

#### **Draft Environmental Media Concentrations for 1,3-Butadiene**

## **Technical Support Document for the Draft Risk Evaluation**

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**CASRN 106-99-0** 



November 2024

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## 125 ABBREVIATIONS AND ACRONYMS

126	AMTIC	Ambient Monitoring Technology Information Center
127	AQS	Air Quality System
128	COU	Condition of use
129	CSATAM	Community-Scale Air Toxics Ambient Monitoring
130	DMR	Discharge Monitoring Report
131	EPA	Environmental Protection Agency
132	GC/MS	Gas chromatograph mass spectrometer
133	HAP	Hazardous Air Pollutant
134	Koa	Octanol-Air partition coefficient
135	Koc	Organic Carbon-Water partition coefficient
136	LDAQ	Louisiana Department of Environmental Quality
137	MDL	Method Detection Limit
138	MIT	Massachusetts Institute of Technology
139	NADP	National Atmospheric Deposition Program
140	NATTS	National Air Toxics Trends Sites
141	NEI	National Emissions Inventory
142	NOAA	National Oceanic and Atmospheric Administration
143	PWS	Public water system
144	SDWA	Safe Drinking Water Act
145	SCAQMD	South Coast Air Quality Management District
146	TCEQ	Texas Commission on Environmental Quality
147	TRI	Toxic Release Inventory
148	TSCA	Toxic Substances Control Act
149	UATMP	Urban Air Toxics Monitoring Program
150	UCMR3	Third Unregulated Contaminant Monitoring Rule
151	WQP	Water Quality Portal

## 152

## 153 ACKNOWLEDGEMENTS

154	The Assessment Team gratefully acknowledges the participation, review, and input from U.S.
155	Environmental Protection Agency (EPA or the Agency) Office of Pollution Prevention and Toxics
156	(OPPT) and Office of Chemical Safety and Pollution Prevention (OCSPP) senior managers and science
157	advisors.
158	
159	Docket
160	Supporting information can be found in the public docket, Docket ID: <u>EPA-HQ-OPPT-2024-0425</u> .
161	
162	Disclaimer
163	Reference herein to any specific commercial products, process, or service by trade name, trademark,
164	manufacturer, or otherwise does not constitute or imply its endorsement, recommendation, or favoring
165	by the United States Government.
166	
167	Authors: Aderonke Adegbule and Kiet Ly
168	
169	Contributors: Yousuf Ahmad, Sheila Healy and Ed Lo
170	
171	Technical Support: Mark Gibson, Hillary Hollinger, and Grace Kaupas.
172	
173	This draft technical support document was reviewed and cleared for release by OPPT and OCSPP

174 leadership.

## 175 SUMMARY

- 176 This draft technical document is in support of the Toxic Substances Control Act (TSCA) *Draft Risk*
- 177 *Evaluation for 1,3-Butadiene*. See the draft risk evaluation for a complete list of all the technical support 178 documents for 1,3-butadiene.
- 179

## 1,3-Butadiene – Environmental Media Concentration: Key Points

EPA evaluated the reasonably available information for various exposure media pathways (air, water, and land) to quantify the presence of 1,3-butadiene in the environment. This draft evaluation aids in determining the type of exposure assessment (quantitative or qualitative) that would be appropriate for each exposure pathway. The key points are summarized below:

- EPA assessed environmental concentrations of 1,3-butadiene for the air (ambient and indoor air), water (surface water, drinking water), and land pathways (groundwater) to inform the environmental exposure and general population exposure assessment.
  - $\circ$  For the air pathway, 1,3-butadiene releases are not expected to undergo air deposition or long-range transport based on a low log K<sub>OA</sub> of 1.5 and a short half-life in the atmosphere.
  - For the water pathway, biodegradation is expected to be fast in aerobic sediments and slower under anaerobic conditions. Based on a high Henry's Law constant (0.076 atm·m<sup>3</sup>/mol at 25 °C) and a high vapor pressure (1,900 mm Hg at 20 °C), volatilization is expected to be the most important removal process for 1,3-butadiene in aquatic environments. Due to its high volatility and low sorption potential to organic matter, 1,3-butadiene is not expected to significantly partition into sediments in water.
  - For the land pathway, 1,3-butadiene is not reported to be appreciably released to soil, and air to soil deposition is not expected to be significant because 1,3-butadiene is volatile and has low affinity for organic matter with a log K<sub>OC</sub> of 1.73 and log K<sub>OW</sub> of 1.99. Any release of 1,3-butadiene to soil (such as air deposition) is expected to volatilize rapidly. Therefore, soil and groundwater concentrations were not quantified but are discussed qualitatively.
- Based on the physical and chemical properties, as well as concentrations reported from databases and scientific literature, the air pathway is expected to be the main pathway contributing to general population exposures to 1,3-butadiene. Therefore, a quantitative assessment was conducted for the air pathway.
- Based on the low amounts of releases to land and water and high frequencies of non-detects reported in several federal databases—(Water Quality Portal (WQP), Discharge Monitoring Report (DMR), and Third Unregulated Contaminant Monitoring Rule (UCMR3—and scientific literature, the water and land pathways are not expected to contribute to general population exposures to 1,3-butadiene. Therefore, a qualitative assessment was conducted for the water and land pathways.

## 182 **1 INTRODUCTION**

- 183 1,3-Butadiene is a colorless gas that is produced during petroleum processing and is primarily used for
- 184 synthetic rubber production; however, small amounts can be found in plastics and fuel. It is released into
- 185 the environment by industrial operations involved with manufacturing, processing, formulation,
- disposal, and other practices. 1,3-Butadiene is also released into the ambient air by constant motor
- 187 vehicle emissions as well as from combustion and other activities related to fuel use and products. For 188 detailed information on 1,3-butadiene releases to the environment, see the *Draft Environmental Release*
- 189 Assessment for 1,3-Butadiene (U.S. EPA, 2024c).
  - 190

191 1,3-Butadiene is a high priority chemical undergoing the Toxic Substances Control Act (TSCA) risk

- 192 evaluation process for existing chemicals, following passage of the Frank R. Lautenberg Chemical
- 193 Safety for the 21st Century Act in 2016, and is subject to the TSCA regulations as described in Table 2-
- 194 1 of the *Draft Risk Evaluation for 1,3-Butadiene* (U.S. EPA, 2024g). This draft assessment/TSD serves
- the risk management needs of EPA's Office of Pollution Prevention and Toxics (OPPT) and is one of many documents, spreadsheets, and other files supporting the *Draft Risk Evaluation for 1,3-Butadiene*
- 197 (<u>U.S. EPA, 2024g</u>).

