 and 2002). Halogens associated with steam electric power plant discharges also raise ecological and human health concerns. Halogens in source water for drinking water treatment plants (DWTPs) can interact with disinfection processes to form halogenated disinfection byproducts (DBPs), which can pose a risk to human health (Cantor et al., 2010; Chisholm et al., 2008; Dong et al., 2019; Hanigan et al., 2017; National Toxicology Program, 2018; Regli et al., 2015; Richardson et al., 2007 and 2008; Richardson and Plewa, 2020; U.S. EPA, 2016a; Villanueva et al., 2004, 2007, and 2015; Wagner and Plewa, 2017; Wei et al., 2013; Yang et al., 2014).

The EPA documented environmental and human health concerns from steam electric power plant discharges in the 2015 final environmental assessment, or 2015 EA (U.S. EPA, 2015a) and the 2020 EA (U.S. EPA, 2020a). For this EA, the EPA conducted a supplemental literature review in 2022 that consisted of identifying and evaluating peer-reviewed journal articles and other materials published since its last full literature review (2010) that focus on current environmental, ecological, and human health impacts resulting from discharges of pollutants in the evaluated wastestreams. The EPA also incorporated relevant articles submitted with public comments and published since the 2022 review into its analysis for the final rule. This section summarizes relevant findings from the EPA’s literature reviews, including an overview of the pollutants discharged in the evaluated wastestreams and their associated environmental concerns. Some of the articles documented impacts of steam electric power plant discharges but did not provide specific wastestream details. When such details were documented in reviewed articles, the EPA included details on applicable wastestreams. See the memorandum *Literature Review for the 2024 Steam Electric Supplemental Rule Environmental Assessment* (U.S. EPA, 2024e) for details.

2.1 Pollutants Discharged in the Evaluated Wastestreams

Several variables can affect the composition of steam electric power plant wastewater, including fuel composition (*e.g.*, parent coal composition varies by coal type and geographic region and inclusion of other fuels in the combustion process), air pollution control technologies (*e.g.*, use of dry versus wet systems), and management techniques used to dispose of the wastewater (*e.g.*, whether the plant commingles its wastestreams) (Carlson and Adriano, 1993; Rowe et al., 2002). Commingling steam electric power plant wastewaters in surface impoundments can result in a complex mixture of pollutants in the effluent that is released to the environment (Rowe et al., 2002).

2.1.1 Metals and Toxic Bioaccumulative Pollutants

Studies commonly cite metals and toxic bioaccumulative pollutants (*e.g.*, arsenic, mercury, and selenium) as the primary cause of ecological damage following exposure to steam electric power plant wastewater (U.S. EPA, 2015a). An important consideration in evaluating these pollutants is their bioavailability, defined as the ability of a particular contaminant to be assimilated into the tissues of exposed organisms.

A pollutant's bioavailability is affected by the characteristics of both the pollutant (*e.g.*, speciation, particle size) and the surrounding environment (*e.g.*, temperature, pH, salinity, oxidation-reduction potential, total organic content, suspended particulate content, and water velocity). Metals and toxic bioaccumulative pollutants in steam electric power plant wastewater are present in both soluble (*i.e.*, dissolved) and particulate (*i.e.*, suspended) form. For example, the EPA collected sampling data for FGD wastewater in support of the steam electric effluent limitations guidelines and standards. These data show that some pollutants, such as arsenic, are present mostly in particulate form while other pollutants, such as selenium and boron, are present mostly in soluble form (ERG, 2012). Environmental conditions influence the tendency of a dissolved pollutant to remain in solution or precipitate out of solution, sorb to either organic or inorganic suspended matter in the water column, or sorb to the mixture of materials (*e.g.*, clays and humic matter) found in sediments (U.S. EPA, 2007). Pollutants that precipitate out of solution can become concentrated in the sediments of a waterbody. Organisms will bioaccumulate pollutants by consuming pollutant-enriched sediments and suspended particles, filtering ambient water containing dissolved pollutants, or both.

Appendix A of the 2020 EA (U.S. EPA, 2020a) provides examples of potential adverse impacts to humans, wildlife, and aquatic organisms resulting from exposure to metals and toxic bioaccumulative pollutants in the evaluated wastestreams and provides the minimal risk level (MRL) for human oral exposure (or similar benchmark value) for reference. Adverse impacts from steam electric power plant discharges of these pollutants are discussed further in the 2015 EA (U.S. EPA, 2015a).

2.1.2 Nutrients

Nutrients (*e.g.*, phosphorus and nitrogen) are essential components for plants and animals to grow and develop; however, increased nutrient concentrations can upset the delicate balance of nutrient supply and demand required to maintain aquatic life in surface waters. For example, excess nutrients can cause harmful algal blooms and low oxygen (hypoxia) in surface waters. These are primarily problems for estuaries, such as the Chesapeake Bay, and coastal waters, such as the Gulf of Mexico. Nutrient loadings from multiple power plants are especially a concern for waterbodies that are nutrient-impaired or in watersheds that have nutrient problems downstream. Nutrient concentrations present in steam electric power plant wastewater are primarily attributed to the fuel composition and air pollution controls in the combustion process.

Nutrient loadings to surface waters can affect the ecological stability of freshwater and saltwater aquatic systems. For example, elevated levels of nutrients can stimulate rapid growth of plants, algae, and cyanobacteria on or near the waterbody surface, which in turn can obstruct sunlight penetration, increase turbidity, and decrease dissolved oxygen levels (U.S. EPA, 2015b). Adverse impacts from steam electric power plant discharges of nutrients are discussed further in the 2015 EA (U.S. EPA, 2015a).

2.1.3 TDS and Salinity

TDS represents the concentration of combined dissolved organic and inorganic matter, whereas salinity represents the total concentration of dissolved inorganic salts. Common inorganic salts found in TDS can include cations (positively charged ions), such as calcium, magnesium, potassium, and sodium, and anions (negatively charged ions), such as carbonates, nitrates, bicarbonates, chlorides, and sulfates. TDS concentrations in steam electric power plants wastestreams include contributions from dissolved metals and halogens (*e.g.*, chlorides, bromides, and iodides).

Salts can enter water naturally through erosion of soils and geologic formations and introduction of their dominant ions to local freshwater systems (Hem, 1985; Olson and Hawkins, 2012; Pond, 2004; U.S. EPA, 2011). In addition to steam electric power plants, other sources of TDS are widespread in the environment, making it more likely that receiving waters for the discharges of the evaluated wastestreams already carry excessive TDS loadings. These other sources include mining activities, use of road salt for de-icing, and discharge of sewage and industrial wastewater (Cañedo-Argüelles et al., 2013; Corsi et al., 2010). Once salinity has increased in freshwater systems, the effect can be persistent. In lentic waters such as lakes and ponds, even small increases in salt levels can result in long-term increases in

salinity, lasting months or years (Evans and Frick, 2001). Kaushal et al. (2005) reported that, after application of deicing salts in winter, chloride concentrations in urban streams remain elevated into spring, summer, and fall and contribute to an accumulation of salts in groundwater and aquifers that may persist over several decades.

Harb et al. (2021) studied how changes in freshwater salinity can have environmental impacts on (1) spray aerosol generation from the breaking of waves and (2) diversity of aquatic bacteria. As waves break, aquatic bacteria can be aerosolized (*i.e.*, transferred from water to air). Changes in the bacteria being transferred from water to air could affect regional climate by altering aerosolized bacteria that act as cloud condensation nuclei (*i.e.*, particles in the air onto which water vapor will condense) and ice-nucleating particles (*i.e.*, particles for formation of cloud ice crystals). In addition, alterations in the aerosolized bacteria could affect public health by increasing inhalation exposure to airborne pathogens (Harb et al., 2021). Harb et al. (2021) sought to understand how increased freshwater salinity can impact the abundance and diversity of aerosolized aquatic bacteria. In freshwater salinity ranges, researchers found that aerosolization of bacteria increased as salinity increased. The study found that salinity altered the transfer of some bacterial families to an aerosol, with some families exhibiting enhanced, diminished, or no change in water to air transfer (Harb et al., 2021).

Exposure to dissolved bioaccumulative pollutants and halogens found in the evaluated wastestreams may cause human health and ecological effects. Researchers have documented the potential consequences of elevated salinity on aquatic ecosystems. Increased salinity has been linked to adverse effects including increases in invasive species, lower rates of organic matter processing, changes in biogeochemical cycles, decreased riparian vegetation, and altered composition of primary producers (*i.e.*, plants, bacteria, and algae) (Cañedo-Argüelles et al., 2013). Increases in aquatic salinity may cause shifts in biotic communities, limit biodiversity, exclude less-tolerant species, and result in acute or chronic effects at specific life stages (Weber-Scannell and Duffy, 2007). Salt additions can lead to loss of exchangeable cations in soil, and the mobility and toxicity of some pollutants, especially metals, can be enhanced at high salt concentrations (Stets et al., 2020). Because interactions between ions can affect the bioavailability and toxicity of individual TDS constituents, the net ecological effect of elevated TDS levels in the aquatic environment depends on its ionic composition (Moore et al., 2017; Mount et al., 1993 and 1997). The 2020 EA (U.S. EPA, 2020a) provides further details on adverse impacts from discharges of TDS and increased salinity in freshwater systems.

2.1.4 Bromine/Bromide

Bromine is naturally present in coal. Some coal-fired steam electric power plants also add bromine, in the form of bromide compounds, to their combustion processes to enhance mercury emissions control or burn refined coal amended with bromide compounds (U.S. EPA, 2020b). After combustion, bromine partitions in part to FGD wastewater and BA transport water in its anion form, known as bromide (EPRI, 2014; Peng et al., 2013). Documented bromide levels in FGD wastewater vary widely and can exceed 175 milligrams per liter (mg/L) (EPRI, 2009; Good, 2018; U.S. EPA, 2015c and 2020b). Average bromide levels of 5.1 mg/L have been documented in BA transport wastewaters (U.S. EPA, 2020b). These levels are higher than the average levels of 0.014 mg/L to 0.2 mg/L reported for freshwater surface waters (Flury and Papritz, 1993; Health Canada, 2015; McGuire et al., 2002). Field-based and modeling studies document elevated bromide levels in surface waters downstream of steam electric power plants and identify FGD wastewater discharges as a substantial source of bromide loadings from the plants (Cornwell et al., 2018; Good and VanBriesen, 2016, 2017, and 2019; Kolb et al., 2020; McTigue et al., 2014; Ruhl et al., 2012; States et al., 2013; U.S. DOJ, 2015; U.S. EPA, 2015c).

Bromide has a low toxicity in freshwater aquatic environments compared to substances such as copper or cadmium cations. Flury and Papritz (1993) present the results from two previous studies on the median lethal toxic concentration (LC₅₀) of bromide compared to other chemicals.

- For golden orfe (*Leuciscus idus melanotus*), the LC₅₀ for bromide is greater than 7,765 mg/L, compared to 0.32 mg/L for copper and 4.5 to 35.4 mg/L for cadmium (Juhnke and Lüdemann, 1978).

- For fathead minnow (*Pimephales promelas*), the LC₅₀ for bromide is greater than 67 mg/L, compared to 0.555 to 1.4 mg/L for copper (Ewell et al., 1986).

Reviews of freshwater aquatic organism toxicology studies cite effect concentrations of bromide that range from 110 to 4,600 mg/L for single-celled organisms, 2.2 to 11,000 mg/L for invertebrates, and 7.8 to 24,000 mg/L for fish (EPRI, 2014; Flury and Papritz, 1993).

The World Health Organization (WHO) estimates that consumption of drinking water supplies with bromide concentrations below 2.0 mg/L would meet acceptable daily intake levels for both children and adults (WHO, 2009). Bromide's toxicity associated with its contribution to DBP formation in drinking water treatment and distribution systems can be of a greater concern (Krasner et al., 2006; Krasner, 2009; Regli et al., 2015; Richardson and Postigo, 2011; U.S. EPA, 2016a; Yang et al., 2014). DBPs are a broad class of compounds that form as byproducts of drinking water disinfection, and some of them have toxic properties. Bromide in source water becomes highly reactive in the presence of commonly used drinking water disinfectants and can form brominated DBPs (Br-DBPs) at low source water concentrations (Bond et al., 2014; Chang et al., 2001; Heeb et al., 2014; Landis et al., 2016; Parker et al., 2014; Richardson et al., 2007; U.S. EPA, 2016a; Wang et al., 2017; Westerhoff et al., 2004). Although multiple factors affect DBP formation, increases and decreases in source water bromide levels are typically associated with concurrent increases and decreases in both total DBP and bromide speciation levels in treated water (AWWARF and U.S. EPA, 2007; Bond et al., 2014; Cornwell et al., 2018; Ged and Boyer, 2014; Hua et al., 2006; Huang et al., 2019; Landis et al., 2016; McTigue et al., 2014; Obolensky and Singer, 2008; Pan and Zhang, 2013; Regli et al., 2015; Sawade et al., 2016; States et al., 2013; Yang and Shang, 2004; Zha et al., 2014).

The 2020 EA (U.S. EPA, 2020a) provides further details on bromide in freshwater systems and adverse impacts in source water for DWTPs.

2.1.5 Iodine/Iodide

Iodine is naturally present in coal.² Some coal-fired steam electric power plants also add iodine, in the form of iodide compounds, to their combustion processes to enhance mercury emissions control or burn refined coal amended with iodide compounds (ADES, 2016; Gadgil, 2016; ICAC, 2019; Sahu, 2017; Senior et al., 2016; Sjostrom et al., 2016; Sjostrom and Senior, 2019; Tinuum, 2020).³ Iodine volatilizes during combustion and partitions to FGD wastewaters and, to a lesser extent, to BA transport waters (ADES, 2016; ICAC, 2019; Meij, 1994; Peng et al., 2013; Sjostrom et al., 2016). In FGD wastewaters, iodine occurs as iodide/triiodide anions and elemental iodine (Sjostrom et al., 2016). Data on typical iodine concentrations in FGD wastewater and BA transport waters are limited. One study (Sjostrom et al., 2016) indicated that iodine concentrations in FGD wastewater should be below about 100 mg/L to ensure normal FGD system operation and to recover iodine for reuse.

Typical iodine levels in freshwater surface waters are less than 0.020 mg/L, though levels ranging from 0.00001 to 0.212 mg/L have been reported.⁴ In freshwater, elemental iodine dissociates to its anionic form and/or reacts with organic material to form iodinated organic compounds. Iodide is highly soluble and exhibits conservative fate and transport in freshwater (Fuge and Johnson, 1986; Moran et al., 2002).

According to available data, iodide has lower ecotoxicity in freshwater aquatic environments than other substances such as copper or cadmium cations. For golden orfe (*Leuciscus idus melanotus*), the LC₅₀ for

² Native iodine levels in coal range from 0.14 to 12.9 parts per million (ppm) (Bettinelli et al., 2002; Gluskoter et al., 1977; Good, 2018). One source states that many coals used by utility plants have iodine levels greater than 3 ppm (Sjostrom et al., 2016).

³ Addition rates are reported to range from 1 to 30 ppm and are typically less than 10 ppm (Gadgil, 2016; ICAC, 2019; Sahu, 2017; Sjostrom et al., 2016).

⁴ The highest measured levels reflect influence of irrigation water return flows in arid areas.

iodide is greater than 4,525 mg/L compared to 0.32 mg/L for copper and 4.5 to 35.4 mg/L for cadmium (Juhnke and Lüdemann, 1978). Estimates of LC₅₀ for iodide range from 860 to 8,230 mg/L for freshwater fish and from 0.17 to 0.83 mg/L for *Daphnia magna*, an aquatic invertebrate (Flury and Papritz, 1993; Laverock et al., 1995). Toxicity to single-celled organisms is reported to be similar to that of bromide (Bringmann and Kühn, 1980; Flury and Papritz, 1993). In comparison, elemental iodine toxicity is higher for freshwater fish, with LC₅₀ concentrations from 0.53 mg/L to greater than 10 mg/L, and is similar to iodide toxicity for *D. magna*, with LC₅₀ concentrations from 0.16 to 1.75 mg/L (Laverock et al., 1995; LeValley, 1982).

For humans, iodine is an essential element for thyroid hormone production and metabolic regulation. Excessive consumption can lead to hypothyroidism (diminished production of thyroid hormones), hyperthyroidism (excessive production and/or secretion of thyroid hormones), or thyroiditis (inflammation of the thyroid gland) (ATSDR, 2004). The MRL for acute and chronic oral exposure to iodide is 0.01 milligrams per kilogram per day based on endocrine effects (ATSDR, 2023).

As with bromide, most toxicity concerns for iodine/iodide are associated with its contribution to DBP formation in drinking water treatment and distribution systems. Iodine in source water becomes reactive during chlorine-, chlorine dioxide-, chloramine-, or ultraviolet (UV)-based disinfection, when it can combine with organic material in source waters to form iodinated DBPs (I-DBPs) (Bichsel and Von Gunten, 2000; Criquet et al., 2012; Dong et al., 2019; Ersan et al., 2019; Hua et al., 2006; Hua and Reckhow, 2007; Krasner, 2009; Krasner et al., 2006; Postigo and Zonja, 2019; Richardson et al., 2008; Tugulea et al., 2018; U.S. EPA, 2016a; Weinberg et al., 2002). Both iodide and iodinated organic compounds in source waters can contribute to I-DBP formation during drinking water disinfection (Ackerson et al., 2018; Dong et al., 2019; Duirk et al., 2011; MacKeown et al., 2020; Pantelaki and Voutsas, 2018; Tugulea et al., 2018). Iodate, a non-toxic iodine compound that can form in the presence of oxidants (including certain DWTP disinfectants), can also contribute to I-DBP formation under certain conditions (Dong et al., 2019; Postigo and Zonja, 2019; Tian et al., 2017; Xia et al., 2017; Yan et al., 2016; Zhang et al., 2016). I-DBP levels are influenced by multiple factors and have been found to increase with iodide or total iodine levels in source water (Criquet et al., 2012; Dong et al., 2019; Gruchlik et al., 2015; Postigo and Zonja, 2019; Tugulea et al., 2018; Ye et al., 2013; Zha et al., 2014).⁵

The 2020 EA (U.S. EPA, 2020a) provides further details on iodine and adverse impacts in source water for DWTPs.

2.2 Potential Impacts from the Evaluated Wastestreams

Changes in surface water chemistry due to contamination from steam electric power plant wastewater can harm all levels of an ecosystem, including organisms at lower trophic levels; this in turn affects the ecosystem's food web and fish inhabiting the surface water. Pollutants in surface water can bioaccumulate in aquatic organisms such as fish. When wildlife or humans ingest these aquatic organisms, they can be exposed to a higher dose of contamination than through direct exposure to the surface water. Surface water impacts associated with discharges of steam electric power plant wastewater include damage to fish populations (*i.e.*, physiological and morphological abnormalities and various behavioral, reproductive, and developmental effects), decreased diversity in insect populations, and decline of aquatic macroinvertebrate population (see Section 2.2.1). Impacts that affect humans include exceedances of National Recommended Water Quality Criteria, fish consumption advisories, designation of surface waters as impaired (limiting recreational activities), and contamination of downstream drinking water sources (see Section 2.2.2 and Section 4).

⁵ Other factors influencing I-DBP formation include pH, temperature, disinfection process type and dosage level, bromide levels, ammonium levels, organic material levels and type, and treatment and distribution system residence time.

This section provides an overview of the environmental impacts caused by exposure to pollutants in discharges of the evaluated wastestreams. It also summarizes additional studies identified as part of the literature review conducted to support this EA and the final supplemental rulemaking. Details of previous literature reviews are included in the 2015 EA (U.S. EPA, 2015a) and the 2020 EA (U.S. EPA, 2020a).

2.2.1 Ecological Impacts

Many of the pollutants in steam electric power plant wastewater (*e.g.*, arsenic, mercury, selenium) readily accumulate in exposed biota. This bioaccumulation is of particular concern due to their impact on higher trophic levels, local terrestrial environments, and transient species, in addition to the aquatic organisms directly exposed to the wastewater. Aquatic systems with long residence times and potential contamination with bioaccumulative pollutants often experience persistent environmental effects following exposure to steam electric power plant wastewater.

Population decline attributed to exposure to steam electric power plant wastewater can alter the structure of aquatic communities and cause cascading effects within the food web that result in long-term impacts to ecosystem dynamics (Rowe et al., 2002). Reductions in organism survival rates from abnormalities caused by exposure to power plant wastewater, and alterations in interspecies relationships, such as declining abundance or quality of prey, can delay ecosystem recovery until key organisms within the food web return to levels prior to power plant wastewater exposure. In a 1980 study of a creek in Wisconsin, fungal decomposition of detritus was limited due to the effects of power plant wastewater. Because of this reduction in available resources, the population of benthic invertebrates (which graze on detrital material) declined, as did benthic fish that prey upon small invertebrates (Magnuson et al., 1980).

Ecological impacts associated with exposure to steam electric power plant wastewater include lethal impacts, such as fish kills, and sublethal impacts, such as teratogenic deformities, oxidative stress, deoxyribonucleic acid (DNA) damage, reduced growth, and genotoxicity (Brandt et al., 2017 and 2019; Carlson and Adriano, 1993; Javed et al., 2016; Lemly, 2018; Rowe et al., 2002). Much of the scientific literature focuses on selenium as a key pollutant of environmental concern in steam electric power plant wastewater. Selenium can bioaccumulate to toxic levels in organisms inhabiting environments with low selenium concentrations. As studied by Lemly (1985), the extent of selenium bioaccumulation depends on the trophic level of the fish present in the water. Lemly observed that selenium accumulation increased as the trophic level increased, which potentially correlates with the observed elimination of multiple higher-trophic-level fish species. The study also found that selenium discharges also affect species diversity in receiving waters (Lemly, 1985). Selenium discharges can lead to long-term issues in ecosystems due to prolonged retention in the environment and cycling and propagation in the food chain (Brandt et al., 2019).

The sublethal effects of selenium vary widely and can affect growth, reproduction, and survival of susceptible organisms. Scientists have demonstrated that various fish and amphibian species are sensitive to elevated selenium concentrations similar to those found in steam electric power plant wastewater. In addition to lethal effects, these fish and amphibian species have developed sublethal symptoms such as accumulation of selenium in tissue (histopathological effects) and in the blood (hematological effects), resulting in decreased growth, changes in weight, abnormal morphology, and reduced hatching success (Coughlan and Velte, 1989; Lemly, 1993 and 2018; Sager and Colfield, 1984; Sorensen, 1988; Sorensen and Bauer, 1984; Sorensen et al., 1982, 1983, 1984). In addition, selenium is highly teratogenic (*i.e.*, able to disturb the growth and development of an embryo or fetus) and readily transferable from mother to egg (Chapman et al., 2009; Janz et al., 2010; Lemly, 1997a; Maier and Knight, 1994).

Although effects documented in the literature primarily focus on selenium, several studies discussed the sublethal effects of other pollutants, such as arsenic, cadmium, chromium, copper, and lead (Rowe et al., 2002), and decreased diversity in receiving water fish species (Javed et al., 2016). Sublethal effects from exposure to pollutants other than selenium in power plant wastewater can include changes to morphology (*e.g.*, fin erosion, oral deformities), behavior (*e.g.*, ability to swim, catch prey, and escape

from predators), and metabolism that can negatively affect long-term survival (Rowe et al., 2002). Vengosh et al. (2019) found concentrations of coal combustion residuals (CCR) pollutants in Sutton Lake, North Carolina, indicating the potential for unmonitored spills of coal ash into nearby receiving waters. From samples taken in 2015 and 2018, researchers found that the lake sediment contained one to two orders of magnitude higher levels of antimony, arsenic, copper, molybdenum, selenium, and thallium compared to a reference lake. Vengosh et al. (2019) noted recent hurricanes across the area may have led to flooding of ash ponds (surface impoundments) and contamination of surface waters. Researchers noted that concentrations in the sediments exceeded freshwater ecological screening standards (Vengosh et al., 2019).

In the literature reviews for this supplemental rule, the EPA identified studies that discussed concerns with bromide and halogenated DBPs' impact on ecological receptors and potential impacts from pollutants in CRL. As noted in Section 2.1.4, bromide is one of the pollutants discharged by steam electric power plants, and the discharge of bromide and iodine can lead to increased DBP formation at downstream DWTPs (see Section 2.1.5).

Since 1994, scientists noted the spread of vacuolar myelinopathy (VM), a neurological disease, in bald eagles, other birds of prey, and waterfowl. At DeGray Lake in Arkansas, more than 70 eagle mortalities were found in two years, and investigators began noticing eagles and other waterbirds with neurological impairments across the southeastern United States (Breinlinger et al., 2021). VM has also been found in other wildlife including amphibians, reptiles, and fish. Field and laboratory studies have shown that VM can be transferred up the food chain from fish to wildlife and birds of prey. Documented cases in avian species have been found near artificial waterbodies with abundant aquatic vegetation located in the southeastern United States. Breinlinger et al. (2021) conducted field studies in southeastern U.S. waters and laboratory studies to identify the causative agent of VM. The scientist showed that a neurotoxin, which they termed aetokthonotoxin (AETX), was the causative agent of VM. AETX is produced by *Aetokthonos hydrillicola* (cyanobacterium) growing on aquatic vegetation (*Hydrilla verticillata*). The researchers noted that AETX's structure has characteristics not previously observed in nature and investigated the biosynthesis of the neurotoxin. Breinlinger et al. (2021) determined that the biosynthesis of AETX depends on the bioavailability of bromide, along with other factors (e.g., temperature).

Cui et al. (2021) investigated the potential toxicity and ecological risk to freshwater organisms from exposure to halogenated DBPs. Research was prompted by the increased use of chlorine as a disinfecting agent due to the severe acute respiratory syndrome coronavirus 2 (SARS-CoV-2) outbreak and increased DBP levels in wastewater treatment effluent. The organisms studied covered three trophic levels: phytoplankton (*Scenedesmus sp.*), zooplankton (*Daphnia magna*), and fish (*Danio rerio*). Cui et al. (2021) found that *Scenedesmus sp.* were most sensitive to haloacetic acids (HAAs) and *Daphnia magna* were most sensitive to haloacetonitriles (HANs) and trihalomethanes (THMs). Cui et al. (2021) cited other research on the toxicity of brominated DBPs to aquatic organisms and findings that DBPs can have reproductive impacts on *Daphnia magna* and adversely affect embryonic development of zebrafish. Observed impacts from the DBP exposure (for most of the DBPs tested) included the following:

- Inhibited growth for phytoplankton (*Scenedesmus sp.*).
- Decreased swimming ability (immobilization) for zooplankton (*Daphnia magna*).
- Induced mortality and abnormal development for fish (*Danio rerio*).

Frankel et al. (2022) conducted a study to determine the potential impact on freshwater snails (*Planorbella duryi*) exposed to CRL trace elements (i.e., aluminum, arsenic, calcium, cadmium, chromium, copper, iron, lead, magnesium, manganese, and selenium). The study found that "exposure to environmentally relevant concentrations" of coal ash leachate caused delays in embryonic development, reduced shell width growth in juveniles, and decrease in egg deposition. Bioaccumulation of arsenic, cadmium, chromium, and lead occurred in the snails studied, with arsenic and cadmium concentrations in the tissue reaching over 85,000 and 170,000 times higher than measured in the leachate solution, respectively.

2.2.2 Human Health Effects

Exposure to pollutants can increase risk for noncancer effects in humans, including damage to the circulatory, respiratory, or digestive systems and neurological and developmental effects. Steam electric power plant wastewater contains toxic pollutants and known or suspected carcinogens (*e.g.*, arsenic and cadmium). Documented exceedances of drinking water maximum contaminant levels (MCLs) downstream of steam electric power plants, and the issuance of fish advisories in receiving waters, indicate an ongoing human health concern caused by power plant wastewater discharges. The primary exposure route investigated in this EA is through fish consumption (see Sections 3 and 4). As noted in Section 2.1, pollutants in steam electric power plant discharges can bioaccumulate in fish that are then consumed by recreational and subsistence fishers. For example, Lemly (2014) studied selenium contamination in fish found in Lake Sutton—a popular fishing location that is also used as a cooling reservoir for discharges from the L.V. Sutton Steam Plant settling pond before the water moves downstream into the Cape Fear River. Based on data collected between 1987 and 2011, the selenium concentration in bluegill (*Lepomis macrochirus*) exceeded the toxic thresholds established by researchers, and physical examination showed elevated deformities in the fish (*e.g.*, skeletal and craniofacial defects) compared to a reference lake (29 percent in Lake Sutton to 0.5 percent in the reference lake). Researchers noted similar results in morphological abnormalities at other lakes that receive power plant discharges (*e.g.*, Belews Lake and Hyco Reservoir).

In addition, groundwater and drinking water supplies can be degraded by pollutants in steam electric power plant wastewater (Cross, 1981). Power plants may dispose of or store CCR, or coal ash, in landfills or surface impoundments. Leachate and legacy wastewater (see Section 1), which contain pollutants from the CCR, can migrate from the power plant landfills and surface impoundments via the groundwater at concentrations that could contaminate public or private drinking water wells and surface waters, even years following disposal of combustion residuals (National Research Council of the National Academies, 2006).

As discussed in Sections 2.1.4 and 2.1.5, the discharge of bromide and iodine into drinking water sources is a concern due to the formation of DBPs in DWTPs and their distribution systems.

- Toxicology and epidemiology studies have documented evidence of genotoxic (including mutagenic), cytotoxic, and carcinogenic properties of DBPs, including Br-DBPs (National Toxicology Program, 2018; Richardson et al., 2007; U.S. EPA, 2016a). Studies have documented evidence of a link between DBP exposure and bladder cancer and, to a lesser degree, colon and rectal cancer, other cancers, and reproductive and developmental effects (Cantor et al., 2010; Chisholm et al., 2008; Regli et al., 2015; Richardson et al., 2007; U.S. EPA, 2016a; Villanueva et al., 2004, 2007, and 2015). Br-DBPs typically have higher toxicity than their chlorinated analogues (Cortés and Marcos, 2018; Plewa et al., 2008; Richardson et al., 2007; Sawade et al., 2016; U.S. EPA, 2016a; Yang et al., 2014). Due to bromide's reactivity and DBP toxicity, elevated bromide levels in source waters have been associated with elevated health risks from disinfected water (Hong et al., 2007; Kolb et al., 2017; Regli et al., 2015; Sawade et al., 2016; Wang et al., 2017; Yang et al., 2014). In a 2022 study, Weisman et al. (2022) estimated that approximately 9,000 of the 79,000 annual bladder cancer cases could potentially be attributed to trihalomethanes in the drinking water, with 84 percent of the approximately 9,000 cases are from drinking water systems with surface water, as opposed to groundwater, as the system's intake source.
- *In vitro* toxicology studies with bacteria and mammalian cells have documented evidence of genotoxic (including mutagenic), cytotoxic, tumorigenic, and developmental toxicity properties of I-DBPs. Individual I-DBP species have higher toxicity than their chlorinated and brominated analogues and are among the most cytotoxic DBPs identified to date (Dong et al., 2019; Hanigan et al., 2017; National Toxicology Program, 2018; Richardson et al., 2007 and 2008; Richardson and Plewa, 2020; U.S. EPA, 2016a; Wagner and Plewa, 2017; Wei et al., 2013; Yang et al., 2014). While studies have documented evidence linking disinfected drinking water and DBP exposure to adverse human health effects (see the 2020 EA: U.S. EPA, 2020a), more research is needed to characterize the contribution of I-DBPs to

these effects (Cortés and Marcos, 2018; Dong et al., 2019; Postigo and Zonja, 2019; U.S. EPA, 2016a). In a 2021 study, Long et al. (2021) concluded that iodoacetic acid exposure results in reproductive and developmental toxicity effects. Because conventional drinking water treatment processes do not effectively remove iodide from source waters and vary in their reduction of organic material levels (U.S. EPA, 2016a; Watson et al., 2015), they have the potential to generate I-DBPs when their source waters contain iodine.

2.2.3 Groundwater Impacts

Pollutants in CCR can leach into groundwater from surface impoundments and landfills. Older surface impoundments and landfills are of particular concern because they were often built without liners and leachate collection systems. Liners are typically made of synthetic material, asphalt, clay, or a composite of materials (e.g., synthetic and clay) and are designed to collect leachate and prevent groundwater contamination. CCR held in unlined surface impoundments can enter the subsurface and contaminate groundwater. Pollutants in unlined landfills, used for the dry disposal of CCRs, can also leach as precipitation flows through the residuals pile and dissolves pollutants; the CRL can eventually migrate into groundwater. The EPA has promulgated a series of rules to mitigate CCR disposal issues (e.g., seeping of pollutants into groundwater, airborne pollutants as dust, and surface impoundment failures resulting in larger coal ash spills), starting with the Disposal of Coal Combustion Residuals from Electric Utilities final rule (80 FR 21301), which established requirements for the safe disposal of CCR nationwide. Even with additional requirements in place, pollutants can still enter the groundwater when liners fail or when a disposal site is situated such that natural groundwater fluctuations come into contact with the disposed waste.

Before the CCR regulations, the EPA identified more than 30 documented cases where groundwater contamination from surface impoundments extended beyond the plant boundaries, illustrating the threat to groundwater and drinking water sources (ERG, 2015a). Based on a review of exceedances of state or federal groundwater quality standards at surface impoundments, exceedances were most often due to boron, sulfate, or arsenic (Lewis et al., 2017). In a 2016 study, Harkness et al. (2016) evaluated pollutant migration from coal ash ponds (surface impoundments) to groundwater and surface waters at sites in the southeastern United States. The evaluation found pollutants above background concentrations at the tested sites, including levels above drinking water and ecological impact standards for some surface waters. The researchers note that the closing of the coal ash surface impoundments did not necessarily stop the migration of pollutants from the surface impoundments (Harkness et. al., 2016).

Landfills pose their own groundwater contamination risks. If the landfills are not properly lined, the pollutants in CCR can leach into the soil during precipitation. In areas with acid rain, the precipitation's low pH can accelerate the leaching of contaminants into groundwater. In addition, heavy precipitation can not only accelerate leaching, but also carry pollutants in stormwater runoff, potentially contaminating groundwater or surface water resources (Andersen and Madsen, 1983). Based on a review of CCR landfill damage cases compiled by the EPA, Lewis et al. (2017) noted that all the landfills were constructed before 1990 (before the Resource Conservation and Recovery Act requirements for liners went into effect), and only four of the 32 cited landfills were fully lined. As with groundwater exceedances from surface impoundments, the most common pollutants with exceedances included boron and sulfate. Iron and manganese had exceedances at more than half of the landfills (Lewis et al., 2017).

Frankel et al. (2023) evaluated potential impacts to Quantico Creek, a tributary to the Chesapeake Bay, from the leakage of a nearby CRL landfill and coal ash surface impoundments. Samples taken from the creek near the CRL landfill and coal ash surface impoundments were compared to upstream and downstream locations. Researchers found elevated concentrations of the parameters in the sediment but not the surface water, with the highest concentrations of pollutants including arsenic, boron, cadmium, chromium, copper, selenium, and zinc in samples adjacent to the coal ash surface impoundments. Ecological impacts included reduced species diversity and increased concentrations of aluminum, cadmium, and zinc in the tissues of banded killfish (*Fundulus diaphanous*) near the coal ash landfill

compared to upstream and downstream sites. Frankel et al. (2023) did not find arsenic, chromium, or selenium in the fish tissue samples.

2.2.4 CCR Surface Impoundments as Attractive Nuisances

An “attractive nuisance” is an area or habitat that attracts wildlife and is contaminated with pollutants at concentrations high enough to potentially harm exposed organisms. Two methods of handling steam electric power plant wastewater, surface impoundments and constructed wetlands, are classified as lentic systems supporting aquatic vegetation and organisms. These methods have been known to attract wildlife from other terrestrial habitats and therefore can be considered attractive nuisances. For example, a surface impoundment can affect local wildlife as well as transient species that might rely on them during critical reproduction periods such as seasonal breeding events (Rowe et al., 2002). Exposure to steam electric power plant wastewater during sensitive life cycle events is a concern, given that it has been associated with complete reproductive failure in various vertebrate species (Cumbie and Van Horn, 1978; Gillespie and Baumann, 1986; Lemly, 1997b; Pruitt, 2000).

Several studies have shown that terrestrial fauna nesting near CCR surface impoundments can have higher levels of arsenic, cadmium, chromium, lead, mercury, selenium, strontium, and vanadium than the same species at reference sites (Bryan et al., 2003; Burger et al., 2002; Hopkins et al., 1997, 1998, 2000, 2006; Nagle et al., 2001; Rattner et al., 2006). Field studies have also documented adverse effects on reproduction for turtles and toads living near selenium-laden CCR surface impoundments (Hopkins et al., 2006; Nagle et al., 2001).

In addition to being attractive nuisances, surface impoundments near surface waters can be a source of coal ash spills that damage the environment, ecosystems, and downstream waters. Concerns with these spills include the large economic loss and costs to remediate, along with ecological damage, potential effects on human health, recreational impacts, and losses of consumptive use and aesthetic value. Researchers and state agencies have monitored the receiving water ecosystems following coal ash spills, notably the 2008 coal ash spill that affected the Emory River and Clinch River and the 2014 coal ash spill to the Dan River.

- Following the 2008 coal ash spill at the Tennessee Valley Authority’s Kingston Plant, the Tennessee Department of Environment and Conservation found exceedances of the more stringent criteria for chronic exposure of fish and aquatic life at least once in January 2009 for several metals (*e.g.*, aluminum, cadmium, iron, and lead). Seven months after the spill, all fish collected had concentrations of selenium above a toxic threshold, and most were still contaminated at that level 14 months after the spill. Twenty-one months after the spill, a high percentage of fish were found with lesions, deformities, and infections, all symptoms of extreme stress. In addition, studies have shown elevated levels of arsenic and mercury in sediments near the ash spill, as well as selenium levels exceeding the MCL in three wells underneath the Kingston Plant’s coal ash disposal area, ash processing area, and gypsum disposal facility (U.S. EPA, 2014). In a study eight years after the coal ash spill, researchers determined downstream sediment concentrations of arsenic and selenium are likely from the coal ash; however, other metals in downstream sediment are likely from other anthropogenic sources (Ramsey et al., 2019).
- In 2011 and 2012, Van Dyke et al. (2017) measured trace contaminant concentrations in freshwater turtles in the Emory River, Clinch River, and a reference (unaffected) river. Turtles in the Emory River and Clinch River had higher concentrations of arsenic, copper, iron, mercury, manganese, selenium, and zinc than turtles in the reference river. However, the concentrations were low relative to values known to be toxic to other vertebrates. Researchers stated that they found little evidence that the residual coal ash in the affected rivers had an effect on contaminant bioaccumulation in turtles.
- Ku et al. (2020) evaluated mercury concentration in the Dan River 17 to 29 months following the coal ash spill, which was much smaller than the spill at the Emory and Clinch rivers. They found that mercury contamination in the Dan River surface sediments (0–16 centimeters) could be accounted for by organic matter, rather than the coal ash spill. The study also examined methylmercury

bioaccumulation in invertebrates and fish and did not find evidence of elevated methylmercury bioaccumulation. The researchers concluded that the mercury contamination from the coal ash spill was largely absent in the surface sediment and biota three years after the spill. Alternatively, they suggested that the mercury from the coal ash spill was not typically bioavailable.

- Silva et al. (2023) studied environmental and ecological contamination from ash surface impoundments at a retired coal-fired power plant and decommissioned nuclear reactor. Researchers sampled beetles associated with carrion in west central South Carolina and found substantial trace elements within the beetles' organs and tissues. Compared to the uncontaminated (control) site, the beetles had higher levels of arsenic, selenium, and thallium. Beetles at the uncontaminated site had higher levels of chromium, copper, and nickel.

3. Environmental Assessment Methodology

This section presents the EPA's evaluation of environmental concerns and potential exposures to pollutants commonly found in wastewater discharges from steam electric power plants. It describes the following:

- Pollutant loadings for the evaluated wastestreams.
- Pollutant exposure pathways.
- Methodologies used to quantify the environmental, ecological, and human health effects of pollutants discharged to surface waters from the evaluated wastestreams.
- Environmental assessment (EA) scope (*i.e.*, plants and immediate receiving waters).

3.1 Pollutant Loadings for the Evaluated Wastestreams

As discussed in Section 2, the pollutants commonly found in steam electric power plant wastewater—such as metals, total dissolved solids (TDS), and halogens—can result in impacts to water quality, aquatic life, wildlife, and human health. The EPA analyzed three regulatory options for the final supplemental rule, as shown in Table VII-1 of the rule's preamble. The EPA estimated pollutant loadings for the evaluated wastestreams considered as part of the supplemental rule as described in Section 6 of the technical development document (TDD) (U.S. EPA, 2024a). The EPA calculated plant-specific and receiving-water specific *baseline* and *regulatory option* pollutant loadings (in pounds per year) for flue gas desulfurization (FGD) wastewater, bottom ash (BA) transport water, combustion residual leachate (CRL), and legacy wastewater being discharged to surface water or through publicly owned treatment works (POTWs) to surface water.

Most steam electric power plants (over 95 percent) evaluated for the supplemental rule discharge directly to surface water. Six plants reported transferring BA transport water, FGD wastewater, or CRL to a POTW rather than discharging directly to surface water.⁶ For these POTW transfers, the EPA adjusted the baseline and regulatory option loadings to account for pollutant removals expected during treatment at the POTW for each analyte. See Section 6 of the TDD for industry-wide annual baseline pollutant loadings for the evaluated wastestreams, as well as the reductions in pollutant loadings (relative to baseline) for each of the regulatory options.

The EPA used these pollutant loadings as inputs to support the quantitative evaluation of environmental impacts via the surface water exposure pathway (see Section 3.2). Table 2 presents baseline pollutant loadings and the estimated reduction in pollutant loadings under the evaluated regulatory options for select pollutants. The memorandum *Pollutant Loadings Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024h) discusses the EPA's methodology for estimating pollutant loadings for each immediate receiving water.

⁶ The EPA excluded CRL discharges at one plant from the EA that indirectly discharges to a POTW that does not discharge to any receiving waters, and one indirect discharging plant is only included in the proximity analysis (see U.S. EPA, 2024f).

Table 2. Estimated Annual Baseline Mass Pollutant Loadings and Estimated Reduction in Loadings Under Regulatory Options for the Evaluated Wastestreams^a

Pollutant	Estimated Baseline Pollutant Loadings (lb/year)	Estimated Reduction in Pollutant Loadings Relative to Baseline (lb/year)		
		Option A	Option B	Option C
Aluminum	60,400	45,000	58,600	59,300
Arsenic	777	513	700	726
Boron	7,140,000	5,450,000	5,770,000	6,492,000
Bromide (min) ^b	1,310,000	1,150,000	1,150,000	1,310,000
Bromide (max) ^b	6,810,000	6,380,000	6,380,000	6,810,000
Cadmium	553	152	512	529
Chlorides	223,000,000	175,000,000	180,000,000	203,000,000
Chromium	21,100	20,800	21,000	21,000
Copper	398	181	348	365
Iodine (min) ^b	86,100	76,200	76,200	86,100
Iodine (max) ^b	269,000	250,000	250,000	269,000
Iron	300,000	287,000	299,000	299,000
Lead	230	138	187	200
Magnesium	103,000,000	81,700,000	82,900,000	93,500,000
Manganese	648,000	301,000	565,000	606,000
Mercury	40.0	11.5	38.5	38.8
Molybdenum	22,500	19,700	21,300	31,800
Nickel	3,430	693	3,320	3,350
Nitrogen, total ^c	522,000	194,000	194,000	218,000
Phosphorus, total	12,100	8,930	8,930	9,980
Selenium	4,810	205	1,970	2,080
Thallium	781	245	664	695
Total dissolved solids	806,000,000	588,000,000	656,000,000	734,000,000
Vanadium	19,600	19,400	19,500	19,600
Zinc	6,570	2,040	6,310	6,400

Sources: U.S. EPA, 2024a, 2024g, and 2024h.

Abbreviations: lb/year (pounds per year).

Note: Pollutant loadings and removals are rounded to three significant figures.

a—Includes a subset of all steam electric power-generating pollutants of concern. The EPA selected the pollutants listed in this table based on the following factors: presence of the pollutant in the evaluated wastestreams; documented elevated levels of the pollutant in surface waters or wildlife from exposure to steam electric power plant wastewater; and magnitude of the pollutant loadings to receiving waters.

b—The EPA did not identify data indicating the specific halogen additive (*i.e.*, bromine or iodine) used at each plant to reduce mercury emissions. Therefore, the EPA estimated potential ranges of bromide and iodine loadings.

c—Total nitrogen loadings are the sum of ammonia and nitrate-nitrite (as N) loadings from FGD wastewater, nitrate-nitrite (as N) and total Kjeldahl nitrogen (TKN) loadings from BA transport water, and nitrate-nitrite (as N) loadings from legacy wastewater.

The pollutants with the greatest estimated reductions in annual mass loadings under the final rule (Option B) are TDS (656 million pounds per year, or lb/year, decrease relative to baseline), chlorides (180 million lb/year decrease), magnesium (83 million lb/year decrease), bromide (between 1.15 and 6.38 million lb/year decrease),⁷ and boron (5.77 million lb/year decrease).

Implementation timing under the final rule for each plant varies by wastestream, subcategorization, and the plant's permit renewal schedule. See the preamble for further discussion of the regulatory options and associated deadlines. Due to the differing timelines for individual wastestreams and plants, the net reduction in pollutant loadings and corresponding environmental changes will be staggered over time as the plants implement control technologies. The EA presents the EPA's estimates of environmental improvements associated with each regulatory option using steady-state annual average pollutant loadings reflecting full implementation of the effluent limitations guidelines and standards. Therefore, the results presented in the EA may underestimate short-term environmental impacts for the period before full implementation of the final rule during which plants transition from current discharges to discharges associated with full implementation. In addition, the EA did not evaluate the impacts of any discharges other than the four evaluated wastestreams; therefore, the pollutant loadings and subsequent quantitative analyses do not represent a complete assessment of environmental impacts from steam electric power plants.

3.2 Pollutant Exposure Pathways

An exposure pathway is defined as the route a pollutant takes from its source (*e.g.*, combustion residual surface impoundments) to its endpoint (*e.g.*, a surface water), and how receptors (*e.g.*, fish, wildlife, or people) can come into contact with it. Exposure pathways are typically described in terms of five components:

- Source of contamination (*e.g.*, steam electric power plant wastewater).
- Environmental pathway—the environmental medium or transport mechanism that moves the pollutant away from the source through the environment (*e.g.*, discharges to surface waters).
- Point of exposure—the place (*e.g.*, private drinking water well) where receptors (*e.g.*, people) come into contact with a pollutant from the source of contamination.
- Route of exposure—the way (*e.g.*, ingestion, skin contact) receptors come into contact with the pollutant.
- Receptor population—the aquatic life, wildlife, or people exposed to the pollutant.

⁷ The EPA did not identify data indicating the specific halogen additive (*i.e.*, bromine or iodine) used at each plant to reduce mercury emissions. Therefore, the EPA estimated potential ranges of bromide and iodine loadings. The EPA defined the ranges' lower and upper bounds as follows (U.S. EPA, 2024a and 2024g):

- Bromide (min): Bromide loadings in BA transport water and FGD wastewater from native coal content and the addition of bromide in the flue gas (*i.e.*, as brominated activated carbon). The EPA analyzed additional CRL data that included bromide concentrations in CRL at five plants; however, more than half of the samples were nondetect values. Therefore, the EPA did not estimate bromide loadings in CRL. See the memorandum *2024 Final Rule - Combustion Residual Leachate Analytical Data Evaluation* (U.S. EPA, 2024m).
- Bromide (max): Same as "Bromide (min)" plus bromide loadings due to the use of refined coal or halogen addition at the EGU. Assumes all plants burning refined coal or adding halogens at the EGU use bromine additives.
- Iodine (min): Iodine loadings in FGD wastewater from native coal content only. The EPA had insufficient data to estimate iodine loadings in other receiving waters.
- Iodine (max): Same as "Iodine (min)" plus iodine loadings due to the use of refined coal or halogen addition at the EGU. Assumes all plants burning refined coal or adding halogens at the EGU use iodine additives.

The exposure pathway plays an important role in determining the potential effects of steam electric power plant wastewater on the environment. For example, the physical and chemical characteristics of receiving waters can affect the fate and transport of pollutants from combustion residual surface impoundments to the environment and ultimately impact how the pollutants interact with the biological community.

The EPA identified four primary exposure pathways of concern for steam electric power plant wastewater entering the environment. Table 3 presents the environmental pathways, routes of exposure, and environmental concerns identified from the literature review and the types of analyses conducted to determine the impacts under baseline and potential environmental improvements under the regulatory options. In its analyses to determine environmental impacts and improvements, the EPA evaluated each environmental concern via a given route of exposure and pathway individually (*i.e.*, the combined impact of multiple routes of exposure were not jointly evaluated).

Table 3. Steam Electric Power Plant Wastewater Environmental Pathways and Routes of Exposure Evaluated in the Environmental Assessment for the Final Supplemental Rule

Environmental Pathway	Route of Exposure	Environmental Concern	Analysis to Determine Environmental Impact
Steam electric power plant wastewater discharges to surface waters	Direct contact with surface water	Toxic effects on aquatic organisms ^a	Water quality impacts analysis (quantitative)—see Sections 4.1.1 and 4.3
	Ingestion of surface water	Degradation of surface water quality used as intake to drinking water plants	
	Direct contact with sediment	Toxic effects on benthic organisms ^a	Wildlife impacts analysis (quantitative)—see Sections 4.1.2 and 4.3
	Consumption of aquatic organisms	Bioaccumulation of contaminants and resulting toxic effects on wildlife ^a	
		Toxic effects on humans consuming contaminated fish ^a	
Degradation of fish availability for recreational and subsistence fishers	Human health impacts analysis (quantitative)—see Sections 4.1.3 and 4.2		
Uncollected CRL infiltration to nearby surface waters from combustion residual landfill	Direct contact with surface water or sediment	Toxic effects on humans and aquatic wildlife ^a	Groundwater quality impacts (qualitative)—see Section 2.2.3
Uncollected CRL entering groundwater from combustion residual landfill	Ingestion of groundwater	Changes in groundwater quality	
		Contaminated private drinking water wells	
Combustion residual surface impoundment	Direct contact with or ingestion of surface water	Toxic effects on wildlife ^a	Attractive nuisances (qualitative)—see Section 2.2.4
		Bioaccumulation of contaminants in wildlife	

a—The term “toxic effects” refers to impacts upon exposure, ingestion, inhalation, or assimilation into any organism, either directly from the environment or indirectly by ingestion through food chains. These effects can include death, disease, behavioral abnormalities, cancer, genetic mutations, physiological malfunctions (including malfunctions in reproduction), or physical deformations, in receptors (*e.g.*, aquatic organisms, wildlife, humans) or their offspring.

3.3 Environmental Impacts Selected for Qualitative and Quantitative Assessments in the EA

The EPA used both qualitative and quantitative assessments to describe the potential environmental impacts of the evaluated wastestreams (*i.e.*, FGD wastewater, BA transport water, CRL, and legacy wastewater) from steam electric power plants:

- Qualitative analysis focused on the impacts of uncollected CRL on groundwater quality and the potential for combustion residual surface impoundments to serve as attractive nuisances. Section 2.2.3 describes the EPA's findings on the potential for uncollected CRL to cause changes in groundwater quality and contaminate drinking water sources. Section 2.2.4 presents the EPA's findings on the potential toxic effects and bioaccumulation of contaminants in wildlife exposed to combustion residual surface impoundments.
- Quantitative analyses focused on the surface water exposure pathway. The EPA conducted a proximity analysis to determine whether evaluated wastestreams discharge into sensitive environments. See Section 3.5.

The EPA also evaluated the following wildlife and human health impacts caused by discharges of the evaluated wastestreams to surface waters under baseline as well as the potential reductions in those impacts under the regulatory options:

- Wildlife impacts:
 - Potential toxic effects to aquatic life based on changes in surface water quality—specifically, exceedances of the acute and chronic National Recommended Water Quality Criteria (NRWQC) for freshwater aquatic life.
 - Potential toxic effects on sediment biota based on changes in sediment quality within surface waters—specifically, exceedances of threshold effect concentrations (TECs) for sediment biota.
 - Bioaccumulation of contaminants and potential toxic effects on wildlife from consuming contaminated aquatic organisms—specifically, exceedances of no effect hazard concentrations (NEHCs), indicating a potential risk of reduced reproduction rates in piscivorous wildlife.
- Human health impacts:
 - Exceedances of the human health NRWQC based on two standards: (1) the standard for the consumption of water and organisms and (2) the standard for the consumption of organisms only.
 - Exceedances of drinking water maximum contaminant levels (MCLs). Although MCLs apply to drinking water produced by public water systems and not surface waters themselves, the EPA identified the extent to which immediate receiving waters exceeded an MCL as an indication of the degradation of the overall water quality following exposure to the evaluated wastestreams.
 - Elevated cancer risk due to consuming fish caught from contaminated receiving waters—specifically, instances where the calculated lifetime excess cancer risk due to inorganic arsenic is greater than one excess cancer case risk per one million lifetimes (also expressed as 10^{-6}).
 - Elevated noncancer health risks (*e.g.*, reproductive or neurological impacts) due to consuming fish caught from contaminated receiving waters—specifically, instances where the calculated average daily dose of a pollutant exceeds the oral reference dose (RfD) for that pollutant.

The EPA used its Immediate Receiving Water (IRW) Model to perform the quantitative assessment. Section 3.4 provides an overview of the modeling. Section 3 and Appendices C, D, and E of the 2020 EA (U.S. EPA, 2020a) provide more details on the IRW Model.

The EPA also evaluated additional wildlife and human health impacts resulting from changes in surface water quality, including impacts on threatened and endangered species, changes in ecosystem services, and neurological effects from exposure to lead and mercury. The methodologies and results of these analyses are presented in the BCA Report (U.S. EPA, 2024b). All analyses compare reductions under the regulatory options to baseline.

3.4 Overview of the IRW Model

The Immediate Receiving Water (IRW) Model is an integrated series of modules that utilize existing peer-reviewed methodologies and datasets to estimate environmental and human health risk resulting from wastewater releases. The EPA used the IRW Model to conduct the quantitative assessment of potential wildlife and human health impacts described in Section 3.3. This is the same model—including parameters and benchmark values—described in the 2020 EA (U.S. EPA, 2020a). It is a steady-state equilibrium-partitioning model that evaluates impacts within the immediate surface water⁸ where discharges occur. An equilibrium-partitioning model assumes that dissolved and sorbed pollutants in a receiving water will quickly attain equilibrium in the immediate vicinity of the discharge point because they dissolve or sorb in the surface water faster than they can be transported or dispersed outside that area. The model also assumes that the equilibrium state for each pollutant can be represented by a partition coefficient that divides the total mass of a pollutant in the waterbody into four compartments:

- Constituents dissolved in the water column.
- Constituents sorbed onto suspended solids in the water column.
- Constituents sorbed onto sediments at the bottom of the waterbody.
- Constituents dissolved in pore water in the sediments at the bottom of the waterbody.

As described in Section 5 of the 2015 EA (U.S. EPA, 2015a), the EPA developed the IRW Model to quantify the environmental impacts to surface waters, wildlife, and human health from the wastestreams evaluated for the regulatory options. In developing the model, the EPA considered the type of receiving waters commonly affected by steam electric power plants and the pollutants typically found in the evaluated wastestreams. The IRW Model quantified the environmental risks within rivers/streams and lakes/ponds/reservoirs and evaluated impacts from nine toxic, bioaccumulative pollutants: arsenic, cadmium, copper, lead, mercury, nickel, selenium, thallium, and zinc. Section 4.1 presents the results of the IRW Model analyses based on baseline and regulatory option pollutant loadings for the evaluated wastestreams, along with the limitations and uncertainties of the IRW Model.

3.4.1 Structure of the IRW Model

The IRW Model has three interrelated modules: the Water Quality Module, the Wildlife Module, and the Human Health Module, which are described in further detail in this section. Figure 1 provides an overview of the model's inputs and the connections among the three modules.