## 1981.1 Risk Evaluation Scope

199 The TSCA draft risk evaluation (RE) of 1,3-butadiene includes several human health, environmental,

200 fate, and exposure assessment modules, and a risk evaluation document. A diagram showing the

201 relationships between assessments is provided in Figure 1-1. This draft environmental media

202 concentrations assessment is one of the five technical support documents that are outlined in green.



1,3-butadiene systematic review protocol and data extraction files



203

## 204 Figure 1-1. Draft Risk Assessment Document Map Summary

The purpose of this draft TSD is to (1) evaluate concentrations of 1,3-butadiene in the following
environmental media pathways—air, water, and land—based on measurements reported in peerreviewed literature, gray literature, and available databases ; and (2) determine whether EPA needs to
conduct a quantitative or qualitative assessment accordingly for the *Draft Risk Evaluation for 1,3-*

## 210 Butadiene (U.S. EPA, 2024g). Comparisons to modeled data are provided for context.

## 1.2 Summary of the Chemistry, Fate, and Transport Assessment

212 1,3-Butadiene is a colorless gas with a mildly aromatic gasoline-like odor. 1,3-Butadiene is primarily 213 transformed in ambient air by indirect photolysis through reaction with ozone, nitrates, and hydroxy 214 radicals. Studies indicate 1,3-butadiene has a half-life range of 0.76 to 9 hours, with the longer half-lives 215 corresponding to periods without sunlight. Industrial releases and vehicular emissions are major sources 216 of 1,3-butadiene. It can also be formed in ambient air as a product of combustion of organic matter such as petroleum products, wood, and coal. 1,3-Butadiene is volatile and will evaporate from water and soil 217 218 and does not sorb to sediment. Therefore, air is expected to be the major pathway of concern for 1,3-219 butadiene in the environment, while water, sediment and soil are expected to comprise minor pathways. 220 Figure 1-2 below depicts the transportation and partitioning of 1,3-butadiene in the environment. For 221 more details, see the Draft Physical Chemistry and Fate Assessment for 1,3-Butadiene (U.S. EPA, 222 2024f).





## Figure 1-2. Transport, Partitioning, and Degradation of 1,3-Butadiene in the Environment

- 226 The diagram depicts the distribution (grey arrows), transport and partitioning (black arrows) as well as the
- transformation and degradation (white arrows) of 1,3-butadiene in the environment. The width of the arrow is a
- 228 qualitative indication of the likelihood that the indicated partitioning will occur or the rate at which the indicated
- degradation will occur (*i.e.*, wider arrows indicate more likely partitioning or more rapid degradation).

## 231 2 APPROACH AND METHODOLOGY

- EPA conducted a systematic review to identify peer-reviewed, gray literature and database sources of 232 233 1,3-butadiene measured concentrations in various media pathways (air, water, and land) to characterize 234 environmental media concentrations. Modeled data is acknowledged when available. Environmental 235 media concentration data from studies and databases identified through systematic review were 236 evaluated and assigned an overall quality determination according to the 2021 Draft Systematic Review 237 Protocol Supporting TSCA Risk Evaluations for Chemical Substances, Version 1.0: A Generic TSCA 238 Systematic Review Protocol with Chemical-Specific Methodologies (also called the "2021 Draft Systematic Review Protocol Supporting TSCA Risk Evaluations" (U.S. EPA, 2021b). Studies assigned 239
- to medium and high quality determinations were extracted and integrated as part of the draft risk
- evaluation. For more details on the systematic review process as applied to 1,3-butadiene, see Section
- 242 2.3 of the *Draft Risk Evaluation of 1,3-Butadiene* (U.S. EPA, 2024g) and the *Draft Systematic Review*
- 243 Protocol for 1,3-Butadiene (U.S. EPA, 2024h). Several references reported environmental media
- concentrations outside the United States and are included for global context in Figure 3-1 through Figure
- 245 3-4 below and Table\_Apx A-1 in Appendix A. All studies identified, U.S. and non-U.S., support the
- conclusion that air is the major exposure pathway for 1,3-butadiene.

## 248 **3 AIR PATHWAY**

EPA searched peer-reviewed and gray literature published through 2019, and databases through 2021, to 249 250 obtain concentrations of 1,3-butadiene in ambient air. Sections 3.1 through 3.5 describe the aggregated results of measured and modeled concentrations of 1,3-butadiene in air. While the systematic review 251 252 included studies conducted outside of the United States, discussion herein is focused on U.S. studies as 253 these studies are most informative for this draft assessment. Studies conducted outside of the United States are presented for context but are not discussed further. Measured data from the EPA Ambient 254 255 Monitoring Technology Information Center (AMTIC) (U.S. EPA, 2022a) is presented in Section 3.1.2. 256 Table\_Apx B-1 summarizes the 1,3-butadiene concentrations extracted from systematic review for the air pathway (ambient air, indoor air, landfill gas and personal exposure monitoring). For more details on 257 the systematic review for 1,3-butadiene, see the Draft Systematic Review Protocol for 1,3-Butadiene 258 259 (U.S. EPA, 2024h).

## **3.1 Measured Concentrations in Ambient Air**

## 261 **3.1.1 Peer-Reviewed Literature**

## 262

## 3.1.1.1 Measured Concentrations

263 Measured concentrations of 1,3-butadiene in ambient air were extracted from five U.S. studies published 264 between 1999 and 2015. These measured concentrations were classified as representing general population exposures (*i.e.*, ambient measurements taken in areas near residential populations with no 265 266 known facility sources nearby) or near facility (*i.e.*, measurements collected in areas with industrial 267 and/or commercial activities), denoted by blue and red, respectively, in Figure 3-1. Within the U.S., Logue, et al. (2010) collected measurements in Pittsburgh, Pennsylvania, and reported measurements 268 from both near facility (0.01 to 0.33  $\mu$ g/m<sup>3</sup>) and general population (0.04 to 0.35  $\mu$ g/m<sup>3</sup>; downtown 269 270 Pittsburgh) sites. For near facility measurements, 1,3-butadiene concentrations were attributed to emissions from chemical plants and metallurgic coke plants, while 1,3-butadiene concentrations at the 271 272 downtown, general population sites were attributed to mobile emissions from vehicular traffic. 273 Measurements collected at an exurban background site, outside of Pittsburgh, had a range of 0.0 to 0.15  $\mu g/m^3$ . Bereznicki, et. al (2012) reported a residential general population concentration range of non-274 275 detect (ND) to  $1.91\mu g/m^3$  in Detroit, Michigan. Yu et al. (2014) reported a range of 0.01 to  $1.35 \mu g/m^3$ 276 near industrial areas in Paterson, New Jersey.