- The Water Quality Module uses plant-specific input data (annual average pollutant loadings and cooling water flow rates) and receiving-water-specific input data (*e.g.*, annual average flow rate, lake volume) to calculate annual average total and dissolved pollutant concentrations in the water column

⁸ The lengths of the immediate receiving waters for the EA, as defined in the National Hydrography Dataset Plus (NHDPlus) Version 2, range from about 0.20 to 18 miles. The upstream and downstream boundaries are defined in NHDPlus Version 2, and each plant outfall is located somewhere along the associated immediate receiving water (*i.e.*, the outfalls are not specifically indexed to the upstream end, midpoint, or downstream end). See the memorandum *Receiving Waters Characteristics Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024f) for details on the immediate discharge zone and length of stream reach represented at each discharge location.

and sediment. The module compares these concentrations to selected water quality benchmark values (NRWQC and MCLs) as an indicator of potential impacts on aquatic life and human health.

- The Wildlife Module uses the annual average water column pollutant concentrations from the Water Quality Module to calculate the bioaccumulation of pollutants in fish tissue, providing results for both trophic level 3 (T3) and trophic level 4 (T4) fish.⁹ The module compares these concentrations, and the sediment concentrations calculated by the Water Quality Module, to benchmark values that represent potential impacts on exposed sediment biota (TECs)¹⁰ and piscivorous wildlife (NEHCs). The EPA chose minks and eagles as representative piscivorous wildlife that consume T3 and T4 fish, respectively.
- The Human Health Module uses the fish tissue concentrations from the Wildlife Module to calculate noncancer and cancer risks to human populations from consuming fish caught from contaminated receiving waters. The EPA performed this analysis using two sets of fish consumption rates:¹¹
 - A “standard cohort” data set with consumption rates for recreational fishers and subsistence fishers (and their families), with separate age categories for adult and child fishers. Subsistence fishers are people who rely on self-caught fish for a larger share of their food intake than recreational fishers.
 - A data set with consumption rates for recreational and subsistence fishers in different race/ethnicity categories (non-Hispanic White; non-Hispanic Black; Mexican-American; other Hispanic; and other, including multiple races). The EPA used this data set to evaluate whether the human health impacts under baseline or reductions under the regulatory options (relative to baseline) will disproportionately affect minority groups.¹²

Appendices C, D, and E to the 2020 EA (U.S. EPA, 2020a) describe the IRW Model equations, input data, and environmental parameters in detail. The appendices also describe the limitations and assumptions for each module. Section 5.1 of the 2015 EA (U.S. EPA, 2015a) provides more information on the IRW Model, including a detailed discussion of the equilibrium-partition modeling methodology used in the Water Quality Module.

⁹ T3 fish (*e.g.*, carp, smelt, perch, catfish, sucker, bullhead, sauger) are those that primarily consume invertebrates and plankton, while T4 fish (*e.g.*, salmon, trout, walleye, bass) are those that primarily consume other fish.

¹⁰ In the case of the TEC for selenium, exceedances of the TEC represent potential impacts on higher trophic levels due to consumption of sediment biota with elevated levels of selenium.

¹¹ See the memorandum *Fish Consumption Rates Used in the EA Human Health Module* (ERG, 2015b) for details on the selection of fish consumption rates for these analyses.

¹² The EPA also conducted an environmental justice (EJ) analysis using data from the EPA’s EJSscreen, the EA, and the benefits analysis. See *Environmental Justice Analysis for Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category* (U.S. EPA, 2024d) for more details.

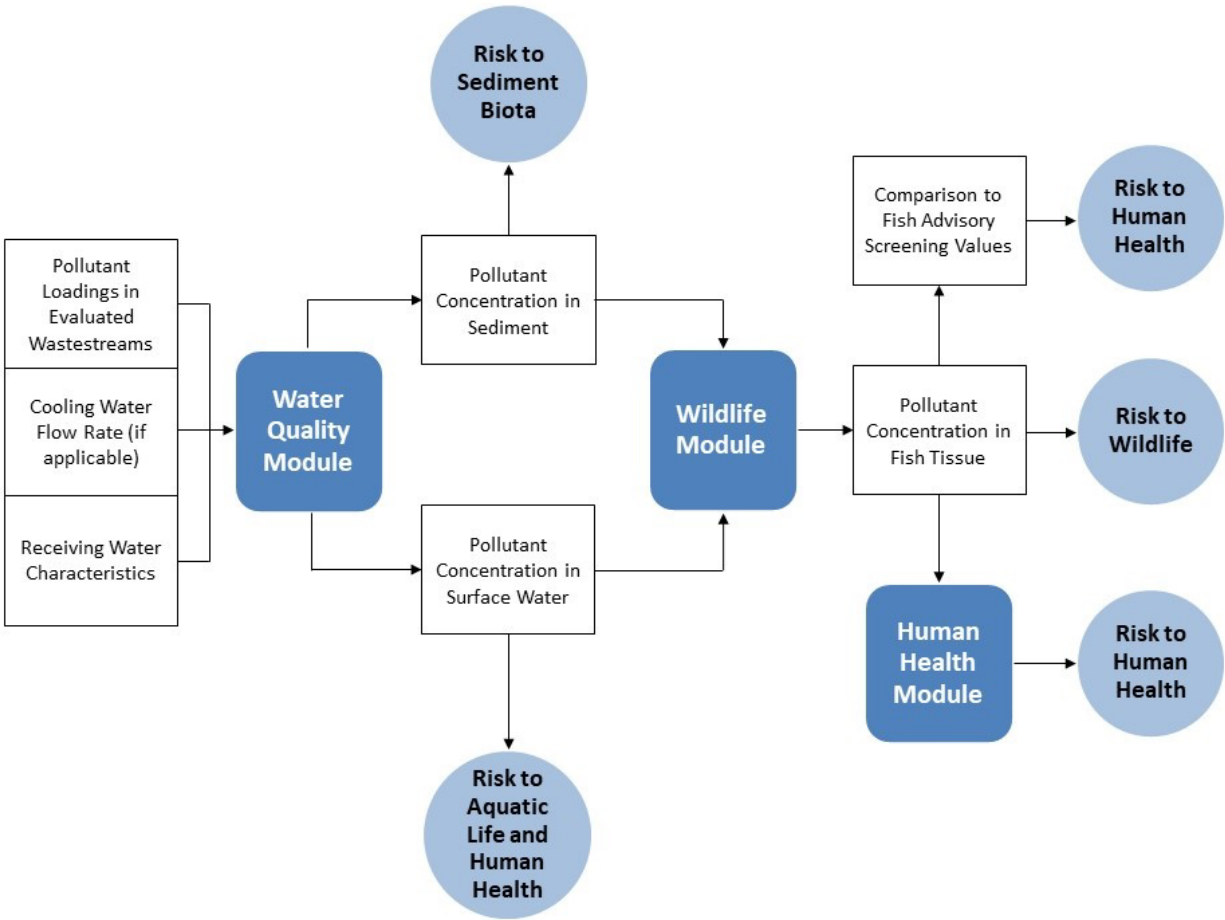


Figure 1. Overview of the IRW Model

3.4.2 Pollutants Evaluated by the IRW Model

The IRW Model analyzed nine toxic pollutants, all of which can bioaccumulate in fish and impact wildlife and human receptors via fish consumption. These pollutants were arsenic, cadmium, copper, lead, mercury, nickel, selenium, thallium, and zinc. The EPA evaluated the same pollutants in the 2020 EA. Table 4 through Table 6 include the benchmarks used in the IRW Model. The EPA identified two updates to benchmarks and incorporated these revised values in the IRW Model:

1. The EPA vacated the cadmium aquatic life criteria (freshwater, chronic) as documented in U.S. EPA (2016c). For the final rule EA, the EPA revised the benchmark as documented in U.S. EPA (2001), which is also the value used in the 2015 EA.
2. The Agency for Toxic Substances and Disease Registry *Minimal Risk Levels (MRLs)* included an updated oral reference dose for copper. The EPA revised the benchmark to be 0.02 milligrams per kilogram body weight per day (mg/kg-day) (ATSDR, 2023).

Table 4. Water Quality Benchmarks: NRWQC and MCLs

Pollutant	FW Acute NRWQC ^{a,b,c} (mg/L)	FW Chronic NRWQC ^{a,b,c} (mg/L)	HH WO NRWQC ^{a,b} (mg/L)	HH O NRWQC ^{a,b} (mg/L)	MCL ^{a,d} (mg/L)
Arsenic	0.34	0.15	0.000018 ^e	0.00014 ^e	0.01
Cadmium	0.0018 ^{f,g}	0.00025 ^{f,g}	—	—	0.005
Copper	0.014 ^h	0.009 ^h	1.3	—	1.3 (action level); 1.0 ⁱ
Lead	0.065 ^f	0.0025 ^f	—	—	0.015 (action level)
Mercury	0.0014	0.00077	—	—	0.002 ^e
Nickel	0.47 ^f	0.052 ^f	0.61	4.6	—
Selenium	Lentic: 0.045 ^j Lotic: 0.094 ^j	Lentic: 0.0015 ^k Lotic: 0.0031 ^k	0.17	4.2	0.05
Thallium	—	—	0.00024	0.00047	0.002
Zinc	0.12 ^f	0.12 ^f	7.4	26	5 ^l

Sources: U.S. EPA, 2001, 2009a, 2009b, 2016b, 2016c, and 2020c.

Abbreviations: FW (freshwater); HH O (human health organisms only); HH WO (human health water and organisms); MCL (maximum contaminant level); mg/L (milligrams per liter); NRWQC (National Recommended Water Quality Criteria).

a—"—" designates instances where a benchmark value does not exist for the pollutant.

b—Unless otherwise noted, pollutant concentrations were compared to NRWQC from the EPA's *National Recommended Water Quality Criteria* (U.S. EPA, 2009b).

c—Benchmark value is expressed in terms of the dissolved pollutant in the water column. For all pollutants except selenium, this is calculated using a total-to-dissolved conversion factor (U.S. EPA, 2009b).

d—Unless otherwise noted, pollutant concentrations were compared to the MCL from the EPA's *National Primary Drinking Water Regulations* (U.S. EPA, 2009a).

e—Benchmark value is for inorganic form of pollutant.

f—The FW NRWQC for this metal is expressed as a function of hardness (mg/L) in the water column. The values given here correspond to a hardness of 100 mg/L.

g—The cadmium benchmark values are from the EPA's *Aquatic Life Ambient Water Quality Criteria for Cadmium—2016* (U.S. EPA, 2016c) for FW acute NRWQC and the EPA's *Update of Ambient Water Quality Criteria for Cadmium* (U.S. EPA, 2001) for FW chronic NRWQC.

h—For this analysis, the EPA calculated FW NRWQC for copper using the Biotic Ligand Model and input water quality data that are representative of the ecoregions containing surface waters that receive discharges of the evaluated wastestreams (and their downstream waters) (U.S. EPA, 2020c).

i—The EPA evaluated both the action level of 1.3 mg/L and the secondary (nonenforceable) drinking water standard of 1.0 mg/L for copper (U.S. EPA, 2020d). The results presented in Section 4 and Attachment A are based on the number of immediate receiving waters with exceedances of the lower secondary drinking water standard (1.0 mg/L).

j—The selenium benchmark values are based on the NRWQC from the EPA's *Aquatic Life Ambient Water Quality Criteria for Selenium—Freshwater 2016* (U.S. EPA, 2016b). The selenium acute NRWQC, as calculated here, assumes a background selenium concentration of zero and an intermittent exposure duration of one day, which is the shortest exposure period to be used when applying the criterion. This serves as an intermittent exposure element of the chronic water quality criterion, intended to address short-term exposures that contribute to chronic effects through selenium bioaccumulation. "Lentic" pertains to still or slow-moving water, such as lakes or ponds. "Lotic" pertains to flowing water, such as streams and rivers.

k—The selenium benchmark values are based on the NRWQC from the EPA's *Aquatic Life Ambient Water Quality Criteria for Selenium—Freshwater 2016* (U.S. EPA, 2016b). The selenium chronic water column NRWQC applies only in the absence of fish tissue measurements. Use of this water column benchmark value may therefore over- or underestimate the number of exceedances.

l—The EPA has not defined an MCL or action level for zinc. This benchmark value represents the secondary (nonenforceable) drinking water standard for zinc (U.S. EPA, 2020d).

Table 5. Sediment Biota and Wildlife Benchmarks: TECs and NEHCs

Pollutant	TEC (mg/kg) ^a	NEHC for Minks (T3 Fish) (µg/g) ^b	NEHC for Eagle (T4 Fish) (µg/g) ^b
Arsenic	9.79	7.65	22.4
Cadmium	0.99	5.66	14.7
Copper	31.6	41.2	40.5
Lead	35.8	34.6	16.3
Mercury/methylmercury	0.18	0.37 ^c	0.5 ^c
Nickel	22.7	12.5	67.1
Selenium	2	1.13	4
Thallium	— ^d	— ^d	— ^d
Zinc	121	904	145

Abbreviations: mg/kg (milligrams per kilogram); NEHC (no effect hazard concentration); T3 (trophic level 3); T4 (trophic level 4); TEC (threshold effect concentration); µg/g (micrograms per gram).

a—Sources: Lemly (2018) for selenium; MacDonald et al. (2000) for all other pollutants.

b—Source: USGS, 2008.

c—No NEHC benchmark for methylmercury. The EPA compared the modeled methylmercury concentrations to the total mercury NEHC, which may underestimate the impact to wildlife.

d—No benchmark value identified; pollutant excluded from evaluation.

Table 6. Human Health Benchmarks: Oral RfDs and CSFs

Pollutant	Oral RfD (mg/kg-day)	CSF (mg/kg-day) ⁻¹	Notes
Arsenic, inorganic	3.00×10^{-4}	1.50	Oral RfD and CSF for drinking water ingestion
Cadmium	1.00×10^{-3}	— ^a	Oral RfD for food consumption
Copper	2.00×10^{-2}	— ^a	Used the intermediate oral MRL as the oral RfD (ATSDR, 2023)
Lead, total	— ^b	— ^a	
Methylmercury	1.00×10^{-4}	— ^a	Oral RfD for fish consumption only
Nickel	2.00×10^{-2}	— ^a	Oral RfD for soluble salts; used for food consumption
Selenium	5.00×10^{-3}	— ^a	Oral RfD for food consumption
Thallium	1.00×10^{-5}	— ^a	Used value cited in U.S. EPA (2012), for soluble thallium as the oral RfD; used for chronic oral exposure
Zinc	3.00×10^{-1}	— ^a	Oral RfD for food consumption

Sources: ATSDR (2023) for copper, U.S. EPA (2012) for thallium, and U.S. EPA (2019) for all other pollutants.

Abbreviations: CSF (cancer slope factor); mg/kg-day (milligrams per kilogram body weight per day); MRL (minimal risk level); RfD (reference dose).

a—No benchmark value identified; pollutant excluded from evaluation.

b—As documented in IRIS (<https://www.epa.gov/iris>), the EPA concluded that it was inappropriate to develop an RfD as some of the effects from lead exposure, “particularly changes in the levels of certain blood enzymes and in aspects of children’s neurobehavioral development, may occur at blood lead levels so low as to be essentially without a threshold.” The CDC identified 10 micrograms per deciliter (µg/dL) as the blood lead level of concern in children; see the *BCA Report* (U.S. EPA, 2024b) for the EPA’s analysis of lead impacts.

Like the 2020 EA, this EA did not use water quality modeling to assess the impacts associated with discharges of TDS, bromides, chlorides, or nutrients (total nitrogen and total phosphorus). The EPA did not have partition coefficients needed to model the pollutants in receiving water using the equilibrium-partition equations presented in Appendix C of the 2020 EA (U.S. EPA, 2020a). The EPA did include some of these pollutants in the surface water quality modeling of immediate and downstream waters, which was performed for the economic benefits analysis (see the BCA Report, U.S. EPA, 2024b).

3.5 Proximity Analysis

The pollutant loadings, ecological impacts, and human health concerns discussed in Section 2 and Section 3.2 are also of concern due to the proximity of many steam electric power plants to sensitive environments where the characteristics of plant wastewater may contribute to the impairment of water quality (e.g., 303(d)-listed waters and waters with fish advisories) or pose a threat to threatened and endangered species (see the BCA Report, U.S. EPA, 2024b). The EPA identified the number of surface waters that receive discharges of the evaluated wastestreams and are located near the following sensitive environments:

- Immediate receiving waters that states, territories, and authorized tribes have identified, pursuant to section 303(d) of the Clean Water Act (CWA), as impaired waterbodies that can no longer meet their designated uses (e.g., drinking, recreation, aquatic habitat) due to pollutant concentrations above water quality standards. These are also known as “CWA section 303(d)-listed waterbodies.”
- Immediate receiving waters for which states, territories, and authorized tribes have issued fish consumption advisories, which indicates that pollutant concentrations in the tissues of fish inhabiting those waters are considered unsafe for human consumption at any or some consumption levels.
- Immediate receiving waters within five miles of drinking water resources, including intakes and reservoirs, public wells, and sole-source aquifers.

The EPA also assessed the potential for discharges of the evaluated wastestreams to cause or contribute to fish advisories, thereby posing a human health risk. The EPA compared the T4 fish tissue concentrations from the Wildlife Module to fish consumption advisory screening values. Screening values are concentrations of target analytes in fish or shellfish tissue that are of potential public health concern; they are used as threshold values to which levels of contamination in similar tissue collected from the ambient environment can be compared. Exceedance of screening values indicates that more intensive site-specific monitoring and/or evaluation of human health risks should be conducted (U.S. EPA, 2000, Table 5-3).¹³

The EPA’s memorandum *Proximity Analyses and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024j) describes the methodology used to evaluate the proximity of steam electric power plant discharges to sensitive environments. Section 4.2 of this report presents the results of the proximity analysis.

The EPA also performed further spatial analyses to identify public drinking water supply intakes downstream from discharges of the evaluated wastestreams. See the BCA Report (U.S. EPA, 2024b) for details on the methodology and results of that analysis.

3.6 Downstream Analysis

As part of the economic benefits analysis, the EPA used a separate pollutant fate and transport model (Downstream Fate and Transport Equations, or D-FATE) to calculate the concentrations of pollutants in surface waters downstream from the immediate receiving water for each plant that discharges the

¹³ See the memorandum *IRW Model: Water Quality, Wildlife, and Human Health Analyses and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024i) for documentation of the fish advisory screening level analysis.

evaluated wastestreams. See the BCA Report (U.S. EPA, 2024b) for a detailed discussion of the D-FATE model and the analysis, which uses annual average pollutant loadings and surface water flow rates.

The EPA used these downstream concentrations from D-FATE as inputs for an analysis that identified which downstream reaches would have at least one exceedance of a water quality, wildlife, or human health benchmark value under baseline or regulatory option loadings. The EPA used this approach to estimate the extent (in river miles) of impacts in downstream surface waters under baseline and the changes in these impacts under the regulatory options evaluated. Results are presented in Section 4.3 of this report. See the memorandum *Downstream Modeling Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024i) for details on the methodology for this analysis.

3.7 Scope of the Evaluated Plants and Immediate Receiving Waters

The EPA estimates that 277 coal-fired electric generating units (EGUs) operated at 148 plants will be operating after December 31, 2028. Section 3 of the TDD (U.S. EPA, 2024a) describes how the EPA updated the industry profile to reflect changes since the 2020 rule. Section 5 and Section 6 of the TDD describe the population of plants and EGUs that the EPA estimated compliance costs and pollutant loadings under baseline (for 246 coal-fired EGUs operated at 110 plants)¹⁴ and the regulatory options.

The scope of the EA includes the 110 plants and their discharges of one or more of the evaluated wastestreams (FGD wastewater, BA transport water, CRL, or legacy wastewater) directly or indirectly to surface waters under baseline and/or one or more regulatory options.¹⁵ The EPA performed quantitative assessments to support the EA using its IRW Model, described in Section 3.4. The IRW Model, which excludes discharges to the Great Lakes and estuaries, encompasses 100 plants that discharge to 114 immediate receiving waters.¹⁶ The IRW Model excludes Great Lake and estuarine immediate receiving waters because the specific hydrodynamics and scale of the analysis required to appropriately model and quantify pollutant concentrations in these types of waterbodies are more complex than can be represented in the IRW Model. The excluded waterbodies include Lake Erie, Lake Michigan (three stream reaches), Lake Superior, Escambia River, Hillsborough Bay, Big Lake, and Sutherland Reservoir. These nine immediate receiving waters (stream reaches) receive evaluated wastestream discharges from ten plants; see *Receiving Waters Characteristics Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024f) for further details.

Table 7 presents the number of plants, generating units, and immediate receiving waters evaluated in the EA. Figure 2 shows the locations of the immediate receiving waters evaluated in the EA proximity analysis and indicates those that are included in the IRW Model. See the memorandum *Receiving Waters Characteristics Analysis and Supporting Documentation for the Environmental Assessment of the Final*

¹⁴ The EPA made plant adjustments after running the final rule analyses, and two plants (and their respective receiving waters) were not included in the pollutant loadings presented in this report or in the IRW Model. The EPA did include the receiving waters in the proximity analysis. Both plants are expected to retire or undergo fuel conversion by 2034. See *Updates to Estimated Compliance Costs and Pollutant Loadings* (U.S. EPA, 2024n) and *Pollutant Loadings Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024h).

¹⁵ Of the 110 plants in the EA, 106 discharge directly to surface water, three discharge indirectly to POTWs, and one discharges wastestreams both directly and indirectly. The EPA excluded CRL discharges at one plant from the EA that indirectly discharges to a POTW that does not discharge to any receiving waters (U.S. EPA, 2024f). Discharges from two additional plants, not included in the count of 110 plants, were not included in the pollutant loadings analysis or IRW Model (only the proximity analysis); see U.S. EPA (2024h and 2024n). One plant is a direct discharging plant and the other is an indirect discharging plant.

¹⁶ Ten of the 110 plants included in the EA discharge to more than one immediate receiving water.

Supplemental Steam Electric Rule (U.S. EPA, 2024f) for the list of immediate receiving waters and details on the EPA’s methodology for identifying them.

The number of evaluated plants and generating units, and the number of the associated immediate receiving waters, vary across baseline and the regulatory options evaluated for the final rule. This is due to differences in the stringency of controls, applicability of these controls based on subcategorization, and estimates of the control technologies that plants would implement to meet requirements (see the preamble for details). Table 8 presents the number of plants, generating units, and immediate receiving waters with nonzero pollutant loadings for baseline and each regulatory option evaluated.

Table 7. Plants, Generating Units, and Immediate Receiving Waters Evaluated in the Environmental Assessment for the Final Supplemental Rule

Category	Number Evaluated in Pollutant Loadings Analysis	Number Evaluated in the Proximity Analysis	Number Evaluated in IRW Model ^a
Plants ^b	110	112	100
Electric generating units ^{b,c}	246	249	222
<i>Immediate Receiving Waters</i>			
River/stream ^b	98	100	98
Lake/pond/reservoir	16	16	16
Great Lakes	5 ^d	5 ^d	—
Estuary/bay/other	4	4	—
Total Immediate Receiving Waters	123^{d,e}	125^{d,e}	114^{d,e}

Sources: U.S. EPA, 2024f and 2024h.

Abbreviations: IRW (immediate receiving water).

a—The IRW Model excludes discharges to nine immediate receiving waters that are one of the Great Lakes and or an estuary because the specific hydrodynamics and scale of the analysis required to appropriately model and quantify pollutant concentrations in these types of waterbodies are more complex than can be represented in the IRW Model.

b—The EPA made plant adjustments after running the final rule analyses, and two plants (and their respective receiving waters) were not included in the pollutant loadings presented in the report or in the IRW Model. The EPA did include the receiving waters in the proximity analysis. Both plants are expected to retire or undergo fuel conversion by 2034. See U.S. EPA (2024h and 2024n).

c—Legacy wastewater discharges at two plants are not associated with an active coal-fired generating unit.

d—Ten plants included discharge to more than one immediate receiving water. One Great Lake immediate receiving water receives discharges from two plants.

e—One plant discharges CRL to a zero-discharge publicly owned treatment works; therefore, no immediate receiving water is associated with the plant’s pollutant loadings from that wastestream. The plant’s legacy wastewater loadings are included in the EA analyses.

Table 8. Plants, Generating Units, and Immediate Receiving Waters with Pollutant Loadings Under Baseline and Regulatory Options for the Final Supplemental Rule

Category	Baseline	Option A	Option B	Option C
<i>Downstream and Proximity Analyses^a</i>				
Plants	112	97	54	17
Electric generating units ^b	249	219	123	33
Immediate receiving waters	125	105	57	18
<i>Subset Also Evaluated in Pollutant Loadings^a</i>				
Plants	110	97	54	17
Electric generating units ^b	246	219	123	33
Immediate receiving waters	123	105	57	18
<i>Subset Also Evaluated in IRW Model^{a,c}</i>				
Plants	100	89	47	16
Electric generating units ^b	222	198	103	29
Immediate receiving waters	114	97	50	17

Sources: U.S. EPA, 2024f and 2024h.

Abbreviations: IRW (immediate receiving water).

a—The EPA made plant adjustments after running the final rule analyses, and two plants (and their respective receiving waters) are not included in the pollutant loadings presented in the report or in the IRW Model. The EPA did include the receiving waters in the proximity analysis. Both plants are expected to retire or undergo fuel conversion by 2034. See U.S. EPA (2024h and 2024n).

b—Legacy wastewater discharges at two plants are not associated with an active coal-fired generating unit.

c—The IRW Model excludes discharges to nine immediate receiving waters that are one of the Great Lakes and or an estuary because the specific hydrodynamics and scale of the analysis required to appropriately model and quantify pollutant concentrations in these types of waterbodies are more complex than can be represented in the IRW Model.

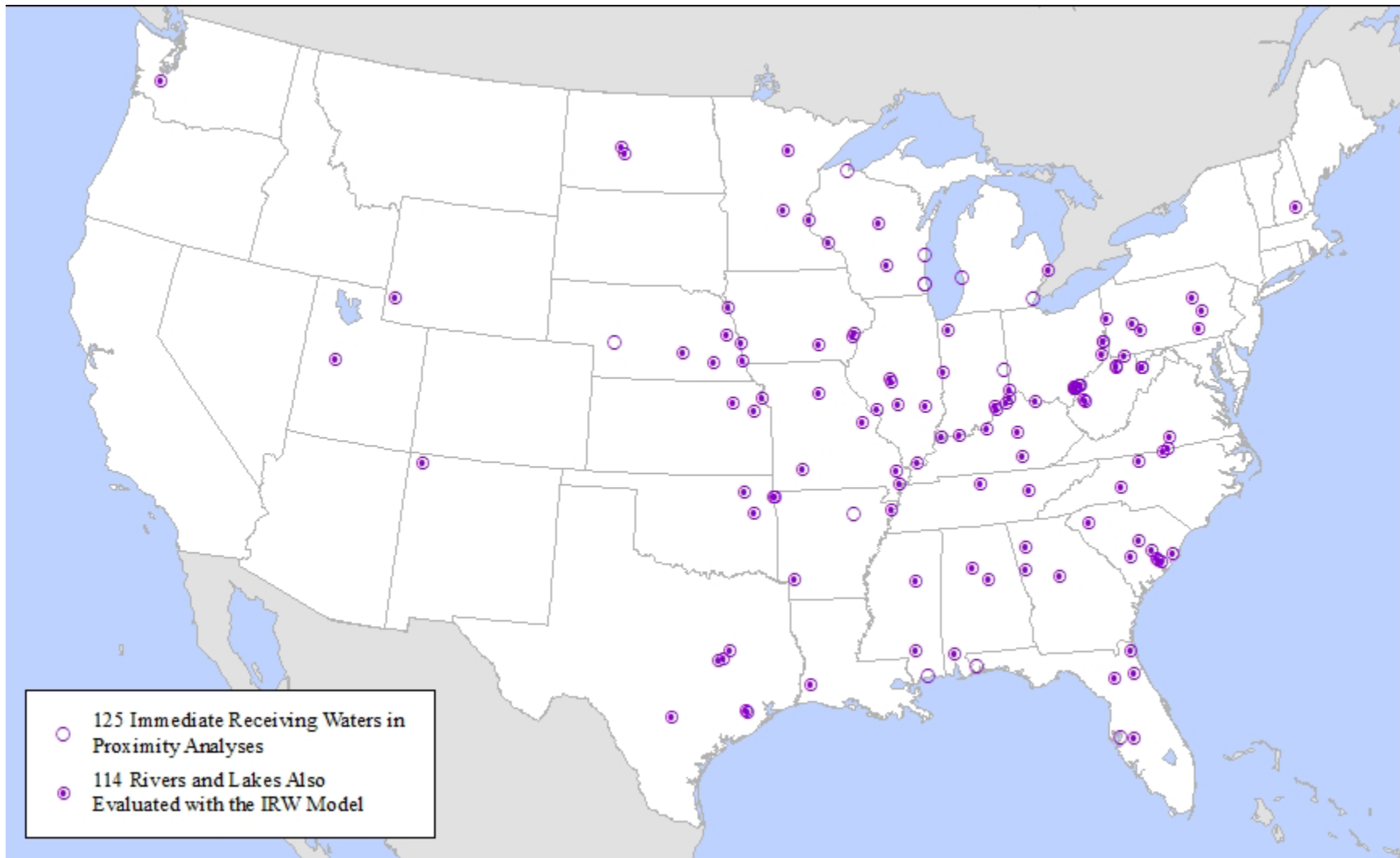


Figure 2. Locations of Immediate Receiving Waters Evaluated in the Environmental Assessment for the Final Supplemental Rule

4. Results of the Quantitative Environmental Assessment for the Final Supplemental Rule

The EPA used the plant-specific and receiving-water-specific pollutant loadings, described in Section 3.1, to determine the environmental impacts of the evaluated wastestreams—*i.e.*, flue gas desulfurization (FGD) wastewater, bottom ash (BA) transport water, combustion residual leachate (CRL), and legacy wastewater—from steam electric power plants. This section presents the results of the quantitative analyses described in Sections 3.3 through 3.6, which include the following:

- Use of the EPA’s Immediate Receiving Water (IRW) Model to:
 - Estimate the annual average pollutant concentrations in immediate receiving waters due to discharges of the evaluated wastestreams under baseline and the regulatory options, estimate the bioaccumulation of pollutants in fish tissue within those waters, and estimate the daily and lifetime pollutant exposure doses among humans who consume those fish.
 - Compare the estimated concentrations and estimated exposure doses to various benchmark values as indicators of potential water quality, wildlife, and human health impacts.
 - Evaluate the estimated changes in those impacts under the regulatory options, as compared to baseline.
- A proximity analysis to identify immediate receiving waters that are designated as Clean Water Act (CWA) section 303(d)–listed impaired waterbodies; have been issued fish consumption advisories; or are within five miles of drinking water resources, including intakes and reservoirs, public wells, and sole-source aquifers.
- Use of pollutant fate and transport model (D-FATE) outputs to estimate potential water quality, wildlife, and human health impacts in downstream surface waters under baseline and evaluate the estimated changes in those impacts under the regulatory options.

The BCA Report (U.S. EPA, 2024b) discusses the EPA’s evaluation of other impacts that were not quantified in the environmental assessment.

4.1 Environmental Impacts Identified by the IRW Model

The IRW Model includes modules assessing potential changes in impacts on water quality, wildlife, and human health in waters receiving discharges of the evaluated wastestreams from steam electric power plants.¹⁷ See Section 3.4 of this document and Appendices C, D, and E of the 2020 environmental assessment (EA) (U.S. EPA, 2020a) for details on the IRW Model’s structure and methodology, including equations, input data, and environmental parameters.

The following sections present the environmental impact results estimated from each module for the nine modeled pollutants: arsenic, cadmium, copper, lead, mercury, nickel, selenium, thallium, and zinc. The results identify modeled exceedances of water quality, wildlife, and human health benchmark values under baseline and the reduction in those exceedances under each regulatory option. Appendix A includes additional IRW Model outputs.

¹⁷ The EA encompasses a total of 125 immediate receiving waters and 112 plants (some of which discharge to multiple receiving waters). The EPA made plant adjustments after running final rule analyses, and two plants and their respective receiving waters were only included in the proximity analysis. Both plants are expected to retire or undergo fuel conversion by 2034. See U.S. EPA (2024h and 2024n). The IRW Model, which excludes the Great Lakes and estuaries, analyzes a total of 114 immediate receiving waters and loadings from 100 plants.

4.1.1 Water Quality Impacts

The IRW Water Quality Module assesses the quality of surface waters that receive discharges of the evaluated wastestreams by comparing estimated pollutant concentrations in the water column to the National Recommended Water Quality Criteria (NRWQC) and drinking water maximum contaminant levels (MCLs)¹⁸ under baseline and each regulatory option. The Water Quality Module results described in this section are based on estimated annual average pollutant loadings and flow rates. The module considers modeled exceedances of the freshwater acute NRWQC, freshwater chronic NRWQC, human health water and organism (HH WO) NRWQC, human health organism only (HH O) NRWQC, and drinking water MCL.

The EPA compared the modeled receiving water concentrations to the water quality benchmarks presented in Table 4. Table 9 summarizes the number of immediate receiving waters exceeding the water quality benchmarks. Table 10 presents the number of immediate receiving waters with exceedances of any NRWQC or MCL by pollutant. The EPA identified water quality benchmark exceedances for all nine pollutants evaluated for one or more immediate receiving waters. Pollutants with exceedances in multiple receiving waters included arsenic, cadmium, copper, lead, selenium, and thallium. Under baseline, the EPA estimated that 38 of the 114 immediate receiving waters (33 percent) exceeded one or more water quality benchmark. Under the final rule (Option B), the number of immediate receiving waters exceeding a benchmark will decrease by 24 immediate receiving waters.

Table 9. Modeled IRWs with Exceedances of NRWQC and MCLs Under Baseline and Regulatory Options

Water Quality Evaluation Benchmark	Pollutant	Number of Modeled IRWs Exceeding Benchmark Value (Difference Relative to Baseline) ^a			
		Baseline	Option A	Option B	Option C
Freshwater acute NRWQC	Any pollutant	3	2 (-1)	2 (-1)	2 (-1)
	Cadmium	3	2 (-1)	1 (-2)	1 (-2)
	Copper	1	1 (0)	0 (-1)	0 (-1)
	Nickel	1	1 (0)	0 (-1)	0 (-1)
	Selenium	1	1 (0)	1 (0)	1 (0)
	Zinc	1	1 (0)	0 (-1)	0 (-1)
Freshwater chronic NRWQC	Any pollutant	12	11 (-1)	5 (-7)	5 (-7)
	Cadmium	8	5 (-3)	2 (-6)	2 (-6)
	Copper	2	2 (0)	0 (-2)	0 (-2)
	Lead	1	1 (0)	0 (-1)	0 (-1)
	Mercury	1	1 (0)	0 (-1)	0 (-1)
	Nickel	1	1 (0)	0 (-1)	0 (-1)
	Selenium	12	11 (-1)	5 (-7)	5 (-7)
	Zinc	1	1 (0)	0 (-1)	0 (-1)
HH WO NRWQC	Any pollutant	38	28 (-10)	14 (-24)	7 (-31)
	Arsenic	38	28 (-10)	14 (-24)	7 (-31)
	Nickel	1	1 (0)	0 (-1)	0 (-1)
	Selenium	1	1 (0)	1 (0)	1 (0)
	Thallium	8	7 (-1)	4 (-4)	3 (-5)

¹⁸ Table 4 in Section 3 presents the benchmarks values for the pollutants evaluated.

Table 9. Modeled IRWs with Exceedances of NRWQC and MCLs Under Baseline and Regulatory Options

Water Quality Evaluation Benchmark	Pollutant	Number of Modeled IRWs Exceeding Benchmark Value (Difference Relative to Baseline) ^a			
		Baseline	Option A	Option B	Option C
HH O NRWQC	Any pollutant	21	14 (-7)	4 (-17)	3 (-18)
	Arsenic	21	14 (-7)	4 (-17)	3 (-18)
	Selenium	1	1 (0)	1 (0)	1 (0)
	Thallium	7	5 (-2)	3 (-4)	3 (-4)
Drinking water MCL	Any pollutant	5	4 (-1)	3 (-2)	3 (-2)
	Arsenic	4	2 (-2)	2 (-2)	2 (-2)
	Cadmium	3	2 (-1)	1 (-2)	1 (-2)
	Lead	2	2 (0)	1 (-1)	1 (-1)
	Mercury	1	1 (0)	0 (-1)	0 (-1)
	Selenium	3	3 (0)	2 (-1)	2 (-1)
	Thallium	2	2 (0)	2 (0)	2 (0)
Zinc	1	1 (0)	0 (-1)	0 (-1)	
Total Number of Unique Immediate Receiving Waters ^b		38	28 (-10)	14 (-24)	7 (-31)

Source: U.S. EPA, 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); MCL (maximum contaminant level); NRWQC (National Recommended Water Quality Criteria).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Total may not equal the sum of the individual values because some immediate receiving waters have multiple types of exceedances.

Table 10. Modeled IRWs with Exceedances of NRWQC and MCLs, by Pollutant, Under Baseline and Regulatory Options

Pollutant	Number of Modeled IRWs Exceeding Benchmark Value (Difference Relative to Baseline) ^a			
	Baseline	Option A	Option B	Option C
Arsenic	38	28 (-10)	14 (-24)	7 (-31)
Cadmium	8	5 (-3)	2 (-6)	2 (-6)
Copper	2	2 (0)	0 (-2)	0 (-2)
Lead	2	2 (0)	1 (-1)	1 (-1)
Mercury	1	1 (0)	0 (-1)	0 (-1)
Nickel	1	1 (0)	0 (-1)	0 (-1)
Selenium	12	11 (-1)	5 (-7)	5 (-7)
Thallium	8	7 (1)	4 (-4)	3 (-5)
Zinc	1	1 (0)	0 (-1)	0 (-1)
Any Pollutant^b	38	28 (-10)	14 (-24)	7 (-31)

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); MCL (maximum contaminant level); NRWQC (National Recommended Water Quality Criteria).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Total may not equal the sum of the individual values because some immediate receiving waters have multiple types of exceedances.

In the 2020 EA, the EPA conducted a water quality analysis using estimated monthly pollutant loadings and flow rates to assess the significance of monthly variability in the modeled water quality impacts. The results were similar to those using the annual average analysis, and the EPA determined the following key takeaways:

- Most worst-case months occur during the summer, whereas most best-case months occur during the winter and early spring.
- There is potential for impacts on aquatic life during certain periods characterized by low flows, high loadings, or a combination of the two.
- Certain geographic areas could experience adverse seasonal cumulative effects due to concurrent, or nearly concurrent, discharges of evaluated wastestreams from multiple plants.

These results suggest that seasonal water quality impacts from discharges of the evaluated wastestreams may be more prevalent than indicated by the annual average analysis. Seasonal cumulative effects in affected watersheds could be particularly pronounced during summer and early autumn. The EPA expects that swimming, fishing, and boating in local waterways are more common during these seasons, potentially increasing opportunities for exposure to degraded water quality conditions in the immediate receiving waters. In addition, fish species that spawn in the affected waterways during these periods (including federally threatened or endangered species) could have an increased potential for adverse impacts from pollutant exposure, since the timing of their sensitive life stages would align with worst-case water quality conditions. See the 2020 EA (U.S. EPA, 2020a) for more details.

Appendix C of the 2020 EA (U.S. EPA, 2020a) provides details on the following limitations and uncertainties of the IRW Water Quality Module:

- Estimated pollutant loadings are based on data from a subset of steam electric power plants.
- It uses annual-average pollutant loadings and flow rates.
- It does not consider temporal variability and pollutant speciation.
- It does not account for ambient background pollutant concentrations or contributions from other point and nonpoint sources.
- It assumes that equilibrium is quickly attained within the waterbody following discharge and is consistently maintained between the water column and surficial bottom sediments.
- It assumes that pollutants dissolved or sorbed within the water column and bottom sediments can be described by a partition coefficient and other calculation assumptions.
- It assumes that pollutants sorbed to bottom sediments are considered a net loss from the water column and assumes a pollutant burial rate of zero within the bottom sediment.

4.1.2 Wildlife Impacts

As described in Section 3.4, the IRW Wildlife Module assesses impacts to sediment biota, minks, and eagles. This analysis expands on the evaluation of potential wildlife impacts based on the Freshwater Chronic and Acute NRWQC in the Water Quality Module. Table 11 presents the number of immediate receiving waters with modeled exceedances of the threshold effect concentrations (TECs) and no effect hazard concentrations (NEHCs)¹⁹ under baseline and reduction in those exceedances under the regulatory options. Results are presented for all pollutants in aggregate and individually for pollutants with exceedances. The EPA did not have benchmark data to compare thallium concentrations in the immediate receiving water; therefore, that pollutant is excluded from the wildlife impacts analysis.

Under baseline, the EPA estimated that all eight evaluated pollutants had one or more immediate receiving water that exceeded sediment TECs. Pollutants with exceedances in multiple receiving waters included arsenic, cadmium, copper, mercury, nickel, selenium, and zinc. Lead had an exceedance under baseline and all the regulatory options for one receiving water. Under the final rule (Option B), the number of immediate receiving waters with exceedances of TECs decreases by at least 70 percent for five of the eight pollutants (arsenic, cadmium, mercury, nickel, and zinc). Copper and selenium had smaller improvements under the final rule, with respective reductions of 50 and 54 percent of immediate receiving waters exceeding the TEC.

Four pollutants (cadmium, mercury, selenium, and zinc) exceeded the NEHCs for minks and eagles under baseline and the regulatory options. Under the final rule (Option B), the EPA calculated that the number of immediate receiving waters exceeding the NEHC for minks decreased by 14 immediate receiving waters for mercury and nine immediate receiving waters for selenium. The number of immediate receiving waters exceeding the NEHC for eagle decreased by 19 immediate receiving waters for mercury and nine immediate receiving waters for selenium under the final rule. Under baseline, cadmium and zinc exceeded NEHC for minks and eagles at one receiving water; the final rule will eliminate these exceedances.

¹⁹ Table 5 in Section 3 presents the benchmarks values for the pollutants evaluated.

Table 11. Modeled IRWs with Exceedances of TECs and NEHCs Under Baseline and Regulatory Options

Wildlife Evaluation Benchmark	Pollutant ^a	Number of Modeled IRWs Exceeding Benchmark Value (Difference Relative to Baseline) ^b			
		Baseline	Option A	Option B	Option C
Sediment TEC	Any pollutant	24	24 (0)	11 (-13)	7 (-17)
	Arsenic	3	2 (-1)	0 (-3)	0 (-3)
	Cadmium	8	5 (-3)	2 (-6)	2 (-6)
	Copper	2	2 (0)	1 (-1)	1 (-1)
	Lead	1	1 (0)	1 (0)	1 (0)
	Mercury	19	9 (-10)	2 (-17)	2 (-17)
	Nickel	14	6 (-8)	2 (-12)	2 (-12)
	Selenium	24	24 (0)	11 (-13)	7 (-17)
	Zinc	7	4 (-3)	2 (-5)	2 (-5)
Fish ingestion NEHC for minks	Any pollutant	16	16 (0)	6 (-10)	5 (-11)
	Cadmium	1	1 (0)	0 (-1)	0 (-1)
	Mercury	16	7 (-9)	2 (-14)	2 (-14)
	Selenium	15	15 (0)	6 (-9)	5 (-10)
	Zinc	1	1 (0)	0 (-1)	0 (-1)
Fish ingestion NEHC for eagles	Any pollutant	22	17 (-5)	6 (-16)	5 (-17)
	Cadmium	1	1 (0)	0 (-1)	0 (-1)
	Mercury	22	15 (-7)	3 (-19)	2 (-20)
	Selenium	15	15 (0)	6 (-9)	5 (-10)
	Zinc	1	1 (0)	0 (-1)	0 (-1)
Any Wildlife Pollutant Benchmark for Any Pollutant ^c		24	24 (0)	11 (-13)	7 (-17)

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); TEC (threshold effect concentration); NEHC (no effect hazard concentration).

a—Thallium excluded from the analysis (no benchmarks for comparison). No immediate receiving waters exceeded the TEC for copper and lead. No immediate receiving waters exceeded NEHC benchmarks for arsenic, cadmium, copper, lead, nickel, or zinc.

b—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

c—Total may not equal the sum of the individual values because some immediate receiving waters have multiple types of exceedances.

Appendix D of the 2020 EA (U.S. EPA, 2020a) provides details on the following limitations and uncertainties of the IRW Wildlife Module:

- Impact estimates are based on an individual exposure pathway and individual pollutant exposure rather than cumulative risks across exposure pathways and the interaction of multiple pollutants.
- Bioaccumulation factors are not available for all pollutants (use of bioconcentration factors does not account for the accumulation of pollutants via the food web).
- It does not consider indirect ecological effects such as depletion of food sources.

- It assumes the selected receptor species and receiving water occur together (*i.e.*, all immediate receiving waters are habitats for the receptor species).
- It assumes the diet of the receptor species consists of fish inhabiting the immediate receiving water.
- It assumes all forms of a pollutant are equally bioavailable to ecological receptors.
- Modeling assumes that the receiving water is fully mixed; however, water in lakes might stratify and affect chemical speciation by stratum.

4.1.3 Human Health Impacts

The IRW Human Health Module evaluates noncancer and cancer human health impacts among various human cohorts (recreational and subsistence fishers; children and adults; and different race/ethnicity categories) from consuming fish caught from immediate receiving waters that are contaminated by discharges of the evaluated wastestreams. The module uses oral reference doses (RfDs) to evaluate changes in noncancer health risks and a lifetime excess cancer risk (LECR) benchmark value of one-in-a-million, or 10^{-6} , to evaluate changes in cancer risk. This analysis expands on the evaluation of potential human health impacts based on the NRWQC and MCLs in the Water Quality Module.

Under baseline, the EPA estimated the average daily dose of one or more individual pollutant from fish consumption among subsistence fishers exceed the oral RfDs (noncancer) in 31 to 39 (27 to 34 percent) of immediate receiving waters, depending on the age group evaluated. Average daily doses among recreational fishers exceeded oral RfDs in 26 to 28 (23 to 25 percent) of immediate receiving waters. The lower prevalence of exceedances among recreational fishers is primarily due to their lower average fish tissue consumption rates. These results suggest that fish in immediate receiving waters can have health effects on surrounding fisher populations.

As shown in Table 12, the exceedances are primarily driven by mercury (as methylmercury), selenium, and thallium. The EPA calculated no exceedances for arsenic (inorganic) or nickel (total) under baseline and the regulatory options. The EPA estimated that the number of immediate receiving waters contributing to oral RfD (noncancer) exceedances decreased for all standard cohorts (*i.e.*, cohorts that are not split into different race/ethnicity categories) under all regulatory options. Under the final rule (Option B), the EPA estimated the following decreases in number of immediate receiving waters with fish that, if consumed, would exceed oral RfDs:

- Methylmercury—decrease by at least 20 immediate receiving waters for all standard cohorts.
- Selenium—decrease by at least seven immediate receiving waters for all standard cohorts.
- Thallium—decrease by at least eight immediate receiving waters for all standard cohorts.

Although the EPA did not directly assess the potential health effects posed by lead in this EA, the final rule decreases the annual loadings of lead to the environment by 187 pounds per year compared to baseline.²⁰ The monetized human health effects associated with changes in lead discharges are discussed in the BCA Report (U.S. EPA, 2024b).

As part of this rulemaking, the EPA evaluated the joint toxic action of multiple pollutants discharged into the evaluated wastestreams from steam electric power plants to determine potential cumulative human health impacts at the immediate receiving waters. See the memorandum *Assessment of Human Health Impacts from Multiple Pollutants in Steam Electric Power Plant Discharges* (U.S. EPA, 2024k) for a summary of the results.

²⁰ For comparison, the 2015 rule reduced lead discharges by 19,200 pounds per year (U.S. EPA, 2015a).

Table 12. Modeled IRWs with Exceedances of Oral RfD (Noncancer Human Health Effects) Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Pollutant	Number of Modeled IRWs Exceeding Oral RfD (Difference Relative to Baseline) ^a			
		Baseline	Option A	Option B	Option C
Child—recreational	Any pollutant	28	22 (-6)	9 (-19)	6 (-22)
	Cadmium	3	2 (-1)	1 (-2)	1 (-2)
	Methylmercury	28	22 (-6)	8 (-20)	5 (-23)
	Selenium	15	15 (0)	6 (-9)	5 (-10)
	Thallium	16	15 (-1)	6 (-10)	5 (-11)
	Zinc	1	1 (0)	0 (-1)	0 (-1)
Child—subsistence	Any pollutant	39	28 (-11)	15 (-24)	8 (-31)
	Cadmium	4	4 (0)	2 (-2)	2 (-2)
	Copper	1	1 (0)	0 (-1)	0 (-1)
	Methylmercury	38	28 (-10)	15 (-23)	8 (-30)
	Selenium	22	22 (0)	8 (-14)	5 (-17)
	Thallium	24	19 (-5)	10 (-14)	7 (-17)
	Zinc	1	1 (0)	0 (-1)	0 (-1)
Adult—recreational	Any pollutant	26	18 (-8)	6 (-20)	5 (-21)
	Cadmium	1	1 (0)	1 (0)	1 (0)
	Methylmercury	25	17 (-8)	5 (-20)	4 (-21)
	Selenium	12	12 (0)	5 (-7)	4 (-8)
	Thallium	13	9 (-4)	5 (-8)	4 (-9)
	Zinc	1	1 (0)	0 (-1)	0 (-1)
Adult—subsistence	Any pollutant	31	23 (-8)	9 (-22)	6 (-25)
	Cadmium	4	3 (-1)	2 (-2)	2 (-2)
	Methylmercury	31	23 (-8)	9 (-22)	6 (-25)
	Selenium	15	15 (0)	6 (-9)	5 (-10)
	Thallium	16	15 (-1)	6 (-10)	5 (-11)
	Zinc	1	1 (0)	0 (-1)	0 (-1)
Any Pollutant and Age/Fishing Mode Cohort^b		39	28 (-11)	15 (-24)	8 (-31)

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Total may not equal the sum of the individual values because some immediate receiving waters have multiple types of exceedances.

Under baseline, the EPA estimated that nine immediate receiving waters (eight percent) could contain fish contaminated with inorganic arsenic that present cancer risks greater than the LECR benchmark value of one-in-a-million for the most sensitive, standard cohort (adult subsistence fishers). Under the final rule (Option B), the number of immediate receiving waters whose fish exceed this cancer risk threshold will

decrease by seven (78 percent) for this cohort. Table 13 presents the number of immediate receiving waters where the LECR for inorganic arsenic exceeds one-in-a-million.

Table 13. Modeled IRWs with LECR Greater Than One-in-a-Million (Cancer Human Health Effects) Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Number of Modeled IRWs with LECR Greater than One-in-a-Million (Difference Relative to Baseline) ^a			
	Baseline	Option A	Option B	Option C
Child—recreational	0	0 (0)	0 (0)	0 (0)
Child—subsistence	3	2 (-1)	1 (-2)	1 (-2)
Adult—recreational	4	2 (-2)	2 (-2)	2 (-2)
Adult—subsistence	9	3 (-6)	2 (-7)	2 (-7)
Total Number of Immediate Receiving Waters^b	9	3 (-6)	2 (-7)	2 (-7)

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); LECR (lifetime excess cancer risk).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Total may not equal the sum of the individual values because some immediate receiving waters have multiple types of exceedances.

The EPA also performed an analysis using fish consumption rates for recreational and subsistence fishers in different race/ethnicity categories to assess whether the steam electric power plant wastewater discharges disproportionately affect minority groups. Table 14 presents the number of immediate receiving waters in which the modeled average daily dose of any pollutant exceeds the oral RfD. Table 15 presents the number of immediate receiving waters that could contain fish contaminated with inorganic arsenic that present cancer risks greater than the LECR benchmark value of one-in-a-million. Results in the tables are presented by cohort (recreational and subsistence fisher) and race/ethnicity category.

As shown in Table 14, the number of immediate receiving waters where the average daily dose of at least one individual pollutant from fish consumption exceeds the oral RfDs is highest among subsistence fishers (child or adults) that fall in the “Other, Including Multiple Races” category. The increased prevalence of exceedances is primarily due to higher average fish tissue consumption rates for this category and fishing mode. Under the final rule, the EPA estimated reductions in the number of immediate receiving waters with exceedances of human health risk under the final rule to be between 19 and 23 immediate receiving waters, depending on the fisher type and cohort.

Inorganic arsenic concentrations in fish resulted in an estimated cancer risk greater than one-in-a-million to adult subsistence, minority fishers (*i.e.*, excluding the non-Hispanic white cohort) in nine to 11 immediate receiving waters under baseline. Four immediate receiving waters had inorganic arsenic concentrations in fish above the LECR threshold of one-in-a-million for adult recreational, minority fishers under baseline. Cancer risks for the child cohorts are lower. The estimated cancer risk among adult minority fishers is higher than the risk among adult nonminority fishers. The EPA estimated reductions in the number of immediate receiving waters with exceedances of cancer risk under the final rule to be up to eight immediate receiving waters, depending on the fisher type and cohort.

Appendix A presents the IRW Human Health Module results by pollutant for each age group and mode of fishing for both standard and race/ethnicity cohorts.

Table 14. Modeled IRWs with Exceedances of Oral RfDs by Race/Ethnicity Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD (Difference Relative to Baseline) ^a			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	26	18 (-8)	6 (-20)	5 (-21)
	Non-Hispanic Black	26	19 (-7)	7 (-19)	5 (-21)
	Mexican-American	28	20 (-8)	8 (-20)	5 (-23)
	Other Hispanic	26	19 (-7)	7 (-19)	5 (-21)
	Other, Including multiple races	28	20 (-8)	8 (-20)	5 (-23)
Child—subsistence	Non-Hispanic White	29	23 (-6)	9 (-20)	6 (-23)
	Non-Hispanic Black	31	23 (-8)	9 (-22)	6 (-25)
	Mexican-American	32	25 (-7)	12 (-20)	7 (-25)
	Other Hispanic	32	23 (-9)	9 (-23)	6 (-26)
	Other, including multiple races	34	26 (-8)	14 (-20)	8 (-26)
Adult—recreational	Non-Hispanic White	26	18 (-8)	6 (-20)	5 (-21)
	Non-Hispanic Black	26	19 (-7)	7 (-19)	5 (-21)
	Mexican-American	28	20 (-8)	8 (-20)	5 (-23)
	Other Hispanic	26	19 (-7)	7 (-19)	5 (-21)
	Other, including multiple races	28	20 (-8)	8 (-20)	5 (-23)
Adult—subsistence	Non-Hispanic White	29	23 (-6)	9 (-20)	6 (-23)
	Non-Hispanic Black	31	23 (-8)	9 (-22)	6 (-25)
	Mexican-American	32	25 (-7)	12 (-20)	7 (-25)
	Other Hispanic	32	23 (-9)	9 (-23)	6 (-26)
	Other, including multiple races	34	26 (-8)	14 (-20)	8 (-26)

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

Table 15. Modeled IRWs with LECR Greater Than One-in-a-Million (Cancer Human Health Effects) Race/Ethnicity Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs with LECR Above One-in-a-Million (Difference Relative to Baseline) ^a			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	0	0 (0)	0 (0)	0 (0)
	Non-Hispanic Black	0	0 (0)	0 (0)	0 (0)
	Mexican-American	0	0 (0)	0 (0)	0 (0)
	Other Hispanic	0	0 (0)	0 (0)	0 (0)
	Other, including multiple races	0	0 (0)	0 (0)	0 (0)
Child—subsistence	Non-Hispanic White	2	2 (0)	1 (-1)	0 (-2)
	Non-Hispanic Black	3	2 (-1)	1 (-2)	1 (-2)
	Mexican-American	3	2 (-1)	1 (-2)	1 (-2)
	Other Hispanic	3	2 (-1)	1 (-2)	1 (-2)
	Other, including multiple races	3	2 (-1)	1 (-2)	1 (-2)
Adult—recreational	Non-Hispanic White	4	2 (-2)	2 (-2)	2 (-2)
	Non-Hispanic Black	4	2 (-2)	2 (-2)	2 (-2)
	Mexican-American	4	2 (-2)	2 (-2)	2 (-2)
	Other Hispanic	4	2 (-2)	2 (-2)	2 (-2)
	Other, including multiple races	4	2 (-2)	2 (-2)	2 (-2)
Adult—subsistence	Non-Hispanic White	9	3 (-6)	2 (-7)	2 (-7)
	Non-Hispanic Black	9	3 (-6)	2 (-7)	2 (-7)
	Mexican-American	10	3 (-7)	2 (-8)	2 (-8)
	Other Hispanic	10	3 (-7)	2 (-8)	2 (-8)
	Other, including multiple races	11	4 (-7)	3 (-8)	2 (-9)

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); LECR (lifetime excess cancer risk).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

The EPA also compared trophic level 4 (T4) fish tissue pollutant concentrations to fish consumption advisory screening values to assess the potential for discharges of the evaluated wastestreams to cause or contribute to fish advisories and pose a human health risk.²¹ Based on the modeling results, up to 32 immediate receiving waters (28 percent) may contain fish with contamination levels that could trigger advisories for recreational and/or subsistence fishers under baseline; this decreases to 10 immediate receiving waters (9 percent) under the final rule (Option B). Mercury and selenium are the pollutants

²¹ For this analysis, the EPA used the fish consumption advisory screening values from the EPA's *Guidance for Assessing Chemical Contaminant Data for Uses in Fish Advisories, Volume 1* (U.S. EPA, 2000).

most likely to exceed screening values. Table 16 presents the number of immediate receiving waters where the modeled T4 fish tissue concentrations exceed screening values used for fish advisories.²²

Table 16. Comparison of Modeled T4 Fish Tissue Concentrations to Fish Advisory Screening Values Under Baseline and Regulatory Options

Pollutant	Screening Value (ppm)	Number of IRWs with Modeled T4 Fish Tissue Concentrations Exceeding Screening Value (Difference Relative to Baseline) ^a			
		Baseline	Option A	Option B	Option C
<i>Recreational Fishers</i>					
Arsenic (as inorganic arsenic) ^b	0.026	0	0	0	0
Cadmium	4	1	1	0	0
Mercury (as methylmercury)	0.4	22	16	4	3
Selenium	20	8	7	3	3
Total for Any Pollutant in Evaluated Wastestreams^c	—	22	16	4	3
<i>Subsistence Fishers</i>					
Arsenic (as inorganic arsenic) ^b	0.00327	0	0	0	0
Cadmium	0.491	4	3	2	2
Mercury (as methylmercury)	0.049	32	24	10	6
Selenium	2.457	18	18	8	5
Total for Any Pollutant in Evaluated Wastestreams^c	—	32	24	10	6

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); ppm (parts per million); T4 (trophic level 4).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Screening value presented is for carcinogenic effects (lower value than noncarcinogenic effects).

c—Total may not equal the sum of the individual values because some immediate receiving waters are impaired for multiple pollutants.