278

## 279 Figure 3-1. Measured Concentrations of 1,3-Butadiene (µg/m<sup>3</sup>) in Ambient Air from 1998 to 2017

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285

## 3.1.1.1 Modeled Concentrations

Modeled concentrations of 1,3-butadiene in ambient air were also captured during systematic review from five studies (four in the United States and one in Japan) with concentration results of a similar range as measured concentrations (<u>Yu and Stuart, 2016; Loh et al., 2007; Luecken et al., 2006; Suzuki et</u> al., 2004; <u>Radian Engineering, 1997</u>).

## 3.1.2 Monitoring Database

To complement measurements available in peer-reviewed literature, measured ambient air 286 concentrations of 1,3-butadiene were obtained from the EPA's AMTIC<sup>1</sup> Ambient Monitoring Archive 287 (AMA) for Hazardous Air Pollutants (HAPs) database (U.S. EPA, 2022a). The AMTIC-AMA houses 288 data from over 5,000 ambient air monitoring sites across the United States collected from 1990 to 2021. 289 290 Contributing data sources include the EPA's Air Quality System (AQS), the Texas Commission on 291 Environmental Quality (TCEQ), the South Coast Air Quality Management District (SCAQMD), the National Atmospheric Deposition Program (NADP), the National Oceanic and Atmospheric 292 293 Administration (NOAA), the Massachusetts Institute of Technology (MIT), the Louisiana Department of 294 Environmental Quality (LDEQ), other state, local, tribal, and federal monitoring agencies, and 295 academic, community, and short-term studies.

296

To be consistent with the TRI 2016 to 2021 reporting years, EPA evaluated monitoring data for samples collected from January 2016 through December 2021. The 1,3-butadiene AMTIC

- samples collected from January 2016 through December 2021. The 1,3-butadiene AMTIC
   monitoring data included over 55,000 24-hour sampling entries from 12 monitoring programs
- 300 covering 34 states and 331 census tracts. Samples were collected using pressurized canisters over

<sup>&</sup>lt;sup>1</sup> See <u>https://www.epa.gov/amtic/air-toxics-ambient-monitoring</u> for more information.

- 301 **24-hour durations and analyzed using gas chromatograph mass spectrometer (GC/MS). The**
- 302 resulting 1,3-butadiene concentrations were converted to µg/m<sup>3</sup>. Monitored concentrations from
- 303 the AMTIC archive ranged from 0.0 to 122.8 μg/m<sup>3</sup>. The maximum concentration was reported
- 304 from the State of Texas, which reported approximately 9.3 percent of all 1,3-butadiene monitoring
- data. The method detection limit (MDL) ranged from 0.002 to 1.106  $\mu$ g/m<sup>3</sup> with 41.1 to 52.8
- 306 percent of samples collected between 2016 through 2021 reporting 1,3-butadiene concentrations
   307 below the MDL. See
- 308 Table\_Apx C-1 for a summary of the AMTIC monitoring data, which includes other sampling durations
- beyond 24 hours. For more information on 1,3-butadiene ambient air monitoring data, see the
- 310 supplemental file Draft Ambient Monitoring Technology Information Center (AMTIC) Monitoring Data
- 311 2016 to 2021 for 1,3-Butadiene (U.S. EPA, 2024b). Data for this supplemental file was downloaded
- 312 from the AMTIC Ambient Monitoring Archive for HAPs Database (U.S. EPA, 2022a).

## 313 **3.2 Measured Concentrations in Indoor Air**

## 314 3.2.1 Peer-Reviewed Literature

## 3.2.1.1 Measured Concentrations

Measured concentrations of 1,3-butadiene in indoor air were extracted from four U.S. studies published 316 between the years 1999 and 2007 (Figure 3-2). Loh et. al (2006) reported a geometric mean 317 concentration of 1.05  $\mu$ g/m<sup>3</sup> and a maximum concentration of 35.5  $\mu$ g/m<sup>3</sup> from measurements collected 318 319 in a public dining space in Boston, Massachusetts, and a geometric mean concentration of  $0.21 \,\mu g/m^3$ and a maximum concentration of 2.20  $\mu$ g/m<sup>3</sup> in retail stores. The study attributed 1,3-butadiene in the 320 dining spaces to cooking and tobacco smoke. Sax, et. al (2006) reported a mean concentration of 1.01 321  $\mu g/m^3$  and a maximum concentration of 9.02  $\mu g/m^3$  in homes throughout New York City, New York, 322 and a mean concentration of 0.41  $\mu$ g/m<sup>3</sup> and maximum concentration of 1.47  $\mu$ g/m<sup>3</sup> in homes in Los 323 324 Angeles, California.

325



326

## 327 Figure 3-2. Concentrations of 1,3-Butadiene (μg/m<sup>3</sup>) in Indoor Air from 1998 to 2011

328

## 3.2.1.2 Modeled Concentrations

Modeled concentrations of 1,3-butadiene in indoor air were also captured during systematic review from one study in the United States with results that were within range of measured concentrations (Loh et al., 2007).

## **332 3.3 Measured Concentrations in Landfill Gas**

333

## 3.3.1 Peer-Reviewed Literature

Measured concentrations of 1,3-butadiene in landfill gas were extracted from 1 U.S. study with 12 samples collected in 1998 (Figure 3-3). The average concentration of 3.98 ppm (8,800  $\mu$ g/m<sup>3</sup>) was reported from measurements collected at a municipal solid-waste landfill in New York City, New York (Eklund et al., 1998). It is noted in this study that landfill gas is collected by an on-site gas collection system and treated for impurities before being sold to utility services. Therefore, exposure to the general population from landfill gas is not expected.

340

	US Other	-	Near Facility
	1487387 - Eklund et al., 1998 - US		
	0.1		10
341		Concentra	tion (ppm)

## 342 Figure 3-3. Concentrations of 1,3-Butadiene (ppm) in Landfill Gas from 1998

## 343 **3.4 Personal Exposure Monitoring**

## 344 **3.4.1 Peer-Reviewed Literature**

345 In addition to these studies, measured concentrations of 1,3-butadiene in personal exposure monitoring

samples; that is, study participants wearing personal air samplers, were available from four U.S. studies
conducted from 1996 and 2015 (Figure 3-4) (<u>Shin et al., 2015</u>; <u>Dodson et al., 2007</u>; <u>Sax et al., 2006</u>;
<u>Heavner et al., 1996</u>). However, due to the relatively small sample sizes across these studies, which
ranged from 46 to 104 participants, and varying exposure factors; for example, smoking, behavioral, and
activity patterns, it is difficult to extrapolate results from individual exposures to expected populationbased exposures.