Appendix E of the 2020 EA (U.S. EPA, 2020a) details the following limitations and uncertainties of the IRW Human Health Module:

- Impact estimates are based on individual exposure pathway and individual pollutant exposure rather than cumulative risks across exposure pathways and the interaction of multiple pollutants.
- Exposure factors will vary by individual physical characteristics.
- The uncertainties associated with human health benchmark values are present, as described in the EPA’s *Guidelines for Carcinogen Risk Assessment* (U.S. EPA, 2005) and Integrated Risk Information System (IRIS) (U.S. EPA, 2019).

²² As described in Section 4.2.2, none of the immediate receiving waters are under fish consumption advisories for cadmium or selenium; each advisory screening value exceedance shown in Table 16 for these pollutants therefore indicates a “new” receiving water of concern that may warrant additional monitoring and/or evaluation of human health risk.

- The module assumes that the diet of the human health cohorts consists of fish inhabiting the immediate receiving water.
- It assumes all forms of a pollutant are equally bioavailable to human health cohorts.

4.2 Discharges to Sensitive Environments

As discussed in Section 3.5, the EPA evaluated pollutant discharges to sensitive environments (*i.e.*, impaired waters, fish consumption advisory waters, and drinking water resources). Discharges of the evaluated wastestreams to CWA section 303(d) impaired waters and fish consumption advisory waters²³ may contribute to water quality impairments, increased health risk associated with consuming fish, and a reduction in the extent of viable downstream fisheries. Discharges of pollutants in the evaluated wastestreams to drinking water resources would likely be reduced to safe levels as part of intake water treatment; however, these pollutants could affect the effectiveness of the treatment processes, which could increase public drinking water treatment costs.²⁴ Table 17 summarizes the number of immediate receiving waters that are classified as either CWA section 303(d) impaired waters, fish consumption advisory waters, or drinking water resources under baseline and each regulatory option. The EPA evaluated 125 immediate receiving waters that receive discharges of the evaluated wastestreams, either directly or indirectly via POTWs. Of these 125 immediate receiving waters, all 125 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C. Sections 4.2.1 through 4.2.3 present the results of the EPA’s assessment of immediate receiving waters that are sensitive environments.²⁵

Table 17. Modeled IRWs Identified as CWA Section 303(d) Impaired Waters, Fish Consumption Advisory Waters, or Drinking Water Resources Under Baseline and Regulatory Options

Sensitive Environment Category	Number of Modeled IRWs Receiving Discharges of the Evaluated Wastestreams (Difference Relative to Baseline) ^a			
	Baseline	Option A	Option B	Option C
IRWs receiving discharges of the evaluated wastestreams	125	105	57	18
Impaired water	64	55 (-9)	24 (-40)	8 (-56)
Subset impaired for one or more pollutants associated with the evaluated wastestreams ^b	43	37 (-6)	17 (-26)	6 (-37)
Fish consumption advisory water	72	60 (-12)	33 (-39)	12 (-60)
Subset with a fish consumption advisory for one or more pollutants associated with the evaluated wastestreams ^c	50	42 (-8)	23 (-27)	10 (-40)

²³ Fish consumption advisory waters are waterbodies for which states, territories, and authorized tribes have issued fish consumption advisories, indicating that pollutant concentrations in the tissues of fish inhabiting those waters are considered unsafe to consume.

²⁴ For more information on drinking water treatment processes used to reduce or eliminate metals commonly detected in the evaluated wastestreams from steam electric power plants, see the memorandum *Drinking Water Treatment Technologies That Can Reduce Metal and Selenium Concentrations Associated with Discharges from Steam Electric Power Plants* (ERG, 2013).

²⁵ See the memorandum *Proximity Analyses and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024j) for a description of the methodology used to evaluate the proximity of plants to CWA section 303(d) impaired waters, fish consumption advisory waters, and drinking water resources.

Table 17. Modeled IRWs Identified as CWA Section 303(d) Impaired Waters, Fish Consumption Advisory Waters, or Drinking Water Resources Under Baseline and Regulatory Options

Sensitive Environment Category	Number of Modeled IRWs Receiving Discharges of the Evaluated Wastestreams (Difference Relative to Baseline) ^a			
	Baseline	Option A	Option B	Option C
Drinking water resource within five miles ^d	116	97 (-19)	54 (-62)	16 (-100)

Source: U.S. EPA, 2024j.

Abbreviations: IRW (immediate receiving water).

a—For this proximity analysis, the EPA evaluated 125 immediate receiving waters that receive discharges of the evaluated wastestreams, either directly or indirectly via a publicly owned treatment works. Of these 125 immediate receiving waters, all 125 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

b—The subset of immediate receiving waters that were impaired due to one or more of the following pollutants: arsenic, boron, cadmium, chlorides, chromium, copper, lead, manganese, mercury, metals (other than mercury), nitrogen (reported as ammonia, nitrate, or nitrite), nutrients, phosphorus, selenium, total dissolved solids, and zinc.

c—The subset of immediate receiving waters with a fish consumption advisory for one or more of the following pollutants: cadmium, lead, mercury, and selenium.

d—Drinking water resources include intakes and reservoirs, public wells, and sole-source aquifers.

4.2.1 Impaired Waters

The EPA estimated that more than half (64 of 125) of the immediate receiving waters analyzed in this EA are CWA Section 303(d) impaired waters.²⁶ As shown in Table 18, 18 of the immediate receiving waters under baseline are impaired for mercury, 16 are impaired for metals (other than mercury),²⁷ and eight are impaired for nutrients. Figure 3 through Figure 5 present the locations of immediate receiving waters that are classified as impaired by high concentrations of these three impairment categories. A total of 43 immediate receiving waters under baseline (34 percent) are impaired for a pollutant associated with the evaluated wastestreams.

Under the final rule (Option B), 40 immediate receiving waters listed as impaired (62.5 percent) will no longer receive discharges of the evaluated wastestreams.

²⁶ See the memorandum *Proximity Analyses and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024j) for a complete list of the impairment categories identified in the EPA’s CWA section 303(d) waters proximity analysis.

²⁷ The “metals (other than mercury)” impairment category in the EPA’s national CWA section 303(d) impaired waters data set includes impairments caused by metalloids and nonmetals such as arsenic, boron, and selenium.

Table 18. Modeled IRWs Identified as CWA Section 303(d) Impaired Waters for Pollutants Present in the Evaluated Wastestreams Under Baseline and Regulatory Options

Pollutant Causing Impairment	Number of Modeled IRWs Receiving Discharges of the Evaluated Wastestreams (Difference Relative to Baseline) ^a			
	Baseline	Option A	Option B	Option C
Mercury	18	18 (0)	11 (-7)	3 (-15)
Metals, other than mercury ^b	16	12 (-4)	2 (-14)	2 (-14)
Nutrients	8	8 (0)	5 (-3)	1 (-7)
TDS	1	0 (-1)	0 (-1)	0 (-1)
Total for Pollutants Associated with the Evaluated Wastestreams^c	43	37 (-6)	17 (-26)	6 (-37)
Total for Any Impairment Category	64	55 (-9)	24 (-40)	8 (-56)

Source: U.S. EPA, 2024j.

Abbreviations: CWA (Clean Water Act); IRW (immediate receiving water); TDS (total dissolved solids).

a—For this proximity analysis, the EPA evaluated 125 immediate receiving waters that receive discharges of the evaluated wastestreams, either directly or indirectly via a publicly owned treatment works. Of these 125 immediate receiving waters, all 125 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

b—Of the 16 immediate receiving waters classified as impaired for “metals, other than mercury” under baseline, five are specifically listed as impaired for one or more of the following individual pollutants evaluated in this environmental assessment: cadmium (1), copper (1), lead (2), manganese (2), selenium (1), and zinc (1). One additional immediate receiving water is impaired for boron (but not included in the “metals, other than mercury” impairment category).

c—Total may not equal the sum of the individual values because some immediate receiving waters are impaired for multiple pollutants.

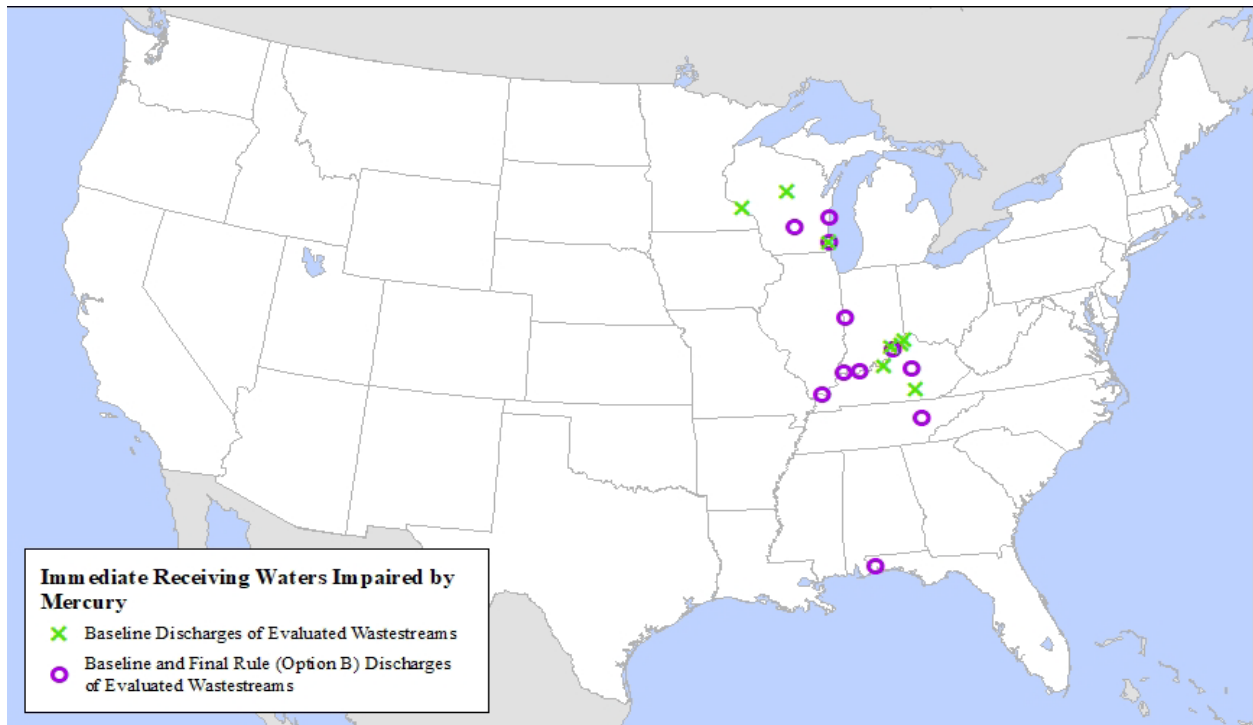


Figure 3. Immediate Receiving Waters Impaired by Mercury

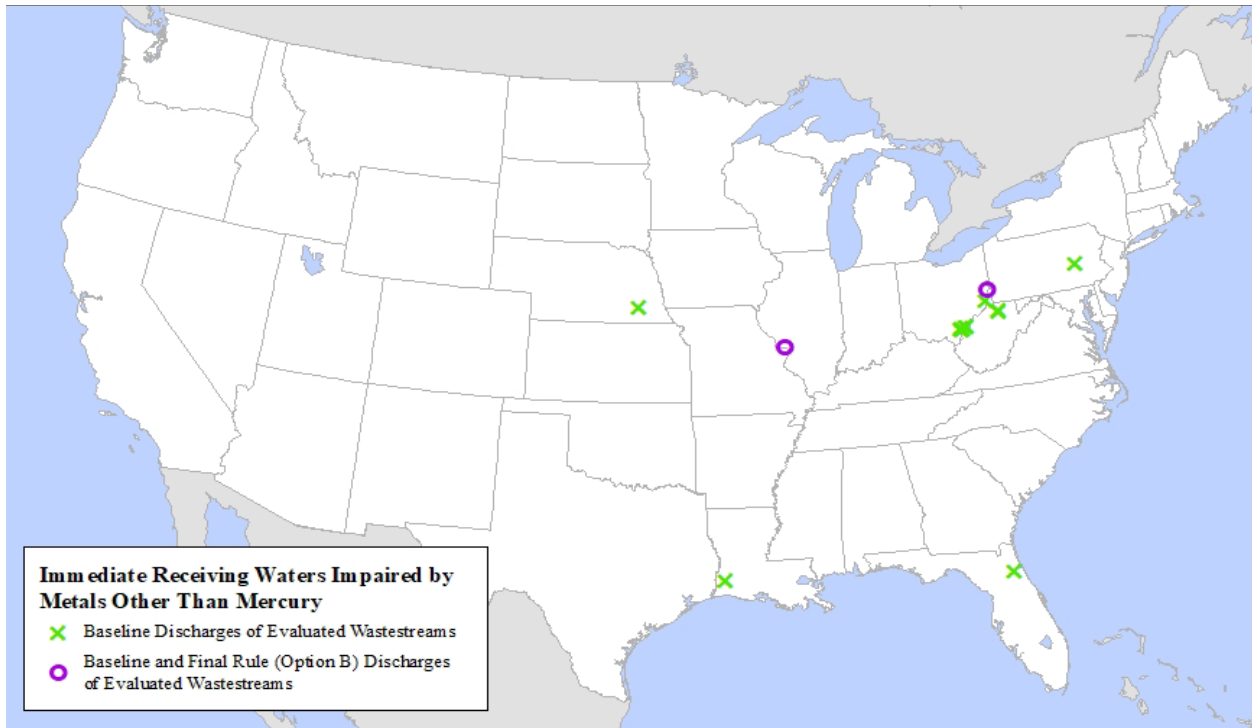


Figure 4. Immediate Receiving Waters Impaired by Metals Other Than Mercury

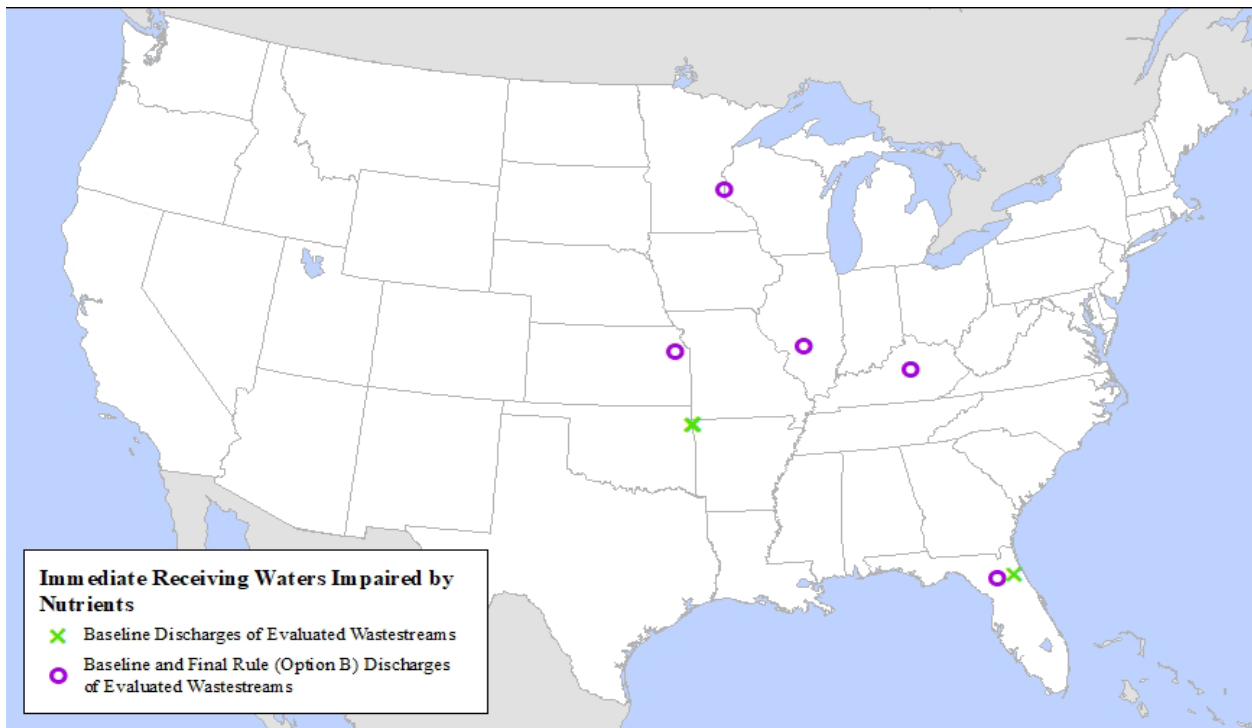


Figure 5. Immediate Receiving Waters Impaired by Nutrients

As shown in Table 2 of this report, the final rule (Option B) results in a decrease in pollutant loadings to the immediate receiving waters, including sensitive environments. The reduction in loadings will help impaired waters to recover; decrease the bioaccumulation of toxic pollutants in fish, thereby reducing the number of fish advisories; and reduce stress on threatened and endangered species and sensitive watersheds such as drinking water resources. The final rule has a net decrease on the loadings of pollutants to waters that are already impaired for those pollutants. The EPA estimated the following net changes relative to baseline in pollutant loadings to impaired waters once requirements under the final rule have been met by the steam electric power plants discharging the evaluated wastestreams to the impaired waterbodies:

- Decrease in nitrogen and phosphorus loadings of 4,910 pounds per year (lb/year) and 220 lb/year, respectively, to nutrient-impaired waters.
- Decrease in phosphorus loadings of 23.0 lb/year to phosphorus-impaired waters.
- Decrease in mercury loadings of 5.90 lb/year to mercury-impaired waters.
- Decrease in loadings to receiving waters impaired for a metal (except mercury), including:
 - Aluminum decrease of 7,190 lb/year.
 - Arsenic decrease of 88.3 lb/year.
 - Boron decrease of 892,000 lb/year.
 - Cadmium decrease of 69.9 lb/year.
 - Chromium decrease of 2,660 lb/year.
 - Copper decrease of 42.8 lb/year.
 - Iron decrease of 37,400 lb/year.
 - Lead decrease of 25.6 lb/year.
 - Magnesium decrease of 12,700,000 lb/year.
 - Manganese decrease of 80,700 lb/year.
 - Nickel decrease of 419 lb/year.
 - Selenium decrease of 267 lb/year.
 - Thallium decrease of 43.2 lb/year.
 - Vanadium decrease of 2,490 lb/year.
 - Zinc decrease of 809 lb/year.
- Decrease in TDS loadings of 135,000 lb/year to one TDS-impaired waterbody.

4.2.2 Fish Consumption Advisories

The EPA estimated that 58 percent (72 of 125) of the immediate receiving waters analyzed in this EA are under a fish consumption advisory.²⁸ As shown in Table 19, 50 of the immediate receiving waters under baseline (40 percent) are under an advisory for a pollutant associated with the evaluated wastestreams. All of these immediate receiving waters are under a fish consumption advisory for mercury, and one is under a fish advisory for lead. Figure 6 presents the locations of immediate receiving waters with fish consumption advisories for mercury.

Under the final rule (Option B), 39 immediate receiving waters with a fish consumption advisory (54 percent reduction) will no longer receive discharges of the evaluated wastestreams. Under the final rule,

²⁸ See the memorandum *Proximity Analyses and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024j) for a complete list of the types of advisories identified in the EPA's fish consumption advisories proximity analysis, including advisories due to pollutants that are not associated with the evaluated wastestreams.

the EPA estimated a decrease in the annual mercury loadings of 22.8 lb/year to immediate receiving waters with a fish consumption advisory for mercury.

Table 19. Modeled IRWs Identified as Fish Consumption Advisory Waters for Pollutants Present in the Evaluated Wastestreams Under Baseline and Regulatory Options

Pollutant Causing Fish Consumption Advisory	Number of Modeled IRWs Receiving Discharges of the Evaluated Wastestreams (Difference Relative to Baseline) ^a			
	Baseline	Option A	Option B	Option C
Lead	1	1 (0)	1 (0)	0 (-1)
Mercury	50	42 (-8)	23 (-27)	10 (-40)
Total for Pollutants Associated with the Evaluated Wastestreams^b	50	42 (-8)	23 (-27)	10 (-40)
Total for Any Fish Advisory	72	60 (-12)	33 (-39)	12 (-60)

Source: U.S. EPA, 2024j.

Abbreviations: IRW (immediate receiving water).

a—For this proximity analysis, the EPA evaluated 125 immediate receiving waters that receive discharges of the evaluated wastestreams, either directly or indirectly via a publicly owned treatment works. Of these 125 immediate receiving waters, all 125 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

b—Total may not equal the sum of the individual values because some immediate receiving waters are under a fish advisory for multiple pollutants.

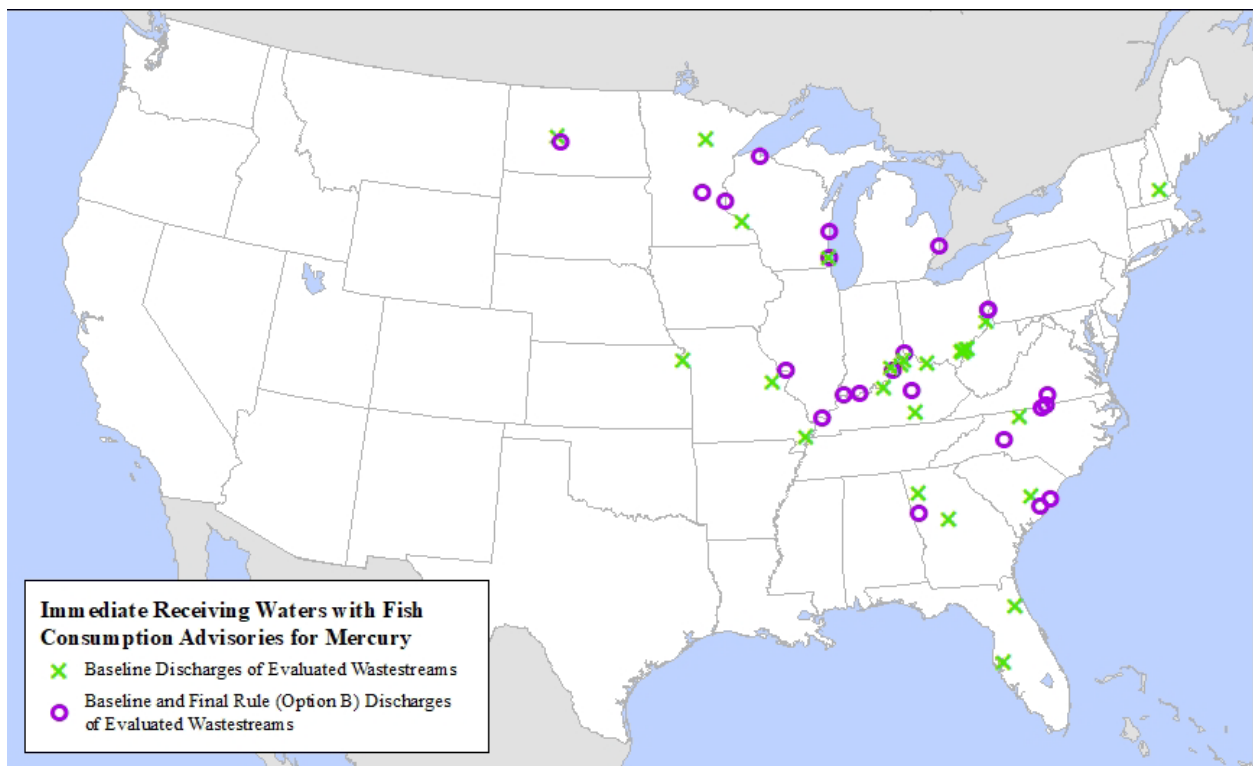


Figure 6. Immediate Receiving Waters with Fish Consumption Advisories for Mercury

4.2.3 Drinking Water Resources

The EPA estimated that 93 percent (116 of 125) of the immediate receiving waters analyzed in this EA are located within 5 miles of a drinking water resource. Under baseline, 103 of the immediate receiving waters (82 percent) are located near public wells, 38 immediate receiving waters (30 percent) are located near drinking water intakes/reservoirs, and two immediate receiving waters (less than 2 percent) are located near sole-source aquifers. Table 20 presents the number of immediate receiving waters evaluated under baseline and the regulatory options and the number of those immediate receiving waters located within 5 miles of a drinking water resource.

Under the final rule (Option B), 62 immediate receiving waters located within 5 miles of a drinking water resource (53 percent reduction) will no longer receive discharges of the evaluated wastestreams.

As discussed in Section 2.2, drinking water supplies can be degraded by pollutants in steam electric power plant wastewater (Cross, 1981), and bromide and iodine discharges are of particular concern due to the formation of disinfection byproducts at drinking water treatment plants and their distribution systems. Under the final rule, the EPA estimated a decrease in bromide loadings of 945,000 to 6.17 million lb/year and a decrease in iodine loadings of 66,900 to 241,000 lb/year to immediate receiving waters located within five miles of drinking water resources.

Table 20. Modeled IRWs Identified as Located Within 5 Miles of a Drinking Water Resource Under Baseline and Regulatory Options

Type of Drinking Water Resource	Number of Modeled IRWs Receiving Discharges of the Evaluated Wastestreams (Difference Relative to Baseline) ^a			
	Baseline	Option A	Option B	Option C
Intakes and reservoirs	38	33 (-5)	24 (-14)	5 (-33)
Public wells	104	87 (-16)	49 (-54)	15 (-88)
Sole-source aquifers	2	2 (0)	1 (-1)	0 (-2)
Total for Any Immediate Receiving Water^b	116	97 (-19)	54 (-62)	16 (-100)

Source: U.S. EPA, 2024j.

Abbreviations: IRW (immediate receiving water).

a—For this proximity analysis, the EPA evaluated 125 immediate receiving waters that receive discharges of the evaluated wastestreams, either directly or indirectly via a publicly owned treatment works. Of these 125 immediate receiving waters, all 125 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

b—Total may not equal the sum of the individual values because some immediate receiving waters are within five miles of multiple drinking water resource types.