351 352



353

Figure 3-4. Personal Exposure Measurements of 1,3-Butadiene (µg/m<sup>3</sup>) from 1996 to 2015

## **355 3.5 Evidence Integration**

Measurements from the AMTIC ambient air monitoring sites database (U.S. EPA, 2022a) and 356 concentrations reported in the literature provide evidence that exposure to 1,3-butadiene from the air 357 358 pathway is expected in ambient air. Therefore, EPA conducted a quantitative assessment of exposure to 1,3-butadiene in ambient air for the general population. The Agency modeled ambient air concentrations 359 based on releases from the TRI 2016 to 2021 database to assess general population exposure to ambient 360 air from facility releases of 1,3-butadiene. For more details on environmental releases and general 361 362 population exposure, see the Draft Environmental Release for 1,3-Butadiene (U.S. EPA, 2024c) and the 363 Draft General Population Exposure Assessment for 1,3-Butadiene (U.S. EPA, 2024d), respectively. 364

## 365 4 WATER PATHWAY

The DMR (U.S. EPA, 2024a) recorded no releases of 1,3-butadiene from wastewater treatment facilities to surface water bodies. The TRI (U.S. EPA, 2022b) only recorded very small releases with the maximum being about 400 kg into Mississippi river in 2020 (see *Draft Environmental Release and Occupational Exposure Assessment for 1,3-Butadiene* (U.S. EPA, 2024e)). With average flow rate in the Mississippi of  $1.15 \times 10^{12}$  L/day, the estimated surface water concentration would be about 0.96 ng/L, not accounting for volatilization that would further drive down the concentration.

372

1,3-Butadiene has a low water solubility of 735 mg/L. In addition, volatilization of 1,3-butadiene from

374 water surfaces is expected to occur rapidly due to a Henry's Law constant of 0.076 atm.m<sup>3</sup>/mol. EPI

375 Suite<sup>TM</sup> estimates volatilization rates of hours to a few days for a surface water body of 1 m depth, wind

376 velocities ranging from 0.5 to 5 m/s, and water current velocity ranging from 0.0 to 1 m/s (see *Draft* 

377 *Chemistry, Fate and Transport Assessment for 1,3-Butadiene* (U.S. EPA, 2024f)).

## **4.1 Measured Concentrations in Surface Water**

The WQP (NWQMC, 2022) is a publicly available resource that integrates water quality data from the 379 USGS National Water Information System (NWIS) (USGS, 2013) and the EPA Water Quality 380 381 Exchange (WQX) Data Warehouse (U.S. EPA, 2019). The NWIS database contains current and historical water data from more than 1.5 million sites across the nation. The WQX is the EPA's 382 383 repository of water quality monitoring data collected by water resource management groups across the 384 nation. The complete set of 1,3-butadiene monitoring results stored in the WQP (NWQMC, 2022) was 385 retrieved in January 2024 with only the chemical name applied as a filter. The raw dataset included 386 10,206 samples from across the United States collected between 2011 and 2023. A higher number of samples were from California, New York, Texas, Georgia, North Carolina, and Florida (where >39% of 387 388 the U.S. population resides) compared to other states. The WQP 1,3-butadiene dataset did not contain samples from Alaska, Delaware, Rhode Island, Hawaii, or Vermont (<2% of the U.S. population live in 389 390 these states).

391

392 While the 1,3-butadiene WQP data set exceeded 10,000 samples, most of the samples were groundwater 393 samples with only 231 samples available for surface water. The surface water samples were sourced 394 from 23 states for the period of 2012 to 2020. Greater than 80 percent of the surface water samples were sourced from North Carolina (81), Georgia (47), Illinois (29), South Carolina (25), and Virginia (13). 395 396 Most of samples were collected between 2012 to 2015 with about 75 percent of the samples being 397 collected in 2014 alone. For more information and details, see the supplemental file Draft Water Quality 398 Portal (WQP) Monitoring Data 2011 to 2023 for 1,3-Butadiene (U.S. EPA, 2024i). Without exception, 399 all 231 surface water samples reported 1,3-butadiene concentrations below the MDL of 0.08 µg/L.

## 400 **4.2 Evidence Integration for Surface Water**

Based on the low reported releases to surface water (U.S. EPA, 2024e), physical and chemical properties
(U.S. EPA, 2024f), and monitoring data reporting a 0 percent detection frequency of 1,3-butadiene
(Section 4.1), EPA decided not to conduct a quantitative assessment of risk for surface water.

404

405 After systematic review and a review of the WQP database (<u>NWQMC, 2022</u>), EPA did not find reported

406 measured sediment concentrations of 1,3-butadiene in the United States. As described in Sections 4.1 407 and 4.2, 1,3-butadiene is not expected to be found in surface water; therefore, there is no route for it to

- 408 get into sediments. Furthermore, 1,3-butadiene has a low estimated organic carbon:water partition
- 409 coefficient (Koc) value of 54 (U.S. EPA, 2012), indicating that if the chemical were to enter surface
- 410 water, it would be unlikely to partition to sediment. As a result, EPA has decided not to conduct a

411 quantitative assessment of risk for sediments.

## 412 **4.3 Measured Concentrations in Drinking Water**

413 Public water systems (PWSs) are regulated under the Safe Drinking Water Act (SDWA)<sup>2</sup> to enforce 414 common standards for drinking water across the country. To assess concentrations of 1,3-butadiene in 415 water known to be distributed as drinking water, monitoring data collected by PWSs were evaluated. Concentrations of 1,3-butadiene found in finished (*i.e.*, treated) drinking water were extracted from the 416 EPA's publicly available UCMR3<sup>3</sup> dataset, which includes samples collected between 2013 to 2015 417 418 (U.S. EPA, 2017). The data covers all PWSs serving more than 10,000 people and 800 representative 419 PWSs serving 10,000 or fewer people within all U.S. states and territories. The data shows that all but 2 420 of 36,839 tested drinking water samples (>99.9%) did not detect 1.3-butadiene (reporting limit [RL] of 421 0.1  $\mu$ g/L). The two samples reported 1,3-butadiene at concentrations of 0.32 and 0.54  $\mu$ g/L in Florida 422 (PWSID FL3480962) and Puerto Rico (PWSID PR0002591), respectively. No additional information about the two detections in drinking water was found. EPA was not able to evaluate drinking water 423 424 sourced outside the PWSs but expects no significant differences from the public drinking water based on

425 the physical and chemical properties of 1,3-butadiene.