4.3 Impacts in Downstream Surface Waters

The EPA performed an analysis of surface waters downstream from the immediate receiving water for each plant that discharges the evaluated wastestreams. The downstream analysis uses the outputs from a separate pollutant fate and transport model (see the BCA Report, U.S. EPA, 2024b, for a description) to assess potential water quality, wildlife, and human health impacts in approximately 17,000 river miles of downstream surface waters. The methodology, which uses estimated annual average pollutant loadings and surface water flow rates, is summarized in Section 3.6 of this report and presented in further detail in the memorandum *Downstream Modeling Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule* (U.S. EPA, 2024l).

Table 21 presents the results of this downstream analysis. This table lists each of the water quality, wildlife, and human health benchmark values used in the IRW Model²⁹ and indicates the total length of downstream surface waters for which the EPA calculated an exceedance of a benchmark value for at least one of the modeled pollutants. Based on the results of the downstream modeling, 777 downstream river miles are affected by steam electric power plant discharges under baseline. Under the final rule (Option B), pollutant concentrations exceeding water quality, wildlife, and/or human health benchmarks will decrease to 411 river miles (47 percent reduction).

Table 21. Modeled Downstream River Miles with Exceedances of Any Pollutant Evaluation Benchmark Value Under Baseline and Regulatory Options

Evaluation Benchmark	Modeled Downstream River Miles Exceeding Benchmark Value (Difference Relative to Baseline) ^a			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC	0	0 (0)	0 (0)	0 (0)
Freshwater chronic NRWQC	16.7	16.1 (-0.607)	2.51 (-14.2)	0 (-16.7)
Human health water and organism NRWQC	363	213 (-149)	104 (-258)	78.0 (-285)
Human health organism only NRWQC	121	29.5 (-91.7)	6.38 (-115)	0 (-121)
Drinking water MCL	1.23	1.23 (0)	0 (-1.23)	0 (-1.23)
<i>Wildlife Results</i>				
Fish ingestion NEHC for minks	40.4	27.5 (-12.9)	4.37 (-36.0)	0 (-40.4)
Fish ingestion NEHC for eagles	121	27.5 (-93.7)	4.37 (-117)	0 (-121)
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	289	186 (-103)	86.0 (-203)	65.1 (-224)
Oral RfD for adult (recreational)	203	94.8 (-108)	54.0 (-149)	41.5 (-162)
Oral RfD for child (subsistence)	688	469 (-219)	333 (-355)	301 (-387)
Oral RfD for adult (subsistence)	420	294 (-126)	193 (-226)	167 (-253)
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	0	0 (0)	0 (0)	0 (0)
LECR for adult (recreational)	1.23	0 (-1.23)	0 (-1.23)	0 (-1.23)
LECR for child (subsistence)	1.23	0 (-1.23)	0 (-1.23)	0 (-1.23)
LECR for adult (subsistence)	13.0	1.23 (-11.8)	0 (-13.0)	0 (-13.0)
Total for Any Benchmark^b	777	547 (-230)	411 (-366)	379 (-398)

Source: U.S. EPA, 2024I.

Abbreviations: LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose).

a—River miles are rounded to three significant figures. As part of this analysis, the EPA evaluated approximately 17,000 river miles of surface waters downstream of immediate receiving waters. For this analysis, the EPA estimated pollutant concentrations in the immediate receiving water and the downstream receiving waters using the D-FATE model.

b—Total may not equal the sum of the individual values because some river miles exceed multiple benchmarks.

²⁹ The water quality outputs used in the downstream analysis were derived from a pollutant fate and transport model that does not simulate pollutant partitioning to the benthic layer; therefore, this analysis does not include comparisons to the sediment TEC.

4.4 Summary of Key Environmental and Human Health Improvements

The EPA estimated that the reduced discharges of pollutants to the immediate receiving waters expected from the final rule will translate into improvements in water quality and reduction in pollutant exposures for wildlife and human health in the immediate receiving waters and further downstream from steam electric power plant discharges. The final supplemental rule will result in the following environmental improvements as estimated by the EA:

- 63 percent reduction in the number of immediate receiving waters exceeding an NRWQC for the protection of human health.
- Over 85 percent reduction in the number of immediate receiving waters that support fish whose tissue pollutant concentrations exceed mercury benchmarks for the protection of piscivorous wildlife (represented by minks and eagles).
- 69 percent reduction in the number of immediate receiving waters that support fish whose tissue pollutant concentrations exceed fish consumption advisories.
- 62 percent reduction in the number of immediate receiving waters that support fish whose tissue pollutant concentrations pose a risk of noncancer health effects in exposed populations.
- 78 percent reduction in the number of immediate receiving waters that support fish whose arsenic tissue concentrations pose a cancer risk to exposed populations.

As shown in the downstream modeling analysis, discharges of the evaluated wastestreams affect surface waters beyond the immediate receiving waters. Pollutant removals associated with the final rule will improve environmental and human health for communities beyond the area immediately surrounding steam electric power plants.

The environmental improvements quantified in the EA do not encompass the full range of improvements that will result from the final supplemental rule. For example, the following improvements are not quantified (or have only limited analysis) in this EA:

- Reducing the loadings of bioaccumulative pollutants to the broader ecosystem, resulting in decrease in long-term exposures and sublethal ecological effects.
- Reducing sublethal chronic effects of toxic pollutants on aquatic life not captured by the NRWQC.
- Mitigating impacts to the population diversity and community structures of aquatic and aquatic-dependent wildlife.
- Reducing loadings of pollutants for which the EPA did not perform water quality modeling in support of the EA (*e.g.*, aluminum, boron, iron, manganese, nutrients, TDS, and vanadium).
 - Reducing loadings of bromide and iodine to drinking water resources.

The EPA expects secondary improvements, associated directly or indirectly, as a result of the final supplemental rule. Pollutant removals not only improve water quality in surface waters but also enhance their aesthetics (*e.g.*, by improving clarity and decreasing odor and discoloration). Improvements in surface water quality may improve the quality of source water for downstream drinking water treatment plants and wells that are influenced by surface water. Such improvements may also improve the quality of water used for irrigation or for industrial uses (lower contaminant levels). Recreational benefits from water quality improvements include more enjoyment from swimming, fishing, and boating and potentially increased revenue from more people partaking of recreational activities. The final rule may also reduce economic impacts such as cleanup and treatment costs for contamination, reduce water usage, reduce potential for algal blooms, and decrease air emissions. The BCA Report (U.S. EPA, 2024b) provides further details on these secondary improvements and other benefits.

5. References

1. Ackerson, N.O.B., E.J. Machek, A.H. Killinger, E.A. Crafton, P. Kumkum, H.K. Liberatore, M.J. Plewa, S.D. Richardson, T.A. Ternes, and S.E. Duirk. 2018. Formation of DBPs and halogen-specific TOX in the presence of iopamidol and chlorinated oxidants. *Chemosphere* 202:349–357. EPA-HQ-OW-2009-0819-7862.
2. ADES. 2016. Advanced Emissions Solutions, Inc. *How the Power of Innovation Became a Corporate Compass for the Environment*. (March 31). EPA-HQ-OW-2009-0819-7937.
3. ATSDR. 2004. Agency for Toxic Substances and Disease Registry. *Toxicological Profile for Iodine*. Department of Health and Human Services. Atlanta, GA. (April). EPA-HQ-OW-2009-0819-8214.
4. ATSDR. 2023. Agency for Toxic Substances and Disease Registry. *Minimal Risk Levels (MRLs)*. Department of Health and Human Services. Atlanta, GA. (September). DCN SE11765.
5. AWWARF and U.S. EPA. 2007. American Water Works Association Research Foundation and U.S. Environmental Protection Agency. *Long Term Variability of BDOM and NOM as Precursors in Watershed Sources*. EPA-HQ-OW-2009-0819-7906.
6. Bettinelli, M., S. Spezia, C. Minoia, and A. Ronchi. 2002. Determination of chlorine, fluorine, bromine, and iodine in coals with ICP-MS and I.C. *Atomic Spectroscopy* 23(4):105–110. EPA-HQ-OW-2009-0819-7863.
7. Bichsel, Y., and U. von Gunten. 2000. Formation of iodo-trihalomethanes during disinfection and oxidation of iodide containing waters. *Environmental Science & Technology* 34:2784–2791. EPA-HQ-OW-2009-0819-7864.
8. Bond, T., J. Huang, N.J.D. Graham, and M.R. Templeton. 2014. Examining the interrelationship between DOC, bromide and chlorine dose on DBP formation in drinking water—A case study. *Science of the Total Environment* 470–471:469–479. EPA-HQ-OW-2009-0819-7866.
9. Brandt, J.E., E.S. Bernhardt, G.S. Dwyer, and R.T. Di Giulio. 2017. Selenium ecotoxicology in freshwater lakes receiving coal combustion residual effluents: A North Carolina example. *Environmental Science & Technology* 51:2418–2426. EPA-HQ-OW-2009-0819-7991.
10. Brandt, J.E., M. Simonin, R.T. Di Giulio, and E.S. Bernhardt. 2019. Beyond selenium: Coal combustion residuals lead to multielement enrichment in receiving lake food webs. *Environmental Science & Technology* 53:4119–4127. EPA-HQ-OW-2009-0819-8734 and EPA-HQ-OW-2009-0819-8734.1.
11. Breinlinger, S., T.J. Phillips, B.N. Haram, J. Mareš, J.A. Martínez Yerena, P. Hrouzek, R. Sobotka, W.M. Henderson, P. Schmieder, S.M. Williams, J.D. Lauderdale, H.D. Wilde, W. Gerrin, A. Kust, J.W. Washington, C. Wagner, B. Geier, M. Liebeke, H. Enke, T.H.J. Niedermeyer, and S.B. Wilde. 2021. Hunting the eagle killer: A cyanobacterial neurotoxin causes vacuolar myelinopathy. *Science* 371(6536). <https://doi.org/10.1126/science.aax9050>. DCN SE10266.
12. Bringmann, G., and R. Kühn. 1980. Comparison of the toxicity thresholds of water pollutants to bacteria, algae, and protozoa in the cell multiplication inhibition test. *Water Research* 14:231–241. EPA-HQ-OW-2009-0819-9008.
13. Bryan, A.L. Jr., W.A. Hopkins, J.A. Baionno, and B.P. Jackson. 2003. Maternal transfer of contaminants to eggs in common grackles (*Quiscalus quiscula*) nesting on coal fly ash basins. *Archives of Environmental Contamination and Toxicology* 45(2):273–277. EPA-HQ-OW-2009-0819-0071.
14. Burger, J., K.F. Gaines, C.G. Lord, I.L. Brisbin, S. Shukla, and M. Gochfield. 2002. Metal levels in raccoon tissues: Differences on and off the Department of Energy’s Savannah River site in South Carolina. *Environmental Monitoring and Assessment* 74:67–84. EPA-HQ-OW-2009-0819-0104.

15. Cañedo-Argüelles, M., B.J. Kefford, C. Piscart, N. Prat, R.B. Schäfer, and C. Schulz. 2013. Salinisation of rivers: An urgent ecological issue. *Environmental Pollution* 173:157–167. EPA-HQ-OW-2009-0819-7975.
16. Cantor, K.P., C.M. Villanueva, C.T. Silverman, J.D. Figueroa, F.X. Real, M. Garcia-Closas, N. Malats, S. Chanock, M. Yeager, A. Tardon, R. Garcia-Closas, C. Serra., A. Carrato, G. Castano-Vinyals, C. Samanic, N. Rothman, and M. Kogevinas. 2010. Polymorphisms in GSTT1, GSTZ1, and CYP2E1, disinfection by-products, and risk of bladder cancer in Spain. *Environmental Health Perspectives* 118(11):1545–1550. EPA-HQ-OW-2009-0819-0931.
17. Carlson, C.L., and D.C. Adriano. 1993. Environmental impacts of coal combustion residues. *Journal of Environmental Quality* 22:227–247. EPA-HQ-OW-2009-0819-0947.
18. Chang, E.E., Y.P. Lin, and P.C. Chiang. 2001. Effects of bromide on the formation of THMs and HAAs. *Chemosphere* 43:1029–1034. EPA-HQ-OW-2009-0819-7842.
19. Chapman, P., W. Adams, M. Brooks, C. Delos, S. Luoma, W. Maher, H. Ohlendorf, T. Presser, and P. Shaw. 2009. *Pellston Workshop on Ecological Assessment of Selenium in the Aquatic Environment*. Society of Environmental Toxicology and Chemistry. Pensacola, FL. EPA-HQ-OW-2009-0819-5698.
20. Chisholm, K., A. Cook, C. Bower, and P. Weinstein. 2008. Risk of birth defects in Australian communities with high levels of brominated disinfection by-products. *Environmental Health Perspectives* 116(9):1267–1273. EPA-HQ-OW-2009-0819-7877.
21. Cornwell, D.A., B.K. Sidhu, R. Brown, and N.E. McTigue. 2018. Modeling bromide river transport and bromide impacts on disinfection byproducts. *Journal AWWA* 110(1):1–23. EPA-HQ-OW-2009-0819-7856.
22. Corsi, S.R., D.J. Graczyk, S.W. Geis, N.L. Booth, and K.D. Richards. 2010. A fresh look at road salt: Aquatic toxicity and water-quality impacts on local, regional, and national scales. *Environmental Science & Technology* 44(19):7376–7382. EPA-HQ-OW-2009-0819-8735.
23. Cortés, C., and R. Marcos. 2018. Genotoxicity of disinfection byproducts and disinfected waters: A review of recent literature. *Mutation Research—Genetic Toxicology and Environmental Mutagenesis* 831:1–12. EPA-HQ-OW-2009-0819-7878.
24. Coughlan, D.J., and J.S. Velte. 1989. Dietary toxicity of selenium-contaminated red shiners to striped bass. *Transactions of the American Fisheries Society* 118:400–408. EPA-HQ-OW-2009-0819-0085.
25. Criquet, J., S. Allard, E. Salhi, C.A. Joll, A. Heitz, and U. von Gunten. 2012. Iodate and iodo-trihalomethane formation during chlorination of iodide-containing waters: Role of bromide. *Environmental Science & Technology* 46:7350–7357. EPA-HQ-OW-2009-0819-7954.
26. Cross, F.L. 1981. Coal pile environmental impact problems. *Pollution Engineering* 13(7):35–37. EPA-HQ-OW-2009-0819-0116.
27. Cui, H., B. Chen, Y. Jiang, Y. Tao, X. Zhu, and Z. Cai. 2021. Toxicity of 17 disinfection by-products to different trophic levels of aquatic organisms: Ecological risks and mechanisms. *Environmental Science & Technology* 55(15):10534–10541. <https://doi.org/10.1021/acs.est.0c08796>. DCN SE10267.
28. Cumbie, P.M., and S.L. Van Horn. 1978. Selenium accumulation associated with fish mortality and reproductive failure. *Proceedings of Annual Conference of the Southeastern Association of Fish and Wildlife* 32:612–624. EPA-HQ-OW-2009-0819-0086.
29. Dong, H., Z. Qiang, and S.D. Richardson. 2019. Formation of iodinated disinfection byproducts (I-DBPs) in drinking water: Emerging concerns and current issues. *Accounts of Chemical Research* 52:896–905. EPA-HQ-OW-2009-0819-7881.

30. Duirk, S.E., C. Lindell, C.C. Cornelison, J. Kormos, T.A. Ternes, M. Attene-Ramos, J. Osiol, E.D. Wagner, M.J. Plewa, and S.D. Richardson. 2011. Formation of toxic iodinated disinfection by-products from compounds used in medical imaging. *Environmental Science & Technology* 45:6845–6854. EPA-HQ-OW-2009-0819-7955.
31. EPRI. 2009. Electric Power Research Institute. *Multimedia Fate of Trace Elements at a Full-Scale Bituminous Coal Power Plant with an SCR and Wet FGD*. (May). EPA-HQ-OW-2009-0819-7882.
32. EPRI. 2014. Electric Power Research Institute. *Program on Technology Innovation: Bromine Usage, Fate, and Potential Impacts for Fossil Fuel–Fired Power Plants: A Literature Review*. (July). EPA-HQ-OW-2009-0819-7355.
33. ERG. 2012. Eastern Research Group, Inc. *Final Power Plant Monitoring Data Collected Under Clean Water Act Section 308 Authority (“CWA 308 Monitoring Data”)*. (May 30). EPA-HQ-OW-2009-0819-0885.
34. ERG. 2013. Eastern Research Group, Inc. *Drinking Water Treatment Technologies That Can Reduce Metal and Selenium Concentrations Associated with Discharges from Steam Electric Power Plants*. (April). EPA-HQ-OW-2009-0819-2231.
35. ERG. 2015a. Eastern Research Group, Inc. *Damage Cases and Other Site Impacts*. (September). EPA-HQ-OW-2009-0819-6421.
36. ERG. 2015b. Eastern Research Group, Inc. *Fish Consumption Rates Used in the EA Human Health Module*. (September 21). EPA-HQ-OW-2009-0819-5788.
37. Ersan, M.S., C. Liu, G. Amy, M.J. Plewa, E.D. Wagner, and T. Karanfil. 2019. Chloramination of iodide-containing waters: Formation of iodinated disinfection byproducts and toxicity correlation with total organic halides of treated waters. *Science of the Total Environment* 697:134142. (December 20). EPA-HQ-OW-2009-0819-8737.
38. Evans, M., and C. Frick. 2001. *The Effects of Road Salts on Aquatic Ecosystems*. (August). EPA-HQ-OW-2009-0819-7984.
39. Ewell, W.S., J.W. Gorsuch, R.O. Kringle, K.A. Robillard, and R.C. Spiegel. 1986. Simultaneous evaluation of the acute effects of chemicals on seven aquatic species. *Environmental Toxicology and Chemistry* 5:831–840. (As cited in Flury and Papritz, 1993.)
40. Flury, M., and A. Papritz. 1993. Bromide in the natural environment: Occurrence and toxicity. *Journal of Environmental Quality* 22(4):747–758. EPA-HQ-OW-2009-0819-7883.
41. Frankel, T.E., C. Crowell, L. Giancarlo, D.L. Hydorn, and B.K. Odhiambo. 2022. Investigating the potential impacts of coal ash runoff on the freshwater Seminole ramshorn snail (*Planorbella duryi*) under laboratory conditions. *Chemosphere*. October. 10.1016/j.chemosphere.2022.136815. DCN SE11732.
42. Frankel, T.E., E. Tyler, C. Willmore, B.K. Odhiambo, and L. Giancarlo. 2023. Assessing the presence, concentration, and impacts of trace element contamination in a Chesapeake Bay tributary adjacent to a coal ash landfill (Possum Point, VA). *Environmental Pollution* 339: 122768. <https://doi.org/10.1016/j.envpol.2023.122768>. DCN SE11733.
43. Fuge, R., and C.C. Johnson. 1986. The geochemistry of iodine—A review. *Environmental Geochemistry and Health* 8(2):31–54. EPA-HQ-OW-2009-0819-7884.
44. Gadgil, M. 2016. 20 years of mercury re-emission—What do we know? Presented at the *Power Plant Pollutant Control and Carbon Management “MEGA” Symposium*. Baltimore, MD. (August 16–18). EPA-HQ-OW-2009-0819-7885.
45. Ged, E.C., and T.H. Boyer. 2014. Effect of seawater intrusion on formation of bromine containing trihalomethanes and haloacetic acids during chlorination. *Desalination* 345:85–93. EPA-HQ-OW-2009-0819-7886.

46. Gillespie, R.B., and P.C. Baumann. 1986. Effects of high tissue concentrations of selenium on reproduction by bluegills. *Transactions of the American Fisheries Society*, 115:208–213. EPA-HQ-OW-2009-0819-0087.
47. Gluskoter, H.J., R.R. Ruch, W.G. Miller, R.A. Cahill, G.B. Dreher, and J.K. Kuhn. 1977. Trace elements in coal: Occurrence and distribution. *Illinois State Geological Survey Circular No. 499*, Urbana, IL. 1–166. (June). EPA-HQ-OW-2009-0819-8959.
48. Good, K.D. 2018. *Coal-fired power plant wastewater contributions to bromide concentrations in drinking water sources*. Dissertation. Carnegie Mellon University. Pittsburgh, PA. (December). EPA-HQ-OW-2009-0819-7860.
49. Good, K.D., and J.M. VanBriesen. 2016. Current and potential future bromide loads from coal-fired power plants in the Allegheny River Basin and their effects on downstream concentrations. *Environmental Science & Technology* 50:9078–9088. EPA-HQ-OW-2009-0819-7711.
50. Good, K.D., and J.M. VanBriesen. 2017. Power plant bromide discharges and downstream drinking water systems in Pennsylvania. *Environmental Science & Technology* 51:11829–11838. EPA-HQ-OW-2009-0819-7831.
51. Good, K.D., and J.M. VanBriesen. 2019. Coal-fired power plant wet flue gas desulfurization bromide discharges to U.S. watersheds and their contributions to drinking water sources. *Environmental Science & Technology* 53:213–223. EPA-HQ-OW-2009-0819-7888.
52. Gruchlik, Y., A. Heitz, C. Joll, U. von Gunten, S. Allard, J. Criquet, S. McDonald, J. Tan, L. Breckler, F. Bradder, G. Roeszler, D. Halliwell. 2015. *Novel treatment technologies for the minimisation of bromide and iodide in drinking water*. Water Corporation of Western Australia and Water Research Australia. EPA-HQ-OW-2009-0819-7844.
53. Hanigan, D., L. Truong, M. Simonich, R. Tanguay, and P. Westerhoff. 2017. Zebrafish embryo toxicity of 15 chlorinated, brominated, and iodinated disinfection by-products. *Journal of Environmental Sciences* 58:302–310. EPA-HQ-OW-2009-0819-7889.
54. Harb, C., J. Pan, S. DeVilbiss, B. Badgley, L.C. Marr, D.G. Schmale, and H. Foroutan. 2021. Increasing Freshwater Salinity Impacts Aerosolized Bacteria. *Environmental Science & Technology* 55(9):5731–5741. <https://doi.org/10.1021/acs.est.0c08558>. DCN SE10268.
55. Harkness, J.S., B. Sulkin, and A. Vengosh. 2016. Evidence for coal ash ponds leaking in the southeastern United States. *Environmental Science and Technology* 50(12): 6583–6592. <https://doi.org/10.1021/acs.est.6b01727>. DCN SE11736.
56. Health Canada. 2015. *Bromate in Drinking Water* (document for public consultation). (July). EPA-HQ-OW-2009-0819-7870.
57. Heeb, M.B., J. Criquet, S.G. Zimmermann-Steffens, and U. von Gunten. 2014. Oxidative treatment of bromide-containing waters: Formation of bromine and its reactions with inorganic and organic compounds—A critical review. *Water Research* 48(1):15–42. EPA-HQ-OW-2009-0819-7892.
58. Hem, J.D. 1985. *Study and Interpretation of the Chemical Characteristics of Natural Water*. Supply Paper 2254. U.S. Geological Survey. Richmond, VA. EPA-HQ-OW-2009-0819-8005.
59. Hong, H.C., Y. Liang, B.P. Han, A. Mazumder, and M.H. Wong. 2007. Modeling of trihalomethane (THM) formation via chlorination of the water from Dongjiang River (source water for Hong Kong's drinking water). *Science of the Total Environment* 385:48–54. EPA-HQ-OW-2009-0819-7895.
60. Hopkins, W.A., M.T. Mendonça, and J.D. Congdon. 1997. Increased circulating levels of testosterone and corticosterone in southern toads, *Bufo terrestris*, exposed to coal combustion waste. *General and Comparative Endocrinology* 108:237–246. EPA-HQ-OW-2009-0819-0949.

61. Hopkins, W.A., M.T. Mendonça, C.L. Rowe, and J.D. Congdon. 1998. Elevated trace element concentrations in southern toads, *Bufo terrestris*, exposed to coal combustion waste. *Archives of Environmental Contamination and Toxicology* 35:325–329. EPA-HQ-OW-2009-0819-0943.
62. Hopkins, W.A., J.W. Snodgrass, J.H. Roe, B.P. Jackson, J.C. Gariboldi, and J.D. Congdon. 2000. Detrimental effects associated with trace element uptake in lake chubsuckers *Erimyzon sucetta* exposed to polluted sediments. *Archives of Environmental Contamination and Toxicology* 39:193–199. EPA-HQ-OW-2009-0819-0076.
63. Hopkins, W.A., S.E. Durant, B.P. Staub, C.L. Rowe, and B.P. Jackson. 2006. Reproduction, embryonic development, and maternal transfer of contaminants in the amphibian *Gastrophryne carolinensis*. *Environmental Health Perspectives* 114(5):661–666. EPA-HQ-OW-2009-0819-0074.
64. Hua, G., D.A. Reckhow, and J. Kim. 2006. Effect of bromide and iodide ions on the formation and speciation of disinfection byproducts during chlorination. *Environmental Science & Technology* 40(9):3050–3056. EPA-HQ-OW-2009-0819-7897.
65. Hua, G., and D.A. Reckhow. 2007. Comparison of disinfection byproduct formation from chlorine and alternative disinfectants. *Water Research* 41:1667–1678. EPA-HQ-OW-2009-0819-7898.
66. Huang, K.Z., Y.F. Xie, and H.L. Tang. 2019. Formation of disinfection by-products under influence of shale gas produced water. *Science of the Total Environment* 647:744–751. EPA-HQ-OW-2009-0819-7900.
67. ICAC. 2019. Institute of Clean Air Companies. *ICAC Technical Briefing on Mercury Control Technologies*. (March 21). EPA-HQ-OW-2009-0819-7371.1.
68. Janz, D.M., D.K. DeForest, M.L. Brooks, P.M. Chapman, G. Gilron, D. Hoff, W.A. Hopkins, D.O. McIntyre, C.A. Mebane, V.P. Palace, J.P. Skorupa, and M. Wayland. 2010. Selenium toxicity to aquatic organisms. (As cited in Chapman et al., 2009.)
69. Javed, M., I. Ahmad, N. Usmani, and M. Ahmad. 2016. Bioaccumulation, oxidative stress and genotoxicity in fish (*Channa punctatus*) exposed to a thermal power plant effluent. *Ecotoxicology and Environmental Safety* 127:163–169. EPA-HQ-OW-2009-0819-7990.
70. Juhnke, I., and D. Lüdemann. 1978. Ergebnisse der Untersuchung von 200 chemischen Verbindungen auf akute Fischtoxizität mit dem Goldorfentest. *Zeitschrift für Wasser und Abwasserforschung* 11:161–165. (As cited in Flury and Papritz, 1993.)
71. Kaushal, S.S., P.M. Groffman, G.E. Likens, K.T. Belt, W.P. Stack, V.R. Kelly, L.E. Band, and G.T. Fisher. 2005. Increased salinization of fresh water in the northeastern United States. *Proceedings of the National Academy of Sciences* 102(38):13517–13520. EPA-HQ-OW-2009-0819-8740.
72. Kolb, C., M. Pozzi, C. Samaras, and J.M. VanBriesen. 2017. Climate change impacts on bromide, trihalomethane formation, and health risks at coastal groundwater utilities. *ASCE-ASME Journal of Risk and Uncertainty in Engineering Systems, Part A: Civil Engineering* 3(3):04017006. EPA-HQ-OW-2009-0819-7713.
73. Kolb, C., K.D. Good, and J.M. VanBriesen. 2020. Modeling trihalomethane increases associated with source water bromide contributed by coal-fired power plants in the Monongahela River basin. *Environmental Science & Technology* 54:726–734. EPA-HQ-OW-2009-0819-8743 and EPA-HQ-OW-2009-0819-8743.1.
74. Krasner, S.W. 2009. The formation and control of emerging disinfection by-products of health concern. *Philosophical Transactions of the Royal Society A* 367:4077–4095. EPA-HQ-OW-2009-0819-7956.