## 426 **4.4 Evidence Integration for Drinking Water**

427 Based on the physical and chemical properties of 1,3-butadiene—that is, its low water solubility and

high tendency to volatilize from water as well as the monitored data showing that 1,3-butadiene is not

429 detected in drinking water from PWSs—EPA has decided not to conduct a quantitative assessment of

430 risk for drinking water.

<sup>&</sup>lt;sup>2</sup> See <u>https://www.epa.gov/sdwa</u> for more information.

<sup>&</sup>lt;sup>3</sup> See <u>https://www.epa.gov/dwucmr/third-unregulated-contaminant-monitoring-rule</u> for more information.

## 431 5 LAND PATHWAY

- Land contamination from 1,3-butadiene may be expected to occur from releases to land, which mainly
  includes discharge of wastewater into underground injection wells and disposal of waste into landfills
  (see *Draft Environmental Release and Occupational Exposure Assessment for 1,3-Butadiene* (U.S. EPA,
  2024e)).
  According to reports in the 2016 CDR, the use of plastic and rubber products, including synthetic
- According to reports in the 2016 CDR, the use of plastic and rubber products, including synthetic
   rubbers, were identified as consumer conditions of use for 1,3-butadiene. EPA has determined, however
   that 1,3-butadiene, a monomer used in polymer-derived consumer products such as synthetic rubbers, is
   stable in these products and not expected to degrade and expose the consumer to the 1,3-butadiene
   monomer (U.S. EPA, 2024g). Furthermore, rubber and plastic products in landfills would not be
   expected to form leachates containing any significant amount of 1,3-butadiene.
- 443
- 444 Sludge from water treatment systems in polymer production facilities are not expected to contain 445 significant amounts of 1,3-butadiene due to its volatility and low affinity for organic matter. Any release 446 of 1,3-butadiene to soil (such as air deposition) would volatilize rapidly and is not expected to remain in 447 soil for any significant amount of time.
- 448
- 449 Releases to land make up less than 1 percent of releases of 1,3-butadiene to the environment (U.S. EPA, 450 2024e). Most of the releases of 1,3-butadiene to land are to class I underground injection wells. 451 Oversight of these wells requires that the wells are designed and constructed to prevent the movement of 452 injected waste streams into drinking water systems. Wells typically consist of three or more concentric 453 layers of pipe including surface casing, long string casing, and injection tubing. In addition, wells must 454 be sited at locations with geologies that mitigate any movement of contaminants outside of a confined 455 layer if there were a well failure. Extensive pre-siting geological tests confirm that the injection zone is 456 of sufficient lateral extent and thickness and is sufficiently porous so that fluids injected through the well 457 can enter the rock formation without extensive buildup of pressure or possible displacement of injected 458 fluids outside of the intended zone. Hence, it is unlikely that this disposal pathway could contaminate a drinking water source.
- 459 460
- Releases to land comprise less than 1 percent of releases of 1,3-butadiene to the environment (U.S. EPA, 461 462 2024e). Most of the releases of 1,3-butadiene to land are to class I underground injection wells. 463 Oversight of these wells requires that the wells are designed and constructed to prevent the movement of 464 injected waste streams into drinking water systems. Wells typically consist of three or more concentric 465 layers of pipe, including surface casing, long string casing, and injection tubing. In addition, wells must be sited at locations with geologies that mitigate any movement of contaminants outside of a confined 466 467 layer if there were a well failure. Extensive pre-siting geological tests confirm that the injection zone is 468 of sufficient lateral extent and thickness and is sufficiently porous so that fluids injected through the well 469 can enter the rock formation without extensive buildup of pressure or possible displacement of injected 470 fluids outside of the intended zone. Thus, it is unlikely that this disposal pathway could contaminate a 471 drinking water source.

## 472 **5.1 Measured Concentrations in Groundwater**

The complete set of 1,3-butadiene monitoring results stored in the WQP (<u>NWQMC</u>, 2022) were retrieved in January 2024 with only the chemical name applied as a filter. The raw dataset included 10206 samples from across the United States collected between 2011 and 2023. A higher number of samples were from California, New York, Texas, Georgia, North Carolina, and Florida (where >39% of the U.S. population resides) compared to other states and it did not contain samples from Alaska,

Delaware, Rhode Island, Hawaii, and Vermont (<2% of the U.S. population live in these states). When</li>
samples of other types of water were excluded, 9,378 groundwater samples from 46 states for the period
of 2011 to 2023 remained in the dataset. Greater than 50 percent of the samples were from Arizona with
another 18 percent from California. None of the groundwater samples reported concentrations of 1,3butadiene above the MDL. The MDL ranged from 0.81 to 40 ug/lFor more details on the monitoring
data, see the supplemental file *Draft Water Quality Portal (WQP) Monitoring Data 2011 to 2023 for 1,3-Butadiene* (U.S. EPA, 2024i).

485

486 The complete set of 1,3-butadiene monitoring results stored in the WQP (<u>NWQMC, 2022</u>) were

retrieved in January 2024 with only the chemical name applied as a filter. The raw dataset included
10,206 samples from across the United States collected between 2011 and 2023. A higher number of
samples were from California, New York, Texas, Georgia, North Carolina, and Florida compared to

490 other less populated states and it did not contain samples from Alaska, Delaware, Rhode Island, Hawaii,

and Vermont. When samples of other types of water were excluded, 9,378 groundwater samples from 46
 states for the period of 2011 to 2023 remained in the dataset. Greater than 50 percent of the samples

- 493 were from Arizona with another 18 percent from California. None of the groundwater samples reported
- 494 concentrations of 1,3-butadiene above the MDL. The MDL ranged from 0.81 to 40  $\mu$ g/L For additional
- details on the monitoring data, see the supplemental file *Draft Water Quality Portal (WQP) Monitoring Data 2011 to 2023 for 1,3-Butadiene (U.S. EPA, 2024i).*

## 497 **5.2 Evidence Integration**

Based on the low volume of releases to land, the low risk of failure of the predominant release scenario,
 the physical and chemical properties of 1,3-butadiene, as well as monitoring data indicating no

detections of 1,3-butadiene, EPA did not perform a quantitative analysis for the land pathway because
 exposure to the general population is not expected to occur.