75. Krasner, S.W., H.S. Weinberg, S.D. Richardson, S.J. Pastor, R. Chinn, M.J. Sclimenti, G.D. Onstad, and A.D. Thruston. 2006. Occurrence of a new generation of disinfection byproducts. *Environmental Science & Technology* 40(23):7175–7185. EPA-HQ-OW-2009-0819-7901.
76. Ku, P., M.T. Tsui, S. Liu, K.B. Corson, A.S. Williams, M.R. Monteverde, G.E. Woerndle, A.E. Hershey, and P.A. Rublee. 2020. Examination of mercury contamination from a recent coal ash spill into the Dan River, North Carolina, United States. *Ecotoxicology and Environmental Safety* 208:111469. <https://doi.org/10.1016/j.ecoenv.2020.111469>. DCN SE10269.
77. Landis, M.S., A.S. Kamal, K.D. Kovalcik, C. Croghan, G.A. Norris, and A. Bergdale. 2016. The impact of commercially treated oil and gas produced water discharges on bromide concentrations and modeled brominated trihalomethane disinfection byproducts at two downstream municipal drinking water plants in the upper Allegheny River, Pennsylvania. *Science of the Total Environment* 542:505–520. EPA-HQ-OW-2009-0819-8048.
78. Laverock, M.J., M. Stephenson, and C.R. Macdonald. 1995. Toxicity of iodine, iodide, and iodate to *Daphnia magna* and rainbow trout (*Oncorhynchus mykiss*). *Archives of Environmental Contamination and Toxicology* 29:344–350. EPA-HQ-OW-2009-0819-8049.
79. Lemly, A.D. 1985. Toxicology of selenium in a freshwater reservoir: Implications for environmental hazard evaluation and safety. *Ecotoxicology and Environmental Safety* 10:314–338. EPA-HQ-OW-2009-0819-0077.
80. Lemly, A.D. 1993. Teratogenic effects of selenium in natural populations of freshwater fish. *Ecotoxicology and Environmental Safety* 26(2):181–204. EPA-HQ-OW-2009-0819-5742.
81. Lemly, A.D. 1997a. A teratogenic deformity index for evaluating impacts of selenium on fish populations. *Ecotoxicology and Environmental Safety* 37:259–266. EPA-HQ-OW-2009-0819-5743.
82. Lemly, A.D. 1997b. Ecosystem recovery following selenium contamination in a freshwater reservoir. *Ecotoxicology and Environmental Safety* 36:275–281. EPA-HQ-OW-2009-0819-0940.
83. Lemly, A.D. 2014. Teratogenic effects and monetary cost of selenium poisoning of fish in Lake Sutton, North Carolina. *Ecotoxicology and Environmental Safety* 104:160–167. <https://doi.org/10.1016/j.ecoenv.2014.02.022>. DCN SE10270.
84. Lemly, A.D. 2018. Selenium poisoning of fish by coal ash wastewater in Herrington Lake, Kentucky. *Ecotoxicology and Environmental Safety* 150:49–53. EPA-HQ-OW-2009-0819-7949.
85. LeValley, M.J. 1982. Acute toxicity of iodine to channel catfish (*Ictalurus punctatus*). *Bulletin of Environmental Contamination and Toxicology* 29:7–11. EPA-HQ-OW-2009-0819-8050.
86. Lewis, A., A. Bittner, K. Radloff, and B. Hensel. 2017. Chapter 20: Storage of coal combustion products in the United States: Perspectives on potential human health and environmental risks. In: Robl, T., A. Oberlink, and R. Jones (Eds.). *Coal Combustion Products (CCP's): Characteristics, Utilization and Beneficiation*. 481–507. Woodhead Publishing. <https://doi.org/10.1016/B978-0-08-100945-1.00020-4>. DCN SE10271.
87. Long, K., Y. Sha, Y. Mo, S. Wei, H. Wu, D. Lu, Y. Xia, Q. Yang, W. Zheng, and X. Wei. 2021. Androgenic and teratogenic effects of iodoacetic acid drinking water disinfection byproduct *in vitro* and *in vivo*. *Environmental Science & Technology* 55(6):3827–3835. <https://doi.org/10.1021/acs.est.0c06620>. DCN SE10272.
88. MacDonald, D.D., C.G. Ingersoll, and T.A. Berger. 2000. Development and evaluation of consensus-based sediment quality guidelines for freshwater ecosystems. *Archives of Environmental Contamination and Toxicology* 39(1):20. EPA-HQ-OW-2009-0819-7905.
89. MacKeown, H., J.A. Gyamfi, V.K.M. Schoutteten, D. Dumoulin, L. Verdickt, B. Ouddane, and J. Criquet. 2020. Formation and removal of disinfection by-products in a full scale drinking water treatment plant. *Science of the Total Environment* 704(2020):135280. EPA-HQ-OW-2009-0819-8758.

90. Magnuson, J.J., A.M. Forbes, D.M. Harrell, and J.D. Schwarzmeier. 1980. *Responses of Stream Invertebrates to an Ashpit Effluent Wisconsin Power Plant Impact Study*. EPA-600-3-80-081. U.S. Environmental Protection Agency. EPA-HQ-OW-2009-0819-5745.
91. Maier, K.J., and A.W. Knight. 1994. Ecotoxicology of selenium in freshwater systems. *Reviews of Environmental Contamination and Toxicology* 134:31–48. (As cited in Chapman et al., 2009.)
92. McGuire, M.J., J.L. McLain, and A. Obolensky. 2002. *Information Collection Rule Data Analysis*. American Water Works Association Research Foundation. Denver, CO. EPA-HQ-OW-2009-0819-8052.
93. McTigue, N., D.A. Cornwell, K. Graf, and R. Brown. 2014. Occurrence and consequences of increased bromide in drinking water sources. *Journal AWWA* 106(11):E492–E508. EPA-HQ-OW-2009-0819-5796.
94. Meij, R. 1994. Trace element behavior in coal-fired power plants. *Fuel Processing Technology* 39(1–3):199–217. EPA-HQ-OW-2009-0819-7833.
95. Moore, J., D.L. Bird, S.K. Dobbis, and G. Woodward. 2017. Nonpoint source contributions drive elevated major ion and dissolved inorganic carbon concentrations in urban watersheds. *Environmental Science & Technology Letters* 4:198–204. EPA-HQ-OW-2009-0819-8747.
96. Moran, J.E., S.D. Oktay, and P.H. Santschi. 2002. Sources of iodine and iodine 129 in rivers. *Water Resources Research* 38(8):1149–1159. EPA-HQ-OW-2009-0819-8055.
97. Mount, D.R., D.D. Gulley, and J.M. Evans. 1993. Salinity/toxicity relationship to predict the acute toxicity of produced waters to freshwater organisms. *Proceedings, 1st Society of Petroleum Engineers/U.S. Environmental Protection Agency Environmental Conference*. San Antonio, TX (March 7–10). 605–614. EPA-HQ-OW-2009-0819-7979.
98. Mount, D.R., D.D. Gulley, J.R. Hockett, T.D. Garrison, and J.M. Evans. 1997. Statistical models to predict the toxicity of major ions to *Ceriodaphnia dubia*, *Daphnia magna* and *Pimephales promelas* (fathead minnows). *Environmental Toxicology and Chemistry* 16(10):2009–2019. EPA-HQ-OW-2009-0819-7980.
99. Nagle, R.D., C.L. Rowe, and J.D. Congdon. 2001. Accumulation of selective maternal transfer of contaminants in the turtle *Trachemys scripta* associated with coal ash disposition. *Archives of Environmental Contamination and Toxicology* 40:531–536. EPA-HQ-OW-2009-0819-0078.
100. National Toxicology Program. 2018. *Report on Carcinogens: Monograph on Haloacetic Acids Found as Water Disinfection By-Products*. U.S. Department of Health and Human Services. (March 30). EPA-HQ-OW-2009-0819-8056.
101. National Research Council of the National Academies. 2006. *Managing Coal Combustion Residues in Mines*. National Academies Press. Washington, DC. EPA-HQ-OW-2009-0819-0939.
102. Obolensky, A., and P.C. Singer. 2008. Development and interpretation of disinfection byproduct formation models using the Information Collection Rule database. *Environmental Science & Technology* 42(15):5654–5660. EPA-HQ-OW-2009-0819-8057.
103. Olson, J.R., and C.P. Hawkins. 2012. Predicting natural base-flow stream water chemistry in the western United States. *Water Resources Research* 48:W02504. EPA-HQ-OW-2009-0819-8006.
104. Pan, Y., and X. Zhang. 2013. Four groups and new aromatic halogenated disinfection byproducts: Effect of bromide concentration on their formation and speciation in chlorinated drinking water. *Environmental Science & Technology* 47(3):1265–1273. EPA-HQ-OW-2009-0819-8058.
105. Pantelaki, I., and D. Voutsas. 2018. Formation of iodinated THMs during chlorination of water and wastewater in the presence of different iodine sources. *Science of the Total Environment* 613–614:389–397. EPA-HQ-OW-2009-0819-8059.

106. Parker, K.M., T. Zeng, J. Harkness, A. Vengosh, and W.A. Mitch. 2014. Enhanced formation of disinfection byproducts in shale gas wastewater-impacted drinking water supplies. *Environmental Science & Technology* 48(19):11161–11169. EPA-HQ-OW-2009-0819-8060.
107. Peng, B., L. Li, and D. Wu. 2013. Distribution of bromine and iodine in thermal power plant. *Journal of Coal Science & Engineering* 3:387–391. EPA-HQ-OW-2009-0819-8024.
108. Plewa, M.J., E.D. Wagner, M.G. Muellner, K.-M. Hsu, and S.D. Richardson. 2008. Comparative mammalian cell toxicity of N-DBPs and C-DBPs. In: A. Karanfil, S.W. Krasner, P. Westerhoff, and Y. Xie (Eds.). *Disinfection By-products in Drinking Water*. 36–50. American Chemical Society. EPA-HQ-OW-2009-0819-8062.
109. Pond, G.J. 2004. *Effects of Surface Mining and Residential Land Use on Headwater Stream Biotic Integrity in the Eastern Kentucky Coalfield Region*. Kentucky Department for Environmental Protection. Frankfort, KY. EPA-HQ-OW-2009-0819-8007.
110. Postigo, C., and B. Zonja. 2019. Iodinated disinfection byproducts: Formation and concerns. *Current Opinion in Environmental Science & Health* 7:19–25. EPA-HQ-OW-2009-0819-8064.
111. Pruitt, L., 2000. Indiana’s first HCP conserves least tern—brief article. *Endangered Species Bulletin*. (July). EPA-HQ-OW-2009-0819-0951
112. Ramsey, A.B., A.M. Faiia, and A. Szykiewicz. 2019. Eight years after the coal ash spill—Fate of trace metals in the contaminated river sediments near Kingston, Eastern Tennessee. *Applied Geochemistry* 104:158–167. <https://doi.org/10.1016/j.apgeochem.2019.03.008>. DCN SE11740.
113. Rattner, B.A., M.A. McKernan, K.M. Esereich, W.A. Link, G.H. Olsen, D.J. Hoffman, K.A. Knowles, and P.C. McGowan. 2006. Toxicity and hazard of vanadium to mallard ducks (*Anas platyrhynchos*) and Canada geese (*Branta canadensis*). *Journal of Toxicology and Environmental Health, Part A* 69:331–351. EPA-HQ-OW-2009-0819-0070.
114. Regli, S., J. Chen, M. Messner, M.S. Elovitz, F.L. Letkiewicz, R.A. Pegram, T.J. Pepping, S.D. Richardson, and J.M. Wright. 2015. Estimating potential increased bladder cancer risk due to increased bromide concentrations in sources of disinfected drinking waters. *Environmental Science & Technology* 49:13094–13102. EPA-HQ-OW-2009-0819-7848.
115. Richardson, S.D., and M.J. Plewa. 2020. To regulate or not to regulate? What to do with more toxic disinfection byproducts? *Journal of Environmental Chemical Engineering* 8:103939. Advance online publication. EPA-HQ-OW-2009-0819-8749.
116. Richardson, S.D., and C. Postigo. 2011. Drinking water disinfection by-products. In: D. Barceló (Ed.). *Emerging Organic Contaminants and Human Health*. 93–137. Springer. New York, NY. EPA-HQ-OW-2009-0819-7913.
117. Richardson, S.D., M.J. Plewa, E.D. Wagner, R. Schoeny, and D.M. DeMarini. 2007. Occurrence, genotoxicity, and carcinogenicity of regulated and emerging disinfection by-products in drinking water: A review and roadmap for research. *Mutation Research* 636:178–242. EPA-HQ-OW-2009-0819-7850.
118. Richardson, S.D., F. Fasano, J.J. Ellington, F.G. Crumley, K.M. Buettner, J.J. Evans, B.C. Blount, L.K. Silva, T.J. Waite, G.W. Luther, A.B. McKague, R.J. Miltner, E.D. Wagner, and M.J. Plewa. 2008. Occurrence and mammalian cell toxicity of iodinated disinfection byproducts in drinking water. *Environmental and Science Technology* 42:8330–8338. EPA-HQ-OW-2009-0819-7912.
119. Rowe, C.L., O.M. Kinney, and J.D. Congdon. 1996. Oral deformities in tadpoles (*Rana catesbeiana*) associated with coal ash deposition: Effects on grazing ability and growth. *Freshwater Biology* 36(3):723–730. EPA-HQ-OW-2009-0819-0945.
120. Rowe, C.L., W.A. Hopkins, and J.D. Congdon. 2002. Ecotoxicological implications of aquatic disposal of coal combustion residues. *Environmental Monitoring & Assessment* 80(3):207–276. EPA-HQ-OW-2009-0819-0946.

121. Ruhl, L., A. Vengosh, G.S. Dwyer, H. Hsu-Kim, G. Schwartz, A. Romanski, and S.D. Smith. 2012. The impact of coal combustion residue effluent on water resources: A North Carolina example. *Environmental Science & Technology* 46:12226–12233. EPA-HQ-OW-2009-0819-1978.
122. Sager, D.R., and C.R. Colfield. 1984. Differential accumulation of selenium among axial muscle, reproductive and liver tissues of four warmwater fish species. *Journal of the American Water Resources Association* 20(3):359–363. EPA-HQ-OW-2009-0819-0095.
123. Sahu, R. 2017. *Expert Report Relating to Potential Emissions Impacts from the Installation of a Refined Coal System at the PacifiCorp Hunter Power Plant*. EPA-HQ-OW-2009-0819-8065.
124. Sawade, E., R. Fabris, A. Humpage, and M. Drikas. 2016. Effect of increasing bromide concentration on toxicity in treated drinking water. *Journal of Water and Health* 14(2):183–191. EPA-HQ-OW-2009-0819-7947.
125. Senior, C., S. Sjostrom, G. Mitchell, and D. Boll. 2016. Reducing operating costs and risks of Hg control with fuel additives. Paper #89, presented at the *Power Plant Pollutant Control and Carbon Management “MEGA” Symposium*. Baltimore, MD (August 16–19). EPA-HQ-OW-2009-0819-8066.
126. Silva, A.E., R.J. Speakman, B.F. Barnes, D.R. Coyle, J.C. Leaphart, E.F. Abernethy, K.L. Turner, O.E. Rhodes Jr., J.C. Beasley, and K.J.K. Gandhi. 2023. Bioaccumulation of contaminants in *Scarabaeidae* and *Silphidae* beetles at sites polluted by coal combustion residuals and radiocesium. *Science of the Total Environment* 904: 166821. <https://doi.org/10.1016/j.scitotenv.2023.166821>. DCN SE11743.
127. Sjostrom, S., and C. Senior. 2019. Overview of mercury control approaches used by U.S. plants. *Worldwide Pollution Control Association (WPCA) News*. 27–34. (February). EPA-HQ-OW-2009-0819-8068.
128. Sjostrom, S., K.E. Baldrey, and C. Senior. 2016. *Control of wet scrubber oxidation inhibitor and byproduct recovery*. U.S. Patent Application Publication US2016/0375403A1. (December 29). U.S. Patent and Trademark Office. EPA-HQ-OW-2009-0819-8067.
129. Sorensen, E.M.B. 1988. Selenium accumulation, reproductive status, and histopathological changes in environmentally exposed redear sunfish. *Archives of Toxicology* 61:324–329. EPA-HQ-OW-2009-0819-0096.
130. Sorensen, E.M.B., and T.L. Bauer. 1984. A correlation between selenium accumulation in sunfish and changes in condition factor and organ weight. *Environmental Pollution Series A* 32:357–366. EPA-HQ-OW-2009-0819-0121.
131. Sorensen, E.M.B., C.W. Harlan, and J.S. Bell. 1982. Renal changes in selenium-exposed fish. *The American Journal of Forensic Medicine and Pathology* 3:123–129. EPA-HQ-OW-2009-0819-0123.
132. Sorensen, E.M.B., J.S. Bell, and C.W. Harlan. 1983. Histopathological changes in selenium-exposed fish. *The American Journal of Forensic Medicine and Pathology* 4(2):111–123. EPA-HQ-OW-2009-0819-0097.
133. Sorensen, E.M.B., P.M. Cumbie, T.L. Bauer, J.S. Bell, and C.W. Harlan. 1984. Histopathological, hematological, condition-factor, and organ weight changes associated with selenium accumulation in fish from Belews Lake, North Carolina. *Archives of Environmental Contamination and Toxicology* 13:153–162. EPA-HQ-OW-2009-0819-0075.
134. States, S., G. Cyprych, M. Stoner, F. Wydra, J. Kuchta, J. Monnell, and L. Casson. 2013. Marcellus Shale drilling and brominated THMs in Pittsburgh, PA, drinking water. *Journal AWWA* 105(8):E432–E448. EPA-HQ-OW-2009-0819-6284.

135. Stets, E.G., L.A. Sprague, G.P. Oelsner, H.M. Johnson, J.C. Murphy, K. Ryberg, A.V. Vecchia, R.E. Zuellig, J.A. Falcone, and M.L. Riskin. 2020. Landscape drivers of dynamic change in water quality of U.S. rivers. *Environmental Science & Technology* 54:4336–4343. EPA-HQ-OW-2009-0819-8752.
136. Tian, F.-X., X.-J. Hu, B. Xu, T.-Y. Zhang, and Y.-Q. Gao. 2017. Phototransformation of iodate by UV irradiation: Kinetics and iodinated trihalomethane formation during subsequent chlor(am)ination. *Journal of Hazardous Materials* 326:138–144. EPA-HQ-OW-2009-0819-8070.
137. Tinuum. 2020. Tinuum Group, LLC. *Comments of the Tinuum Group, LLC on the 2019 Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*. (January 21). EPA-HQ-OW-2009-0819-8306-A1.
138. Tugulea, A-M., R. Aranda-Rodriguez, D. Bérubé, M. Giddings, F. Lemieux, J. Hnatiw, L. Dabeka, and F. Breton. 2018. The influence of precursors and treatment process on the formation of Iodo-THMs in Canadian drinking water. *Water Research* 130:215–223. EPA-HQ-OW-2009-0819-8071.
139. U.S. DOJ. 2015. U.S. Department of Justice. *United States of America v. Duke Energy Business Services LLC, et al. Joint Factual Statement*. (February 20). EPA-HQ-OW-2009-0819-8218.
140. U.S. EPA. 2000. U.S. Environmental Protection Agency. *Guidance for Assessing Chemical Contaminant Data for Uses in Fish Advisories, Volume 1*. EPA 823-B-00-007. (November). EPA-HQ-OW-2009-0819-0157.
141. U.S. EPA. 2001. U.S. Environmental Protection Agency. *2001 Update of Ambient Water Quality Criteria for Cadmium*. EPA-822-R-01-001. (April). DCN SE11766.
142. U.S. EPA. 2005. U.S. Environmental Protection Agency. *Guidelines for Carcinogen Risk Assessment*. EPA-630-P-03-001F. Washington, DC. (March). EPA-HQ-OW-2009-0819-7934.
143. U.S. EPA. 2007. U.S. Environmental Protection Agency. *Framework for Metals Risk Assessment*. EPA 120/R-07/001. Washington, DC. (March). EPA-HQ-OW-2009-0819-0162.
144. U.S. EPA. 2009a. U.S. Environmental Protection Agency. *National Primary Drinking Water Regulations*. EPA 816-F-09-004. Washington, DC. (May). EPA-HQ-OW-2009-0819-0144.
145. U.S. EPA. 2009b. U.S. Environmental Protection Agency. *National Recommended Water Quality Criteria*. Washington, DC. EPA-HQ-OW-2009-0819-0143.
146. U.S. EPA. 2011. U.S. Environmental Protection Agency. *The Effects of Mountaintop Mines and Valley Fills on Aquatic Ecosystems of the Central Appalachian Coalfields*. EPA-600-R-09-138F. EPA-HQ-OW-2009-0819-8009.
147. U.S. EPA. 2012. U.S. Environmental Protection Agency. *Provisional Peer-Reviewed Toxicity Values for Thallium and Compounds*. Cincinnati, OH. (September). EPA-HQ-OW-2009-0819-0174.
148. U.S. EPA. 2014. U.S. Environmental Protection Agency. *Damage Case Compendium Technical Support Document, Volume I: Proven Damage Cases*. Washington, DC. EPA-HQ-OW-2009-0819-5765.
149. U.S. EPA. 2015a. U.S. Environmental Protection Agency. *Environmental Assessment for the Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*. EPA-821-R-15-006. (September). EPA-HQ-OW-2009-0819-6427.
150. U.S. EPA. 2015b. U.S. Environmental Protection Agency. *Preventing Eutrophication: Scientific Support for Dual Nutrient Criteria*. EPA 820-S-15-001. Washington, DC. (February). EPA-HQ-OW-2009-0819-5660.
151. U.S. EPA. 2015c. U.S. Environmental Protection Agency. *Sources Contributing Inorganic Species to Drinking Water Intakes During Low Flow Conditions on the Allegheny River in Western Pennsylvania*. EPA-600-R-14-430. (May). EPA-HQ-OW-2009-0819-8226.

152. U.S. EPA. 2016a. U.S. Environmental Protection Agency. *Six-Year Review 3 Technical Support Document for Disinfectants/Disinfection Byproducts Rules*. EPA 810-R-16-012. (December). EPA-HQ-OW-2009-0819-7903.
153. U.S. EPA. 2016b. U.S. Environmental Protection Agency. *Aquatic Life Ambient Water Quality Criterion for Selenium—Freshwater 2016*. EPA 822-R-16-006. (June). EPA-HQ-OW-2009-0819-7828.
154. U.S. EPA. 2016c. U.S. Environmental Protection Agency. *Aquatic Life Ambient Water Quality Criteria for Cadmium—2016*. EPA 820-R-16-002. (March). EPA-HQ-OW-2009-0819-7803.
155. U.S. EPA. 2019. U.S. Environmental Protection Agency. Integrated Risk Information System (IRIS). Washington, DC. EPA-HQ-OW-2009-0819-9002.
156. U.S. EPA. 2020a. U.S. Environmental Protection Agency. *Supplemental Environmental Assessment for Revisions to the Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*. EPA-821-R-20-002. (August). EPA-HQ-OW-2009-0819-9012.
157. U.S. EPA. 2020b. U.S. Environmental Protection Agency. *Supplemental Technical Development Document for Revisions to the Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*. EPA-821-R-20-001. (August). EPA-HQ-OW-2009-0819-8935.
158. U.S. EPA. 2020c. U.S. Environmental Protection Agency. *Estimation of Acute and Chronic Aquatic Life Ambient Freshwater Water Quality Criteria for Copper (for Use in Analyses Supporting the Revised Steam ELG)*. (August). EPA-HQ-OW-2009-0819-8995.
159. U.S. EPA. 2020d. U.S. Environmental Protection Agency. Secondary drinking water standards: Guidance for nuisance chemicals (website). EPA-HQ-OW-2009-0819-8755.
160. U.S. EPA. 2024a. U.S. Environmental Protection Agency. *Technical Development Document for the Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*. EPA-821-R-24-004.
161. U.S. EPA. 2024b. U.S. Environmental Protection Agency. Benefit and Cost Analysis for the Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category. EPA-821-R-24-006.
162. U.S. EPA. 2024c. U.S. Environmental Protection Agency. *Regulatory Impact Analysis for the Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*. EPA-821-R-24-007.
163. U.S. EPA. 2024d. U.S. Environmental Protection Agency. *Environmental Justice Analysis for the Final Supplemental Effluent Limitations Guidelines and Standards for the Steam Electric Power Generating Point Source Category*. EPA-821-R-24-008.
164. U.S. EPA. 2024e. U.S. Environmental Protection Agency. *Literature Review for the 2024 Steam Electric Supplemental Rule Environmental Assessment*. (April). DCN SE11698.
165. U.S. EPA. 2024f. U.S. Environmental Protection Agency. *Receiving Waters Characteristics Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule*. (April). DCN SE11624.
166. U.S. EPA. 2024g. U.S. Environmental Protection Agency. *FGD Halogen Loadings from Steam Electric Power Plants—2024 Final Rule*. (April). DCN SE11703.
167. U.S. EPA. 2024h. U.S. Environmental Protection Agency. *Pollutant Loadings Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule*. (April). DCN SE11701.
168. U.S. EPA. 2024i. U.S. Environmental Protection Agency. *IRW Model: Water Quality, Wildlife, and Human Health Analyses and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule*. (April). DCN SE11700.

169. U.S. EPA. 2024j. U.S. Environmental Protection Agency. *Proximity Analyses and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule*. (April). DCN SE11699.
170. U.S. EPA. 2024k. U.S. Environmental Protection Agency. *Assessment of Human Health Impacts from Multiple Pollutants in Steam Electric Power Plant Discharges*. (April). DCN SE11767.
171. U.S. EPA. 2024l. U.S. Environmental Protection Agency. *Downstream Modeling Analysis and Supporting Documentation for the Environmental Assessment of the Final Supplemental Steam Electric Rule*. (April). DCN SE11702.
172. U.S. EPA. 2024m. U.S. Environmental Protection Agency. *2024 Final Rule - Combustion Residual Leachate Analytical Data Evaluation*. (April). DCN SE11715.
173. U.S. EPA. 2024n. U.S. Environmental Protection Agency. *Updates to Estimated Compliance Costs and Pollutant Loadings*. (April). DCN SE11780.
174. USGS. 2008. U.S. Geological Survey. Environmental contaminants in freshwater fish and their risk to piscivorous wildlife based on a national monitoring program. *Environmental Monitoring and Assessment* 152:469–494. EPA-HQ-OW-2009-0819-0128.
175. Van Dyke, J.U., C.M. Bodinof Jachowski, D.A. Steen, B.P. Jackson, and W.A. Hopkins. 2017. Spatial differences in trace element bioaccumulation in turtles exposed to a partially remediated coal fly ash spill. *Environmental Toxicology and Chemistry* 36:201–211. DCN SE10273.
176. Vengosh, A., E.A. Cowan, R.M. Coyte, A.J. Kondash, Z. Wang, J.E. Brandt, and G.S. Dwyer. 2019. Evidence for unmonitored coal ash spills in Sutton Lake, North Carolina: Implications for contamination of lake ecosystems. *Science of the Total Environment* 686:1090–1103. <https://doi.org/10.1016/j.scitotenv.2019.05.188>. DCN SE11745.
177. Villanueva, C.M., K.P. Cantor, S. Cordier, J.J.K. Jaakkola, W.D. King, C.F. Lynch, S. Porru, and M. Kogevinas. 2004. Disinfection byproducts and bladder cancer: A pooled analysis. *Epidemiology* 15(3):357–367. EPA-HQ-OW-2009-0819-0933.
178. Villanueva, C.M., K.P. Cantor, J.O. Grimalt, N. Malats, D. Silverman, A. Tardon, R. Garcia-Closas, C. Serra, A. Carrato, G. Castaño-Vinyals, R. Marcos, N. Rothman, F.X. Real, M. Dosemeci, and M. Kogevinas. 2007. Bladder cancer and exposure to water disinfection by-products through ingestion, bathing, showering, and swimming in pools. *American Journal of Epidemiology* 165(2):148–156. EPA-HQ-OW-2009-0819-7921.
179. Villanueva, C.M., S. Cordier, L. Font-Ribera, L.A. Salas, and P. Levallois. 2015. Overview of disinfection by-products and associated health effects. *Current Environmental Health Reports* 2(1):107–115. EPA-HQ-OW-2009-0819-7852.
180. Wagner, E.D., and M.J. Plewa. 2017. CHO cell cytotoxicity and genotoxicity analyses of disinfection by-products: An updated review. *Journal of Environmental Sciences* 58:64–76. EPA-HQ-OW-2009-0819-7922.
181. Wang, Y., M.J. Small, and J.M. VanBriesen. 2017. Assessing the risk associated with increasing bromide in drinking water sources in the Monongahela River, Pennsylvania. *Journal of Environmental Engineering* 143(3):1–10. EPA-HQ-OW-2009-0819-7853.
182. Watson, K., M.M.J. Farré, and N. Knight. 2015. Enhanced coagulation with powdered activated carbon or MIEX® secondary treatment: A comparison of disinfection by-product formation and precursor removal. *Water Research* 68:454–466. EPA-HQ-OW-2009-0819-7929.
183. Weber-Scannell, P., and L. Duffy. 2007. Effects of total dissolved solids on aquatic organisms: A review of literature and recommendations for salmonid species. *American Journal of Environmental Sciences* 3:1–6. EPA-HQ-OW-2009-0819-7982.