#### WEIGHT OF SCIENTIFIC EVIDENCE 6 502

#### 6.1 Strengths, Limitations, Assumptions, and Key Sources of Uncertainty 503 for Measured Concentrations in Ambient Air 504

505 AMTIC data has been previously reviewed and verified by the AMTICs Ambient Air Monitoring 506 Group, which has taken various quality assurance steps to ensure data quality and has been certified in 507 accordance with 40 CFR Part 58.15. Due to strictly regulated monitoring requirements, EPA has high 508 confidence in the AMTIC ambient air data set (U.S. EPA, 2022a), which also received a high-quality 509 rating from EPA's systematic review process. (U.S. EPA, 2021a).

510

511 A primary limitation of the AMTIC data is that the data have not been annualized and therefore

512 represent a diverse collection of sampling durations (none of which are annual averages) that are not

513 directly comparable to modeled data. Additionally, because monitored data represents a total aggregate

514 concentration from all sources of 1.3-butadiene contributing to ambient air concentrations, the AMTIC 515 data cannot be used to characterize exposures exclusively from TSCA COUs. See the Draft General

516 Population Exposures for 1,3-Butadiene (U.S. EPA, 2024d) for more details comparing modeled and 517 measured data.

518

519 The 1,3-butadiene concentrations from systematic review were extracted from 31 studies from 6 countries. All media types included studies and data collected in the United States with ratings of 520 medium or high quality, which are representative of exposure for the general population in the United 521 522 States. Studies from non-U.S. countries may not be representative of exposure for the general population 523 in the United States; one study was given a rating of low quality from Spain. See Appendix A for study 524 ratings for all peer-reviewed literature from systematic review. Notably, measured data from systematic 525 review vary temporally, air concentrations are especially subject to season variation, and vary 526 geospatially. Methodology for sample collection and analysis are specific to each peer-reviewed literature and vary with instrumentation and analysis. Concentrations of 1,3-butadiene across all media 527 528 are attributed to many different sources, including but not limited to facility emissions, traffic emissions, 529 heating and cooking combustion activities, and environmental tobacco smoke, and are not specific to 530 only TSCA conditions of use.

531

#### 6.2 Strengths, Limitations, Assumptions, and Key Sources of Uncertainty for Measured Concentrations in Water 532

533 The 1,3-butadiene data from the WQP has a strong bias of samples collected from California, New 534 York, Texas, Georgia, North Carolina, and Florida (comprising >39% of the U.S. population) relative to 535 other states and was missing data from Alaska, Delaware, Rhode Island, Hawaii and Vermont 536 (comprising <2% of the U.S. population). The states with a higher number of data points are states where a higher percentage of the U.S. population resides. In addition, states with a concentration of 537 538 facilities releasing 1,3-butadiene are included in the monitoring database. Due to the presence of 1,3-539 butadiene releasing facilities, these states would be expected to have the largest 1,3-butadiene releases. 540 Because data reflects that 1,3-butadiene is typically not detected above the detection limit in water, EPA 541 has robust confidence that in areas with less releases, 1,3-butadiene will not be in the water. In addition, 542 based on the physical and chemical properties of 1,3-butadiene and low release quantities to water and 543 land, EPA has confidence that the WQP data is representative of the entire United States. Again, WQP is 544 not specific to only TSCA conditions of use.

545

546 The 1,3-butadiene data from the UCMR3 covered all PWSs serving more than 10,000 people and 800

- 547 representative PWSs serving 10,000 or fewer people. The PWSs were monitored during a 12-month
- 548 period from January 2013 through December 2015. This data may have a bias towards states with
- 549 greater populations (*e.g.*, California) due to a higher number of PWSs serving 10,000 or more people,
- 550 compared to the representative 800 PWSs serving 10,000 or fewer people, including Tribal nations. In
- addition, the UCMR3 only required samples to be collected during a 12-month period between January
- 552 2013 through December 2015, which is a timespan of approximately 3 years, so it does not consider
- temporal variability beyond the 12-month sampling periods. Lastly, the UCMR3 data is not specific to
- only TSCA conditions of use. Nevertheless, EPA has robust confidence in the representativeness of the
- 555 UCMR3 data.

## 556 **7 CONCLUSIONS**

Based on the (1) physical and chemical properties of 1,3-butadiene (*i.e.*, high volatility, low solubility, 557 and low sorption tendencies) (U.S. EPA, 2024f); (2) low release volume to land and water (U.S. EPA, 558 2024e); and (3) minimal detection of 1.3-butadiene in U.S. surface and groundwaters, EPA has robust 559 confidence that air is the major pathway of exposure for 1,3-butadiene and that contributions to exposure 560 561 from the land and water pathways will be infrequent and at low levels. As a result, air is the only pathway that will be assessed quantitatively. 562 563 564 For regions where monitoring data are available, EPA has robust confidence in the overall characterization of environmental media concentrations for 1,3-butadiene as it relies upon standard 565

reporting databases that are reviewed with quality control and assurance protocols, such as AMTIC, WQP, and UCMR, and extracted data from peer-reviewed literature that received medium- to high-

568 quality ratings from EPA's systematic review process.

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# Appendix A SUPPLEMENTAL INFORMATION OF PEER REVIEWED LITERATURE AND STUDY RATINGS FOR MEASURED CONCENTRATIONS OF 1,3 BUTADIENE

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## Table\_Apx A-1. Summary of Peer-Reviewed Literature that Measured 1,3-Butadiene (µg/m3) Levels in Ambient Air

Citation	Country	Location Type	Sampling Year(s)	Sample Size (Frequency of Detection)	Detection Limit (µg/m <sup>3</sup> )	Overall Quality Level		
		<u>+</u>	Not specified	1		•		
<u>Zhou et al.</u> (2011)	CN	General Population	2008	8 (N/R)	N/R	Medium		
			Vapor/gas					
<u>U.S. EPA</u> (2015)	US	General Population	2010–2012	121 (N/R)	N/R	High		
<u>Logue et al.</u> (2010)	US	General Population	2006–2008	244 (N/R)	N/R	High		
Logue et al. (2010)	US	Near Facility	2006–2008	244 (N/R)	N/R	High		
Dodson et al. (2007)	US	General Population	2007	89 (0.38)	N/R	High		
Bereznicki et al. (2012)	US	General Population	2004–2007	992 (0.79)	0.174	Medium		
<u>Yu et al.</u> (2014)	US	Near Facility	2005–2006	209 (0.84)	0.01	High		
<u>Huang et al.</u> (2019)	CN	General Population	2016–2017	37 (N/R)	N/R	High		
Domingo et al. (2015)	ES	Near Facility	2014	6 (0)	2.78	Low		
<u>Baek et al.</u> (2020)	KR	General Population	2009–2010	384 (0.27)	N/R	Medium		
Baek et al. (2020)	KR	Near Facility	2009–2010	384 (0.27)	N/R	Medium		
<u>Yazar et al.</u> (2011)	SE	General Population	2009	9 (1)	0.01	Medium		
Harrison et al. (2009)	GB	General Population	2005–2007	128 (N/R)	N/R	High		
	No fraction							
<u>Gordon et al.</u> (1999)	US	General Population	1998	14 (0)	0.38	Medium		
N/R = not reported; CN = China; GB = Great Britain; SE = Spain; KR = South Korea; US = United States								