184. Wei, X., S. Wang, W. Zheng, X. Wang, X. Liu, S. Jiang, J. Pi, Y. Zheng, G. He, and W. Qu. 2013. Drinking water disinfection byproduct iodoacetic acid induces tumorigenic transformation of NIH3T3 cells. *Environmental Science & Technology* 47:5913–5920. EPA-HQ-OW-2009-0819-7930.
185. Weinberg, H.S., S.W. Krasner, S.D. Richardson, and A.D. Thruston, Jr. 2002. *The Occurrence of Disinfection By-products (DBPs) of Health Concern in Drinking Water: Results of a Nationwide DBP Occurrence Study*. U.S. Environmental Protection Agency, Office of Research and Development, National Exposure Research Laboratory. EPA 600-R-02-068. Athens, GA. (September). EPA-HQ-OW-2009-0819-7855.
186. Weisman, R., A. Heinrich, F. Letkiewicz, M. Messner, K. Studer, L. Wang, and S. Regli. 2022. Estimating national exposures and potential bladder cancer cases associated with chlorination DBPs in U.S. drinking water. *Environmental Health Perspectives* 130(8). EPA-HQ-OW-2009-0819-9608.
187. Westerhoff, P., P. Chao, and H. Mash. 2004. Reactivity of natural organic matter with aqueous chlorine and bromine. *Water Research* 38(6):1502–1513. EPA-HQ-OW-2009-0819-7931.
188. WHO. 2009. World Health Organization. *Bromide in Drinking-Water: Background document for development of WHO Guidelines for Drinking-Water Quality*. WHO/HSE/WSH/09.01/6. Geneva, Switzerland. EPA-HQ-OW-2009-0819-0155.
189. Xia, Y., Y.-L. Lin, B. Xu, C.-Y. Hu, Z.-C. Gao, W.-H. Chu, and N.-Y. Gao. 2017. Iodinated trihalomethane formation during chloramination of iodate containing waters in the presence of zero valent iron. *Water Research* 124:219–226. EPA-HQ-OW-2009-0819-8075.
190. Yan, M., M. Li, and X. Han. 2016. Behaviour of I/Br/Cl-THMs and their projected toxicities under simulated cooking conditions: Effects of heating, table salt and residual chlorine. *Journal of Hazardous Materials* 314:105–112. EPA-HQ-OW-2009-0819-8076.
191. Yang, X., and C. Shang. 2004. Chlorination byproduct formation in the presence of humic acid, model nitrogenous organic compounds, ammonia, and bromide. *Environmental Science & Technology* 38(19):4995–5001. EPA-HQ-OW-2009-0819-8077.
192. Yang, Y., Y. Komaki, S.Y. Kimura, H.-Y. Hu, E.D. Wagner, B.J.B. Mariñas, and M.J. Plewa. 2014. Toxic impact of bromide and iodide on drinking water disinfected with chlorine or chloramines. *Environmental Science & Technology* 48:12362–12369. EPA-HQ-OW-2009-0819-7858.
193. Ye, T., B. Xu, Y.-L. Lin, C.-Y. Hu, L. Lin, T.-Y. Zhang, and N.-Y. Gao. 2013. Formation of iodinated disinfection by-products during oxidation of iodide-containing waters with chlorine dioxide. *Water Research* 47(9):3006–3014. EPA-HQ-OW-2009-0819-8078.
194. Zha, X.-S., Y. Liu, X. Liu, Q. Zhang, R.-H. Dai, L.-W. Ying, J. Wu, J.-T. Wang, and L. Ma. 2014. Effects of bromide and iodide ions on the formation of disinfection by-products during ozonation and subsequent chlorination of water containing biological source matters. *Environmental Science Pollution Research* 21(4):2714–2723. EPA-HQ-OW-2009-0819-8079.
195. Zhang, T.-Y., Y.-L. Lin, A.-Q. Wang, F.-X. Tian, B. Xu, S.-J. Xia, and N.-Y. Gao. 2016. Formation of iodinated trihalomethanes during UV/chloramination with iodate as the iodine source. *Water Research* 98:199–205. EPA-HQ-OW-2009-0819-8080.

Attachment A. Additional IRW Model Results

This appendix presents pollutant loadings and additional model outputs for all pollutants included in the Immediate Receiving Water (IRW) Model (arsenic, cadmium, copper, lead, mercury, nickel, selenium, thallium, and zinc) beyond those discussed in Section 4 of this environmental assessment. It includes the following tables:

- Table A-1. Modeled IRWs with Exceedances of Benchmark Values for One or More Pollutants Under Baseline and Regulatory Options
- Table A-2. Modeled IRWs with Exceedances of Arsenic Benchmark Values Under Baseline and Regulatory Options
- Table A-3. Modeled IRWs with Exceedances of Cadmium Benchmark Values Under Baseline and Regulatory Options
- Table A-4. Modeled IRWs with Exceedances of Copper Benchmark Values Under Baseline and Regulatory Options
- Table A-5. Modeled IRWs with Exceedances of Lead Benchmark Values Under Baseline and Regulatory Options
- Table A-6. Modeled IRWs with Exceedances of Mercury Benchmark Values Under Baseline and Regulatory Options
- Table A-7. Modeled IRWs with Exceedances of Nickel Benchmark Values Under Baseline and Regulatory Options
- Table A-8. . Modeled IRWs with Exceedances of Selenium Benchmark Values Under Baseline and Regulatory Options
- Table A-9. Modeled IRWs with Exceedances of Thallium Benchmark Values Under Baseline and Regulatory Options
- Table A-10. Modeled IRWs with Exceedances of Zinc Benchmark Values Under Baseline and Regulatory Options
- Table A-11. Modeled IRWs with Exceedances of Arsenic Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options
- Table A-12. Modeled IRWs with Exceedances of Cadmium Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options
- Table A-13. Modeled IRWs with Exceedances of Copper Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options
- Table A-14. Modeled IRWs with Exceedances of Mercury (as Methylmercury) Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options
- Table A-15. Modeled IRWs with Exceedances of Nickel Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options
- Table A-16. Modeled IRWs with Exceedances of Selenium Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options
- Table A-17. Modeled IRWs with Exceedances of Thallium Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options
- Table A-18. Modeled IRWs with Exceedances of Zinc Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options
- Table A-19. Modeled IRWs with Lifetime Excess Cancer Risk for Inorganic Arsenic Exceeding One-in-a-Million by Race/Ethnicity Category Under Baseline and Regulatory Options

Table A-1. Modeled IRWs with Exceedances of Benchmark Values for One or More Pollutants Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Pollutant Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings for the nine modeled pollutants from 110 steam electric power plants in pollutant loadings analysis ^b	17,600	13,400	3,550	3,200
Evaluation Benchmark	Number of Modeled IRWs Exceeding Benchmark Value ^c			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC	3	2	2	2
Freshwater chronic NRWQC	12	11	5	5
HH WO NRWQC	38	28	14	7
HH O NRWQC	21	14	4	3
Drinking water MCL	5	4	3	3
<i>Wildlife Results</i>				
Sediment TEC	24	24	11	7
Fish ingestion NEHC for minks	16	16	6	5
Fish ingestion NEHC for eagles	22	17	6	5
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational)	22	16	4	3
T4 fish tissue concentration screening value (subsistence)	32	24	10	6
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	28	22	9	6
Oral RfD for child (subsistence)	39	28	15	8
Oral RfD for adult (recreational)	26	18	6	5
Oral RfD for adult (subsistence)	31	23	9	6
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	0	0	0	0
LECR for child (subsistence)	3	2	1	1
LECR for adult (recreational)	4	2	2	2
LECR for adult (subsistence)	9	3	2	2

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Values represent the industry loadings and the IRW Model outputs for the following nine evaluated pollutants: arsenic, cadmium, copper, lead, mercury, nickel, selenium, thallium, and zinc. Pollutant loadings are rounded to three significant figures.

b—The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

Table A-2. Modeled IRWs with Exceedances of Arsenic Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Arsenic Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	777	264	77.1	50.8
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC ^e	0	0	0	0
Freshwater chronic NRWQC ^e	0	0	0	0
HH WO NRWQC ^f	38	28	14	7
HH O NRWQC ^f	21	14	4	3
Drinking water MCL	4	2	2	2
<i>Wildlife Results</i>				
Sediment TEC	3	2	0	0
Fish ingestion NEHC for minks	0	0	0	0
Fish ingestion NEHC for eagles	0	0	0	0
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational) ^{f,g}	0	0	0	0
T4 fish tissue concentration screening value (subsistence) ^{f,g}	0	0	0	0
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational) ^f	0	0	0	0
Oral RfD for child (subsistence) ^f	0	0	0	0
Oral RfD for adult (recreational) ^f	0	0	0	0
Oral RfD for adult (subsistence) ^f	0	0	0	0
<i>Human Health Results—Cancer</i>				
LECR for child (recreational) ^f	0	0	0	0
LECR for child (subsistence) ^f	3	2	1	1
LECR for adult (recreational) ^f	4	2	2	2
LECR for adult (subsistence) ^f	9	3	2	2

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b— The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total arsenic concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—Benchmark value is based on dissolved arsenic.

f—Benchmark value is based on inorganic arsenic.

g—Values represent number of immediate receiving waters exceeding either the noncarcinogenic or carcinogenic screening values.

Table A-3. Modeled IRWs with Exceedances of Cadmium Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Cadmium Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	553	401	41.1	23.8
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC ^e	3	2	1	1
Freshwater chronic NRWQC ^e	8	5	2	2
HH WO NRWQC	f	f	f	f
HH O NRWQC	f	f	f	f
Drinking water MCL	3	2	1	1
<i>Wildlife Results</i>				
Sediment TEC	8	5	2	2
Fish ingestion NEHC for minks	1	1	0	0
Fish ingestion NEHC for eagles	1	1	0	0
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational)	1	1	0	0
T4 fish tissue concentration screening value (subsistence)	4	3	2	2
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	3	2	1	1
Oral RfD for child (subsistence)	4	4	2	2
Oral RfD for adult (recreational)	1	1	1	1
Oral RfD for adult (subsistence)	4	3	2	2
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	f	f	f	f
LECR for child (subsistence)	f	f	f	f
LECR for adult (recreational)	f	f	f	f
LECR for adult (subsistence)	f	f	f	f

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b— The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total cadmium concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—Benchmark value is based on dissolved cadmium.

f—A benchmark value is not yet established for this pollutant or was not included in the EPA’s analyses.

Table A-4. Modeled IRWs with Exceedances of Copper Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Copper Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	398	217	49.4	32.9
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC ^e	1	1	0	0
Freshwater chronic NRWQC ^e	2	2	0	0
HH WO NRWQC	0	0	0	0
HH O NRWQC	f	f	f	f
Drinking water MCL	0	0	0	0
<i>Wildlife Results</i>				
Sediment TEC	2	2	1	1
Fish ingestion NEHC for minks	0	0	0	0
Fish ingestion NEHC for eagles	0	0	0	0
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational)	f	f	f	f
T4 fish tissue concentration screening value (subsistence)	f	f	f	f
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	0	0	0	0
Oral RfD for child (subsistence)	1	1	0	0
Oral RfD for adult (recreational)	0	0	0	0
Oral RfD for adult (subsistence)	0	0	0	0
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	f	f	f	f
LECR for child (subsistence)	f	f	f	f
LECR for adult (recreational)	f	f	f	f
LECR for adult (subsistence)	f	f	f	f

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b— The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total copper concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—Benchmark value is based on dissolved copper.

f—A benchmark value is not yet established for this pollutant or was not included in the EPA’s analyses.

Table A-5. Modeled IRWs with Exceedances of Lead Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Lead Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	230	91.6	42.9	29.5
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC ^e	0	0	0	0
Freshwater chronic NRWQC ^e	1	1	0	0
HH WO NRWQC	f	f	f	f
HH O NRWQC	f	f	f	f
Drinking water MCL	2	2	1	1
<i>Wildlife Results</i>				
Sediment TEC	1	1	1	1
Fish ingestion NEHC for minks	0	0	0	0
Fish ingestion NEHC for eagles	0	0	0	0
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational)	f	f	f	f
T4 fish tissue concentration screening value (subsistence)	f	f	f	f
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	f	f	f	f
Oral RfD for child (subsistence)	f	f	f	f
Oral RfD for adult (recreational)	f	f	f	f
Oral RfD for adult (subsistence)	f	f	f	f
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	f	f	f	f
LECR for child (subsistence)	f	f	f	f
LECR for adult (recreational)	f	f	f	f
LECR for adult (subsistence)	f	f	f	f

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b— The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total lead concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—Benchmark value is based on dissolved lead.

f—A benchmark value is not yet established for this pollutant or was not included in the EPA’s analyses.

Table A-6. Modeled IRWs with Exceedances of Mercury Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Mercury Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	40.0	28.5	1.53	1.21
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC ^e	0	0	0	0
Freshwater chronic NRWQC ^e	1	1	0	0
HH WO NRWQC	f	f	f	f
HH O NRWQC	f	f	f	f
Drinking water MCL ^g	1	1	0	0
<i>Wildlife Results</i>				
Sediment TEC	19	9	2	2
Fish ingestion NEHC for minks ^h	16	7	2	2
Fish ingestion NEHC for eagles ^h	22	15	3	2
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational) ^h	22	16	4	3
T4 fish tissue concentration screening value (subsistence) ^h	32	24	10	6
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational) ^h	28	22	8	5
Oral RfD for child (subsistence) ^h	38	28	15	8
Oral RfD for adult (recreational) ^h	25	17	5	4
Oral RfD for adult (subsistence) ^h	31	23	9	6
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	f	f	f	f
LECR for child (subsistence)	f	f	f	f
LECR for adult (recreational)	f	f	f	f
LECR for adult (subsistence)	f	f	f	f

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b—The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total mercury concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—Benchmark value is based on dissolved mercury.

f—A benchmark value is not yet established for this pollutant or was not included in the EPA’s analyses.

g—Benchmark value is based on inorganic mercury.

h—Benchmark value is based on methylmercury.

Table A-7. Modeled IRWs with Exceedances of Nickel Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Nickel Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	3,430	2,740	113	79.4
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC ^e	1	1	0	0
Freshwater chronic NRWQC ^e	1	1	0	0
HH WO NRWQC	1	1	0	0
HH O NRWQC	0	0	0	0
Drinking water MCL	f	f	f	f
<i>Wildlife Results</i>				
Sediment TEC	14	6	2	2
Fish ingestion NEHC for minks	0	0	0	0
Fish ingestion NEHC for eagles	0	0	0	0
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational)	f	f	f	f
T4 fish tissue concentration screening value (subsistence)	f	f	f	f
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	0	0	0	0
Oral RfD for child (subsistence)	0	0	0	0
Oral RfD for adult (recreational)	0	0	0	0
Oral RfD for adult (subsistence)	0	0	0	0
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	f	f	f	f
LECR for child (subsistence)	f	f	f	f
LECR for adult (recreational)	f	f	f	f
LECR for adult (subsistence)	f	f	f	f

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b— The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total nickel concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—Benchmark value is based on dissolved nickel.

f—A benchmark value is not yet established for this pollutant or was not included in the EPA's analyses.

Table A-8. Modeled IRWs with Exceedances of Selenium Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Selenium Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	4,810	4,600	2,840	2,730
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC ^e	1	1	1	1
Freshwater chronic NRWQC ^e	12	11	5	5
HH WO NRWQC	1	1	1	1
HH O NRWQC	1	1	1	1
Drinking water MCL	3	3	2	2
<i>Wildlife Results</i>				
Sediment TEC	24	24	11	7
Fish ingestion NEHC for minks	15	15	6	5
Fish ingestion NEHC for eagles	15	15	6	5
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational)	8	7	3	3
T4 fish tissue concentration screening value (subsistence)	18	18	8	5
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	15	15	6	5
Oral RfD for child (subsistence)	22	22	8	5
Oral RfD for adult (recreational)	12	12	5	4
Oral RfD for adult (subsistence)	15	15	6	5
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	f	f	f	f
LECR for child (subsistence)	f	f	f	f
LECR for adult (recreational)	f	f	f	f
LECR for adult (subsistence)	f	f	f	f

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b— The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total selenium concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—Benchmark value is based on dissolved selenium.

f—A benchmark value is not yet established for this pollutant or was not included in the EPA's analyses.

Table A-9. Modeled IRWs with Exceedances of Thallium Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Thallium Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	781	536	117	85.5
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC	e	e	e	e
Freshwater chronic NRWQC	e	e	e	e
HH WO NRWQC	8	7	4	3
HH O NRWQC	7	5	3	3
Drinking water MCL	2	2	2	2
<i>Wildlife Results</i>				
Sediment TEC	e	e	e	e
Fish ingestion NEHC for minks	e	e	e	e
Fish ingestion NEHC for eagles	e	e	e	e
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational)	e	e	e	e
T4 fish tissue concentration screening value (subsistence)	e	e	e	e
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	16	15	6	5
Oral RfD for child (subsistence)	24	19	10	7
Oral RfD for adult (recreational)	13	9	5	4
Oral RfD for adult (subsistence)	16	15	6	5
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	e	e	e	e
LECR for child (subsistence)	e	e	e	e
LECR for adult (recreational)	e	e	e	e
LECR for adult (subsistence)	e	e	e	e

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b— The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total thallium concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—A benchmark value is not yet established for this pollutant or was not included in the EPA’s analyses.

Table A-10. Modeled IRWs with Exceedances of Zinc Benchmark Values Under Baseline and Regulatory Options

Pollutant Loadings Basis	Industry Zinc Loadings (lb/year) ^a			
	Baseline	Option A	Option B	Option C
Mass loadings from 110 steam electric power plants in pollutant loadings analysis ^b	6,570	4,530	265	174
Evaluation Benchmark ^c	Number of Modeled IRWs Exceeding Benchmark Value ^d			
	Baseline	Option A	Option B	Option C
<i>Water Quality Results</i>				
Freshwater acute NRWQC ^e	1	1	0	0
Freshwater chronic NRWQC ^e	1	1	0	0
HH WO NRWQC	0	0	0	0
HH O NRWQC	0	0	0	0
Drinking water MCL	1	1	0	0
<i>Wildlife Results</i>				
Sediment TEC	7	4	2	2
Fish ingestion NEHC for minks	1	1	0	0
Fish ingestion NEHC for eagles	1	1	0	0
<i>Human Health Results—Fish Consumption Advisories</i>				
T4 fish tissue concentration screening value (recreational)	f	f	f	f
T4 fish tissue concentration screening value (subsistence)	f	f	f	f
<i>Human Health Results—Noncancer</i>				
Oral RfD for child (recreational)	1	1	0	0
Oral RfD for child (subsistence)	1	1	0	0
Oral RfD for adult (recreational)	1	1	0	0
Oral RfD for adult (subsistence)	1	1	0	0
<i>Human Health Results—Cancer</i>				
LECR for child (recreational)	f	f	f	f
LECR for child (subsistence)	f	f	f	f
LECR for adult (recreational)	f	f	f	f
LECR for adult (subsistence)	f	f	f	f

Sources: U.S. EPA, 2024h and 2024i.

Abbreviations: HH O (human health organism only); HH WO (human health water and organism); IRW (immediate receiving water); lb/year (pounds per year); LECR (lifetime excess cancer risk); MCL (maximum contaminant level); NEHC (no effect hazard concentration); NRWQC (National Recommended Water Quality Criteria); RfD (reference dose); TEC (threshold effect concentration); T4 (trophic level 4).

a—Pollutant loadings are rounded to three significant figures.

b—The pollutant loadings analysis includes discharges from 110 plants to 123 immediate receiving waters (some plants discharge to multiple receiving waters). Of these 123 immediate receiving waters, all 123 receive discharges of the evaluated wastestreams under baseline, 105 do under Option A, 57 do under Option B, and 18 do under Option C.

c—All benchmark values are based on total zinc concentration, unless otherwise stated.

d—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

e—Benchmark value is based on dissolved zinc.

f—A benchmark value is not yet established for this pollutant or was not included in the EPA's analyses.

Table A-11. Modeled IRWs with Exceedances of Arsenic Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD ^{a,b}			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Child—subsistence	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Adult—recreational	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Adult—subsistence	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Benchmark value is based on inorganic arsenic.

Table A-12. Modeled IRWs with Exceedances of Cadmium Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD ^{a,b}			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	1	1	1	1
	Non-Hispanic Black	1	1	1	1
	Mexican-American	2	2	1	1
	Other Hispanic	1	1	1	1
	Other, including multiple races	2	2	1	1
Child—subsistence	Non-Hispanic White	3	2	1	1
	Non-Hispanic Black	4	3	2	2
	Mexican-American	4	4	2	2
	Other Hispanic	4	4	2	2
	Other, including multiple races	4	4	2	2
Adult—recreational	Non-Hispanic White	1	1	1	1
	Non-Hispanic Black	1	1	1	1
	Mexican-American	2	2	1	1
	Other Hispanic	1	1	1	1
	Other, including multiple races	2	2	1	1
Adult—subsistence	Non-Hispanic White	3	2	1	1
	Non-Hispanic Black	4	3	2	2
	Mexican-American	4	4	2	2
	Other Hispanic	4	4	2	2
	Other, including multiple races	4	4	2	2

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Benchmark value is based on dissolved cadmium.

Table A-13. Modeled IRWs with Exceedances of Copper Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD ^{a,b}			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Child—subsistence	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Adult—recreational	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Adult—subsistence	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Benchmark value is based on total copper.

Table A-14. Modeled IRWs with Exceedances of Mercury (as Methylmercury) Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD ^a			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	25	17	5	4
	Non-Hispanic Black	26	19	6	4
	Mexican-American	27	20	8	5
	Other Hispanic	26	19	7	5
	Other, including multiple races	27	20	8	5
Child—subsistence	Non-Hispanic White	29	23	9	5
	Non-Hispanic Black	31	23	9	6
	Mexican-American	32	24	10	6
	Other Hispanic	32	23	9	6
	Other, including multiple races	33	26	14	8
Adult—recreational	Non-Hispanic White	25	17	5	4
	Non-Hispanic Black	26	19	6	4
	Mexican-American	27	20	8	5
	Other Hispanic	26	19	7	5
	Other, including multiple races	27	20	8	5
Adult—subsistence	Non-Hispanic White	29	23	9	5
	Non-Hispanic Black	31	23	9	6
	Mexican-American	32	24	10	6
	Other Hispanic	32	23	9	6
	Other, including multiple races	33	26	14	8

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

Table A-15. Modeled IRWs with Exceedances of Nickel Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD ^{a,b}			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Child—subsistence	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Adult—recreational	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Adult—subsistence	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Benchmark value is based on total nickel.

Table A-16. Modeled IRWs with Exceedances of Selenium Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD ^{a,b}			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	12	12	5	4
	Non-Hispanic Black	12	12	5	4
	Mexican-American	13	13	5	4
	Other Hispanic	13	13	5	4
	Other, including multiple races	13	13	5	4
Child—subsistence	Non-Hispanic White	15	15	6	5
	Non-Hispanic Black	15	15	6	5
	Mexican-American	19	19	8	5
	Other Hispanic	17	17	7	5
	Other, including multiple races	19	19	8	5
Adult—recreational	Non-Hispanic White	12	12	5	4
	Non-Hispanic Black	12	12	5	4
	Mexican-American	13	13	5	4
	Other Hispanic	13	13	5	4
	Other, including multiple races	13	13	5	4
Adult—subsistence	Non-Hispanic White	15	15	6	5
	Non-Hispanic Black	15	15	6	5
	Mexican-American	19	19	8	5
	Other Hispanic	17	17	7	5
	Other, including multiple races	19	19	8	5

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Benchmark value is based on total selenium.

Table A-17. Modeled IRWs with Exceedances of Thallium Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD ^{a,b}			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	13	9	5	4
	Non-Hispanic Black	13	12	5	4
	Mexican-American	15	12	5	4
	Other Hispanic	13	12	5	4
	Other, including multiple races	15	12	5	4
Child—subsistence	Non-Hispanic White	16	15	6	5
	Non-Hispanic Black	16	15	6	5
	Mexican-American	20	18	8	6
	Other Hispanic	20	17	6	5
	Other, including multiple races	21	19	10	7
Adult—recreational	Non-Hispanic White	13	9	5	4
	Non-Hispanic Black	13	12	5	4
	Mexican-American	15	12	5	4
	Other Hispanic	13	12	5	4
	Other, including multiple races	15	12	5	4
Adult—subsistence	Non-Hispanic White	16	15	6	5
	Non-Hispanic Black	16	15	6	5
	Mexican-American	20	18	8	6
	Other Hispanic	20	17	6	5
	Other, including multiple races	21	19	10	7

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Benchmark value is based on total thallium.

Table A-18. Modeled IRWs with Exceedances of Zinc Oral Reference Dose Values by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding Oral RfD ^{a,b}			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	1	1	0	0
	Non-Hispanic Black	1	1	0	0
	Mexican-American	1	1	0	0
	Other Hispanic	1	1	0	0
	Other, including multiple races	1	1	0	0
Child—subsistence	Non-Hispanic White	1	1	0	0
	Non-Hispanic Black	1	1	0	0
	Mexican-American	1	1	0	0
	Other Hispanic	1	1	0	0
	Other, including multiple races	1	1	0	0
Adult—recreational	Non-Hispanic White	1	1	0	0
	Non-Hispanic Black	1	1	0	0
	Mexican-American	1	1	0	0
	Other Hispanic	1	1	0	0
	Other, including multiple races	1	1	0	0
Adult—subsistence	Non-Hispanic White	1	1	0	0
	Non-Hispanic Black	1	1	0	0
	Mexican-American	1	1	0	0
	Other Hispanic	1	1	0	0
	Other, including multiple races	1	1	0	0

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); RfD (reference dose).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.

b—Benchmark value is based on total zinc.

Table A-19. Modeled IRWs with Lifetime Excess Cancer Risk for Inorganic Arsenic Exceeding One-in-a-Million by Race/Ethnicity Category Under Baseline and Regulatory Options

Age and Fishing Mode Cohort	Race/Ethnicity Category	Number of Modeled IRWs Exceeding LECR ^a			
		Baseline	Option A	Option B	Option C
Child—recreational	Non-Hispanic White	0	0	0	0
	Non-Hispanic Black	0	0	0	0
	Mexican-American	0	0	0	0
	Other Hispanic	0	0	0	0
	Other, including multiple races	0	0	0	0
Child—subsistence	Non-Hispanic White	2	2	1	0
	Non-Hispanic Black	3	2	1	1
	Mexican-American	3	2	1	1
	Other Hispanic	3	2	1	1
	Other, including multiple races	3	2	1	1
Adult—recreational	Non-Hispanic White	4	2	2	2
	Non-Hispanic Black	4	2	2	2
	Mexican-American	4	2	2	2
	Other Hispanic	4	2	2	2
	Other, including multiple races	4	2	2	2
Adult—subsistence	Non-Hispanic White	9	3	2	2
	Non-Hispanic Black	9	3	2	2
	Mexican-American	10	3	2	2
	Other Hispanic	10	3	2	2
	Other, including multiple races	11	4	3	2

Source: U.S. EPA, 2024i.

Abbreviations: IRW (immediate receiving water); LECR (lifetime excess cancer risk).

a—The IRW Model, which excludes the Great Lakes and estuaries, analyzes 114 total immediate receiving waters and loadings from 100 plants (some of which discharge to multiple receiving waters). Of these 114 immediate receiving waters, all 114 receive discharges of the evaluated wastestreams under baseline, 97 do under Option A, 50 do under Option B, and 17 do under Option C.