## Table\_Apx A-2. Summary of Peer-Reviewed Literature that Measured 1,3-Butadiene (µg/m3) Levels in Indoor Air 723

724

Citation	Country	Location Type	Sampling Year	Sample Size (Frequency of Detection)	Detection Limit (µg/m <sup>3</sup> )	Overall Quality Level		
Not specified								
Zhou et al. (2011)	CN	Public Space	2008	6 (N/R)	N/R	Medium		
Zhou et al. (2011)	CN	Residential	2008	10 (0.90)	N/R	Medium		
Zhou et al. (2011)	CN	Vehicle	2008	6 (N/R)	N/R	Medium		
			Vapor/gas					
Dodson et al. (2007)	US	Public Space	2007	178 (0.28)	N/R	High		
<u>Dodson et al.</u> (2007)	US	Residential	2007	89 (0.31)	N/R	High		
Loh et al. (2006)	US	Public Space	2004	71 (0.89)	0.17	Medium		
Sax et al. (2006)	US	Residential	1999–2000	81 (0.67)	1	High		
<u>Gordon et al.</u> (1999)	US	Residential	1998	24 (0.04)	0.38	Medium		
<u>Huang et al.</u> (2019)	CN	Residential	2016–2017	88 (N/R)	N/R	High		
Delgado-Saborit et al. (2011)	GB	Public Space	2005–2007	40 (N/R)	0.0051	High		
Delgado-Saborit et al. (2011)	GB	Residential	2005–2007	155 (N/R)	0.0051	High		
Harrison et al. (2009)	GB	Public Space	2005–2007	77 (N/R)	N/R	High		
Harrison et al. (2009)	GB	Residential	2005–2007	152 (N/R)	N/R	High		
Harrison et al. (2009)	GB	Vehicle	2005–2007	43 (N/R)	N/R	High		
N/R, Not reported; CN = China; GB = Great Britain; US = United States								

#### 727 Table\_Apx A-3. Summary of Peer-Reviewed Literature that Measured 1,3-Butadiene (ppm) Levels in Landfill Gas

## 728

Citation	Country	Exposure Scenario	Location Type	Sampling Year	Sample Size (Frequency of Detection)	Detection Limit (ppm)	Overall Quality Level		
<u>Eklund et al.</u> (1998)	US	Landfill Gas	Near Facility	1998	12 (N/R)	N/R	Medium		
N/R = Not reported; US = United States									

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## Table\_Apx A-4. Summary of Peer-Reviewed Literature that Measured 1,3-Butadiene (µg/m<sup>3</sup>) Levels in Personal Inhalation 732

Citation	Country	Location Type	Sampling Year	Sample Size (Frequency of Detection)	Detection Limit (µg/m <sup>3</sup> )	Overall Quality Level				
Combined vapor/gas and particulate										
<u>Heavner et al.</u> (1996)	US	General Population	1992	168 (N/R)	N/R	Medium				
Not specified										
Zhou et al. (2011)	CN	General Population	2008	20 (N/R)	N/R	Medium				
			Vapor/gas	•	•					
<u>Dodson et al.</u> (2007)	US	General Population	2007	89 (0.25)	N/R	High				
Shin et al. (2015)	US	Near Facility	2005-2007	239 (N/R)	N/R	Medium				
<u>Sax et al. (2006)</u>	US	General Population	1999–2000	81 (0.70)	1	High				
<u>Yazar et al. (2011)</u>	SE	General Population	2009	39 (1)	0.01	Medium				
<u>Hagenbjörk-</u> <u>Gustafsson et al.</u> (2014)	SE	General Population	2001–2008	275 (N/R)	N/R	Medium				
Saborit et al. (2009)	GB	General Population	2005–2007	500 (N/R)	N/R	Medium				
<u>Delgado-Saborit et</u> <u>al. (2011)</u>	GB	General Population	2005–2007	500 (N/R)	0.0051	High				
Harrison et al. (2009)	GB	General Population	2005–2007	500 (N/R)	N/R	High				
N/R = not reported; CN = China; GB = Great Britain; SE = Spain; US = United States										

# Appendix B SUPPLEMENTAL INFORMATION OF PEER-REVIEWED LITERATURE FOR MEASURED CONCENTRATIONS OF 1,3-BUTADIENE

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## Table\_Apx B-1. Summary Statistics of 1,3-Butadiene Concentrations Extracted from Systematic Review

Matrices	Location	Number of Studies with U.S. Data (N)	Number of Studies with Non-U.S. Data (N)	Total Number of Studies (N)	Unit	Fraction	Average of Arithmeti c Mean Estimates for U.S. Data	Average of 90th Percentile Estimates for U.S. Data	Average of Arithmeti c Mean Estimates for non- U.S. Data	Average of 90th Percentile Estimates for Non- US Data	Average of arithmetic Mean Estimates for All Data	Average of 90th Percentile Estimates for All Data
Ambient Air	General Population	4 (1,183)	3 (54)	7 (1237)	$\mu g/m^3$	Any	4.3E-01	8.0E-01	3.2E00	4.4E00	4.6E-01	7.8E-01
	Near Facility	2 (420)	0	2 (420)	$\mu g/m^3$	Any	1.51E-01	3.21E-01	_	_	1.5E-01	3.2E-01
	Public Spaces	2 (113)	2 (46)	4 (159)	$\mu g/m^3$	Any	1.7E00	3.1E00	3.9E00	6.9E00	1.5E00	2.3E00
Indoor Air	Residential	3 (83)	3 (253)	6 (335)	$\mu g/m^3$	Gas/vapor	8.6E-01	1.9E00	2.5E00	4.4E00	1.5E00	2.4E00
	Vehicles	—	1 (6)	1 (6)	$\mu g/m^3$	Any	—	_	6.2E-01	1.1E00	6.2E-01	1.1E00
Landfill Gas	Near Facility	_	_	_	ppm	Any	_	_	_	_	_	_
Personal Inhalation	General Population	4 (486)	5 (1,334)	9 (1,820)	$\mu g/m^3$	Any	1.5E00	3.5E00	1.5E00	3.5E00	6.9E-01	1.6E00

# Appendix CSUPPLEMENTAL INFORMATION OF AIR MONITORING DATA FOR<br/>MEASURED CONCENTRATIONS OF 1,3-BUTADIENE

Ambient Monitoring Technology Information Center Hazardous Air Pollutants										
Sample Duration	Statistic (µg/m <sup>3</sup> )	Year								
		2016	2017	2018	2019	2020	2021			
	Maximum	1.45E01	1.74E01	1.90E01	4.56E01	5.29E01	1.23E02			
	Mean	1.02E-01	7.99E-02	7.76E-02	6.84E-02	1.09E-01	1.56E-01			
	Without NDs	1.73E-01	1.53E-01	1.53E-01	1.44E-01	2.24E-01	3.30E-01			
	Median	2.49E-02	1.53E-02	8.95E-03	0.00E00	0.00E00	0.00E00			
24 h anna	Without NDs	7.28E-02	6.79E-02	6.95E-02	6.42E-02	6.81E-02	7.87E-02			
24 hours	Minimum	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00			
	Without NDs	2.29E-03	2.26E-03	2.22E-03	2.27E-03	2.35E-03	4.42E-03			
	Total Samples	9931	11211	11457	9799	10271	3219			
	Total NDs (%)	4083	5365	5652	5151	5257	1700			
	MDL	0.002 to 1.106	0.002 to 1.106	0.002 to 1.084	0.008 to 1.084	0.013 to 0.929	0.0044 to 0.553			
	Maximum	8.21E-01	2.35E-01	3.24E00	2.38E-01	8.25E-01	6.58E-01			
	Mean	4.13E-02	4.56E-03	3.43E-02	2.89E-03	1.08E-02	1.07E-01			
	Without NDs	1.80E-01	1.85E-01	8.37E-01	1.85E-01	5.56E-01	1.58E-01			
	Median	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00	6.75E-02			
2 hours	Without NDs	1.19E-01	1.63E-01	2.85E-01	1.85E-01	6.94E-01	1.21E-01			
3 nours	Minimum	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00			
	Without NDs	3.74E-02	1.58E-01	1.39E-01	1.32E-01	1.50E-01	4.15E-02			
	Total Samples	161	122	122	128	155	43			
	Total NDs (%)	124	119	117	126	152	14			
	MDL	0.055 to 0.0685	0.055307027	0.055307027	0.055307027	0.055307027	0.019468073			

## Table\_Apx C-1. Summary of AMTIC Monitoring Data 2016-2021 for 1,3-Butadiene

Ambient Monitoring Technology Information Center Hazardous Air Pollutants											
Sample Duration	Statistic (µg/m <sup>3</sup> )	Year									
		2016	2017	2018	2019	2020	2021				
	Maximum	2.55E00	4.14E00	7.23E01	2.65E02	4.23E02	2.95E02				
	Mean	5.71E-02	6.91E-02	1.10E-01	1.25E-01	1.38E-01	2.27E-01				
	Without NDs	9.42E-02	9.12E-02	1.40E-01	1.53E-01	2.06E-01	3.80E-01				
	Median	3.88E-02	4.90E-02	5.24E-02	5.85E-02	4.64E-02	2.43E-02				
1 hours	Without NDs	6.31E-02	5.99E-02	6.30E-02	7.07E-02	7.09E-02	6.98E-02				
	Minimum	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00				
	Without NDs	5.44E-03	1.15E-03	2.32E-03	3.58E-03	5.23E-03	2.68E-03				
	Total Samples	13,792	49,876	61,421	63,462	105,406	104,276				
	Total NDs (%)	5,427	12,087	13,240	11,970	34,912	41,966				
	MDL	0.055307027	0.055307027	0.055 to 0.221	0.022 to 0.221	0.016 to 0.0276	0.055 to 0.221				
	Maximum	4.27E02	1.49E01	1.49E02	8.90E02	3.41E01	1.97E01				
	Mean	4.53E00	5.50E-01	1.68E00	8.52E00	1.26E00	1.84E00				
	Without NDs	6.02E00	7.91E-01	2.71E00	1.84E01	2.48E00	5.15E00				
	Median	1.59E-01	9.21E-02	8.94E-02	0.00E00	3.41E-02	0.00E00				
25	Without NDs	2.19E-01	2.25E-01	1.89E-01	4.53E-01	2.71E-01	1.72E00				
25 minutes	Minimum	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00	0.00E00				
	Without NDs	4.35E-02	2.20E-02	2.22E-02	4.46E-02	2.36E-02	4.42E-02				
	Total Samples	101	105	153	186	104	28				
	Total NDs (%)	25	32	58	100	51	18				
	MDL	0.176982477	0.176982477	0.176982477	0.176982477	0.176982477	0.176982477				

Ambient Monitoring Technology Information Center Hazardous Air Pollutants											
Sample Duration	Statistic (µg/m <sup>3</sup> )	Year									
		2016	2017	2018	2019	2020	2021				
	Maximum	-	-	-	-	4.25E00	-				
	Mean	-	-	-	-	6.62E-02	-				
	Without NDs	-	-	-	-	1.97E-01	-				
	Median	-	-	-	-	0.00E00	-				
10	Without NDs	-	-	_	-	7.05E-02	-				
10 minutes	Minimum	-	-	-	-	0.00E00	-				
	Without NDs	-	-	-	-	2.79E-02	-				
	Total Samples	-	-	-	-	3,600	-				
	Total NDs (%)	-	-	-	-	2,393	-				
	MDL	-	-	_	-	0.026547372	-				
	Maximum	-	-	-	-	9.29E02	1.74E02				
	Mean	-	-	-	-	3.59E00	3.16E00				
	Without NDs	-	-	-	-	8.16E00	8.56E00				
	Median	-	-	-	-	0.00E00	0.00E00				
5	Without NDs	-	-	-	-	6.02E00	6.90E00				
5 minutes	Minimum	-	-	-	-	0.00E00	0.00E00				
	Without NDs	-	-	-	-	2.12E-02	2.16E-02				
	Total Samples	-	-	-	-	141,890	317,245				
	Total NDs (%)	-	-	-	-	79,566	200,239				
	MDL	-	_	-	-	39.82105637	39.82105637				



Figure\_Apx C-1. Concentrations of Annual Averages (µg/m<sup>3</sup>)